



| | |
|-------------------------------|------|
| <i>Final Technical Report</i> | 2007 |
|-------------------------------|------|

Water Quality Monitoring and Bioassessment in Nine San Francisco Bay Region Watersheds in 2001-2003:

- Walker Creek**
- Lagunitas Creek**
- San Leandro Creek,**
- Wildcat Creek/San Pablo Creek**
- Suisun Creek**
- Arroyo Las Positas**
- Pescadero Creek/Butano Creek**
- San Gregorio Creek**
- Stevens Creek/Permanente Creek**

June 2007



SURFACE WATER AMBIENT MONITORING PROGRAM

SAN FRANCISCO BAY REGION

**WATER QUALITY MONITORING
AND BIOASSESSMENT
IN NINE SAN FRANCISCO BAY REGION
WATERSHEDS**

WALKER CREEK, LAGUNITAS CREEK, SAN LEANDRO CREEK,
WILDCAT CREEK/SAN PABLO CREEK, SUISUN CREEK,
ARROYO LAS POSITAS, PESCADERO CREEK/BUTANO CREEK,
SAN GREGORIO CREEK, AND STEVENS CREEK/PERMANENTE CREEK

2001 – 2003

Final Report

2007

**SAN FRANCISCO BAY REGIONAL WATER QUALITY CONTROL BOARD
UNIVERSITY OF CALIFORNIA, DAVIS**

List of Authors

Patricia Nicely ¹
Matt Cover ²
Karen Taberski ²
Nelia White ²
John Hunt ¹
Peter Krottje ²
Steve Moore ²

¹ Marine Pollution Studies Laboratory at Granite Canyon
Department of Environmental Toxicology
University of California, Davis
34500 Highway 1, Monterey, CA 93940

² San Francisco Bay Regional Water Quality Control Board
1515 Clay St. Suite 1400
Oakland, CA 94612

With assistance from Amara Vandervort and Megan Kilner of the Moss Landing Marine Labs Quality Assurance Research Group; Cassandra Lamerdin and Marco Sigala of the SWAMP Data Management Team; Jeff Kapellas especially for the maps, Terri Fashing, Kimberly Harrison and Karissa Anderson of the San Francisco Bay Regional Water Quality Control Board; and Janet Cox of the San Francisco Estuary Project.

This report should be cited as:

SFBRWQCB. 2007. Water quality monitoring and bioassessment in nine San Francisco Bay Region watersheds: Walker Creek, Lagunitas Creek, San Leandro Creek, Wildcat Creek/San Pablo Creek, Suisun Creek, Arroyo Las Positas, Pescadero Creek/Butano Creek, San Gregorio Creek, and Stevens Creek/Permanente Creek. Oakland, CA: Surface Water Ambient Monitoring Program, San Francisco Bay Regional Water Quality Control Board.

Summary Table of Contents

Chapters

Chapter 1: Introduction, pp. 1-1 through 1-8

Chapter 2: Selection and Description of Sampled Watersheds, pp. 2-1 through 2-11

Chapter 3: Sampling Design, pp. 3-1 through 3-19

Chapter 4: Methods, pp. 4-1 through 4-17

Chapter 5: Quality Assurance & Quality Control, pp. 5-1 through 5-4

Chapter 6: Regional Trends in Water Quality, pp. 6-1 through 6-64

Chapter 7: Results by Watershed, pp. 7-1 through 7-136

Chapter 8: Conclusions and Recommendations, pp. 8-1 through 8-8

Chapter 9: References, pp. 9-1 through 9-8

Appendices

A: Land Uses

B: Station Locations

C: Quality Assurance, Tables C-1 through C-10

D: Continuous Field Measures, Tables D-1 through D-9

E: Rapid Bioassessment, Tables E-1 through E-3

F: Discrete Field Measures

G: Water Chemistry, Tables G-1 through G-6

H: Sediment Chemistry

I: Clam Tissue Chemistry, Tables I-1 through I-4

J: Water and Sediment Toxicity

K: Water Coliforms

Note: Each chapter contains a comprehensive table of contents for that chapter, including tables and figures. Please see Chapter 1: Introduction for detailed explanations of the scope and organization of this report.

Executive Summary

Study Objectives

The goal of the Surface Water Ambient Monitoring Program (SWAMP) in the San Francisco Bay Region is to monitor and assess watersheds in the Region, using a weight-of-evidence approach based on measurement of physical, chemical, and biological water quality parameters. Data developed in this program are intended to be used for evaluating watersheds for Clean Water Act section 305(b) reporting and 303(d) listing. Specific objectives of the monitoring program are:

- Develop new data to support beneficial use protection
- Measure water quality indicators and stressors to evaluate spatial and temporal trends
- Determine relationships between water quality indicators, specific stressors, and land use, including water management
- Identify reference sites
- Evaluate monitoring tools

Watershed monitoring has been implemented over three hydrologic cycles and across nine planning watersheds. The watershed monitoring data reported here, for SWAMP Fiscal Year 1 (2000-2001) and Fiscal Year 2 (2001-2002), were generated in 2001-2002 and 2002-2003, respectively. Watershed data were compared with water quality thresholds for assigned beneficial uses and analyzed for spatial and temporal trends and linkages with land use. The report does not provide an evaluation of beneficial use support, nor does it assess watershed impairment; however, data provided herein can be used in support of such determinations.

Design

In this study water, sediment, and tissue samples were collected in nine major watersheds: Walker Creek, Lagunitas Creek, San Leandro Creek, Wildcat/San Pablo Creek, Suisun Creek, Arroyo Las Positas in 2001-2002; and Pescadero/Butano Creek, San Gregorio Creek, and Stevens/Permanente Creek in 2002-2003. Sampling sites were selected in a deterministic design to target tributaries and specific land uses so that data could be used to help identify general sources of stressors. At selected sites, suites of parameters were sampled, including benthic macroinvertebrate assemblages, temperature, dissolved oxygen, nutrients, trace metals, trace organic compounds, toxicity, and coliforms. Parameters were measured in a two-tiered design in which aquatic bioassessment was the most widely used indicator of stream condition. Other suites of indicators were monitored based on land use and beneficial use (*e.g.*, coliform bacteria for water contact recreation).

Regional Results

Bioassessment data indicated a relationship between biological integrity in streams and land use practices. Stream sites receiving runoff from open space and rural residential areas had the best biological conditions; stream sites draining agricultural, grazing, and mixed use sites had intermediate biological conditions; and stream sites draining urban areas had the most degraded benthic invertebrate communities. Physical habitat conditions, particularly riparian habitat and channel alteration, were associated with the health of benthic communities. One of the most important natural factors affecting benthic communities was flow intermittency.

Elevated temperatures and depressed dissolved oxygen concentrations were common throughout the watersheds, and these conditions were often beyond ranges supportive of salmonid populations. Region-wide there were significant correlations between riparian habitat conditions and measurements of temperature and dissolved oxygen. Sites at which channels were less modified, riparian and stream bank vegetation more extensive, and canopy cover was greatest had significantly lower stream temperatures and significantly higher dissolved oxygen concentrations.

Nutrients were often elevated in urban areas which were most commonly found at the downstream end of watersheds. Nitrate concentrations over all seasons were highest and most variable in streams draining urban areas, followed by mixed land use, agriculture, grazing, rural residential and open space. The mean nitrate level for urban streams was more than twice that for streams draining agricultural areas and nearly ten times that of streams in open space areas. Although nitrate concentrations never exceeded Basin Plan objectives for drinking water or agricultural supply, most measured nitrate values exceeded the EPA reference guideline of 0.155 mg-N/L that was developed to protect against eutrophication.

In general, the number of contaminant exceedances was low. However, urban areas tended to have the highest concentration of contaminants. This was particularly true of polycyclic aromatic hydrocarbons (PAHs). Legacy pesticide and PCB concentrations tended to be highest at the downstream ends of urban watersheds. In general, ambient concentrations of contaminants were highest during the dry season, which may be due to dilution during wet weather. Samples were not collected during storms.

Throughout the region, toxicity was moderate. Out of 59 water samples, algal growth was reduced in 15 samples, invertebrate growth or survival was affected in three samples and fish larval survival was affected in one sample. Sediment toxicity affecting amphipod growth was observed in approximately half of the samples. In general, the level of water and sediment toxicity was moderate, except in San Leandro Creek. Results from this study combined with previous data showed that *Ceriodaphnia* toxicity due to diazinon has decreased in urban creeks, probably as a result of management measures restricting its use.

Watershed Specific Results

Pescadero/Butano and San Gregorio Creeks had the highest water quality and most intact benthic communities of the watersheds surveyed. Arroyo Las Positas had the highest number of temperature and dissolved oxygen measurements outside ranges acceptable for salmonid or coldwater habitat. Stevens Creek had the highest aquatic toxicity. San Leandro Creek had the highest sediment toxicity with 100 percent amphipod mortality. Relationships between toxicity and individual chemicals were not clear. Toxicity is most likely due to the combined action of a variety of contaminants. Arroyo Las Positas and San Pablo Creek had elevated levels of the herbicide oxadiazon.

Conclusions

Most of the watersheds in this study had similar overlapping patterns of water quality stressors. Temperature and dissolved oxygen values often exceeded thresholds for fish and other aquatic life, even in areas with less intensive human land use activities. As healthy riparian habitat correlated significantly with temperature and dissolved oxygen in this study, the primary management focus in these areas is likely to be improvement and maintenance of riparian habitat. Overlapping this basic physical pattern is the influx of nutrients and chemical contaminants from areas of more intensive human activity, primarily urban areas. Management activities here should also include riparian habitat improvement to allow for the absorption and degradation of pollutants through retention in vegetated areas. Designing riparian corridors to limit transport of these materials is likely to be among the easiest and more successful strategies to reduce contaminants, improve temperature and dissolved oxygen conditions, and maintain the essential ecological values of riparian areas. Management measures should also concentrate on preventing pollution from entering creeks by: 1) educating urban citizens on the use of pesticides and fertilizers, pet care and car washing, 2) regulating pollution sources and source materials, and 3) using pervious surface and vegetation in the design of green buildings, parking lots, green swales and decentralized landscape for the treatment of stormwater.

In this study, macroinvertebrate bioassessments proved to be an excellent indicator for evaluating the health of aquatic life. Multi-probe continuous monitoring sondes were extremely valuable in determining whether elevated temperatures and low dissolved oxygen may have an impact on salmonids and other aquatic life. In many cases physical habitat assessments illustrated the probable cause of elevated temperatures and low dissolved oxygen.

A watershed monitoring coalition should be established in the Bay Area to provide the development of more meaningful watershed information and maximize resource efficiency. In a collaborative effort, SWAMP may best serve to provide a regional context for local monitoring programs.

Table of Contents

| | | |
|-----|--|-----|
| 1 | Introduction..... | 1-1 |
| 1.1 | Overview of the Surface Water Ambient Monitoring Program in California | 1-1 |
| 1.2 | Goals and Objectives of SWAMP in the San Francisco Bay Region..... | 1-1 |
| 1.3 | Overview of the San Francisco Bay Region SWAMP Sampling Program | 1-2 |
| 1.4 | Scope of the Report..... | 1-3 |
| 1.5 | Organization of the Report..... | 1-3 |
| 1.6 | Map Notes: How to read the maps in this report | 1-5 |

1 Introduction

1.1 Overview of the Surface Water Ambient Monitoring Program in California

California Assembly Bill 982 (Water Code Section 13192; Statutes of 1999) required that the State Water Resources Control Board (SWRCB) assess and report on State water monitoring programs and prepare a proposal for a comprehensive surface water quality monitoring program. In the SWRCB Report to the Legislature from November 2000 entitled "Proposal for a comprehensive ambient surface water quality monitoring program," the SWRCB proposed to restructure the existing water quality monitoring programs into a new program, the Surface Water Ambient Monitoring Program (SWAMP). This new program consists of statewide environmental monitoring focused on providing the information needed to effectively manage the State's water resources. SWAMP is designed to be consistent, cooperative, adaptable, scientifically sound, and to meet clear monitoring objectives. It will also facilitate reporting and categorizing of the State's water quality under Sections 305 (b) and 303 (d) of the federal Clean Water Act. Details of SWAMP can be found in the SWAMP Quality Assurance Management Plan (QAMP) (Puckett 2002).

Specifically, the statewide SWAMP has been designed to meet four goals:

1. Create an ambient monitoring program that addresses all hydrologic units of the State.
2. Document ambient water quality conditions in potentially clean and polluted areas.
3. Identify specific water quality problems preventing the realization of beneficial uses of water in targeted watersheds.
4. Provide data to evaluate the overall effectiveness of water quality regulatory programs in protecting beneficial uses of waters of the State.

A new statewide monitoring design is currently being developed.

1.2 Goals and Objectives of SWAMP in the San Francisco Bay Region

In October 1999, the San Francisco Bay Regional Water Quality Control Board (Regional Board) developed a Regional Monitoring and Assessment Strategy (RMAS) in order to collect information on all in the San Francisco Bay Region. SWAMP is being used in this Region to implement the RMAS. The three components of SWAMP/RMAS are:

- 1) Regional Board-lead activities, concentrating on monitoring watersheds, lakes/reservoirs and bays and estuaries other than San Francisco Bay, including programs such as State Mussel Watch (SMW), the Toxic Substances Monitoring Program (TSMP) and the Coastal Fish Contamination Program (CFCP)
- 2) Partner-lead watershed monitoring programs that are being conducted by local agencies/groups with similar goals, structure, scope, protocols and quality assurance as the Regional Board-lead activities
- 3) The San Francisco Estuary Regional Monitoring Program (RMP)

Information collected in Component 1 (Regional Board-lead activities) is the subject of this report.

The goal of the SWAMP program in this Region has been to monitor and assess watersheds in the Region, using a weight-of-evidence approach based on measurement of physical, chemical, and biological water quality parameters. Data developed in this program is intended to be useful for evaluating for 305b reporting and 303d listing. Specific objectives of the monitoring program are to:

- 1) Collect and evaluate physical, chemical, and biological water and sediment quality data to determine whether beneficial uses are being protected in watersheds in the Region
- 2) Use a weight-of-evidence approach to measure environmental stressors (pollutants or other water quality parameters), biological effects (e.g., toxicity tests), and ecological indicators (e.g., benthic community analysis) using a study design that allows evaluation of spatial and temporal trends
- 3) Determine relationships between observed biological effects, levels of specific stressors, and land uses including water management
- 4) Identify reference sites
- 5) Evaluate watershed monitoring tools to develop a program that uses the best environmental indicators to achieve the purposes of the program

Due to a reduction in regional SWAMP funding, we plan in the future to meet these objectives in collaboration with other watershed monitoring programs.

1.3 Overview of the San Francisco Bay Region SWAMP Sampling Program

The technical approach in the San Francisco Bay Region for Regional Board lead activities under SWAMP included:

- 1) Monitoring watersheds to assess water quality impacts and establish regional reference sites; and
- 2) Monitoring edible fish for contaminant levels in reservoirs and coastal areas where people catch and consume fish.

Watershed monitoring was implemented by San Francisco Bay Regional Board staff and contract laboratories over three hydrologic cycles and across nine planning watersheds. The watershed monitoring data reported here, for SWAMP Fiscal Year 1 (2000-2001) and Fiscal Year 2 (2001-2002), were generated in 2001-2002 and 2002-2003, respectively.

Contaminant monitoring in edible fish was implemented through the Toxic Substances Monitoring Program (TSMP) and the Coastal Fish Contamination Program (CFCP), which were included under SWAMP. Edible fish sampling occurred in coastal areas and reservoirs popular for fishing. TSMP data were generated from fish collected during 2000 and 2001; CFCP data were generated from fish collected during 1998, 1999 and 2001. The results of fish tissue studies can be found in the report “Chemical Concentrations in Fish Tissues from Selected Reservoirs and Coastal Areas in the San Francisco Bay Region,” (SFBRWQCB 2005). This report is available at <http://www.waterboards.ca.gov/sanfranciscobay/monitoring.html>.

1.4 Scope of the Report

This report provides a data summary for watershed monitoring completed during Fiscal Years 1 and 2 (2001 to 2003). Watershed data were compared with published water quality goals and analyzed for spatial trends, temporal trends, and linkages with land use. The report does not provide an evaluation of beneficial use support, nor does it assess impairment; however, data provided herein can be used in support of such determinations.

1.5 Organization of the Report

Each section, with the exception of Section 9 (References), is prefaced by its own table of contents, followed by lists of the tables and figures in that section.

Section 1. Introduction. Provides the background of SWAMP in California, outlines its goals and sampling program in the San Francisco Bay Region, and defines the scope of the report. The introduction also includes this description of the report’s sections and category definitions for the maps.

Section 2. Selection and Description of Watersheds. Provides a regional map to locate each watershed, explains how the watersheds in this report were selected, lists the sequence in which the watersheds are presented throughout the report, and gives general watershed descriptions.

Section 3. Sampling Design. Reviews how sampling sites were chosen and the overall sampling design, as well as which parameters were sampled and the relationship to land use.

Section 4. Methods. Provides a review of the procedures used in sampling, laboratory analysis, and data analysis, including a background on bioassessment.

Section 5. Quality Assurance & Quality Control. Details the data quality procedures applied to water, sediment, clam tissue, and toxicity test data and summarizes the general QA/QC used throughout the report. For specific QA/QC results, refer to the tables in Appendix C.

Section 6. Regional Trends in Water Quality. Presents spatial and seasonal trends in the data; background on temporal patterns in continuous monitoring data; an extensive discussion of natural variation in reference conditions for bioassessment; and a brief discussion of reference conditions for water and sediment chemistry, and for nutrients.

Section 7. Results by Watershed. Summarizes sites of concern, exceedances, and water quality in relation to land use for each watershed. Presents, by watershed, results and analysis for bioassessment; continuous monitoring; water, sediment, and clam tissue chemistry; water and sediment toxicity; and coliform bacteria.

Section 8. Conclusions and Recommendations. Discusses findings in relation to land use, physical habitat, and seasonal trends; discusses reference conditions; evaluates monitoring tools; and presents recommendations for management and future monitoring and assessment.

Section 9. References. Lists references cited in the report.

If you are interested in one particular watershed, reviewing the following parts of the report will help orient you before reading the results in section 7:

- Section 1: Map Notes
- Section 2: Watershed descriptions
- Section 3: Watershed maps (Figures 3-1 through 3-9) and Table 3-4, which show what was monitored where and general land use categories
- Section 4: Methods; Tables 4-2 and 4-3 for water and sediment thresholds; Table 4-4 for biological metrics
- Section 6: Overview of regional results and background for understanding the watershed-specific results

Station lists

In Section 3, **Table 3-4** is a comprehensive list of the stations in each watershed and what was sampled where, along with land use. Table 3-4 does not include the station name, but the subsequent maps indicate the general location, and **Appendix B** (the Station Location Table) provides a station list with full station codes, station names, waterbodies, and the latitude and longitude. The order of the watersheds follows the sequence listed in Section 2.1 under watershed selection. **Appendix A** lists land use in detail and general category by full station code and station name.

In Section 7, **7.x.3** (the macroinvertebrate assemblage subsection for each watershed) includes a table of bioassessment sites (with short station code and station name) and gives descriptions of the sites by assemblage group.

Full station codes are nine characters (such as 201WLK030), beginning with a three-digit hydrological unit code (HUC) for the basin (201 for Marin), followed by a three-character alpha code for the watershed (WLK for Walker Creek) and a three-digit code for the station (030 for a downstream Keys Creek site). The last three-digit code increases from the mouth, up tributaries, then back up to the mainstem headwaters. In the appendices and several tables, the initial HUC is usually included, but the maps and most of the report narrative use the shortened six-character watershed and station code to identify the station (WLK030).

The appendices provide the detailed data as follows:

Appendix A: Land uses

Appendix B: Station location information

Appendix C: Quality assurance data

Appendix D: Continuous field measures

Appendix E: Rapid bioassessment

Appendix F: Discrete field measures

Appendix G: Water chemistry

Appendix H: Sediment chemistry

Appendix I: Clam tissue chemistry

Appendix J: Water and sediment toxicity

Appendix K: Water coliforms

1.6 Map Notes: How to read the maps in this report








1. Overview map: Section 2, Figure 2-1
2. Watershed index maps: Section 3, Figures 3-1 through 3-9
3. Regional *E. coli* map: Section 6, Figure 6-31
4. Bioassessment maps: Section 7 throughout
5. Continuous Monitoring maps: Section 7 throughout


1. The Overview map (Figure 2-1) locates each watershed within the San Francisco Bay region's boundary.


2. Watershed index maps category definitions


The watershed station index maps (Figures 3-1 through 3-9) are a guide to what sampling was done where and the land use at each station. It is a graphic representation of the information presented in more detail in Table 3-4.


The **Land use** designation is from the simplified table of six land uses:


Urban , Mixed , Agricultural , Grazing , Rural Residential ,
and Open Space . Although not a general land use category,  indicates a quarry or mine.


 **Bioassessment** refers to Rapid Bioassessment of benthic macroinvertebrates using the California Stream Bioassessment Protocol, which includes a physical habitat assessment. The selection includes stations sampled in a National Park Service study on Olema Creek in the Lagunitas Creek watershed and from a separate SFBRWQCB study at selected stations on Wildcat and San Leandro Creeks.


 At stations with **Continuous monitoring**, the standard field parameters (temperature, dissolved oxygen, pH, and specific conductance) were monitored with YSI 6600 sondes for an average of 11 days at a time and up to four times a year.

 **Conventional water** parameters include ammonia, nitrate, nitrite, total kjeldahl nitrogen (TKN), orthophosphate (OP), total phosphate, boron, chloride, sulfate, total dissolved solids (TDS), chlorophyll-*a*, pheophytin-*a*, alkalinity, hardness, suspended sediment concentration (SSC), total organic carbon (TOC), and dissolved organic carbon (DOC) as well as standard field parameters, sampled as grab samples, usually three times a year.

 The **Water contaminants and toxicity** category includes testing for presence of OP and organochloride (OC) pesticides, polycyclic aromatic hydrocarbons (PAHs), polychlorinated biphenyl (PCB) congeners, and trace metals, as well as ELISA tests and three-species toxicity tests. This category aggregates the water metals and toxicity and the water organics categories in Table 3-4 to limit icons represented on the maps; thus not all stations were tested for all parameters.

 **Sediment analysis** includes grain size evaluation and testing for presence of OP and OC pesticides, PAHs, PCB congeners, and trace metals, as well as *Hyaella azteca* toxicity tests.

 **Tissue analysis** includes testing for presence of OP and OC pesticides, PAHs, PCB congeners, and trace metals.






 The **Pathogens** category includes weekly sampling for five weeks in the summer and analysis for total coliform, fecal coliform, and *E. coli*.

3. Regional *E. coli* map (Figure 6-31) represents in red each station that exceeded the *E. coli* steady state limit; other stations tested are represented in grey.

4. Bioassessment maps category definitions





The maps of benthic macroinvertebrate index scores (Figures 7-4, 7-9, 7-18, 7-26, 7-35, 7-44, 7-54, 7-64, and 7-73) report taxa richness and percent sensitive EPT (Ephemeroptera, Plecoptera, Trichoptera). Table 4-4 defines the bioassessment metrics.

Taxa richness

| | | |
|---|---------|--|
|  | <17 | Poor |
|  | 18 - 27 | Fair |
|  | 28 – 38 | Reference condition for intermittent streams |
|  | >38 | Reference conditions for perennial streams |
|  | no data | Incompatible data |

Taxa richness data from Olema Creek (in the Lagunitas Creek watershed) were collected using a variation in the protocol, thus comparisons using those data are inappropriate, and the “no data” symbol is displayed at those sites.






Percent sensitive EPT

| | | |
|---|---------|---|
|  | 0 | Poor |
|  | 1 - 20 | Fair |
|  | 21 - 44 | Reference conditions for intermittent streams |
|  | >44 | Reference conditions for perennial streams |

5. Continuous monitoring maps category definitions

The maps of temperature and dissolved oxygen display exceedances from episodes of continuous monitoring (Figures 7-4, 7-12, 7-20, 7-29, 7-38, 7-48, 7-58, 7-68, 7-76).

Temperature

| | | |
|---|---------|-------------------------------------|
|  | >24°C | Instantaneous maximum for salmonids |
|  | > 17°C | MWAT for steelhead |
|  | >14.8°C | MWAT for coho salmon |
|  | ≤14.8°C | No exceedance |
|  | no data | <7 days or not in critical period |

The limits mapped for temperature include:

- 24°C as an instantaneous maximum for salmonids; this temperature is the lower end of the range of acute thresholds considered lethal to salmonids (U.S. EPA 1977)






- 17°C as Maximum Weekly Average Temperature (MWAT) limit for steelhead (Sullivan *et al.* 2000)
- 14.8°C as an MWAT for coho salmon (Sullivan *et al.* 2000)

The MWATs are limits representing a 10 percent growth reduction. The instantaneous maximum, represented by the dead fish, is independent of the weekly average temperature. Thus, a station could exceed 24°C on a few days, but still have an average weekly temperature below both MWAT standards. Therefore, the expected red dot for poor temperature is replaced by the dead fish, which may be accompanied by an exceedance of either MWAT standard.

All the watersheds in this survey are designated as supporting coldwater habitat except for Wildcat and San Pablo creeks. Lagunitas and Walker creeks are the only creeks to support coho habitat currently, although San Gregorio, Pescadero, and Butano creeks may have historically. While the maps represent limits for coho and coldwater habitat, the watershed may not be designated for that beneficial use. The text will evaluate each watershed appropriately.

The grey circle, meaning no MWAT, indicates that either there were fewer than seven days of continuous monitoring or that the station was not monitored during the critical summer period so that its highest weekly average temperature is not representative of an annual maximum.

Dissolved Oxygen

| | | |
|---|----------|--|
|  | <5 mg/L | warmwater fish minimum |
|  | <7 mg/L | coldwater fish minimum |
|  | < 9 mg/L | critical spawning minimum |
|  | ≥ 9mg/L | no exceedance |
|  | no data | not in critical period or QC rejected data |

The limits mapped for dissolved oxygen reflect tolerances for warmwater fish, coldwater fish, and coho salmon. The Basin Plan minimum dissolved oxygen limit for warmwater fish habitat is 5 mg/L (milligrams per liter); for coldwater fish habitat, it is 7 mg/L. The 9 mg/L limit is a protective limit used by the North Coast RWQCB for spawning during critical periods and applied only to selected streams and those with no specific criterion (NCRWQCB 2001); it is also the level at which slight impairment occurs to embryo-larval production (U.S. EPA 1991). It is mapped here as an indication of DO levels, not as a standard. The text will interpret the levels appropriately for each watershed and season.

The grey square for no data means either the data were rejected because post-calibration indicated drift exceeding limits for quality assurance or that the station was not monitored during the critical summer period so that its dissolved oxygen values are not representative of an annual minimum.

Table of Contents

2 Selection and Description of Sampled Watersheds 2-2

2.1 Watershed Selection..... 2-2

2.2 Watershed Descriptions 2-2

2.2.1 Walker Creek Watershed 2-3

2.2.2 Lagunitas Creek Watershed 2-4

2.2.3 San Leandro Creek Watershed..... 2-5

2.2.4 Wildcat Creek/San Pablo Creek Watershed 2-6

2.2.5 Suisun Creek Watershed 2-7

2.2.6 Arroyo Las Positas Watershed..... 2-8

2.2.7 Pescadero Creek/Butano Creek Watershed 2-9

2.2.8 San Gregorio Creek Watershed 2-10

2.2.9 Stevens Creek/Permanente Creek Watershed..... 2-10

List of Figures

Figure 2-1. Regional overview map showing locations of sampled watersheds 2-1

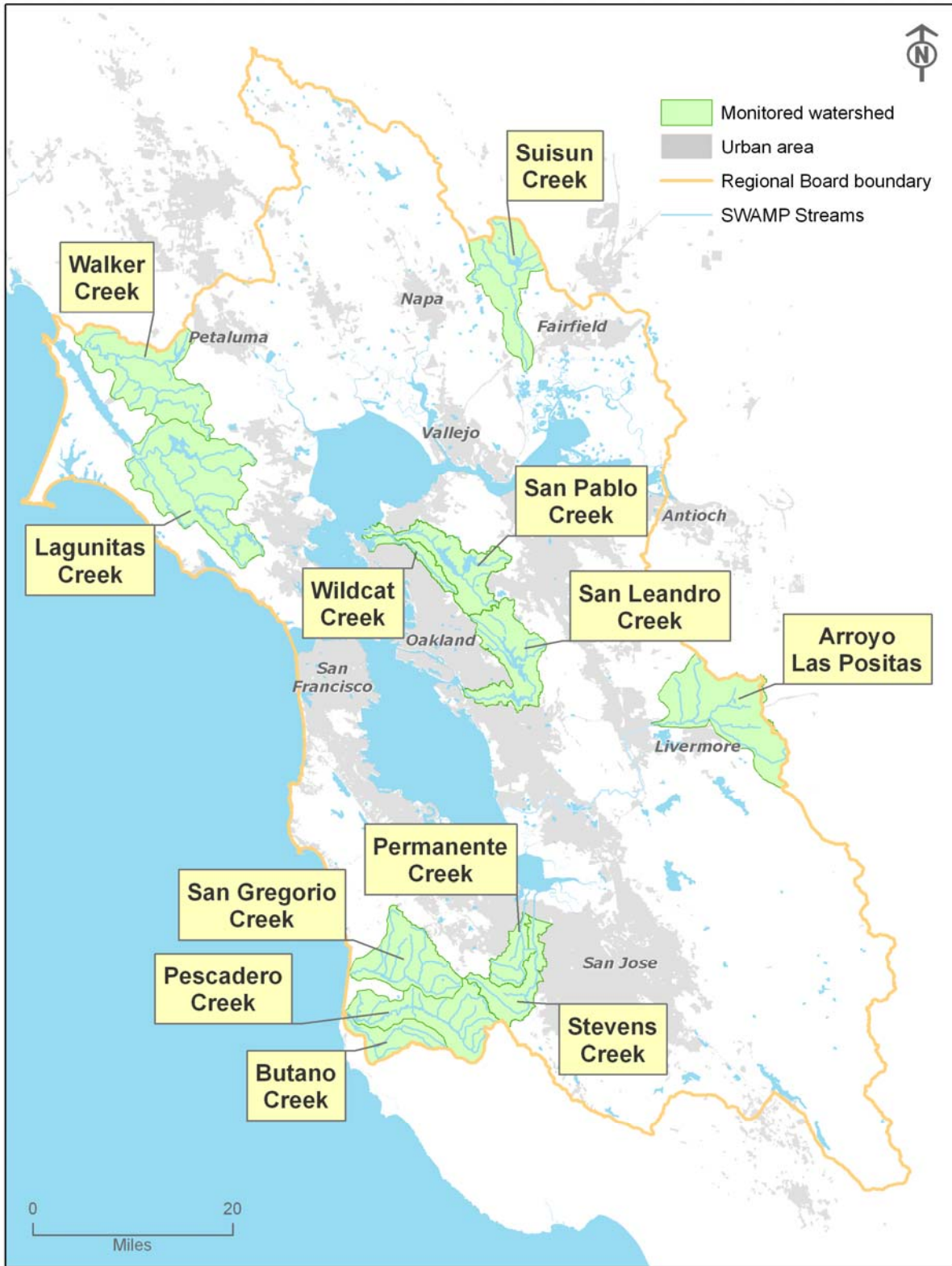


Figure 2-1. Regional overview map showing locations of sampled watersheds

2 Selection and Description of Sampled Watersheds

2.1 Watershed Selection

The Region was divided into 47 planning watersheds, each between 30 and 200 square miles in area, with most between 50 and 100 square miles. Some of these planning watersheds are self-contained hydrologic units that drain to an estuary or the ocean, and others have been either combined with adjacent watersheds or are sub-watersheds within a larger drainage basin. All planning watersheds are fully contained within one of the seven hydrologic basins of the San Francisco Bay Region.

Planning watersheds were prioritized for monitoring and assessment using criteria aimed at generating the most useful and current information with the least amount of new resources and investigations. Criteria also consider time-sensitive issues such as imminent development plans (*e.g.*, major housing or flood control projects), upcoming stream restoration projects, or declining sensitive aquatic resources. The prioritized order of planning watersheds achieves balance geographically, by eco-region, and includes both data-rich and data-poor watersheds as well as a balance of potentially clean and problem watersheds.

The prioritization criteria for planning watersheds include:

1. **Existing local efforts.** Build on existing watershed monitoring and assessment efforts, including citizen monitoring.
2. **Sensitive aquatic resources.** Focus on areas with sensitive aquatic resources or species, such as habitat for steelhead trout, a federally-listed threatened species.
3. **Pre-project information.** Collect pre-project ambient data in areas proposed for urbanization, stream restoration, or hydromodification.
4. **Watersheds with limited information.** Initiate monitoring in areas that have little or no current water quality and habitat information.
5. **Represent all eco-regions.** Fill information gaps in certain eco-regions. For instance, collect stream bioassessment data to support biocriteria development or geomorphic data to support physical criteria development.
6. **Geographic balance.** The list of planning watersheds should be balanced geographically and by eco-region, in order to capture the full range of stream types in the region and to recognize watershed management efforts in all parts of the region.

The prioritized list of all planning watersheds to be monitored under SWAMP in the San Francisco Bay Region is presented in the SWAMP 5-year Workplan (SFBRWQCB 2004). The top nine prioritized planning watersheds were the focus of SWAMP sampling efforts in Year 1 and Year 2 (shown in Figure 2-1) in the following order:

Year 1

- Walker Creek
- Lagunitas Creek
- San Leandro Creek
- Wildcat Creek/San Pablo Creek
- Suisun Creek
- Arroyo Las Positas

Year 2

- Pescadero Creek/Butano Creek
- San Gregorio Creek
- Stevens Creek/Permanente Creek

Descriptions of these watersheds are presented below.

2.2 Watershed Descriptions

The following sections provide descriptions of the watersheds monitored for this report. Each description includes, when available, an overview of the watershed and surrounding geologic and vegetation features, as well as current and past land use activities water quality issues, and existing beneficial uses listed for the watershed. More detailed information, including geologic characteristics and past studies, can be found in the project Workplan (SFBRWQCB 2002).

2.2.1 Walker Creek Watershed

The Walker Creek watershed has a drainage area of 73 square miles, mostly in northwestern Marin County, with a small portion in Sonoma County. Significant tributaries to Walker Creek include Keys Creek (also known as Keyes), which flows through the gentle hills east of Tomales, joining Walker Creek near Tomales Bay; Chileno Creek, which flows through Chileno Valley; and, in the upper watershed, Salmon Creek and Arroyo Sausal Creek, which flow through Hicks Valley. Frink and Verde Canyons each support ephemeral streams that join Walker Creek upstream from Chileno Creek. Soulajule Reservoir impounds the 15 square mile drainage of Arroyo Sausal. Privately-owned Laguna Lake at the headwaters of Chileno Creek is a large, shallow water body.

Annual rainfall varies from 24 to 30 inches (Rantz 1971); Chileno Valley is notably hotter and drier than the coastal areas. The lower watershed is composed of low, rolling hills, while the upper watershed consists of rugged canyons in the southeastern headwaters. Vegetation is 60 percent open grassland and 40 percent shrub and woodland (mixed hardwood and a few conifers) (Zumwalt 1972). Portions of the watershed are susceptible to landslides, erosion, and gullyng.

Land use in the watershed includes grazing, farming, residential, and mining. Grazing is the major land use impact in the Walker Creek watershed, with dairies concentrated in the valleys of

Keys Creek near Tomales and Chileno Valley, and beef or sheep ranches predominating elsewhere (Bush 1995). Until 1991, Laguna Lake was annually drained and planted with corn; a duck farm currently drains to the lake. Historic potato farming, which resulted in increased erosion, peaked in the 1860s around Tomales (Bush 1995). The only residential communities (both in the lower watershed) are the small town of Tomales, and Blue Mountain, a community near Keys Creek. There is a sewage treatment pond adjacent to Keys Creek. Tomales Bay supports commercial oyster farming which, due to coliform inputs from the watershed, closes down during periods of heavy rainfall; a TMDL for pathogens has been developed for Tomales Bay. The Gambonini Mine operated as a large open pit cinnabar mine for mercury from 1964 to 1970, and is a significant source of mercury to Walker Creek and Tomales Bay (Whyte 2002). Several smaller historic mercury mines also exist in the watershed. Walker Creek was once mined for gravel, just downstream of the confluence with Chileno Creek (Bush 1995). Based on land use and soil characteristics, the major water quality issues in the watershed are water diversion, nutrients, coliforms, erosion, sedimentation, and mercury.

Walker Creek is protected habitat for coho salmon, steelhead trout, and California freshwater shrimp (U.S. Fish and Wildlife Service 1997). The beneficial uses listed in the Basin Plan for the San Francisco Bay Region (SFBRWQCB 1995) that cover habitat for these species are: warm and cold freshwater habitat, preservation of rare and endangered species, fish migration and spawning, and wildlife habitat. These beneficial uses are also listed for the entire watershed. Soulajule Reservoir is a municipal/domestic water supply, source of freshwater replenishment, and fishing area that provides warm freshwater habitat, wildlife habitat, and contact/non-contact water recreation. Walker Creek and its watershed are also listed for potential use in contact and non-contact water recreation.

2.2.2 Lagunitas Creek Watershed

The Lagunitas Creek watershed drains 103 square miles of west central Marin County. From the headwaters on the north slope of Mount Tamalpais, Lagunitas Creek flows about 25 miles before emptying into Tomales Bay. The first eight miles of Lagunitas Creek are dammed for municipal drinking water by the MMWD; Peters Dam holds 32,900 acre-feet in Kent Lake, about a mile downstream from three smaller reservoirs near the headwaters—Lake Lagunitas, Bon Tempe Lake, and Alpine Lake. San Geronimo Creek drains San Geronimo Valley and joins Lagunitas Creek at Shafter Bridge, about a mile below Peters Dam. Devils Gulch drains the western slope of Barnabe Mountain and joins Lagunitas Creek about 2.5 miles downstream. The reach between San Geronimo and Tocaloma, especially in Samuel P. Taylor State Park, is prime habitat for coho salmon, steelhead, and California freshwater shrimp. Halleck and Nicasio Creeks drain 35.9 square miles of the gentle grassland hills of Nicasio Valley, joining in Nicasio Reservoir. About a mile below Seeger Dam, Nicasio Creek joins Lagunitas Creek. Olema Creek drains 14.5 square miles of Inverness and Bolinas Ridges; it joins Lagunitas Creek southwest of Point Reyes Station, where tidal influence affects the flow.

Annual rainfall varies from 28 to 52 inches (Rantz 1971), with the most intense rain over Kent Lake and the upper watershed, and the driest portion in Nicasio Valley. Some areas are subject to

landslides, erosion, and gulying. The upper watershed supports a dense forest of redwood, Douglas fir, coast live oak, bay, and alder. The northern uplands of Nicasio and San Geronimo Valleys are grassland, with bay, coast live oak, and Douglas fir on the steeper slopes to the south. The gentle slopes on the east side of Olema Creek and the lower watershed are predominantly grassland and chaparral.

Current and historic land use in the watershed has been highly diverse. Historically, some lowland areas of Olema Creek were farmed for feed crops and vegetables, and extensive logging took place between 1850 and 1964. Grazing is a current and predominant land use in the watershed and includes dairy and beef cattle and horses. Several golf courses exist in the watershed, one of which drains to Bon Tempe Lake. Gravel extraction also occurs at two sites in the watershed. Most of the Lagunitas Creek watershed is publicly owned and protected land, offering numerous recreational opportunities. MMWD owns about 30 percent of the watershed, mostly in the southern portion near the headwaters; the National Park Service (NPS) owns 21 percent of the watershed as Point Reyes National Seashore (PRNS) and Golden Gate National Recreation Area (GGNRA); Marin County Open Space District (MCOSD) has 3.5 percent; and Samuel P. Taylor State Park accounts for 3 percent. Each has trails for hikers, bicycles, and horses. There are also many residential communities in the watershed, all of which use septic systems. Tomales Bay supports commercial oyster farming which, due to coliform inputs from the watershed, closes down during periods of heavy rainfall. The Regional Water Board recently developed a TMDL for pathogens in Tomales Bay. Given these land uses and the instability of some soils, water quality issues in the watershed include habitat impairment, nutrients, herbicides, pesticides, erosion, turbidity, temperature, and coliforms.

Lagunitas Creek is protected habitat for coho salmon, steelhead, and California freshwater shrimp. It is one of the most important coho salmon streams in California, supporting approximately 10 percent of the current population. The beneficial uses listed in the Basin Plan (SFBRWQCB 1995) that cover habitat for these and other species in the creek and watershed are: cold and warm freshwater habitat, preservation of rare or endangered species, fish migration and spawning, and wildlife habitat. Nicasio Reservoir is a source of municipal/domestic water and freshwater replenishment, as well as a fishing area with water contact and non-contact recreation, and provides habitat for fish spawning, warm water species, and wildlife. Nicasio Creek provides habitat for cold-water species, wildlife, and fish spawning and migration (in limited areas), is used for water contact and non-contact recreation, and provides municipal/domestic water and freshwater replenishment. The MMWD reservoirs are each a municipal/domestic water supply, with open trails for non-contact water recreation, and are also listed for water contact recreation, cold and warm water habitat, fish spawning, and wildlife habitat. Olema Creek is listed for cold and warm water habitat, wildlife habitat, fish spawning and migration, navigation and water contact recreation. The entire watershed is listed for agricultural supply.

2.2.3 San Leandro Creek Watershed

San Leandro Creek watershed is a 46.5 square mile drainage basin in Alameda and Contra Costa Counties. Although it is channelized and concrete lined in portions of the lower watershed, San

Leandro Creek is one of the few East Bay creeks entirely above ground. From the headwaters near Round Top Peak, the creek drains to Arrowhead Marsh in San Leandro Bay, just north of the Oakland International Airport. Redwood, Indian, Moraga, Buckhorn, and Kaiser Creeks drain into the Upper San Leandro Reservoir; downstream, Grass Valley Creek drains into Lake Chabot. The Mokelumne Aqueduct delivers drinking water from the Sierra Nevada to the reservoir near the mouth of Moraga Creek. Lower San Leandro Creek, from below Chabot dam to Arrowhead Marsh, runs through residential and urban industrial areas of Oakland and San Leandro. Freshwater input to Lower San Leandro Creek is limited by upstream diversions, and consists primarily of seepage (URS Greiner Woodward Clyde 1999) and occasional winter releases from Lake Chabot; summer flow in the lower watershed is dominated by storm drain runoff.

Annual rainfall in the watershed ranges from 18 to 28 inches per year (Rantz 1971). The landscape of this highly urbanized watershed varies considerably, from gently rolling hills to steep canyons. Grassland is the dominant vegetation type in undeveloped areas, with scattered oak woodlands and coastal chaparral. Redwood Creek runs through a second and third growth redwood forest in Redwood Regional Park. Stands of introduced eucalyptus, Monterey pine, and fruit trees are common. A freshwater marsh of about 18 acres near Upper San Leandro Reservoir supports wetland species.

Land use in the upper watershed includes recreation, residential, and grazing. The drainage of upper San Leandro Creek above Chabot dam is protected watershed of the East Bay Municipal Utility District (EBMUD) and undeveloped parkland for recreational use held by the East Bay Regional Park District (EBRPD). Cattle-grazing occurs on portions of EBMUD and EBRPD lands. The towns of Moraga and Rheem Valley are in the upper drainage above Upper San Leandro Reservoir. Recreational land use in the upper watershed includes horse stables and golf courses. Lake Chabot provides water for golf course irrigation. Land use below Lake Chabot is exclusively urban. Water quality issues in the watershed associated with land use and erodibility include low flow conditions due to water diversions, bacteria toxicity, nutrients, sediment and other contamination due to urban runoff.

The beneficial uses listed in the Basin Plan (SFBRWQCB 1995) for resident species in Lake Chabot and Upper San Leandro Reservoir are cold and warm freshwater habitat, wildlife habitat, and fish spawning. Other beneficial uses for these reservoirs include municipal/domestic supply (emergency only), and water recreation (both contact and non-contact). Existing beneficial uses in San Leandro Creek include freshwater replenishment, cold freshwater habitat and wildlife habitat. Potential beneficial uses associated with San Leandro Creek are fish migration, contact and non-contact recreation, fish spawning and warm freshwater habitat.

2.2.4 Wildcat Creek/San Pablo Creek Watershed

Wildcat and San Pablo Creeks are contiguous watersheds in Contra Costa County, together draining 48 square miles. Wildcat Creek drains the western slopes of Vollmer Peak in the Berkeley Hills, flows north through Wildcat Canyon, turns west through San Pablo, Richmond, and North Richmond, and runs through Wildcat Marsh before entering San Pablo Bay just north

of Point Richmond. Wildcat Creek has a major tributary, Havey Creek, and two small impoundments, Lake Anza and Jewel Lake. The first four miles of its headwaters are within the East Bay Regional Park District (EBRPD). The remaining seven miles of creek are mostly above ground and un-channelized, making it one of the few streams in the San Francisco Bay Area with a nearly continuously vegetated riparian channel.

Adjoining the narrow Wildcat Creek watershed on the north is the wider 39 square mile watershed of San Pablo Creek. Perennial tributaries of San Pablo Creek include Bear Creek, which drains into Briones Reservoir, and Lauterwasser Creek, which joins San Pablo Creek just upstream of San Pablo Reservoir. The Mokelumne Aqueduct imports drinking water from the Sierra Nevada for storage in Briones and San Pablo Reservoirs. Much of the upper watershed is in protected East Bay Municipal Utility District (EBMUD) land or EBRPD land. In the lower watershed below San Pablo Dam, San Pablo Creek runs through the suburban towns of El Sobrante and San Pablo before entering the bay at Wildcat Marsh. Until 1895, San Pablo Creek meandered in Wildcat Marsh where it connected to Wildcat Creek.

Average annual rainfall in the watershed ranges from 20 to 28 inches (Rantz 1971). The landscape varies from gently rolling to steeply sloping hills. Grasslands dominate the upper watershed, with patches of northern coastal scrub and chaparral and woodlands of oak, bay, and buckeye on the south-facing slopes. Monterey pine and eucalyptus were planted in the hills.

Land use in the watershed includes grazing, residential, and industrial. Grazing occurs in the Havey Canyon area on EBRPD land. Urban communities within the watershed include Orinda, El Sobrante, San Pablo, Richmond, and North Richmond; the latter three include heavily industrialized portions. The EBMUD Orinda water treatment plant discharges its filter backwash to San Pablo Creek. New residential developments are under construction just north of San Pablo Dam. The 300-acre salt marsh at the mouth of the watershed is bordered by a landfill, a sewage treatment plant, and Chevron's industrial holding ponds and refineries. Potential water quality issues in the watershed resulting from these land uses include erosion, sedimentation, coliforms, eutrophication, and toxicity (see also Section 2.3).

Beneficial uses in the Wildcat Creek watershed, as listed in the Basin Plan (SFBRWQCB 1995), include fish migration and spawning, warm water habitat, wildlife habitat, and non-contact water recreation. Beneficial uses in the San Pablo Creek watershed include all of the uses in Wildcat Creek except fish spawning. Beneficial uses in San Pablo Reservoir and Briones Reservoir include cold and warm water habitat, fish spawning, wildlife habitat, municipal/domestic supply, and water contact and non-contact recreation. Lake Anza and Jewel Lake are small impoundments used for recreational purposes only. Two endangered species live in Wildcat Marsh, at the outlet of both creeks: the salt marsh harvest mouse and the clapper rail. Native rainbow trout were re-introduced to Wildcat Creek in 1983 and are still present.

2.2.5 Suisun Creek Watershed

The Suisun Creek watershed is a 57 square mile drainage basin in Solano and Napa Counties. Suisun Creek originates in the Vaca Mountains of the central California Coast Range and

empties into Suisun Marsh and Suisun Bay in the San Francisco Estuary. The major tributary to Suisun Creek is Wooden Valley Creek, which flows through Wooden Valley before its confluence with Suisun Creek three miles below Lake Curry. Suisun Creek and Wooden Valley Creek are perennial streams, while many of the smaller tributaries, such as White's Creek, dry up in mid-summer. There are two reservoirs in the watershed: Lake Curry, a large reservoir on the main stem of Suisun Creek; and Suisun Reservoir, a small, dammed lake on a minor tributary to Suisun Creek. Lake Curry has been owned and operated by the city of Vallejo since the 1920s; however, the City of Vallejo has not drawn water from Lake Curry since the early 1990s.

Suisun Creek and Wooden Valley Creek flow through broad alluvial valleys. The hills of the central California Coast Range are highly erosive, with large amounts of sediment being transported to streams during severe winter rain events. Annual precipitation in the watershed ranges from 18 to 32 inches (Rantz 1971).

Land use in Suisun Valley is predominantly agricultural, with numerous vineyards and orchards spread throughout Suisun Valley, Gordon Valley, and Wooden Valley. Population centers include the small communities of Rockville, Wooden Valley, and Gordon Valley. Other activities include grazing, and water release and retention from Lake Curry Dam. Potential water quality issues in Suisun Creek watershed include pesticides, erosion, sedimentation, nutrients, and dissolved oxygen.

Beneficial uses in the Suisun Creek watershed listed in the Basin Plan (SFBRWQCB 1995) are cold and warm water habitat, fish spawning and migration, freshwater replenishment, and wildlife habitat. Beneficial uses in Lake Curry include municipal/domestic supply, contact and non-contact water recreation, fish spawning, warm-water habitat, and wildlife habitat.

2.2.6 Arroyo Las Positas Watershed

The Arroyo Las Positas watershed (also Arroyo de Las Positas), a 77 square mile watershed in eastern Alameda County, is one of the main tributaries to Alameda Creek in the Livermore Valley basin. Arroyo Las Positas originates at the confluence of its two major tributaries, Altamont Creek and Arroyo Seco. From this confluence, Arroyo Las Positas flows west until it meets Arroyo Mocho, just east of the city of Pleasanton. Other tributaries to Arroyo Las Positas include Cottonwood Creek, Collier Canyon, and Cayetano Creek. Arroyo Las Positas, Arroyo Mocho, Arroyo Valle, Tassajara Creek, and Alamo Creek together drain the Livermore Valley watershed before entering Arroyo Laguna and Alameda Creek west of San Antonio reservoir. Altamont Creek drains the Altamont Hills and is joined by several unnamed, second- and third-order tributaries that drain Brushy Peak and other hills to the north. The 31 square-mile Arroyo Seco watershed drains the south-eastern mountains of the Livermore Valley. Arroyo Seco flows through recent residential developments and the Lawrence Livermore National Laboratory before meeting Altamont Creek. Numerous smaller, unnamed tributaries feed Arroyo Seco, including a Zone 7 Water Agency flood control channel (P-1) that drains the central part of eastern Livermore Valley.

The average rainfall in the Arroyo Las Positas watershed ranges from 12 to 16 inches per year (Rantz 1971). Arroyo Las Positas, and the lower portions of both Altamont Creek and Arroyo Seco, are perennial streams. Cottonwood Creek, Collier Canyon, and Cayetano Creek are believed to be intermittent, with groundwater-fed base flow occurring from December through May. The upper portions of all of the creeks in the Arroyo Las Positas watershed are dry for much of the year. During the dry season the primary source of water in Arroyo Seco is wastewater from the Lawrence Livermore National Laboratory, which is drained by the Zone 7 Water Agency P-1 flood-control channel. Upstream of the confluence with this channel, Arroyo Seco is dry for much of the year, beginning soon after the last winter storm.

Land use in the watershed includes residential, recreation, and agriculture. The Livermore Valley is one of the fastest-growing regions of the Bay Area, characterized by rapid construction of residential communities, golf courses, and commercial areas in previously undeveloped open space. Urban areas include the City of Livermore, Lawrence Livermore National Laboratory, Springtown Golf Course, Los Positas Golf Course, Livermore sewage disposal site, and recent residential developments. Much of the upper portion of the watershed is rangeland, farmed, or open space, although substantial land development is predicted in the future. Potential water quality issues in Arroyo Las Positas watershed include pesticides, erosion, sedimentation, nutrients, and dissolved oxygen.

The Basin Plan does not differentiate beneficial uses between the tributaries of the Alameda Creek watershed. Therefore the existing beneficial uses in the Alameda Creek watershed are applied to the Arroyo Las Positas watershed (SFBRWQCB 1995): agricultural supply, warm and cold water habitat, groundwater recharge, fish spawning and migration, water contact and non-contact recreation, and wildlife habitat.

2.2.7 Pescadero Creek/Butano Creek Watershed

The watershed of Pescadero Creek and Butano Creek is an 82 square mile drainage basin in western San Mateo County. The headwaters of both Pescadero Creek and Butano Creek are in the steep, forested Santa Cruz Mountains. Pescadero Creek and Butano Creeks converge in Pescadero Marsh, which drains into the Pacific Ocean. Numerous tributaries feed Pescadero Creek including Honsinger Creek, Hoffman Creek, McCormick Creek, Lambert Creek, Slate Creek, and Oil Creek. The main tributary to Butano Creek is Little Butano Creek. Average annual precipitation in the watershed ranges from 20 inches at the coast to 46 inches at the headwaters (Rantz 1971).

The primary urban center is the town of Pescadero, near the mouth of the creek. Pescadero Creek is listed on the 303(d) list of impaired water bodies for excessive sedimentation and siltation. Flooding and agricultural runoff are also issues of concern in the watershed.

The Basin Plan (SFBRWQCB 1995) identifies 10 beneficial uses in the surface waters of the Pescadero Creek watershed: agricultural supply, cold and warm freshwater habitat, fish migration and spawning, municipal/domestic supply, preservation of rare or endangered species,

contact and non-contact water recreation, and wildlife habitat. Pescadero Marsh serves as estuarine habitat.

2.2.8 San Gregorio Creek Watershed

San Gregorio Creek watershed is a 52 square mile drainage basin in western San Mateo County. San Gregorio Creek has its headwaters in the Santa Cruz Mountains, and drains into the Pacific Ocean. Numerous tributaries feed San Gregorio Creek including El Corte de Madera Creek, Clear Creek, Bogess Creek, Harrington Creek, La Honda Creek, Woodruff Creek, Alpine Creek, and Mindego Creek. The watershed is characterized by very unstable rock types and soils, high relief, and high-intensity rain events. Average annual precipitation ranges from 20 inches at the coast to over 40 inches in the headwaters (Rantz 1971).

Land use in the watershed includes logging, road-building, farming, grazing, and residential. The watershed is generally very sparsely populated; the largest urban center in the watershed is the town of La Honda, just upstream of the confluence of La Honda Creek and Alpine Creek. Other communities and residences include the town of San Gregorio, residences along Alpine Creek east of La Honda, and residences along the mainstem of San Gregorio Creek. Septic tanks are used in some areas. Agriculture is concentrated along the floodplains and main channel of San Gregorio Creek. San Gregorio Creek is listed on the 303(d) list of impaired water bodies for excessive sedimentation. In addition to sedimentation, water quality issues include water diversion, coliforms, and pesticides.

The Basin Plan (SFBRWQCB 1995) identifies nine beneficial uses in the surface waters of the San Gregorio Creek watershed: agricultural supply, cold and warm freshwater habitat, fish migration and spawning, preservation of rare and endangered species, water contact and non-contact recreation, and wildlife habitat. In addition to these uses, San Gregorio Creek serves as a source of drinking water for residents.

2.2.9 Stevens Creek/Permanente Creek Watershed

The Stevens Creek and Permanente Creek watersheds are neighboring drainage basins in the western Santa Clara Basin. The watershed of Stevens Creek is a 38 square mile drainage basin, with its headwaters high in the densely forested Santa Cruz Mountains. Permanente Creek, just north of Stevens Creek, originates in the Los Altos Hills, which are covered by chaparral and oak woodland. Permanente Creek flows through the cities of Los Altos and Mountain View, where the channels have been heavily modified, and drains into South San Francisco Bay at Mountain View Slough. Peak flows from Permanente Creek are diverted, via the Permanente Creek Diversion, to Stevens Creek at rates up to 1500 cubic feet per second.

The upper portions of both watersheds drain upland, mountainous or hilly landscapes where human development is largely absent. The lower portions of the streams flow through western Santa Clara Valley, a large flat alluvial valley draining into South San Francisco Bay. Average annual precipitation in the watershed ranges from 14 inches at the bayshore to over 40 inches in the headwaters (Rantz 1971). Stevens Creek Reservoir was built where Stevens Creek emerges

from a deep canyon between Monte Bello Ridge and Table Mountain, and serves to store winter runoff for the recharge of the Santa Clara Groundwater Basin during the summer months (SCBWMI 2003). Swiss Creek, the largest tributary of Stevens Creek, enters the reservoir from the west. The capacity of the reservoir is 3,465 acre-feet; the upstream drainage area is approximately 17 miles.

Land uses in the watershed include mining, urbanization, forests, and parks. A cement plant and rock quarry occupies a large segment of land along much of the length of upper Permanente Creek. Stevens Creek Quarry, at the base of Stevens Creek Reservoir on Stevens Canyon Road, has been in operation since 1932, producing crushed rock for the building industry. Both watersheds are heavily populated, and industrial and commercial areas are also significant. Potential water quality issues include sedimentation, channel modification, increased stream temperature, dissolved oxygen, metals, coliforms, nutrients, pesticides, organic pollutants, and other toxic substances.

Beneficial uses in both watersheds include cold water and wildlife habitat, fish spawning, and contact and non-contact water recreation (SFBRWQCB 1995). Stevens Creek watershed is also listed for warm water habitat, fish migration, and freshwater replenishment. Stevens Creek Reservoir is listed for cold and warm water habitat, wildlife habitat, fish migration and spawning, non-contact water recreation, municipal/domestic supply, and groundwater recharge.

Table of Contents

3 Sampling Design..... 3-1
3.1 Watershed Site Selection 3-1
3.2 Watershed Sampling Parameters 3-1
3.2.1 Tier 1 Sampling..... 3-1
3.2.2 Tier 2 Sampling..... 3-2
3.2.3 Integrator Sites 3-3

List of Tables

Table 3-1. Land-use classifications..... 3-4
Table 3-2. Land uses and potential land use impacts..... 3-5
Table 3-3. Water quality indicators 3-6
Table 3-4. Watershed stations with parameters sampled and land use categories..... 3-6

List of Figures

Figure 3-1. Walker Creek watershed 3-11
Figure 3-2. Lagunitas Creek watershed 3-12
Figure 3-3. San Leandro Creek watershed..... 3-13
Figure 3-4. Wildcat Creek and San Pablo Creek watershed..... 3-14
Figure 3-5. Suisun Creek watershed 3-15
Figure 3-6. Arroyo Las Positas watershed..... 3-16
Figure 3-7. Pescadero Creek and Butano Creek watershed..... 3-17
Figure 3-8. San Gregorio Creek watershed..... 3-18
Figure 3-9. Stevens Creek and Permanente Creek watershed 3-19

3 Sampling Design

3.1 Watershed Site Selection

Watershed sampling sites were selected deterministically, focused on the following types of areas:

1. Above confluences (to characterize tributaries)
2. Areas of minimal human land use (to characterize reference conditions)
3. Areas where previous monitoring data indicates a potential impact
4. Areas where land use changes (to evaluate the potential impacts of particular land use types)
5. Areas with particular beneficial uses
6. Locations of restoration projects and permitted entities

Major types of land-use activity occurring within each watershed were identified (based on field reconnaissance, land use/land coverage maps, and other sources of information) in order to inform the selection of water quality indicators. Land-use types occurring at, or upstream, of each sampling site are presented in Appendix A, and are summarized in Table 3-1. Land uses were then related to the potential for various types of water quality degradation (Table 3-2). These potential water quality impacts were used to select appropriate monitoring tools for each site and watershed (Table 3-3, Table 3-4).

Sampling sites were distributed fairly evenly throughout each watershed in order to (1) assess conditions across the entire watershed and (2) identify locations where water quality conditions exhibit significant changes between adjacent sites. Watershed sampling sites are listed in Table 3-4, and shown on Figures 3-1 through 3-9; site location information is presented in Appendix B.

3.2 Watershed Sampling Parameters

Watershed monitoring was conducted using a tiered approach, designed: 1) to determine overall patterns of water quality in the Region and 2) to answer basic questions concerning the protection of beneficial uses and potential impacts of land use and water management. The assessment of physical and biological indicators, toxicity, and chemistry constitutes a weight-of-evidence approach for characterizing water quality in the San Francisco Bay Region. Parameters sampled at each site during Year 1 and 2 are listed in Table 3-4 with land use and shown on the maps in Figures 3-1 through 3-9.

3.2.1 Tier 1 Sampling

Tier 1 sampling was designed to optimize spatial coverage in determining the basic water quality of watersheds, to identify reference sites, and to complement the Tier 2 evaluation. Tier 1 sampling included bioassessment, physical habitat assessment, and discrete measurements of basic water quality parameters such as temperature, dissolved oxygen, specific conductance, and

pH. However, due to the temporal variability in these measurements, only data from continuous monitoring probes were used to evaluate these parameters.

Biological community and physical habitat indicators tend to integrate the effects of water quality stressors over time, and are sensitive to various physical, chemical and biological influences. Rapid Bioassessment (RBA), an in stream macroinvertebrate sampling procedure, was conducted at all Tier 1 sites to help determine overall patterns of water quality in the Region, and to help assess the protection of beneficial uses for aquatic life. The California Stream Bioassessment Procedure (CDFG 1999) is the standardized RBA protocol for assessing biological and habitat conditions of wadeable streams in California, as adapted from national RBA protocols outlined by the U.S. Environmental Protection Agency (Barbour *et al.* 1999). The CSBP was used to measure stream benthic macroinvertebrate (BMI) community and physical/habitat characteristics, in order to determine stream biological and physical integrity. RBA sampling occurred once at each site, in spring of 2001 or 2002.

3.2.2 Tier 2 Sampling

Tier 2 sampling was conducted at a subset of the Tier 1 sites in order to assess water quality in more detail, to identify potential stressors, and to provide information for the assessment of beneficial use protection in relation to land use. Water quality parameters measured during Tier 2 sampling included the following:

- Continuous field measurements of basic water quality (temperature, DO, pH, and specific conductance)
- Conventional chemistry, including boron, chloride, sulfate, hardness, alkalinity, total dissolved solids (TDS), suspended sediment concentration or total suspended solids (SSC or TSS), total organic carbon (TOC), dissolved organic carbon (DOC), nitrate, nitrite, total Kjeldahl nitrogen (TKN), orthophosphate, total phosphorous, ammonia, chlorophyll-a and pheophytin-a
- Toxicity to three EPA test species: *Ceriodaphnia dubia*, *Selenastrum capricornutum* (= *Raphidocelis subcapitata*), and *Pimephales promelas*
- Trace metals
- Trace organics, including pesticides, polycyclic aromatic hydrocarbons (PAHs), and polychlorinated biphenyls (PCBs)
- Coliform bacteria, including total coliforms, fecal coliforms, and *E. coli*

Tier 2 sampling occurred during three hydrologic cycles: the wet season (January through March), spring/decreasing hydrograph (April through May) and the dry season (June through October). Coliform sampling, however, was conducted during five consecutive weeks within the dry season only to obtain data comparable to water quality standards. Prevailing seasonal conditions determined the timing of monitoring events, which occurred on the following schedule:

| Sampling Year | Tier 2 Sampling Periods | | | | |
|---------------|----------------------------|-----------------|---------------|--------------|-----------------|
| | September/ October 2001 | January 2002 | April 2002 | June 2002 | January 2003 |
| 1 | X | X* | X | X** | |
| 2 | | | X | X | X |

*Toxicity testing was not performed in January 2002.

**Year 1 sampling at sites that were dry in October 2001 was attempted again in June 2002.

To determine temporal variability in basic water quality, continuous field monitoring of temperature, dissolved oxygen (DO), pH, and specific conductance (electrical conductivity compensated to 25°C) was conducted at a subset of sites characterized by bioassessment. As part of our pilot study, continuous monitoring was scheduled at sites for one to two week periods, three to four times per year. Deployments averaged 11 days, ranging from 2-29, in order to determine the optimum period of deployment and because of field conditions.

3.2.3 Integrator Sites

Integrator sites were established at the bottom of each watershed area above tidal influence. These sites were intended to assess contaminant levels in watersheds at the locations where they discharge into receiving waters (e.g. the bay). In addition to physical, chemical, and toxicity monitoring, *Corbicula fluminea* (a freshwater clam) were deployed, retrieved, and analyzed for chemical bioaccumulation. Sediment samples were collected from integrator sites for analysis of grain size, chemistry, and toxicity to the amphipod *Hyalella azteca*. Sediment collection and clam deployments occurred once for each watershed, in the dry season only. Watershed boundaries, site locations, land use, and parameters measured at each site are shown in Figures 3-1 through 3-9.

Table 3-1. Land-use classifications

| category | code | classification | definition |
|-----------------|-------------|------------------------|---|
| 1.1 | COM | Commercial | commercial, not necessarily high density |
| 1.1 | DOE | Department of Energy | military or government energy sites (Lawrence Livermore/SLAC/UTC) |
| 1.1 | IND | Industrial | light or heavy industry |
| 1.1 | ROD | Road | major paved roadway (e.g., freeway) |
| 1.1 | URB | Urban | high density, commercial, high impervious coverage |
| 1.2 | RES | Residential | suburban low density |
| 1.2 | SWR | Sewage treatment plant | POTWs and community septic systems |
| 2.1 | DAM | Dam | water impoundment |
| 2.2 | CHN | Channel modification | concrete channels |
| 2.2 | DIV | Diversion | flow alteration by diversion or addition |
| 3.1 | AGR | Agriculture | crops, excluding trees and grapes |
| 3.1 | NUR | Nursery | intense planting for wholesale and retail; includes greenhouses |
| 3.1 | ORC | Orchards | tree fruit, nuts |
| 3.1 | VIN | Vineyards | grapes |
| 3.2 | GRZ | Grazing | cattle, horse, or any ruminant (historic grazing included) |
| 3.3 | DRY | Dairy | dairy cows (confined animal facility) |
| 3.3 | HRS | Horse | stables, trails |
| 4.1 | MNG | Mining | mercury, pyrite, coal, magnesite, manganese, chromite |
| 4.2 | QRY | Quarry | gravel, sand |
| 4.3 | LOG | Logging | tree removal, roads (historic logging included) |
| 5.1 | GLF | Golf | golf courses |
| 5.1 | PRK | Park | urban or developed park, landscaped, in developed area. |
| 5.2 | OPS | Open Space | wildland park, not developed, likely reference sites |

Table 3-2. Land uses and potential land use impacts

| Land Use | Watersheds | | | | | | | | Potential Land Use Impacts | | | | | | | |
|------------------------|------------|-----------|-------------|---------------------|--------|--------------------|------------------|--------------|----------------------------|---------------|--------------------------|----------------|---------------|-------|---------------------------|-----------------------------|
| | Walker | Lagunitas | San Leandro | Wildcat / San Pablo | Suisun | Arroyo Las Positas | Pescadero/Butano | San Gregorio | Stevens / Permanente | sedimentation | human / aquatic toxicity | eutrophication | human illness | trash | aquatic community effects | aquatic habitat degradation |
| Commercial | | • | • | • | | • | • | • | • | • | • | | • | • | • | • |
| Department of Energy | | | | | | • | | | | • | • | | | | • | • |
| Industrial | | | • | • | | • | | | • | • | • | | | • | • | • |
| Road | | • | • | • | • | • | | | • | • | • | | | • | • | • |
| Urban | | | • | • | | | | | • | • | • | | | • | • | • |
| Residential | • | • | • | • | • | • | • | • | • | • | • | • | • | • | • | • |
| Sewage treatment plant | • | • | | | | | | | | • | • | • | • | | • | • |
| Dam | • | • | • | • | • | | | | • | • | | | | | • | • |
| Channel modification | | | | | | • | | | • | • | | | | | • | • |
| Diversion | | | | • | | | | | • | • | | | | | • | • |
| Agriculture | • | | | | | | • | • | | • | • | • | | | • | • |
| Nursery | | | | • | | | • | | | • | • | • | | | • | • |
| Orchards | | | | | • | | | | | • | • | • | | | • | • |
| Vineyards | | | | | • | | • | | | • | • | • | | | • | • |
| Grazing | • | • | • | • | • | • | • | | | • | • | • | • | | • | • |
| Dairy | • | | | | | | | | | • | | • | • | | • | • |
| Horse | | • | • | | | | | | | • | | • | • | | • | • |
| Mining | • | | | | | | | | | • | • | | | | • | • |
| Quarry | | • | | | | • | | | • | • | | | | | • | • |
| Logging | | | | | | | • | | | • | | | | | • | • |
| Golf | | • | | • | | • | | | | • | • | | | | • | • |
| Park | | • | • | • | | | • | | • | • | • | | • | • | • | • |
| Open Space | • | • | • | • | • | | • | • | • | | | | | | • | • |

Table 3-3. Water quality indicators

| Potential land use impacts | Water quality indicators |
|-----------------------------|---|
| sedimentation | suspended sediment, turbidity physical habitat assessment (as part of RBA) |
| aquatic and human toxicity | toxicity testing pesticides, PCBs, PAHs, metals conventional chemistry and basic water quality |
| eutrophication | conventional chemistry including nutrients, basic water quality (temperature, dissolved oxygen, and pH), chlorophyll-a and pheophytin |
| human illness | coliforms |
| trash | Rapid Trash Assessment |
| aquatic community effects | Rapid Bioassessment (RBA) |
| aquatic habitat degradation | physical habitat assessment (as part of RBA) |

These water quality indicators were used to evaluate potential land use impacts in monitored watersheds.

Table 3-4. Watershed stations with parameters sampled and land use categories

Parameters sampled: **BMI** = benthic macroinvertebrates and physical habitat; **CM** = continuous monitoring of basic water quality field parameters (temperature, DO, pH, SC); **CWQ** = conventional water quality (includes nutrients, chlorophyll, organic carbon); **WMT** = water metals and toxicity; **SCT** = sediment chemistry and toxicity (includes metals and organics); **TMO** = tissue metals and organics (bioaccumulation); **BAC** = coliform bacteria; **LU** = land-use categories. Land-use categories: A = agriculture; G = grazing; M = mixed; OS = open space; RR = rural residential; U = urban. All sites with water chemistry and sediment chemistry also had discrete measurements of basic water quality field parameters (Appendix F). One station marked * (201WLK090) had mercury analyzed, but no other metals or toxicity tests.

| Site Code | BMI | CM | CWQ | WMT | WO | SCT | TMO | BAC | LU |
|-------------------------------|-----------|----------|----------|----------|----------|----------|----------|----------|----|
| Walker Creek Watershed | | | | | | | | | |
| <i>total # sites</i> | 13 | 5 | 4 | 0 | 0 | 1 | 1 | 1 | |
| 201WLK030 | • | | • | | | | | | RR |
| 201WLK050 | • | | | | | | | | M |
| 201WLK090 | | | • | * | | • | • | | M |
| 201WLK100 | • | • | • | | | | | | G |
| 201WLK120 | • | | | | | | | | M |
| 201WLK130 | • | | | | | | | | M |
| 201WLK140 | • | • | • | | | | | | G |
| 201WLK160 | • | • | | | | | | | M |
| 201WLK162 | | | | | | | | • | RR |
| 201WLK170 | • | | | | | | | | G |
| 201WLK180 | • | • | | | | | | | M |
| 201WLK190 | • | | | | | | | | G |
| 201WLK200 | • | • | | | | | | | M |
| 201WLK230 | • | | | | | | | | G |
| 201WLK240 | • | | | | | | | | G |

| Site Code | BMI | CM | CWQ | WMT | WO | SCT | TMO | BAC | LU |
|------------------------------------|-----------|----------|----------|----------|----------|----------|----------|----------|----|
| Lagunitas Creek Watershed | | | | | | | | | |
| <i>total # sites</i> | 24 | 6 | 8 | 3 | 3 | 2 | 3 | 5 | |
| 201LAG040 | • | | • | • | • | | • | | M |
| 201LAG050 | • | • | | | | | | | M |
| 201LAG060 | • | | | | | | | | G |
| 201LAG075 | • | | | | | | | | G |
| 201LAG085 | • | • | | | | | | | G |
| 201LAG100 | • | | | | | | | | G |
| 201LAG115 | | | | | | | | • | M |
| 201LAG120 | | | | | | • | | • | G |
| 201LAG130 | • | | • | • | • | | • | | M |
| 201LAG150 | • | | • | | | | | | M |
| 201LAG160 | • | • | • | | | | | | RR |
| 201LAG165 | • | • | | | | | | | M |
| 201LAG170 | • | | | | | | | | M |
| 201LAG180 | • | | | | | | | | M |
| 201LAG185 | | | | | | | | • | M |
| 201LAG190 | • | • | • | | | | | | M |
| 201LAG210 | • | | • | | | | | | M |
| 201LAG220 | • | | | | | | | | M |
| 201LAG230 | | | | | | | | • | RR |
| 201LAG240 | • | | | | | | | | RR |
| 201LAG270 | • | • | • | • | • | • | • | | RR |
| 201LAG290 | • | | | | | | | | RR |
| 201LAG300 | • | | | | | | | | RR |
| 201LAG320 | • | | • | | | | | | M |
| 201LAG330 | • | | | | | | | | OS |
| 201LAG335 | • | | | | | | | | OS |
| 201LAG380 | • | | | | | | | | OS |
| 201LAG390 | • | | | | | | | • | OS |
| San Leandro Creek Watershed | | | | | | | | | |
| <i>total # sites</i> | 11 | 7 | 4 | 2 | 1 | 1 | 0 | 5 | |
| 204SLE030 | • | | • | • | • | • | | | U |
| 204SLE050 | • | • | | | | | | | U |
| 204SLE070 | • | • | | | | | | • | U |
| 204SLE090 | • | • | | | | | | • | M |
| 204SLE170 | • | • | | | | | | • | M |
| 204SLE180 | • | • | | | | | | | OS |
| 204SLE190 | • | • | • | | | | | • | RR |
| 204SLE200 | • | | | | | | | • | RR |
| 204SLE210 | • | • | • | | | | | | U |
| 204SLE220 | • | | | | | | | | G |
| 204SLE230 | • | | • | • | | | | | OS |
| San Pablo Creek Watershed | | | | | | | | | |
| <i>total # sites</i> | 13 | 5 | 5 | 4 | 1 | 1 | 1 | 2 | |
| 206SPA020 | • | | • | • | • | • | • | | U |
| 206SPA050 | • | • | | | | | | | U |

| Site Code | BMI | CM | CWQ | WMT | WO | SCT | TMO | BAC | LU |
|-------------------------------------|----------|----------|----------|----------|----------|----------|----------|----------|----|
| 206SPA060 | | | | | | | | • | U |
| 206SPA070 | • | • | • | • | | | | | U |
| 206SPA100 | • | | | | | | | | U |
| 206SPA110 | • | | | | | | | | U |
| 206SPA130 | • | | | | | | | | M |
| 206SPA140 | • | | | | | | | | M |
| 206SPA150 | • | | • | | | | | • | G |
| 206SPA160 | • | | | | | | | | G |
| 206SPA170 | • | | | | | | | | M |
| 206SPA200 | • | • | • | • | | | | | U |
| 206SPA220 | • | • | • | • | | | | | U |
| 206SPA240 | • | • | | | | | | | U |
| Wildcat Creek Watershed | | | | | | | | | |
| <i>total # sites</i> | 8 | 3 | 1 | 1 | 1 | 1 | 0 | 2 | |
| 206WIL020 | | • | • | • | • | • | | | U |
| 206WIL030 | • | | | | | | | | U |
| 206WIL050 | • | | | | | | | | U |
| 206WIL060 | • | | | | | | | | U |
| 206WIL070 | • | • | | | | | | • | M |
| 206WIL100 | • | • | | | | | | | M |
| 206WIL130 | • | | | | | | | • | M |
| 206WIL170 | • | | | | | | | | M |
| 206WIL180 | | | • | | | | | | M |
| 206WIL190 | • | | | | | | | | M |
| Suisun Creek Watershed | | | | | | | | | |
| <i>total # sites</i> | 9 | 7 | 8 | 4 | 3 | 1 | 1 | 0 | |
| 207SUI010 | • | | • | • | • | | | | M |
| 207SUI020 | • | • | • | • | • | • | • | | A |
| 207SUI050 | • | • | | | | | | | A |
| 207SUI060 | • | | • | • | | | | | A |
| 207SUI090 | | • | | | | | | | M |
| 207SUI110 | • | • | • | • | • | | | | M |
| 207SUI125 | | • | • | | | | | | M |
| 207SUI130 | • | | • | | | | | | M |
| 207SUI180 | • | • | • | | | | | | M |
| 207SUI185 | | • | | | | | | | M |
| 207SUI210 | • | | • | | | | | | A |
| 207SUI260 | • | | | | | | | | G |
| Arroyo Las Positas Watershed | | | | | | | | | |
| <i>total # sites</i> | 7 | 5 | 4 | 3 | 1 | 1 | 1 | 0 | |
| 204ALP010 | • | • | • | • | • | • | • | | U |
| 204ALP040 | • | • | | | | | | | U |
| 204ALP070 | • | | | | | | | | U |
| 204ALP080 | • | • | | | | | | | U |
| 204ALP100 | • | | • | • | | | | | U |
| 204ALP105 | | • | | | | | | | U |
| 204ALP110 | • | | • | • | | | | | U |

| Site Code | BMI | CM | CWQ | WMT | WO | SCT | TMO | BAC | LU |
|-------------------------------------|-----------|----------|----------|----------|----------|----------|----------|----------|----|
| 204ALP140 | • | | • | | | | | | M |
| 204ALP150 | | • | | | | | | | M |
| Butano Creek Watershed | | | | | | | | | |
| <i>total # sites</i> | 5 | 3 | 2 | 1 | 1 | 1 | 1 | 2 | |
| 202BUT010 | • | • | • | • | • | • | • | | A |
| 202BUT020 | • | | • | | | | | | A |
| 202BUT030 | • | • | | | | | | | RR |
| 202BUT040 | • | | | | | | | • | OS |
| 202BUT050 | • | • | | | | | | • | OS |
| Pescadero Creek Watershed | | | | | | | | | |
| <i>total # sites</i> | 17 | 9 | 4 | 2 | 2 | 1 | 1 | 6 | |
| 202PES050 | • | | • | • | • | • | • | | M |
| 202PES060 | • | • | | | | | | • | M |
| 202PES070 | • | | • | • | • | | | | A |
| 202PES080 | • | | | | | | | | A |
| 202PES095 | • | | | | | | | | A |
| 202PES100 | • | • | | | | | | | RR |
| 202PES105 | | • | | | | | | | RR |
| 202PES120 | • | | | | | | | | RR |
| 202PES134 | | | | | | | | • | RR |
| 202PES135 | | | | | | | | • | RR |
| 202PES140 | • | • | • | | | | | | M |
| 202PES150 | • | • | | | | | | • | RR |
| 202PES160 | • | | | | | | | | M |
| 202PES170 | • | • | | | | | | | M |
| 202PES180 | • | • | | | | | | | OS |
| 202PES190 | • | • | • | | | | | | OS |
| 202PES193 | | | | | | | | • | OS |
| 202PES194 | | | | | | | | • | OS |
| 202PES200 | • | | | | | | | | OS |
| 202PES210 | • | | | | | | | | OS |
| 202PES230 | • | | | | | | | | M |
| 202PES240 | • | • | | | | | | | OS |
| San Gregorio Creek Watershed | | | | | | | | | |
| <i>total # sites</i> | 11 | 6 | 4 | 2 | 2 | 1 | 1 | 2 | |
| 202SGR010 | • | • | • | • | • | • | • | | M |
| 202SGR020 | | • | | | | | | | RR |
| 202SGR030 | • | | | | | | | | RR |
| 202SGR040 | • | • | • | | | | | | M |
| 202SGR060 | • | | | | | | | | RR |
| 202SGR075 | • | | | | | | | | RR |
| 202SGR079 | | | | | | | | • | RR |
| 202SGR080 | • | | • | • | • | | | | RR |
| 202SGR090 | • | • | • | | | | | | RR |
| 202SGR100 | | • | | | | | | • | RR |
| 202SGR110 | • | | | | | | | | RR |
| 202SGR120 | • | • | | | | | | | RR |

| Site Code | BMI | CM | CWQ | WMT | WO | SCT | TMO | BAC | LU |
|-----------------------------------|------------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|----|
| 202SGR130 | • | | | | | | | | RR |
| 202SGR150 | • | | | | | | | | RR |
| Permanente Creek Watershed | | | | | | | | | |
| <i>total # sites</i> | 7 | 2 | 2 | 2 | 2 | 1 | 1 | 1 | |
| 205PER010 | • | • | • | • | • | • | • | | U |
| 205PER020 | • | | | | | | | | U |
| 205PER030 | • | | | | | | | | U |
| 205PER040 | • | | | | | | | | U |
| 205PER050 | • | • | | | | | | | U |
| 205PER070 | • | | • | • | • | | | | M |
| 205PER080 | • | | | | | | | • | M |
| Stevens Creek Watershed | | | | | | | | | |
| <i>total # sites</i> | 8 | 3 | 3 | 2 | 2 | 1 | 1 | 4 | |
| 205STE020 | • | • | • | • | • | • | • | • | U |
| 205STE030 | • | | | | | | | | U |
| 205STE040 | • | | | | | | | | U |
| 205STE060 | • | • | • | • | • | | | | U |
| 205STE070 | • | • | | | | | | • | M |
| 205STE080 | | | | | | | | • | RR |
| 205STE090 | | | | | | | | • | RR |
| 205STE100 | • | | • | | | | | | RR |
| 205STE110 | • | | | | | | | | RR |
| 205STE120 | • | | | | | | | | RR |
| Year 1 & 2 Total | 125 | 61 | 45 | 26 | 19 | 13 | 12 | 30 | |

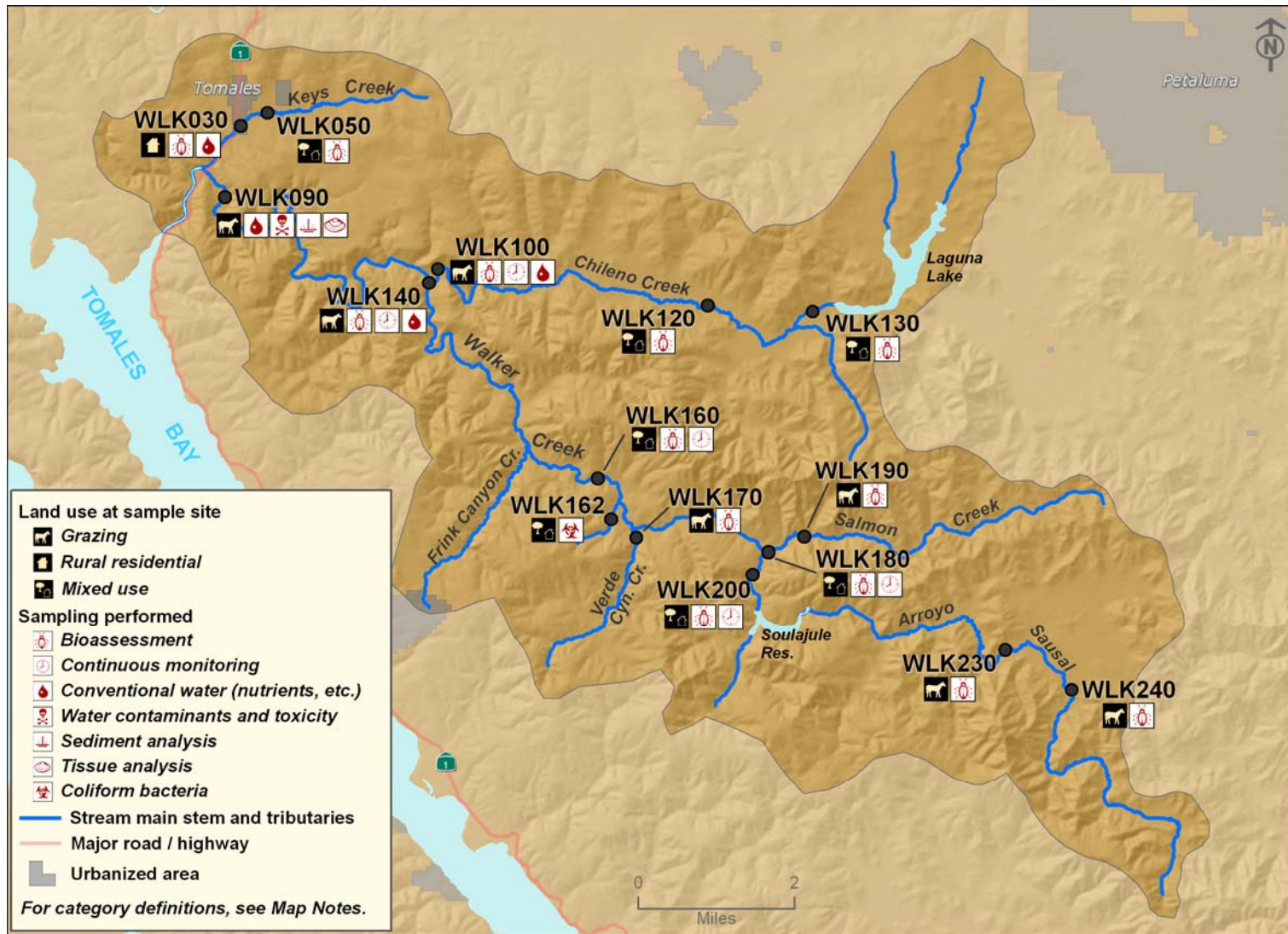


Figure 3-1. Walker Creek watershed

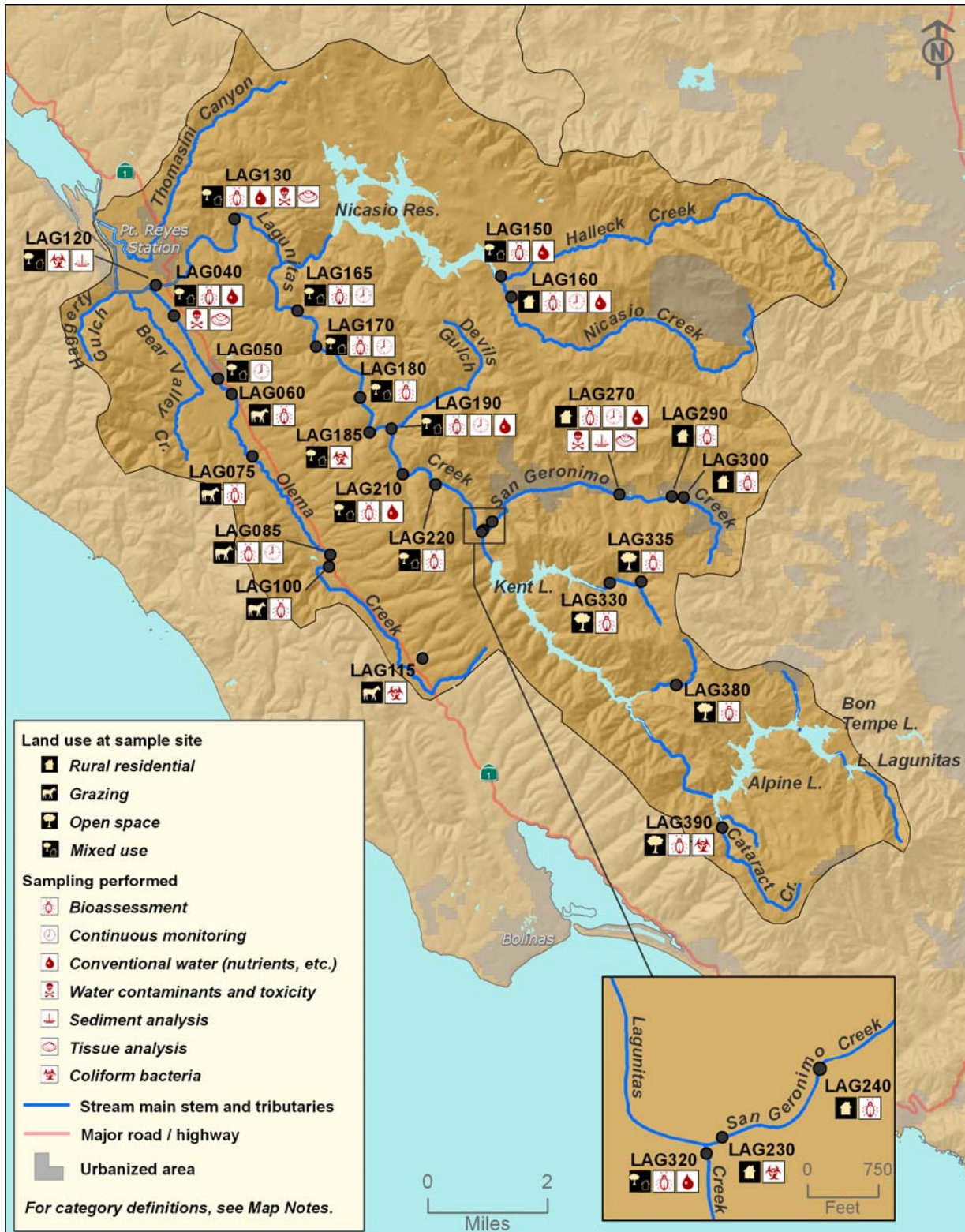


Figure 3-2. Lagunitas Creek watershed

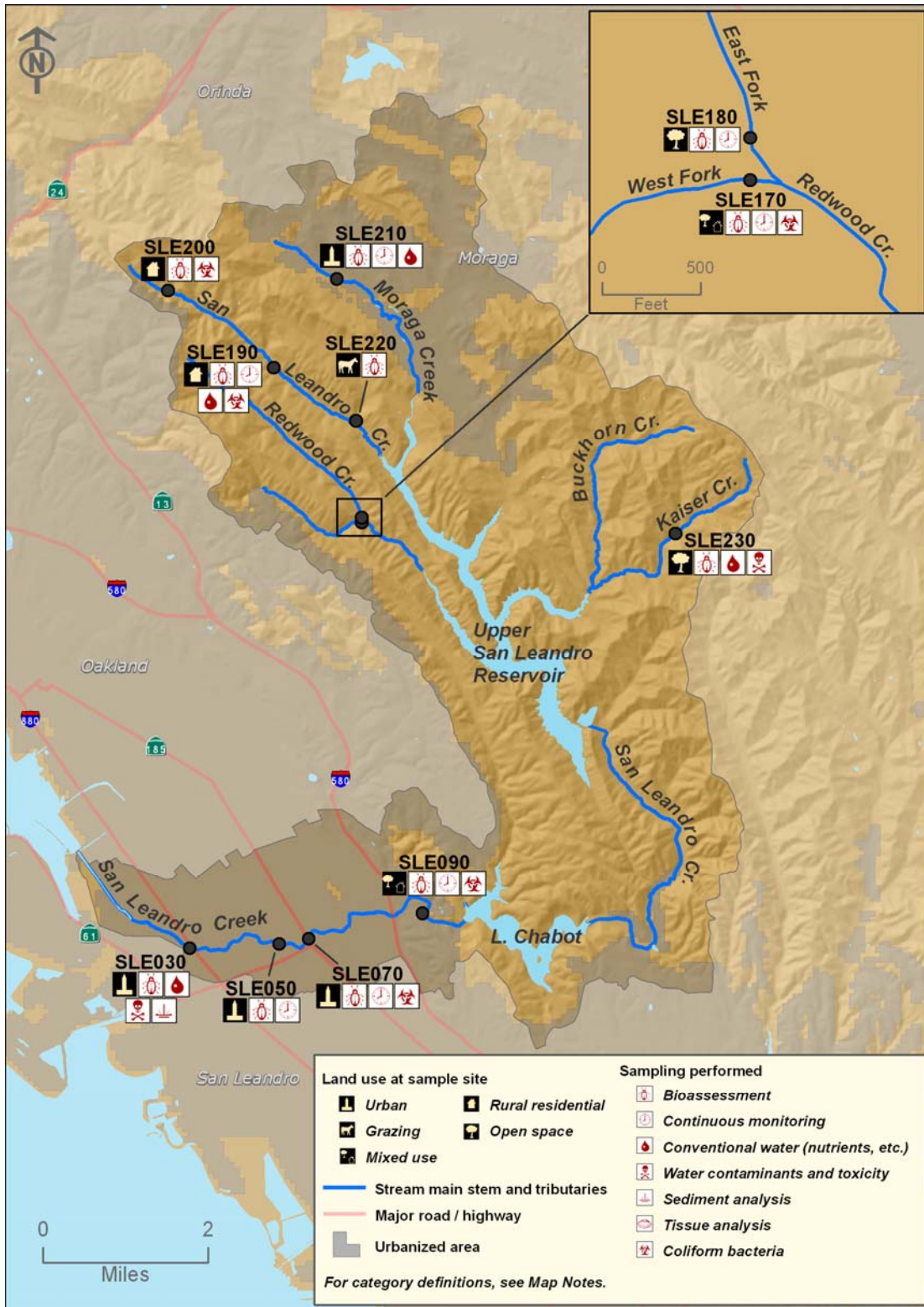


Figure 3-3. San Leandro Creek watershed

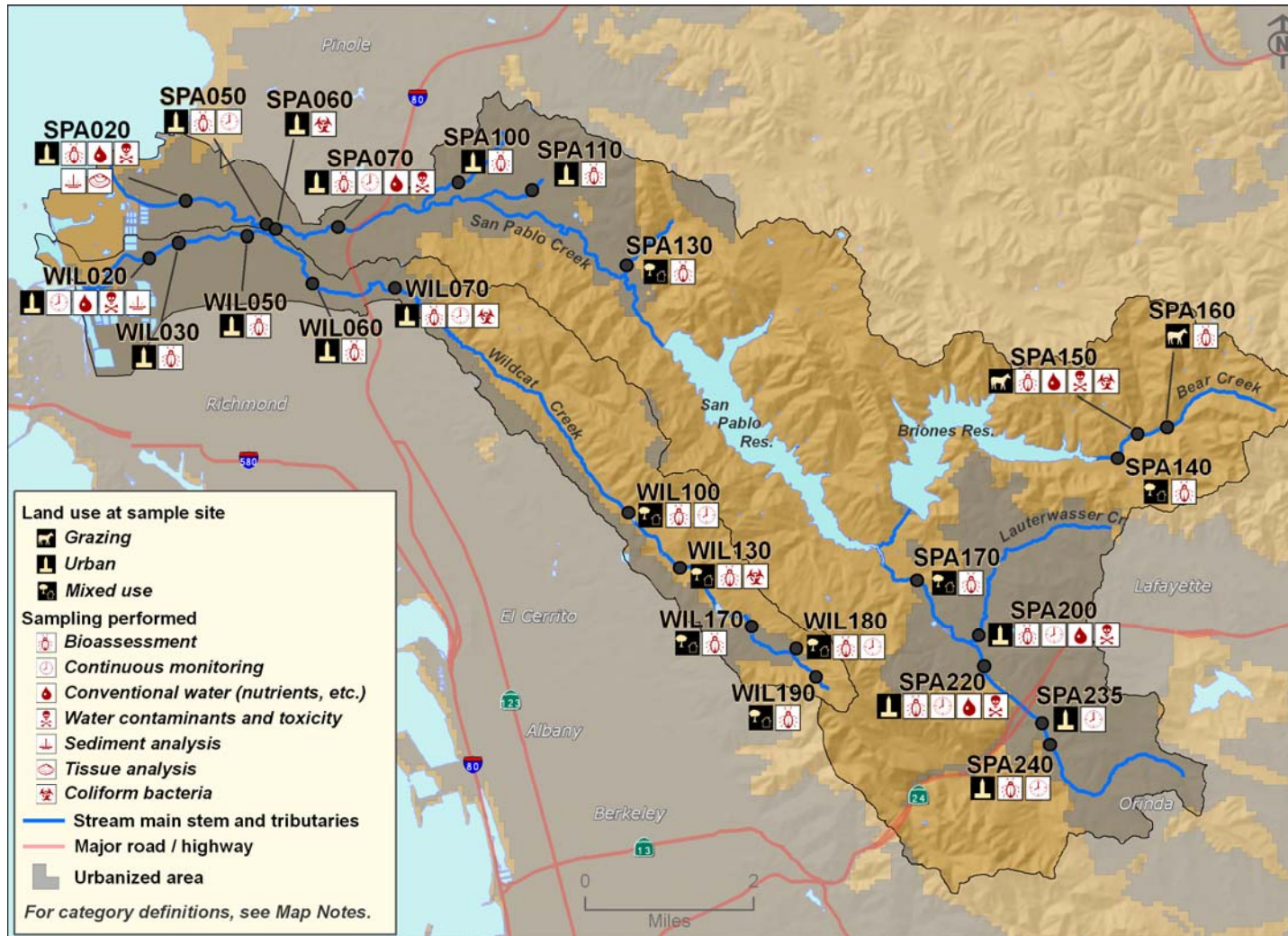


Figure 3-4. Wildcat Creek and San Pablo Creek watershed

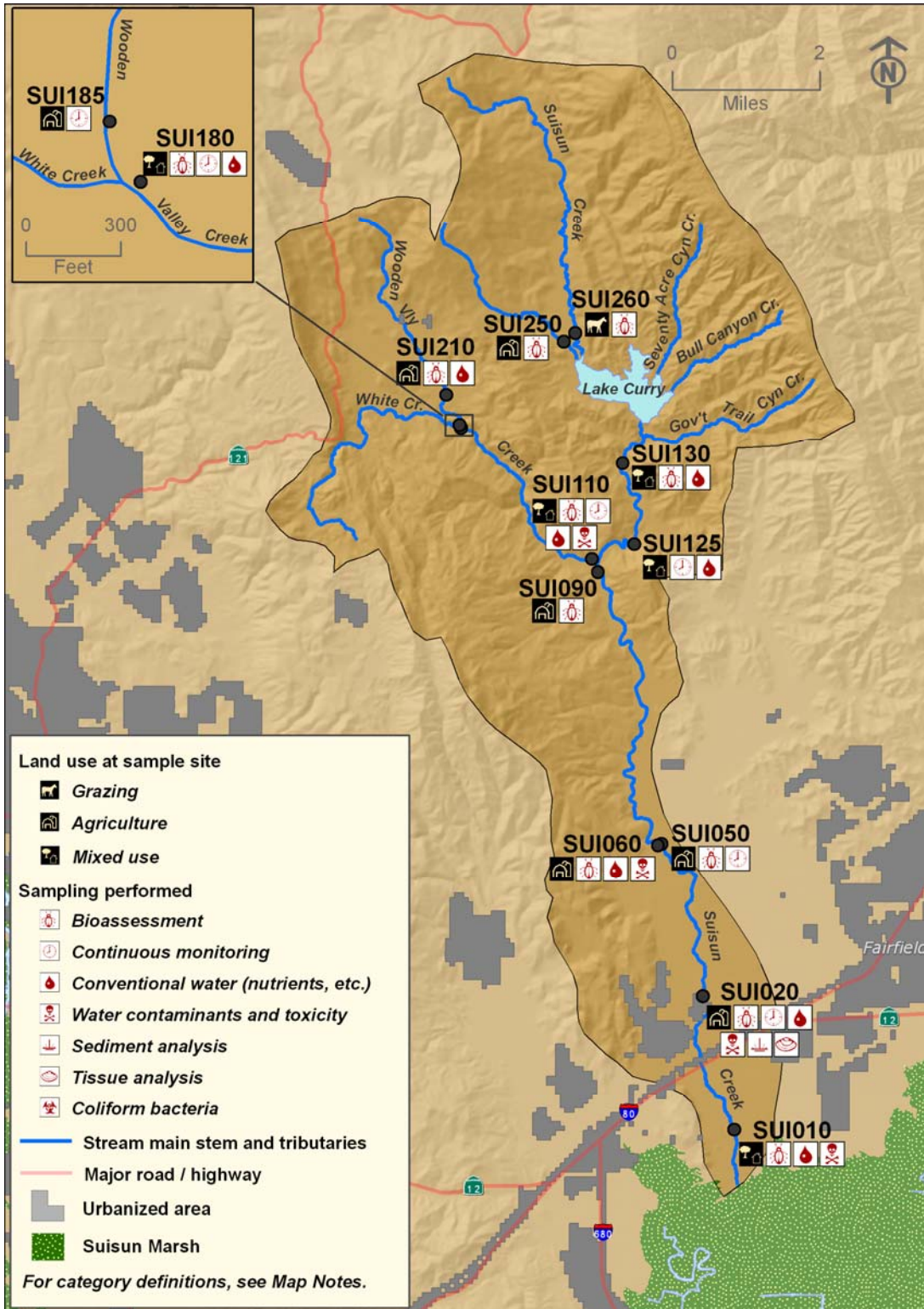


Figure 3-5. Suisun Creek watershed

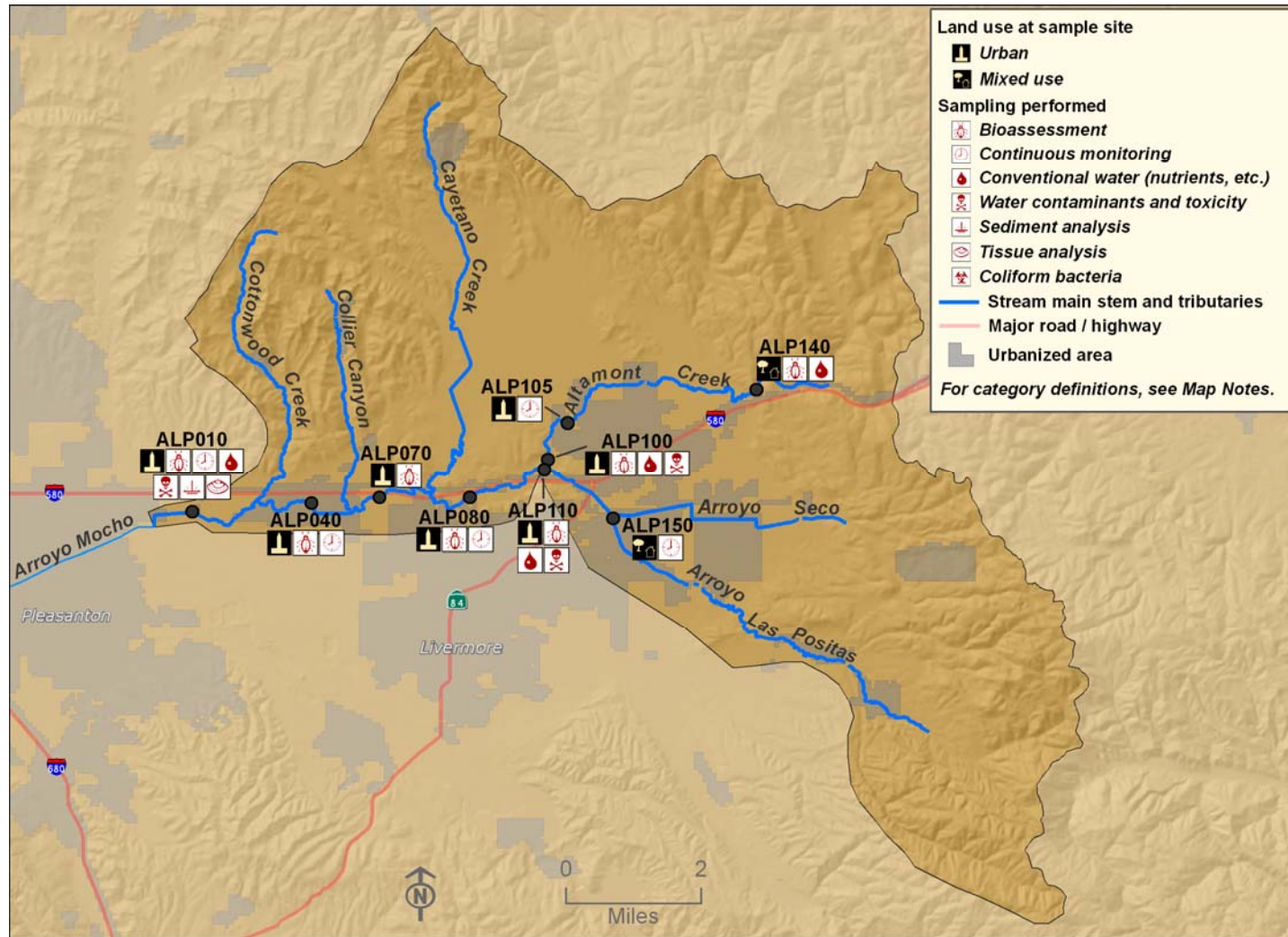


Figure 3-6. Arroyo Las Positas watershed

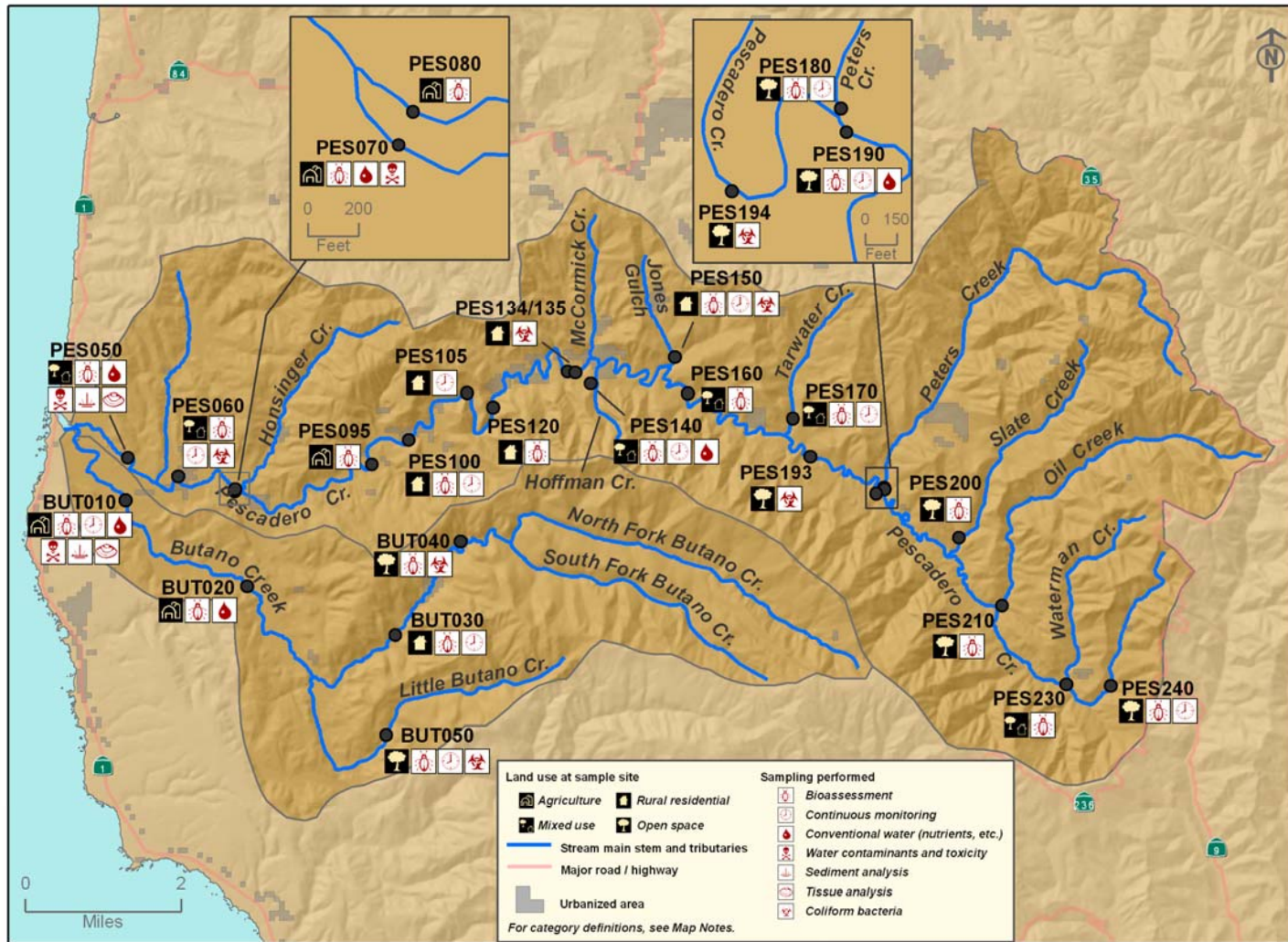


Figure 3-7. Pescadero Creek and Butano Creek watershed

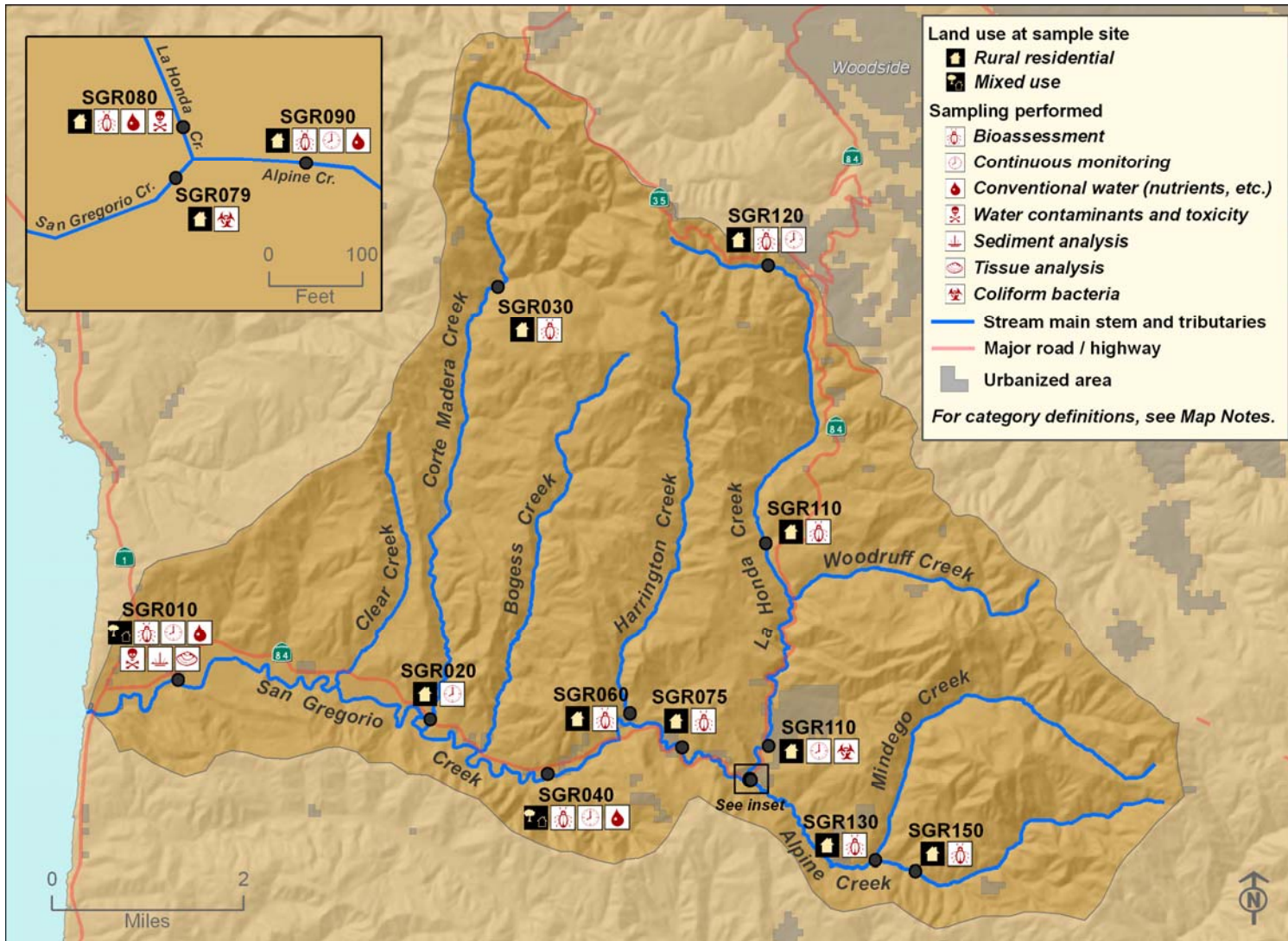


Figure 3-8. San Gregorio Creek watershed

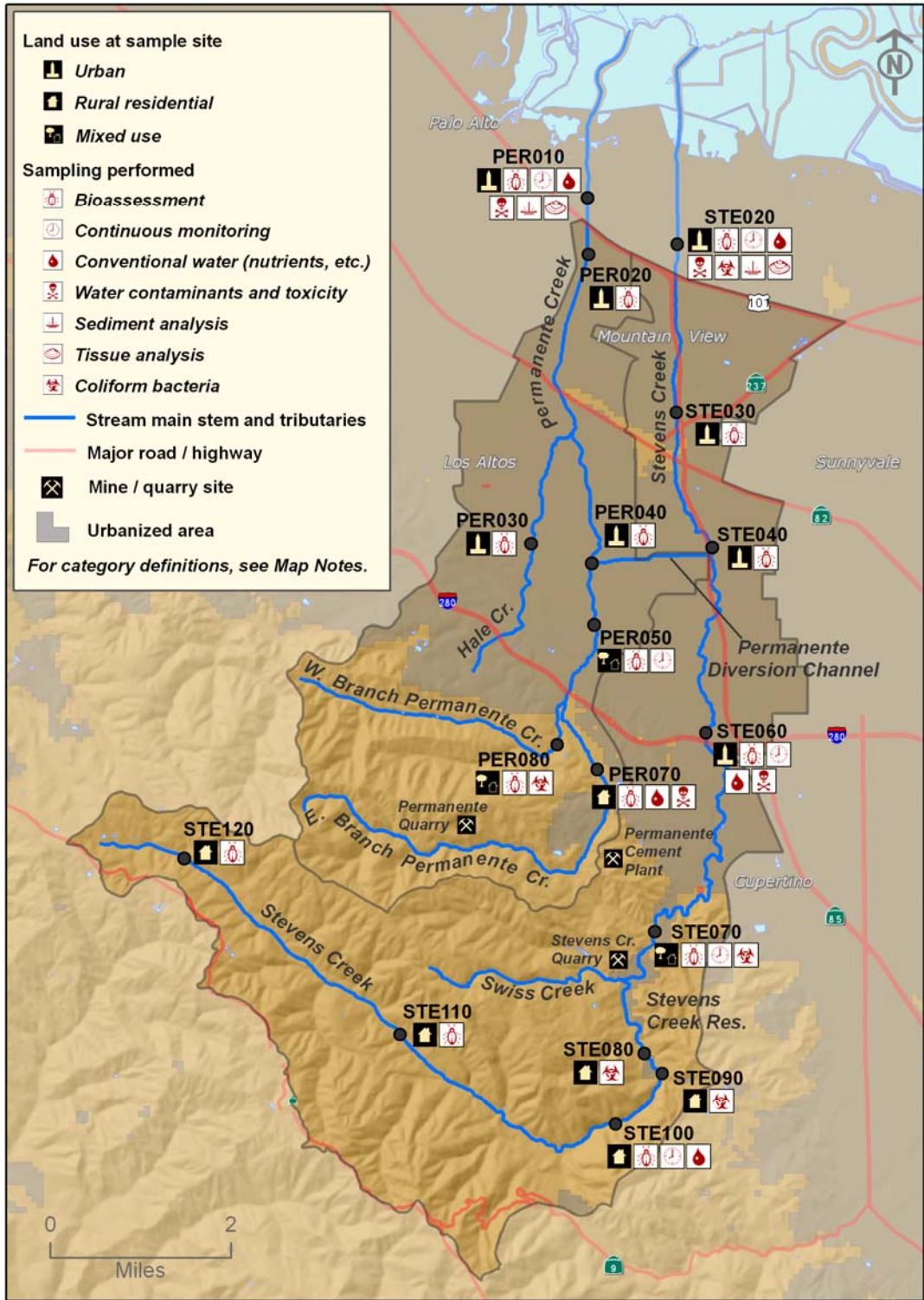


Figure 3-9. Stevens Creek and Permanente Creek watershed

Table of Contents

| | | |
|-------|--|------|
| 4 | Methods..... | 4-1 |
| 4.1 | Rapid Bioassessment | 4-1 |
| 4.1.1 | Background and Rationale for Bioassessment..... | 4-1 |
| 4.1.2 | Field Sampling | 4-2 |
| 4.1.3 | Laboratory Analysis..... | 4-3 |
| 4.2 | Water Chemistry, Basic Water Quality, and Toxicity | 4-3 |
| 4.2.1 | Water Sampling | 4-3 |
| 4.2.2 | Field Measurements of Basic Water Quality | 4-3 |
| 4.2.3 | Chemical Analyses..... | 4-4 |
| 4.2.4 | Toxicity Testing..... | 4-4 |
| 4.3 | Coliforms in Water | 4-4 |
| 4.4 | Sediment Chemistry and Toxicity | 4-5 |
| 4.4.1 | Sediment Sampling | 4-5 |
| 4.4.2 | Chemical Analyses..... | 4-5 |
| 4.4.3 | Toxicity Testing | 4-6 |
| 4.5 | Bioaccumulation Studies with <i>Corbicula</i> | 4-6 |
| 4.6 | Approach to Data Analysis and Interpretation..... | 4-6 |
| 4.6.1 | Analysis of Bioassessment Data | 4-7 |
| 4.6.2 | Basic Water Quality | 4-9 |
| 4.6.3 | Water, Sediment, and Clam Tissue Chemistry | 4-9 |
| 4.6.4 | Toxicity | 4-10 |
| 4.6.5 | Coliforms | 4-10 |

List of Tables

| | | |
|------------|--|------|
| Table 4-1. | Analyte, laboratory, and method for water, sediment, and clam tissue chemistry... 4-11 | |
| Table 4-2. | Water quality thresholds for aquatic life | 4-15 |
| Table 4-3. | Sediment quality guidelines (SQGs) for aquatic life..... | 4-16 |
| Table 4-4. | Biological metrics..... | 4-17 |

4 Methods

A summary of sampling, laboratory analysis, and data analysis procedures is provided in this section. All sampling, chain of custody procedures, and laboratory analyses were performed according to objectives and procedures outlined in the SWAMP QAMP and its appendices (Puckett 2002), unless otherwise noted. Regional Board contacts, as well as participants contracted for field and laboratory studies, are also provided in the QAMP (Puckett 2002). All field activities occurred at GPS-located sites.

4.1 Rapid Bioassessment

4.1.1 Background and Rationale for Bioassessment

Benthic macroinvertebrates are commonly used to assess environmental conditions in freshwater ecosystems around the world, and are the preferred biological assemblage for bioassessment in streams in California (Harrington and Born 1999). Benthic macroinvertebrate assemblages in streams are dominated by insects (Class: Insecta), although other invertebrates such as aquatic worms (Class: Oligochaeta) and snails (Class: Gastropoda) are common. Major orders of benthic insects include mayflies (Ephemeroptera), stoneflies (Plecoptera), caddisflies (Trichoptera), true flies (Diptera), aquatic beetles (Coleoptera), and dragonflies and damselflies (Odonata). Most aquatic insects live in the benthos as larvae and have terrestrial adult life stages, although some groups have complete aquatic life cycles.

Benthic macroinvertebrates are extremely diverse and exhibit a wide range of tolerance to pollution and habitat alteration, making them ideal indicators of water quality (Rosenberg and Resh 1993). Additionally, unlike water chemistry samples, benthic assemblages integrate environmental conditions over time, as the lifespan of aquatic invertebrates ranges from several months to several years. In general, members of the insect orders Ephemeroptera, Plecoptera, and Trichoptera (hereafter EPT) are intolerant of degraded water quality and stream conditions relative to many other macroinvertebrates. Other taxonomic groups, such as Oligochaetes and many taxa of Diptera, respond favorably to certain types of pollution such as nutrient enrichment and fine sediment deposition.

Benthic macroinvertebrates have been used throughout the world to set management objectives and guide policy decisions for freshwater ecosystems. For example, in Great Britain, Canada, New Zealand, and Australia, multivariate statistical models that use correlations between habitat variables and species occurrences have been used to predict the composition of macroinvertebrate assemblages in streams (Wright *et al.* 1984). Differences between the predicted and expected assemblages are representative of the amount of environmental degradation due to altered water quality and habitat conditions. Using a different approach, the California Department of Fish and Game's Aquatic Bioassessment Laboratory (DFG ABL) has developed an Index of Biological Integrity (IBI) for coastal Southern California (Ode *et al.* 2005). An IBI is an assessment tool for ranking the biological condition of stream systems using observed relationships between environmental degradation and the composition of the benthic

community. The DFG ABL is planning to develop additional region-specific IBIs for the Sierra Nevada, North Coast, and Central Valley of California (Ode 2005).

In the San Francisco Bay Area, aquatic bioassessment with benthic macroinvertebrates is used by many watershed monitoring programs, including municipal stormwater monitoring programs, citizen monitoring groups, and government agencies. Despite the high level of interest in using bioassessment for water quality monitoring and assessment, an assessment tool specific to the Bay Area has not yet been developed. The Bay Area Macroinvertebrate Bioassessment Information Network (BAMBI), organized in 2001, is a group of individuals representing government agencies, private consulting firms, watershed groups, and researchers whom are interested in furthering the utility of bioassessment in the Bay Area. The Regional Water Quality Control Board is working closely with BAMBI in order to help develop water quality assessment and management tools on a collaborative basis. Until information on minimally disturbed conditions and natural variability is available, only coarse relative comparisons between monitoring sites are possible. Accordingly, one of the major goals of this report is to analyze bioassessment data collected by the Regional Board with a focus on identifying assessment tools to maximize the utility of aquatic bioassessment for monitoring watershed conditions.

4.1.2 Field Sampling

Bioassessment sampling was conducted at 133 sites in April 2001 and April 2002. Six of those sites were sampled by the National Park Service in Olema Creek using a point-source method; those results were used when metrics were comparable. At the other 127 sites, bioassessment monitoring was performed with a consistent sampling protocol, the 1999 version of the California Stream Bioassessment Procedure (CSBP), developed and conducted by the DFG-ABL (Harrington 1999, Puckett 2002). At each sampling site, a reach of stream containing riffles was selected for sampling. Riffles are the most taxonomically diverse microhabitats within streams, and can be used to determine the relative community composition of streams. Three riffles within each stream reach were randomly selected for sampling. A D-shaped kick net (0.5 mm mesh) was used to collect macrofaunal samples at three sites within each riffle. A 1 by 2-foot portion of substrate upstream of the kick-net was located and disturbed to approximately 4-6 inches in depth for 1-3 minutes at each site. The three site collections per riffle were composited to create one sample that was sieved to 0.5 mm and preserved in 95 percent ethanol.

In addition to the benthic sampling, field crews performed a physical habitat (P-HAB) assessment at each site. The physical habitat assessment component of the CSBP is a qualitative visual assessment of various habitat attributes, modeled after the EPA's Rapid Bioassessment Protocol (Barbour *et al.* 1999a, 1999b). This consisted of generating scores for the following habitat parameters: epifaunal substrate, embeddedness, velocity and depth regimes, sediment deposition, channel flow, channel alteration, riffle frequency, bank vegetation, bank stability, and riparian zone.

4.1.3 Laboratory Analysis

All samples were sorted and identified by the DFG ABL in accordance with the CSBP and the California Aquatic Macroinvertebrate Laboratory Network (CAMLnet) Standard Taxonomic Effort (STE). Macrofaunal samples were randomly subsampled and sorted to obtain 300 individuals per sample. These individuals were stored in an ethanol-glycerin solution, identified to genus or the lowest possible taxonomic unit, and enumerated. In general, insects were identified to genus, except chironomid midges which were identified to tribe/sub-family. Non-insects were identified to order, family, or genus, depending on the taxonomic group (see the CAMLnet STE: <http://www.dfg.ca.gov/cabw/camlnetste.pdf>).

4.2 Water Chemistry, Basic Water Quality, and Toxicity

4.2.1 Water Sampling

Water sampling was conducted by Marine Pollution Studies Laboratory (MPSL)-DFG at Tier 2 sampling sites. Grab samples were collected in appropriately pre-cleaned containers (amber glass, polyethylene or Teflon™, as required). Samples were collected in well-mixed stream sections by uncapping, filling and recapping the container just below the water surface. Samples collected for dissolved metals analyses were collected into a clean syringe and syringe-filtered into the sample container. Some chlorophyll-a samples were filtered in the field and frozen prior to analysis. Techniques used to insure collection of a valid water sample included sampling before other field activities, use of appropriate gloves, and implementation of the “clean hands-dirty hands” technique. Once collected, samples were preserved as needed, labeled, stored at a nominal 4°C, and shipped with appropriate chain-of-custody and handling procedures to the analytical laboratories. Field data sheets were completed for each sampling event to document conditions and sampling details. Regular field duplicates (for all sample types) and field blanks (for mercury, metals, and DOC) were also collected for quality control. For details on water sampling, see the QAMP (Puckett 2002) and Appendix G.

4.2.2 Field Measurements of Basic Water Quality

During Tier 2 sampling, field crews from MPSL-DFG also performed discrete measurements of temperature, DO, pH, specific conductance, salinity, velocity, and turbidity. All measurements were made just below the water surface and according to protocols outlined in the QAMP (Puckett 2002, Appendix F). Bioassessment field crews also performed discrete field measurements of water quality; however, these data were not evaluated in this report because of the problem with temporal variability inherent in this kind of measurement.

Regional Board Staff used YSI 6600 sondes with multi-parameter probes at selected Tier 1 sites to continuously monitor temperature, DO, pH, and specific conductance. Instruments were secured to stationary objects, camouflaged if necessary, and deployed for an average of 11 days (ranging from 2 to 29 days), recording water quality measurements every 15 minutes. Data were

collected at selected sites in each watershed during each hydrologic season. For details on continuous monitoring of water quality, see the QAMP (Puckett 2002) and Appendix D.

4.2.3 Chemical Analyses

Water samples were analyzed for organics, TOC, DOC, total and dissolved metals, sulfate, chloride, boron, hardness, alkalinity, TDS, SSC or TSS, TOC, DOC, nitrate, nitrite, TKN, orthophosphate, total phosphorous, ammonia, chlorophyll a, and pheophytin a. Methods used in these analyses are presented in Table 4-1. Water chemistry data were assessed using the water quality guidelines in Table 4-2, as described in Section 4.6.

Unionized ammonia was calculated from total ammonia using the equations of Whitfield (1974). Field-collected temperature and pH measurements were entered into the equations with laboratory-measured total ammonia. Where necessary, pK values were interpolated from known values.

4.2.4 Toxicity Testing

Water toxicity testing was performed at the UC Davis Marine Pollution Studies Laboratory at Granite Canyon (UCD-GC) to assess the toxicity of waters collected from watershed sites. Three standardized test protocols were used: *Pimephales promelas* 7-day survival and growth, *Ceriodaphnia dubia* 7-day survival and reproduction, and *Selenastrum capricornutum* (*Raphidocelis subcapitata*) 96-hour growth test. Water quality was monitored daily, and temperature monitored continuously. All tests were conducted according to U.S. EPA (1994) guidelines. Negative controls were included with each test batch, and monthly reference toxicant tests (positive controls) with cupric chloride were performed for each species. Approximately every one in ten samples tested was a blind field duplicate. Details of toxicity testing methods can be found in the SWAMP QAMP (Puckett 2002) and Appendix J.

4.3 Coliforms in Water

A coliform assessment was conducted for screening purposes at sites with the potential for recreational water contact. Each site was sampled five times within a 30-day period to satisfy sampling requirements for evaluation of bacteria in water as stipulated in the Basin Plan (SFBRWQCB 1995) and U.S. EPA guidance (U.S. EPA 1986). Sampling took place during the summer months, when water contact recreation is most frequent. Samples from 2001 were collected between August 7 and September 4 at 15 sites; samples from 2002 were collected between July 22 and August 18 at a different set of 15 sites.

Sampling was performed according to U.S. EPA methods for volunteer stream monitoring (U.S. EPA 1997). Samples were collected in lab-sealed, sterile, 100-ml polyethylene bottles, using techniques to exclude bacteria associated with the surface film and bottom sediment. Filled sample bottles were placed in zip-lock bags and stored in a cooler with wet ice. Most samples

were delivered to the laboratory within 6 hours of collection; the longest delivery time was 9:35 hours. All samples were held at the laboratory at 5°C, and analyzed within 24 hours of collection. Field replicates were also collected: in 2001, one duplicate sample was collected; in 2002, six triplicate samples were collected (two during each of three weeks). Replicates were collected simultaneously and adjacent to one another. For details on coliform sample collection, see the QAMP (Puckett 2002) and Appendix K.

Water samples were analyzed in accordance with Standard Method 9221 (multiple-tube fermentation; APHA 1998) for total coliforms, fecal coliforms, and *Escherichia coli*, which are used as indicators for the presence of human pathogens. Dilutions of 1, 10, and 100 yielded minimum reporting limits from 2 to 200 MPN/100 mL (most probable number per 100 milliliters) and reporting ceilings from 1,600 to 160,000 MPN/100 mL. This variation typically affected only the total coliform counts.

4.4 Sediment Chemistry and Toxicity

4.4.1 Sediment Sampling

Bed sediment (hereafter termed "sediment") samples were collected by MPSL-DFG at watershed integrator sites. Sediment collection was carried out after collection of other samples and measurements. Samples were ideally collected from undisturbed, low-energy areas devoid of hard clay, bank deposits, fill, and gravel; where conditions did not allow for collection of a fine-grained sample, sampling was rescheduled when possible. A pre-cleaned Teflon™ scoop was used to spoon the top 2 cm of sediment from five or more sub-sites into a pre-cleaned glass composite jar. After an adequate amount of sediment was collected, it was homogenized thoroughly and aliquoted into pre-cleaned, pre-labeled sample jars (glass or polyethylene, as appropriate) for chemical or toxicological analysis. Once collected, samples were stored at a nominal 4°C and shipped to the analytical laboratories. Field data sheets were completed for each sampling event to document conditions and sampling details. Regular field duplicates and initial field blanks were also collected for quality control. For details on sediment sampling, refer to the bed sediment sampling procedures outlined in the QAMP (Puckett 2002) and Appendix H.

4.4.2 Chemical Analyses

Sediments were analyzed for organics, metals, total organic carbon, moisture, and grain size according to methods presented in Table 4-1. Sediment chemistry data were assessed using the consensus-based sediment quality guidelines (SQGs) for freshwater (MacDonald *et al.* 2000a), as shown in Table 4-3 and as described in Section 4.6. Measured sediment chemical concentrations below threshold effects concentrations (TECs) are considered unlikely to contribute to adverse effects in sediment-dwelling organisms. In contrast, sediment chemical concentrations above probable effects concentrations (PECs) are considered likely to be toxic to sediment-dwelling organisms.

4.4.3 Toxicity Testing

Sediment toxicity was tested at UCD-GC with the *Hyalella azteca* 10-day survival and growth test according to standard U.S. EPA guidelines (U.S. EPA 2000). Granite Canyon well water was used as the overlying water. Overlying water quality was monitored daily, and temperature monitored continuously. Negative controls were included with each test batch, and reference toxicant tests (positive controls) with cadmium chloride were performed monthly. Field duplicates were also tested. Details of toxicity testing methods can be found in the SWAMP QAMP (Puckett 2002) and Appendix J.

4.5 Bioaccumulation Studies with *Corbicula*

Deployment, collection and preparation of *Corbicula fluminea* were performed by MPSL-DFG at watershed integrator sites. Clams were collected in Lake Isabella or the Sacramento River, and tested for contamination prior to deployment. Clams were deployed for one month in anchored polypropylene mesh bags, approximately 15 cm above the stream bed. Approximately 25 to 50 clams, 20 to 30 mm in diameter, were deployed at each site for each analysis (organics and metals). After a month-long deployment, clams were collected and sent to the laboratory for analysis. Clams intended for metals analysis were transported in plastic bags; clams intended for organics analysis were bagged in aluminum, then plastic. All clam handling was performed with methods designed to minimize contamination. Details of clam collection, handling, deployment and retrieval can be found in the SWAMP QAMP (Puckett 2002) and Appendix I.

Clam tissues were analyzed for organics, metals, lipids, and percent moisture according to methods presented in Table 4-1.

4.6 Approach to Data Analysis and Interpretation

Water and sediment quality guidelines (Table 4-2 and Table 4-3) were used to evaluate the condition of streams in each watershed. Water and sediment chemistry values were compared to guidelines established for the protection of aquatic life beneficial uses. Pathogen values were compared to guidelines established for the protection of human health in relation to contact and non-contact recreation beneficial uses. These assessments were done to address two primary questions:

- 1) Is the water suitable for aquatic life?
- 2) Is the water useable for recreation (including contact and non-contact water recreation)?

Numerous other guidelines have been established for protection of other beneficial uses, such as drinking, agricultural supply, fishing, etc. Fishing beneficial uses were addressed in a separate report (SFBRWQCB 2005). Drinking water is provided to customers from distribution systems that include treatment to remove harmful levels of chemicals and pathogens. Other beneficial uses are generally protected by standards developed to protect aquatic life.

To examine the relationships between watershed land use and water quality parameters, sampling sites were categorized into discrete land use classes using a conservative classification scheme. Watershed land use data was evaluated by extensive field reconnaissance and verification, as well as from digital topographic maps (TOPO! 1999) and other published sources. Sites were categorized into six land use types, based on land use in the upstream drainage area:

1. Agriculture: row crop agriculture, including orchards, vineyards, and outdoor nurseries,
2. Grazing: grazing of livestock,
3. Open Space/Parkland: protected open space with minimal human land use other than minor park facilities, roads, and trails,
4. Rural Residential: low intensity residential land use associated with small unincorporated communities with no commercial or industrial land use,
5. Urban: High density residential, commercial, and/or industrial land use,
6. Mixed: Multiple land uses, or watersheds that did not clearly represent one land use.

Streams that drained exclusively protected open space and parkland, with no agricultural or urban land use present, were classified as open space/parkland. If any amount of human land use was present in the upstream watershed, sites were classified according to the type of land use present. For example, a site draining primarily parkland and open space with several residences within the drainage basin was classified as rural residential. Only sites that exclusively drained one land use type, other than open space, were classified as grazing, agriculture, rural residential, or urban. Sites that drained multiple land uses or basins that contained other confounding factors (e.g. mines) were classified as mixed land use.

This analysis should not be interpreted as an examination of direct cause-effect relationships between land use and stream condition, because correlations between land use and natural environmental gradients may confound the analysis. For example, urban areas are often located in low gradient valleys and bay margins, which naturally support different benthic communities and water quality conditions than upland hills and mountains, where most open space land is now found.

4.6.1 Analysis of Bioassessment Data

Bioassessment and habitat assessment parameters were evaluated for potential degradation of wildlife habitat and invertebrate communities, in order to answer the question, “Is the water suitable for aquatic life?”

Data from the 2001 and 2002 sampling events were combined into one dataset by standardizing taxonomic levels and functional feeding groups to the 2003 version of the CAMLnet STE (CAMLnet, www.dfg.ca.gov/cabw/camlnetste.pdf). Bioassessment data from other sources were included in portions of the analysis for the purposes of examining (1) natural geographic and temporal variability and (2) trends within specific watersheds. All data were collected and analyzed according to the CSBP, and were standardized according to the 2003 CAMLnet STE.

Samples from Olema Creek (National Park Service) were collected using the CSBP point-source sampling method, which involves sampling along three transects in one riffle. Although the area sampled (18 sq. ft.) is the same as the non-point source sampling design, only one riffle is sampled, so the richness metrics calculated from this data are believed to underestimate reach-scale values and are not quantitatively comparable to data collected with the non-point source sampling design. For the purposes of making general basin-wide comparisons, the Olema Creek data has been analyzed with data from Lagunitas Creek.

Samples from the Napa River (Friends of the Napa River) were collected using the CSBP, but 500 organisms were subsampled from each riffle, meaning that the total number of organisms sampled was 1500 rather than 900. A randomized subsampling procedure was performed to reduce the number of organisms to make the data comparable to other data analyzed in this report. Additionally, Napa River samples were identified to a lower taxonomic level than is required for the CAMLnet STE, so taxonomic levels were standardized to the 2003 STE. Data collected in the Wildcat Creek and San Leandro Creek watersheds for a separate monitoring study by the Regional Board (Breux *et al.* 2004) were also included in the watershed analyses.

In order to standardize the data to the 2003 CAMLnet STE, several significant changes were made to much of the raw taxonomic data: chironomid midges were classified at the family level (Chironomidae) rather than at the sub-family level, and oligochaetes were classified at class level (Oligochaeta). To compare data to the results described in this report, data must be standardized to this taxonomic level. For data analysis purposes, three riffle samples for each site were combined into a composite sample of approximately 900 organisms. Once standardized, 23 common metrics were recalculated based on the 900 count composite sample (Table 4-2).

Variability in metrics within a site group or between years was examined with the coefficient of variation (C.V.), calculated as:

$$\text{C.V.} = \text{Standard Deviation} / \text{Mean.}$$

A coefficient of variation less than 0.25 for inter-riffle variation has been suggested as a criterion for biological metrics (Harrington and Born 1999). Using the same criterion for inter-site and inter-annual variation provides a conservative indicator of low natural variability.

Several multivariate statistical analyses are presented in this report. Multivariate analyses were performed using PC-ORD for Windows (v. 4.20, MjM Software). Ordination is a technique whereby multiple variables are reduced and expressed in a small number of dimensions. Ordinations were performed on sites using log-transformed taxa abundance data. Sites that are close together in ordination space exhibit similar benthic assemblages; increasing distance between sites indicates that a greater number of different taxa were present at the sites. Non-metric multidimensional scaling (NMS) is the most generally effective ordination technique for ecological community data (McCune and Grace 2002). We used NMS with the Sorensen distance measure, 400 iterations, an instability criterion of 0.00001, 6 potential axes, 40 real runs and 50 randomized runs (NMS autopilot slow and thorough setting). The same ordination graph is used to plot both the regional analyses and the individual watershed analyses.

Cluster analysis defines groups of sites based on the similarity of taxa present. In contrast with the ordination approach, where log-transformed abundance data was used, presence/absence data was used for the cluster analysis in order to examine the results of a variety of analyses. The analysis is graphically represented as a cluster dendrogram, in which sites connected by shorter horizontal lines are more similar than sites connected by longer horizontal lines. Hierarchical cluster analyses were performed using the Sorenson (Bray-Curtis) distance measure and the flexible beta linkage method ($b = -0.25$). Clusters of site-groups were chosen in order to balance (1) group size (i.e. ideally 2-5 sites), (2) geographic similarity of groups (e.g. all mainstem river sites), and (3) moderate to high levels of information remaining (i.e. 50-75%).

4.6.2 Basic Water Quality

Discrete field measurements were taken to calculate concentrations of chemicals (e.g., unionized ammonia), and used to determine the bioavailability of trace metals so that they could be compared to guidelines. However, discrete temperature and dissolved oxygen (DO) measurements are insufficient for evaluating potential impacts since they don't take into account diurnal fluctuations, nor can they be compared to the most appropriate guidelines for salmonids. Therefore, continuous monitoring results were used for comparison to water quality guidelines and objectives and conditions of concern were identified.

Basic water quality measurements were compared to water quality objectives from the San Francisco Bay Regional Water Quality Control Board's Basin Plan (Basin Plan) developed to protect aquatic life, agriculture, and drinking water (Table 4-20). Measurements specific to the support of salmonid populations were also evaluated using guidelines presented in Table 4-2: maximum temperature, Maximum Weekly Average Temperature (MWAT; U.S. EPA 1977), and minimum DO. Highest weekly average temperatures (HWATs) were calculated at sites with at least seven consecutive days of temperature measurement, by averaging the daily temperatures, calculating the seven-day average of these daily temperatures for each seven-day period, and selecting the highest seven-day average as the HWAT. The HWAT is the MWAT for the period sampled. The annual MWAT for most stations in Region 2 occurs between June and October, typically in August; however not all stations were monitored during this period and not all stations were monitored for seven or more days. Continuous field measurements consisted of numerous 24-hour periods and sometimes an additional period of less than 24 hours; where this additional period was less than 21 hours, it was not used in the calculation of HWATs.

4.6.3 Water, Sediment, and Clam Tissue Chemistry

Water Chemistry: Water chemistry was compared to water quality objectives from the San Francisco Bay Regional Water Quality Control Board's Basin Plan developed to protect recreation and aquatic life, as well as to other selected guidelines (Table 4-2).

Sediment Chemistry: The two primary guidelines (Table 4-3) used to evaluate sediment chemistry for potential aquatic life impacts were the threshold effects concentration (TEC), below which harmful biological effects are considered unlikely to be observed, and the probable effects concentration (PEC), above which harmful effects are considered likely to be observed (MacDonald *et al.* 2000). Sediment Quality Guideline Quotient (SQGQ) values were calculated

to estimate the synergistic effect of a mixture of contaminants. SQGQ are calculated by dividing the measured sediment concentration of a chemical by its PEC value. The SQGQ values for a suite of chemicals are then averaged to give a mean value for each sediment sample. These SQGQ values can be compared to ranges evaluated by MacDonald et al. (2000), who found that 85 percent of sediment samples with mean PEC quotients greater than 0.5 were toxic to sediment-dwelling organisms. Similarly, 92 percent of sediment samples with mean PEC quotients greater than 1.0 were toxic to one or more species of aquatic organisms.

Clam Tissue Chemistry: Wet weight concentrations were compared to State Mussel Watch 85th percentile Elevated Data Levels (EDL 85) calculated for the period 1977 through 1997 (Rasmussen 2000). These EDLs are presented in Appendix I along with the clam tissue data evaluation. The EDL 85 is only an indication that a tissue chemical concentration is notably elevated above this value, and is not a measure or predictor of potentially adverse human or animal health effects.

4.6.4 Toxicity

Toxicity test data were evaluated by comparing mean organism response in samples and negative controls, using two methods. In the first method, a separate-variance t-test was conducted to determine whether statistically significant differences (with $\alpha = 0.05$) existed between samples and negative controls; a significant difference would indicate possible toxicity. In the second method, sample response was compared to 80 percent of the negative control response; anything less than 80 percent could indicate toxicity. This second method was chosen to emulate the use of protocol-specific 90th percentile minimum significant difference (MSD) values (Thursby *et al.* 1997). Once MSD-based criteria are generated for all SWAMP test organisms, they will replace the 80 percent criterion. Potentially toxic conditions were identified where both methods indicated a difference in toxicity between samples and controls.

4.6.5 Coliforms

For purposes of comparison with water quality guidelines, various statistics were calculated for the five sample results from each site. Depending on the coliform measure, these included the median, geometric mean (log mean), average, and 90th percentile. Non-detect values were replaced with the reporting limit for that sample in order to perform calculations. Where one sample was not collected due to lack of water, statistics were calculated using four values.

Coliform data were compared with water quality guidelines for protection of recreational use. These guidelines are presented in Table 4-2, as described in Section 4.2.2. It is important to note that coliform counts are extremely variable and can become elevated in the presence of sewage inputs, pets, cattle, wildlife, and even plant material. Because the presence of *E. coli* is directly related to fecal contamination and has been shown to correlate with incidences of human illness in recreational waters (U.S. EPA 1986), it is the most relevant of the three bacterial indicators measured here in terms of assessing the potential for human illness at the sites sampled.

Table 4-1. Analyte, laboratory, and method for water, sediment, and clam tissue chemistry

| Analyte | Analytical Lab | Water Method | Sediment Method | Clam Tissue Method |
|-----------------------|--------------------|-------------------------------|-------------------|-----------------------------|
| Pesticides | | | | |
| Aldrin | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Ametryn | DFG-WPCL | EPA 619M | | |
| Aspon | DFG-WPCL | EPA 8141AM | | |
| Atraton | DFG-WPCL | EPA 619M | | |
| Atrazine | DFG-WPCL | EPA 619M | | |
| Azinphos ethyl | DFG-WPCL | EPA 8141AM | | |
| Azinphos methyl | DFG-WPCL | EPA 8141AM | | |
| Bolstar | DFG-WPCL | EPA 8141AM | | |
| Carbophenothion | DFG-WPCL | EPA 8141AM | | |
| Chlordane, cis | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Chlordane, trans | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Chlordene, alpha | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Chlordene, gamma | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Chlorfenvinphos | DFG-WPCL | EPA 8141AM | | |
| Chlorpyrifos | DFG-WPCL UCD-GC | EPA 8141AM ELISA SOP 3.3** | EPA 8081AM n/a | EPA 3545_3640A_3620B n/a |
| Chlorpyrifos methyl | DFG-WPCL | EPA 8141AM | | |
| Ciodrin (Crotoxyphos) | DFG-WPCL | EPA 8141AM | | |
| Coumaphos | DFG-WPCL | EPA 8141AM | | |
| Dacthal | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| DBCE | DFG-WPCL | | | EPA 3545_3640A_3620B |
| DBOB | DFG-WPCL | | | EPA 3545_3640A_3620B |
| DCBP (p,p') | DFG-WPCL | | EPA 8081AM | EPA 3545_3640A_3620B |
| DDD (o,p') | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| DDD (p,p') | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| DDE (o,p') | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| DDE (p,p') | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| DDMU (p,p') | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| DDT (o,p') | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| DDT (p,p') | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Demeton-s | DFG-WPCL | EPA 8141AM | | |
| Diazinon | DFG-WPCL UCD-GC | EPA 8141AM ELISA SOP 3.3** | EPA 8081AM n/a | EPA 3545_3640A_3620B n/a |
| Dichlofenthion | DFG-WPCL | EPA 8141AM | | |
| Dichlorvos | DFG-WPCL | EPA 8141AM | | |
| Dicrotophos | DFG-WPCL | EPA 8141AM | | |
| Dieldrin | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Dimethoate | DFG-WPCL | EPA 8141AM | | |
| Dioxathion | DFG-WPCL | EPA 8141AM | | |
| Disulfoton | DFG-WPCL | EPA 8141AM | | |
| Endosulfan I | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Endosulfan II | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Endosulfan sulfate | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Endrin | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Endrin Aldehyde | DFG-WPCL | EPA 8081AM | | |
| Endrin Ketone | DFG-WPCL | EPA 8081AM | | |
| Ethion | DFG-WPCL | EPA 8141AM | | |
| Ethoprop | DFG-WPCL | EPA 8141AM | | |
| Famphur | DFG-WPCL | EPA 8141AM | | |
| Fenchlorphos | DFG-WPCL | EPA 8141AM | | |
| Fenitrothion | DFG-WPCL | EPA 8141AM | | |
| Fensulfothion | DFG-WPCL | EPA 8141AM | | |
| Fenthion | DFG-WPCL | EPA 8141AM | | |
| Fonofos (Dyfonate) | DFG-WPCL | EPA 8141AM | | |
| HCH, alpha | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| HCH, beta | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| HCH, delta | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| HCH, gamma | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Heptachlor | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Heptachlor epoxide | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Hexachlorobenzene | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |

| Analyte | Analytical Lab | Water Method | Sediment Method | Clam Tissue Method |
|---------------------------|----------------|------------------|-----------------|----------------------|
| Leptophos | DFG-WPCL | EPA 8141AM | | |
| Malathion | DFG-WPCL | EPA 8141AM | | |
| Merphos | DFG-WPCL | EPA 8141AM | | |
| Methodathion | DFG-WPCL | EPA 8141AM | | |
| Methoxychlor | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Mevinphos | DFG-WPCL | EPA 8141AM | | |
| Mirex | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Molinate | DFG-WPCL | EPA 8141AM | | |
| Naled (Dibrom) | DFG-WPCL | EPA 8141AM | | |
| Nonachlor, cis | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Nonachlor, trans | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Oxadiazon | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Oxychlorane | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Parathion, Ethyl | DFG-WPCL | EPA 8141AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Parathion, Methyl | DFG-WPCL | EPA 8141AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Phorate | DFG-WPCL | EPA 8141AM | | |
| Phosmet | DFG-WPCL | EPA 8141AM | | |
| Phosphamidon | DFG-WPCL | EPA 8141AM | | |
| Prometon | DFG-WPCL | EPA 619M | | |
| Prometryn | DFG-WPCL | EPA 619M | | |
| Propazine | DFG-WPCL | EPA 619M | | |
| Secbumeton | DFG-WPCL | EPA 619M | | |
| Simazine | DFG-WPCL | EPA 619M | | |
| Simetryn | DFG-WPCL | EPA 619M | | |
| Sulfotep | DFG-WPCL | EPA 8141AM | | |
| Tedion | DFG-WPCL | EPA 8081AM | EPA 8081AM | EPA 3545_3640A_3620B |
| Terbufos | DFG-WPCL | EPA 8141AM | | |
| Terbutylazine | DFG-WPCL | EPA 619M | | |
| Terbutryn | DFG-WPCL | EPA 619M | | |
| Tetrachlorvinphos | DFG-WPCL | EPA 8141AM | | |
| Thiobencarb | DFG-WPCL | EPA 8141AM | | |
| Thionazin | DFG-WPCL | EPA 8141AM | | |
| Tokuthion | DFG-WPCL | EPA 8141AM | | |
| Toxaphene | DFG-WPCL | | EPA 8081AM | EPA 3545_3640A_3620B |
| Trichlorfon | DFG-WPCL | EPA 8141AM | | |
| Trichloronate | DFG-WPCL | EPA 8141AM | | |
| PAHs | | | | |
| Acenaphthene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | |
| Acenaphthylene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | |
| Anthracene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | |
| Benz(a)anthracene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | EPA 8270M |
| Benzo(a)pyrene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | |
| Benzo(b)fluoranthene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | EPA 8270M |
| Benzo(e)pyrene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | EPA 8270M |
| Benzo(g,h,i)perylene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | EPA 8270M |
| Benzo(k)fluoranthene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | |
| Biphenyl | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | |
| Chrysene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | EPA 8270M |
| Chrysenes, C1- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Chrysenes, C2- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Chrysenes, C3- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Dibenz(a,h)anthracene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | EPA 8270M |
| Dibenzothiophene | DFG-WPCL | EPA 8270M | EPA 8270M | |
| Dibenzothiophenes, C1- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Dibenzothiophenes, C2- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Dibenzothiophenes, C3- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Dimethylnaphthalene, 2,6- | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | |
| Fluoranthene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | EPA 8270M |
| Fluoranthene/Pyrenes, C1- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Fluorene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | |
| Fluorenes, C1- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Fluorenes, C2- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Fluorenes, C3- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |

| Analyte | Analytical Lab | Water Method | Sediment Method | Clam Tissue Method |
|---|----------------------|---------------------------------------|--|----------------------|
| Indeno(1,2,3-c,d)pyrene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | EPA 8270M |
| Methylnaphthalene, 1- | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | |
| Methylnaphthalene, 2- | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | |
| Methylphenanthrene, 1- | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | |
| Naphthalene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | EPA 8270M |
| Naphthalenes, C1- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Naphthalenes, C2- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Naphthalenes, C3- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Naphthalenes, C4- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Perylene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | |
| Phenanthrene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | EPA 8270M |
| Phenanthrene/Anthracene, C1- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Phenanthrene/Anthracene, C2- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Phenanthrene/Anthracene, C3- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Phenanthrene/Anthracene, C4- | DFG-WPCL | EPA 8270M | EPA 8270M | EPA 8270M |
| Pyrene | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | EPA 8270M |
| Trimethylnaphthalene, 2,3,5- | DFG-WPCL | EPA 8310M, 8270M | EPA 8270M | |
| PCBs | | | | |
| Aroclor 1248, 1254, 1260 | DFG-WPCL | | 1988 | EPA 3545_3640A_3620B |
| Various congeners | DFG-WPCL | EPA 8082M | EPA 8082M | EPA 3545_3640A_3620B |
| Metals | | | | |
| Aluminum* | MPSL-DFG | EPA 1638M | EPA 200.8 | EPA 200.7 |
| Arsenic (organic + inorganic)* | MPSL-DFG | EPA 1638M | EPA 200.8 | EPA 200.7 |
| Cadmium* | MPSL-DFG | EPA 1638M | EPA 200.8 | EPA 200.7 |
| Chromium* | MPSL-DFG | EPA 1638M | EPA 200.8 | EPA 200.7 |
| Copper* | MPSL-DFG | EPA 1638M | EPA 200.8 | EPA 200.7 |
| Lead* | MPSL-DFG | EPA 1638M | EPA 1638 | EPA 200.7 |
| Manganese* | MPSL-DFG | EPA 1638M | EPA 200.8 | EPA 200.7 |
| Mercury (Total, organic + inorganic) | MPSL-DFG | EPA 1631B EPA 1631EM | DFG SOP 103** | EPA 3052 |
| Nickel* | MPSL-DFG | EPA 1638M | EPA 200.8 | EPA 200.7 |
| Selenium* | MPSL-DFG | EPA 1638M | | EPA 200.7 |
| Silver* | MPSL-DFG | EPA 1638M | EPA 200.8 | EPA 200.7 |
| Zinc* | MPSL-DFG | EPA 1638M | EPA 200.8 | EPA 200.7 |
| Medium characteristics | | | | |
| Grain Size, ASTM (Fines, Sand, Gravel) | AMS | | ASTM D422 | |
| Moisture, Percent | DFG-WPCL MPSL-DFG | | EPA 8081AM, 8082M, 8270M DFG SOP 103**, EPA 200.8 | DFG SOP |
| Lipid, Percent | DFG-WPCL | | | EPA 3545_3640A_3620B |
| Conventionals | | | | |
| Alkalinity as CaCO3 | DFG-WPCL | SM 2320B, QC 10303311A | | |
| Ammonia as N | DFG-WPCL | EPA 350.3 | | |
| Boron | SFL Babcock | SM 4500BB EPA 200.7 | | |
| Chloride | DFG-WPCL | EPA 300.0 | | |
| Chlorophyll a | MPSL-DFG SFL | EPA 445.0 SM 10200H-2b | | |
| Coliforms (Total, Fecal, <i>E. coli</i>) | SAL | SM 9221 | | |
| Hardness as CaCO3 | DFG-WPCL | SM 2340C | | |
| Nitrate as N | DFG-WPCL | EPA 300.0, EPA 353.3, QC 10107041B | | |
| Nitrite as N | DFG-WPCL | FR 8507, QC 10107041B | | |
| Nitrogen, Total Kjeldahl | DFG-WPCL | EPA 351.3, QC 10107062E | | |
| Organic Carbon, Dissolved | AMS | EPA 415.1 | | |
| Organic Carbon, Total | AMS | EPA 415.1 | EPA 9060 | |

| Analyte | Analytical Lab | Water Method | Sediment Method | Clam Tissue Method |
|-----------------------------------|-----------------|------------------------------------|-----------------|--------------------|
| Orthophosphate as P | DFG-WPCL | EPA 300.0, EPA 365.3, QC 10115011M | | |
| Pheophytin a | SFL | SM 10200H-2a | | |
| Phosphorus as P, Total | DFG-WPCL | EPA 365.3, QC 10115011D | | |
| Sediment Concentration, Suspended | AMS MPSL-DFG | ASTM D3977 SM 2540 B | | |
| Solids, Total Dissolved | DFG-WPCL | SM 2540 | | |
| Solids, Total Suspended | DFG-WPCL | EPA 160.2 | | |
| Sulfate | DFG-WPCL | EPA 300.0 | | |

*These metals were measured as total and dissolved in water, and total elsewhere.

**These methods for chlorpyrifos and diazinon are described in the QAMP (Puckett 2002)

Babcock: E.S. Babcock & Sons, Inc.

DFG-WPCL: Department of Fish and Game Water Pollution Control Laboratory

MPSL-DFG: Marine Pollution Studies Laboratory, Department of Fish and Game

SFL: Sierra Foothill Laboratory

AMS: Applied Marine Sciences

UCD-GC: University of California at Davis, Granite Canyon Laboratory

SAL: Sequoia Analytical Laboratories, Inc.

Tables 4-2 and 4-3 list the thresholds used to evaluate data in this report. If there were no objectives for an analyte in the Basin Plan, we used limits from the California Toxics Rule (CTR). If there were no thresholds in either of these documents, we used the most appropriate threshold from California Department of Fish and Game, TMDLs, U.S. EPA criteria, or peer reviewed literature. Results from grab samples are appropriate to compare to acute water quality objectives; chronic water quality objectives are used as a screening tool, although they are not applicable to grab samples.

Table 4-2. Water quality thresholds for aquatic life

| Analyte | Description of Standard | Numeric Limit | Units | Reference |
|--|--|---------------|-------------|-------------------------------|
| <i>Field measures</i> | | | | |
| Temperature | Maximum, salmonid | 24 | ° C | USEPA, 1977 |
| | MWAT, steelhead | 17 | ° C | Sullivan <i>et al.</i> , 2000 |
| | MWAT, coho | 14.8 | ° C | Sullivan <i>et al.</i> , 2000 |
| Oxygen, dissolved | Minimum, warmwater | 5 | mg/L | Basin Plan, 2005 |
| | Minimum, coldwater | 7 | mg/L | Basin Plan, 2005 |
| | 3-month median | 80 | % | Basin Plan, 2005 |
| pH | Range | 6.5 to 8.5 | S.U. | Basin Plan, 2005 |
| <i>Nutrients</i> | | | | |
| Ammonia, unionized | Annual median | 0.025 | mg/L | Basin Plan, 2005 |
| Nitrate as N | Maximum | 0.16 | mg/L | USEPA, 2000 |
| Phosphorus, total as P | Maximum | 30 | µg/L | USEPA, 2000 |
| <i>Metals</i> | | | | |
| <i>Cadmium, copper, nickel, silver, and zinc values assume a hardness of 100 mg/L CaCO3. Samples at other hardness levels must be calculated using formulas in the Basin Plan.</i> | | | | |
| Arsenic, dissolved | 1-hour average WQO | 340 | µg/L | Basin Plan, 2005 |
| | 4-day average WQO | 150 | | |
| Cadmium, total | 1-hour average WQO | 3.9 | µg/L | Basin Plan, 2005 |
| | 4-day average WQO | 1.1 | | |
| Chromium VI, dissolved | 1-hour average WQO | 16 | µg/L | Basin Plan, 2005 |
| | 4-day average WQO | 11 | | |
| Copper, dissolved | 1-hour average WQO | 13 | µg/L | Basin Plan, 2005 |
| | 4-day average WQO | 9 | | |
| Lead, dissolved | 1-hour average WQO | 65 | µg/L | Basin Plan, 2005 |
| | 4-day average WQO | 2.5 | | |
| Mercury, total | 1-hour average WQO | 2.4 | µg/L | Basin Plan, 2005 |
| | 4-day average WQO | 0.025 | | |
| Nickel, dissolved | 1-hour average WQO | 470 | µg/L | Basin Plan, 2005 |
| | 4-day average WQO | 52 | | |
| Selenium, total | 1-hour average WQO | 20 | µg/L | Basin Plan, 2005 |
| | 4-day average WQO | 5 | | |
| Silver, dissolved | 1-hour average WQO | 3.4 | µg/L | Basin Plan, 2005 |
| Zinc, dissolved | 1-hour average WQO | 120 | µg/L | Basin Plan, 2005 |
| | 4-day average WQO | 120 | | |
| <i>Organics</i> | | | | |
| PCBs | Continuous 4-day average | 0.014 | µg/L | CTR |
| Chlorpyrifos | Continuous 4-day average | 0.015 | µg/L | CVRWQCB, 2006 |
| Dacthal (DCPA) | Instantaneous maximum AWQC | 14300 | µg/L | USEPA, 1987 |
| Diazinon | 1-hour average | 0.1 | µg/L | SFBRWQCB, 2005 |
| Disulfoton (Disyston) | Instantaneous maximum AWQC | 0.05 | µg/L | USEPA, 1973 |
| Endosulfan | Continuous 4-day average | 0.056 | µg/L | CTR |
| | Instantaneous maximum | 0.22 | | |
| HCH, gamma- (gamma-BHC, Lindane) | Maximum 1-hour average | 0.95 | µg/L | CTR |
| Parathion, methyl | Instantaneous maximum AWQC | 0.08 | µg/L | CDFG |
| Thiobencarb | Instantaneous maximum AWQC | 3.1 | µg/L | CDFG |
| <i>Coliforms</i> | | | | |
| E. coli | steady state | 126 | MPN /100 mL | Basin Plan, 2005 |
| | designated beach | 235 | | |
| Fecal coliform | log mean | 200 | MPN /100 mL | Basin Plan, 2005 |
| | 90th percentile | 400 | | |
| Total coliform | median | 240 | MPN /100 mL | Basin Plan, 2005 |
| | maximum | 10000 | | |
| <i>Toxicity</i> | Water and sediment samples were considered toxic when results were: 1) significantly different than control values and 2) less than 80% of control values. | | | |

Table 4-3. Sediment quality guidelines (SQGs) for aquatic life

| Analyte | SQG type: | | TEC | |
|-----------------------|------------------|--------------|--------------|--------------|
| | PEC | | | |
| | <i>mg/kg</i> | <i>µg/kg</i> | <i>mg/kg</i> | <i>µg/kg</i> |
| <u>Metals</u> | | | | |
| Arsenic | 33 | | 9.79 | |
| Cadmium | 4.98 | | 0.99 | |
| Chromium | 111 | | 43.4 | |
| Copper | 149 | | 31.6 | |
| Lead | 128 | | 35.8 | |
| Mercury | 1.06 | | 0.18 | |
| Nickel | 48.6 | | 22.7 | |
| Zinc | 459 | | 121 | |
| <u>Organics</u> | | | | |
| Anthracene | | 845 | | 57.2 |
| Fluorene | | 536 | | 77.4 |
| Naphthalene | | 561 | | 176 |
| Phenanthrene | | 1170 | | 204 |
| Benz(a)anthracene | | 1050 | | 108 |
| Benzo(a)pyrene | | 1450 | | 150 |
| Chrysene | | 1290 | | 166 |
| Dibenz(a,h)anthracene | | | | 33 |
| Fluoranthene | | 2230 | | 423 |
| Pyrene | | 1520 | | 195 |
| PAH (total) | | 22800 | | 1610 |
| PCB (total) | | 676 | | 59.8 |
| Chlordane | | 17.6 | | 3.24 |
| Dieldrin | | 61.8 | | 1.9 |
| DDD (sum op + pp) | | 28 | | 4.88 |
| DDE (sum op + pp) | | 31.3 | | 3.16 |
| DDT (sum op + pp) | | 62.9 | | 4.16 |
| DDT (total) | | 572 | | 5.28 |
| Endrin | | 207 | | 2.22 |
| Heptachlor epoxide | | 16 | | 2.47 |
| HCH, gamma | | 4.99 | | 2.37 |

Source: MacDonald DD, Ingersoll CG, Berger TA. 2000. "Development and evaluation of consensus-based sediment quality guidelines for freshwater ecosystems." *Archives of Environmental Contamination and Toxicology*. 39:20-31.

Notes: All guidelines are on a dry-weight basis. PEC = probable effect concentration; TEC = threshold effect concentration.

Table 4-4. Biological metrics

| Metric | Definition |
|---|---|
| Richness Metrics | |
| Coleoptera Taxa | Number of taxa in the Coleoptera insect order |
| Diptera Taxa | Number of taxa in the Diptera insect order |
| Ephemeroptera Taxa | Number of taxa in the Ephemeroptera insect order |
| Plecoptera Taxa | Number of taxa in the Plecoptera insect order |
| Trichoptera Taxa | Number of taxa in the Trichoptera insect order |
| Non-Insect Taxa | Number of non-insect taxa |
| Taxa Richness | Total number of taxa (aka cumulative taxa richness) |
| EPT Taxa | Number of taxa in the Ephemeroptera, Plecoptera, and Trichoptera insect orders |
| Composition Metrics | |
| % EPT | Percentage of total individuals belonging to the Ephemeroptera, Plecoptera, and Trichoptera insect orders |
| % Sensitive EPT | Percentage of total individuals belonging to the Ephemeroptera, Plecoptera, and Trichoptera insect orders with tolerance values less than or equal to 3 |
| % Chironomidae | Percentage of total individuals belonging to the Chironomidae family (Diptera) |
| % Coleoptera | Percentage of total individuals belonging to the Coleoptera insect order |
| % Oligochaeta | Percentage of total individuals belonging to the Oligochaeta sub-class |
| % Non-insect | Percentage of total individuals that are not insects |
| Tolerance Metrics | |
| % Intolerant | Percentage of total individuals that have tolerance values greater or equal than 7 |
| % Tolerant | Percentage of total individuals that have tolerance values less than or equal to 3 |
| Tolerance Value | Average tolerance value for all individuals |
| Functional Feeding Group (FFG) Metrics | |
| % Predator | Percentage of all individuals belonging to the predator FFG |
| % Collector-filterer | Percentage of all individuals belonging to the collector-filterer FFG |
| %Collector-gatherer | Percentage of all individuals belonging to the collector-gatherer FFG |
| % Scraper | Percentage of all individuals belonging to the scraper FFG |
| % Shredder | Percentage of all individuals belonging to the shredder FFG |
| % Other | Percentage of all individuals belonging to other FFGs |

Biological metrics were calculated from composited (900 count) raw taxonomic data.

Table of Contents

5 Quality Assurance & Quality Control 5-1

5.1 Laboratory Method Blanks 5-1

5.2 Surrogate Spikes 5-1

5.3 Matrix Spikes and Matrix Spike Duplicates 5-2

5.4 Certified Reference Materials and Laboratory Controls..... 5-2

5.5 Field Blind Duplicates 5-3

5.6 Laboratory Replicates 5-3

5.7 Toxicity Tests..... 5-3

5.8 QA/QC Summary..... 5-3

5 Quality Assurance & Quality Control

The data generated by SWAMP is used to determine the status of beneficial uses throughout the state. Monitoring data are used to assess trends, make regulatory and management decisions, and to support enforcement of policies. Thorough objectives for achieving quality data are outlined in the Quality Assurance Management Plan (Puckett 2002). In general, data quality is demonstrated through analysis of:

- Laboratory method blanks
- Surrogate spikes
- Matrix spikes and matrix spike duplicates
- Certified reference materials/laboratory control samples
- Laboratory replicates
- Field blind duplicates

5.1 Laboratory Method Blanks

Laboratory method blanks were used to assess laboratory contamination introduced during sample preparation and analysis. The method blanks were processed in a manner identical to the associated field samples. According to the QAMP for both organic and inorganic analyses, at least one laboratory method blank should be analyzed per 20 samples or one per batch, whichever is more frequent. However, there were some batches where blanks were not performed at the required frequency or were not performed at all. These batches are listed in Appendix C-1 and will be classified as estimated.

Water and sediment laboratory method blanks are considered acceptable if they remain below the method detection limit (MDL) of each target analyte. All laboratory method blanks were acceptable with the exception of those specified within Appendix C-2. One hundred eighteen laboratory blanks had concentrations that were above the MDL but were less than the reporting limit (RL). Twenty blanks had detectable levels of analyte where the concentrations were above the RL.

5.2 Surrogate Spikes

Surrogate spikes are used to assess analyte losses during sample extraction and clean-up procedures, and must be added to every field and quality control sample prior to extraction. Whenever possible, isotopically-labeled analogs of the analytes should be used.

All surrogate percent recoveries were within the acceptance criteria with the exception of five polycyclic aromatic hydrocarbon batches. The affected data sets are summarized in Appendix C-3.

5.3 Matrix Spikes and Matrix Spike Duplicates

A laboratory-fortified sample matrix (matrix spike, or MS) and a laboratory-fortified sample matrix duplicate (matrix spike duplicate or MSD) are both used to evaluate the effect of the sample matrix on the recovery of the target analyte(s). Individually, these samples are used to assess the bias from an environmental sample matrix plus normal method performance. In addition, these duplicate samples can be used collectively to assess analytical precision.

Aliquots of randomly selected field samples were spiked with known amounts of target analytes. The percent R of each spike was calculated as follows:

$$\%R = (\text{MS Result} - \text{Sample Result}) / (\text{Expected Concentration} - \text{Sample Result}) * 100$$

This process was repeated for a subset of field samples to create MSDs. According to the QAMP for both organic and inorganic analyses, at least one MS/MSD pair should be performed per 20 samples or one per batch, whichever is more frequent. However, some batches did not include MS/MSDs performed at the required frequency. These batches are listed in Appendix C-4.

The MS/MSD relative percent differences (RPDs) were calculated as follows:

$$\text{RPD} = (|(\text{Concentration1} - \text{Concentration2})| / (\text{AVERAGE}(\text{Concentration1} + \text{Concentration2}))) * 100$$

where

Concentration1 = matrix spike concentration

Concentration2 = matrix spike duplicate concentration

Data batches with MS/MSD percent R and RPD concentrations outside of acceptance criteria are summarize in Appendix C-5. All other MS/MSD percent Rs and RPDs were acceptable.

5.4 Certified Reference Materials and Laboratory Controls

Certified reference materials (CRMs), laboratory control samples (LCSs), and laboratory control materials (LCMs) were analyzed to assess the accuracy of a given analytical method. As required by the QAMP, one CRM, LCS, or LCM should be analyzed per 20 samples or one per batch, whichever is more frequent. However, there were some batches where CRMs, LCSs, or LCMs were not performed at this required frequency. These batches are listed in Appendix C-6.

Unacceptable CRM, LCS, and LCM recoveries are presented in Appendix C-7. All other recoveries were acceptable.

5.5 Field Blind Duplicates

Field blind duplicates were analyzed to assess field homogeneity and field sampling procedures. Field blind duplicates were sampled at sites 204ALP010, 205PER010, 206PET310 and 206WIL020. Field duplicate concentrations were compared to field sample concentrations from each site and relative percent difference (RPD) was calculated as follows:

$$\text{RPD} = \left(\frac{|\text{Concentration1} - \text{Concentration2}|}{\text{AVERAGE}(\text{Concentration1} + \text{Concentration2})} \right) * 100$$

where

Concentration1 = field sample concentration

Concentration2 = duplicate sample concentration

If either Concentration 1 or Concentration 2 was less than three times the MDL, the RPD was not calculated, because the concentrations were too low to produce a statistically valid result. RPDs less than 25 percent were considered acceptable as specified in the QAMP. RPDs greater than 25 percent are presented in the Appendix C-8. All other RPDs were acceptable.

5.6 Laboratory Replicates

Laboratory replicates were analyzed to assess laboratory precision. A replicate of at least one field sample per batch was processed and analyzed. However, the batches listed in Appendix C-9 did not include replicates.

The replicates were compared and an RPD was calculated as described in above in subsection 5.5. If either the sample result (Concentration1) or the replicate result (Concentration2) are less than three times the MDL, the RPD is not calculated as these concentrations would be too low to calculate a meaningful difference between them. RPDs less than 25 percent were considered acceptable as specified in the QAMP. RPDs greater than 25 percent are presented in Appendix C-10.

5.7 Toxicity Tests

There were minor deviations in water quality parameters or test conditions (conductivity, dissolved oxygen, temperature, light) in some replicates, and incoming sample temperature or holding times were exceeded in some cases. However, the data should be considered acceptable for their intended purpose.

5.8 QA/QC Summary

All data meeting specified control limits are considered usable without further evaluation. If data fail any portion of these criteria, they can be cross-checked against other quality control samples. If two of the following criteria are met, then the data are acceptable: laboratory replicate RPD,

MS/MSD recovery and RPD, or CRM/LCS/LCM recovery. Therefore, if the laboratory replicate RPD is greater than 25 percent but the MS/MSD and the CRM for that analyte are acceptable, or if a MS/MSD is unacceptable but the laboratory replicate RPD and CRM for that analyte are acceptable, then the data are acceptable and can be used.

Data that meet all SWAMP method quality objectives (MQOs) as specified in the QAMP are classified as “SWAMP-compliant.” Data that fail to meet all program MQOs specified in the SWAMP QAMP, have analytes not covered in the SWAMP QAMP, or are insufficiently documented such that supplementary information is required for them to be used in reports are classified as “Estimated.” “Historical” data batches are generally missing the necessary quality control (QC) samples. During the data quality assessment phase of reporting, end users may find estimated and/or historical data batches meet project data quality objectives.

Table of Contents

| | | |
|-------|--|------|
| 6 | Regional Trends in Water Quality | 6-1 |
| 6.1 | Macroinvertebrate Assemblages and Physical Habitat | 6-2 |
| 6.1.1 | Watershed Land Use | 6-6 |
| 6.2 | Field Measurements of Basic Water Quality | 6-14 |
| 6.2.1 | Regional Trends | 6-14 |
| 6.2.2 | Temporal Patterns | 6-18 |
| 6.3 | Water, Sediment, and Clam Tissue Chemistry | 6-26 |
| 6.3.1 | Water Chemistry | 6-26 |
| 6.3.2 | Sediment Chemistry | 6-36 |
| 6.3.3 | Clam Tissue Chemistry | 6-37 |
| 6.4 | Water and Sediment Toxicity | 6-42 |
| 6.5 | Coliform Bacteria..... | 6-43 |
| 6.6 | Regional Reference Conditions for Bay Area Streams..... | 6-48 |
| 6.6.1 | Benthic Macroinvertebrate Assemblages | 6-48 |
| 6.6.2 | Water and Sediment Chemistry | 6-62 |
| 6.6.3 | Nutrients Reference Conditions..... | 6-63 |

List of Tables

| | | |
|-------------|---|------|
| Table 6-1. | Dominant taxa in assemblage groups | 6-3 |
| Table 6-2. | Biological metrics in different land use classes | 6-7 |
| Table 6-3. | Correlation of P-HAB scores with macroinvertebrate ordination axes..... | 6-10 |
| Table 6-4. | Biological metrics upstream and downstream of dams..... | 6-12 |
| Table 6-5. | Annual variability for select metrics | 6-50 |
| Table 6-6. | Minimally disturbed sites used in the analysis of geographic variability | 6-53 |
| Table 6-7. | Minimally disturbed intermittent and perennial sites' metric scores | 6-55 |
| Table 6-8. | Coefficients of Variation from perennial and intermittent stream classes | 6-57 |
| Table 6-9. | Discrimination efficiencies of 26 metrics of inferred biological impairment | 6-58 |
| Table 6-10. | Metrics for minimally disturbed streams in the San Francisco Bay Area..... | 6-59 |
| Table 6-11. | Occurrence of tolerant taxa at sites in different land use classes | 6-60 |
| Table 6-12. | List of common intolerant taxa: perennial streams | 6-61 |
| Table 6-13. | List of common intolerant taxa: intermittent streams..... | 6-62 |

List of Figures

| | | |
|-------------|---|------|
| Figure 6-1. | Cluster dendrogram of bioassessment sites..... | 6-4 |
| Figure 6-2. | NMS ordination of bioassessment sites, labeled by assemblage group | 6-5 |
| Figure 6-3. | NMS ordination of bioassessment sites labeled by land use class..... | 6-8 |
| Figure 6-4. | Boxplot of EPT taxa richness by land-use class | 6-9 |
| Figure 6-5. | NMS ordination of bioassessment sites, with bi-plot of P-HAB variables..... | 6-11 |
| Figure 6-6. | HWAT and minimum DO versus channel alteration..... | 6-15 |
| Figure 6-7. | HWAT and minimum DO versus canopy cover | 6-16 |
| Figure 6-8. | Regional HWAT versus land use..... | 6-17 |
| Figure 6-9. | Regional minimum DO versus land use..... | 6-18 |

| | |
|---|--------|
| Figure 6-10. BUT010 and PES060 temperature and DO patterns..... | 6-22 |
| Figure 6-11. LAG050 temperature and DO patterns | 6-23 |
| Figure 6-12. LAG190 temperature and DO patterns | 6-23 |
| Figure 6-13. SGR120 temperature and DO patterns..... | 6-25 |
| Figure 6-14. SGR120 specific conductance and DO patterns | 6-25 |
| Figure 6-15. Wet season nutrients and chlorophyll | 6-27 |
| Figure 6-16. Spring season nutrients and chlorophyll | 6-27 |
| Figure 6-17. Dry season nutrients and chlorophyll..... | 6-28 |
| Figure 6-18. Nitrate (NO ₃ -N)..... | 6-30 |
| Figure 6-19. Orthophosphate (dry season)..... | 6-31 |
| Figure 6-20. Total phosphorus (dry season) | 6-32 |
| Figure 6-21. Diazinon in water | 6-3835 |
| Figure 6-22. PAH's in water..... | 6-3935 |
| Figure 6-23. Copper in tissue..... | 6-38 |
| Figure 6-24. Mercury in tissue..... | 6-39 |
| Figure 6-25. Selenium in tissue | 6-39 |
| Figure 6-26. PCBs in tissue | 6-40 |
| Figure 6-27. Chlordanes in tissue | 6-40 |
| Figure 6-28. DDT in tissue | 6-41 |
| Figure 6-29. Other pesticides in tissue..... | 6-41 |
| Figure 6-30. <i>E. coli</i> counts..... | 6-44 |
| Figure 6-31. Fecal coliform bacteria counts | 6-45 |
| Figure 6-32. Total coliform bacteria counts | 6-46 |
| Figure 6-33. Map of <i>E. coli</i> bacteria exceedances | 6-47 |
| Figure 6-34. Taxa presence in Wildcat and San Leandro Creeks..... | 6-51 |
| Figure 6-35. Correlation of environmental variables with taxa presence..... | 6-56 |

6 Regional Trends in Water Quality

While each of the nine major watersheds has unique characteristics, some regional trends exist in terms of seasonality, relationships between water quality and land use, and in exceedances of water and sediment quality guidelines. These are discussed in detail in the sections that follow and are summarized briefly below.

When assessing streams in terms of aquatic habitat, the data indicate that elevated temperatures and depressed DO concentrations were the stressors with the greatest potential to negatively impact salmonids and other aquatic life (Section 6.2). These two parameters, in turn, appeared to be influenced by the overall condition of the riparian corridors through which the streams flow. Physical habitat features such as channel modification, width of the riparian zone, and canopy cover were related with significant trends in temperature and DO (*e.g.*, Figure 6-6 and Figure 6-7). These relationships may be of added importance because streams in areas with good riparian condition tended to have acceptable values for temperature and DO, while streams with poorer or more highly modified physical characteristics tended to have temperature and DO values that exceeded thresholds for salmonids and Basin Plan Objectives.

Chemical contaminant concentrations measured in grab samples tended to be relatively low, and observed water and sediment toxicity tended to be moderate. There were some exceptions, such as 100 percent amphipod mortality in sediments collected near the mouth of San Leandro Creek, and a number of chemicals did exceed guideline values for aquatic life protection in water or sediment. There were some significant seasonal trends in water chemistry, with concentrations of many trace metals and organics being highest in the dry season. Many of the highest chemical measurements were taken from samples collected at sites draining urban areas.

All measured nitrate concentrations were below the numeric Basin Plan Objective for drinking water and agricultural water supply of 10 mg-N/L. Currently there is no numeric nitrate Basin Plan Objective for aquatic life, but most measured nitrate values exceeded the U.S. EPA reference guideline of 0.155 mg/L, which was developed to protect against eutrophication. All reported total phosphorus and most orthophosphate concentrations were above the U.S. EPA guideline of 0.030 mg/L (Section 6.3.1.1). However, the reporting limit for total P was 0.040 mg/L—higher than the guideline concentration. Region-wide nitrate concentrations did not vary significantly with season, but total phosphorus and orthophosphate were significantly higher in the dry season. Nutrient concentrations were highest in samples collected at sites draining urban areas. Average nitrate concentrations at sites draining urban areas were more than twice those for sites draining agricultural areas, and nearly ten times those from streams in open space areas. No water samples collected during these watershed surveys measured concentrations of unionized ammonia above the Basin Plan objective of 0.025 mg/L.

Biological assessments of invertebrate communities showed a clear relationship with land use (Figure 6-4), and also appeared to be affected by channel modification (Section 6.1.1.2). In the absence of major human disturbance, invertebrate communities were most greatly affected by the intermittency of stream flow.

6.1 Macroinvertebrate Assemblages and Physical Habitat

Three major benthic assemblages were identified based on classification and ordination analyses of log transformed taxa abundances at sites. The first three groups of sites identified by the cluster analysis (Figure 6-1) are also identified on the ordination graph (Figure 6-2). Ordination graphs in this report are used to show biological difference or similarity among sampling sites. Non-metric multidimensional scaling (NMS) is a type of ordination—or multivariate analysis—that reduces the complexity of a data set in order to reveal the most important general trends or patterns. Ordination techniques are useful because, like metrics, they reduce the complexity in a list of taxa for numerous sites into data that can be easily visualized. Unlike metrics, ordination techniques summarize all of the taxa and sites into one easy-to-visualize graph.

An ordination graph consists of one, two, or three axes on which individual sites are plotted. The axes represent the most important gradients in the data set, representing the most variation in taxa presence between the sites. The proximity of sites to one another on the ordination graph is usually interpreted as an indication of their similarity. For example, sites close to one another would indicate that they tended to share similar taxa. Sites that are furthest away from one another on the graph indicate that they share very few or no taxa.

Although more than three groups can be identified in the dendrogram (Figure 6-1), this level of grouping provides a good tradeoff between retention of basic information (10 percent of total) and delineation of broad categories. As discussed below, the three groups generally represent broad land use classes that are likely to be an important control on benthic assemblages.

Assemblage Group 1 was numerically dominated by the pollution tolerant taxa Chironomidae (Diptera), Oligochaeta (aquatic worms), *Baetis* sp. (Ephemeroptera: Baetidae), and *Simulium* sp. (Diptera: Simuliidae) (Table 6-1). Together these taxa accounted for 89 percent of organisms found at sites in Assemblage Group 1. Only one other taxa, *Hyaella* (Amphipoda: Hyaellidae), accounted for more than 2 percent of all organisms. The total taxonomic richness of Assemblage Group 1 was 115 taxa. Several taxa were present in this assemblage that were not present in the other two assemblages, including *Corbicula* sp. (Veneroidea: Corbiculidae) and *Psychoda* sp. (Diptera: Psychodidae).

Assemblage Group 2 was dominated by many of the same taxa as Assemblage Group 1 (Table 6-1), but taxonomic diversity was somewhat greater (129 taxa) even though the number of sites in this group was much less than Assemblage Group 1 (23 vs. 38). Assemblage Group 2 represents assemblages of intermediate composition between Assemblages Groups 1 and 3. The four dominant species in Assemblage Group 1 made up only 65 percent of individuals in Assemblage Group 2. Oligochaetes were much less abundant in Assemblage Group 2, making up less than 3 percent of total individuals. Seven taxa each accounted for more than 2 percent of the total individuals, including *Malenka* sp. (Plecoptera: Nemouridae), *Dipheter hageni* (Ephemeroptera: Baetidae), and *Cinygmula* sp. (Ephemeroptera: Heptageniidae).

Table 6-1. Dominant taxa in assemblage groups

These ten most abundant taxa (as measured by percent composition) in each of the three broad assemblage groups are identified in the multivariate cluster dendrogram.

| Abundance Rank | Assemblage Group 1 (n = 38) | | Assemblage Group 2 (n = 23) | | Assemblage Group 3 (n = 57) | |
|----------------|--------------------------------|---------|--------------------------------|---------|--------------------------------|---------|
| | Taxa | % Comp. | Taxa | % Comp. | Taxa | % Comp. |
| 1 | Chironomidae | 50.7 | Chironomidae | 35.8 | Chironomidae | 18 |
| 2 | Oligochaeta | 16.1 | <i>Baetis</i> sp. | 13.9 | <i>Baetis</i> sp. | 9.4 |
| 3 | <i>Baetis</i> sp. | 13.5 | <i>Simulium</i> sp. | 12.1 | <i>Optioservus</i> sp. | 5.2 |
| 4 | <i>Simulium</i> sp. | 8.7 | <i>Malenka</i> sp. | 3.8 | <i>Ephemerella</i> sp. | 5.2 |
| 5 | <i>Hyaella</i> | 2 | Oligochaeta | 2.8 | <i>Malenka</i> sp. | 5.2 |
| 6 | Turbellaria | 1.4 | <i>Dipheter hageni</i> | 2.6 | <i>Simulium</i> sp. | 5.1 |
| 7 | <i>Paraleptophlebia</i> sp. | 1.1 | <i>Cinygmula</i> sp. | 2.5 | <i>Cinygmula</i> sp. | 5 |
| 8 | <i>Physa/ Physella</i> | 0.6 | <i>Paraleptophlebia</i> sp. | 1.8 | <i>Dipheter hageni</i> | 3.3 |
| 9 | <i>Sperchontidae</i> | 0.4 | <i>Lepidostoma</i> sp. | 1.6 | <i>Epeorus</i> sp. | 2.6 |
| 10 | <i>Isoperla</i> sp. | 0.4 | <i>Ephemerella</i> sp. | 1.5 | <i>Rhyacophila</i> sp. | 2.5 |

Assemblage Group 3 was the most diverse benthic assemblage, with 199 taxa. This group also contained the most sites (57). Sites in this group are closely clustered in ordination space, suggesting that the sites are more similar taxonomically than those within the other two assemblages. The four dominant taxa from Assemblage Group 1 accounted for only 33 percent of individuals in Assemblage Group 3. Oligochaetes made up less than 1 percent of total individuals in this assemblage. Eleven taxa each accounted for more than 2 percent of the total individuals, including eight taxa from the EPT insect orders. Many taxa were found exclusively in this assemblage group, such as *Micrasema* sp. (Trichoptera: Brachycentridae), *Maruina lanceolata* (Diptera: Psychodidae), *Narpus* sp. (Coleoptera: Elmidae), and *Neothremma* sp. (Trichoptera: Uenoidae).

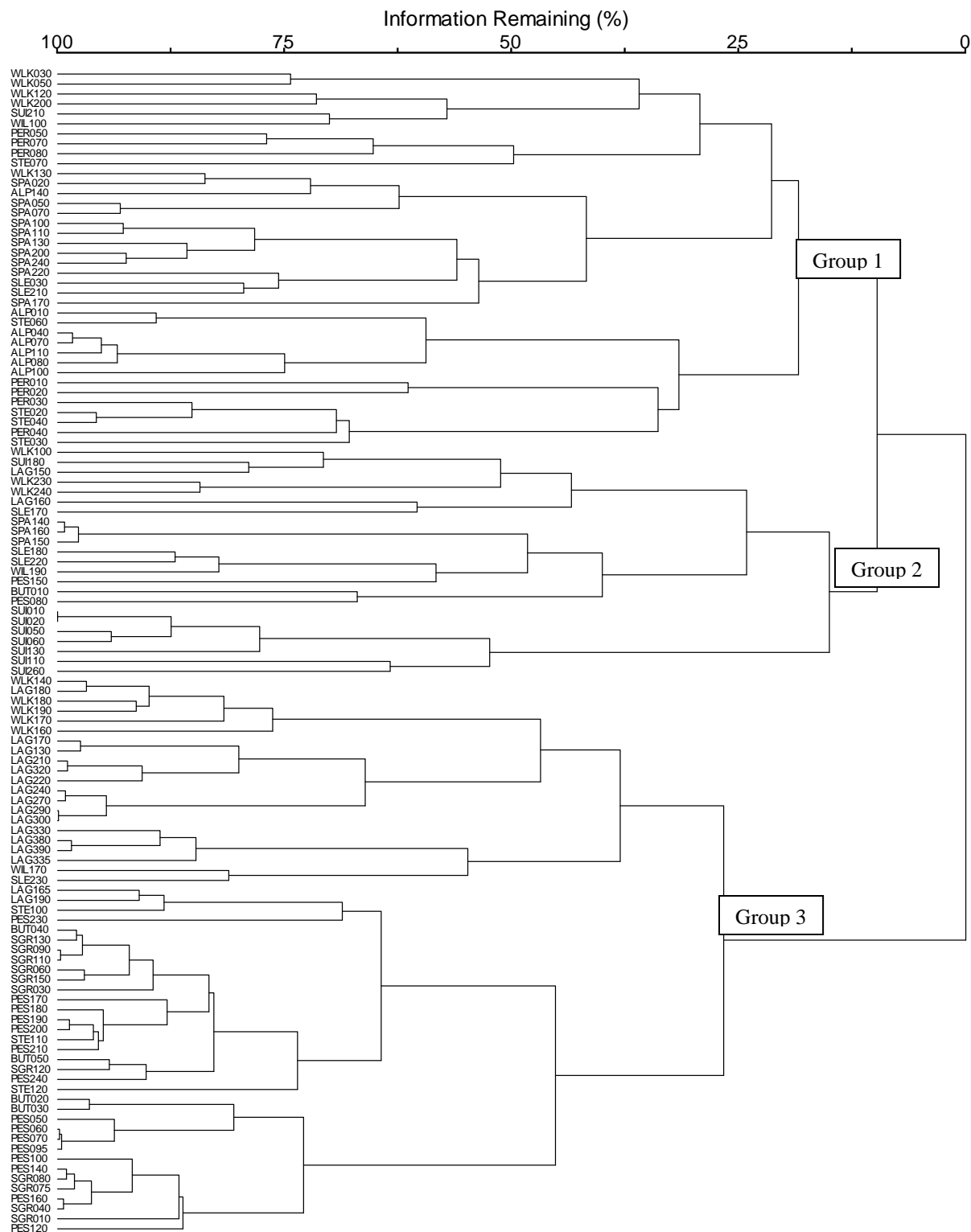


Figure 6-1. Cluster dendrogram of bioassessment sites

Bioassessment sites were classified into three broad assemblage groups based on taxa presence.

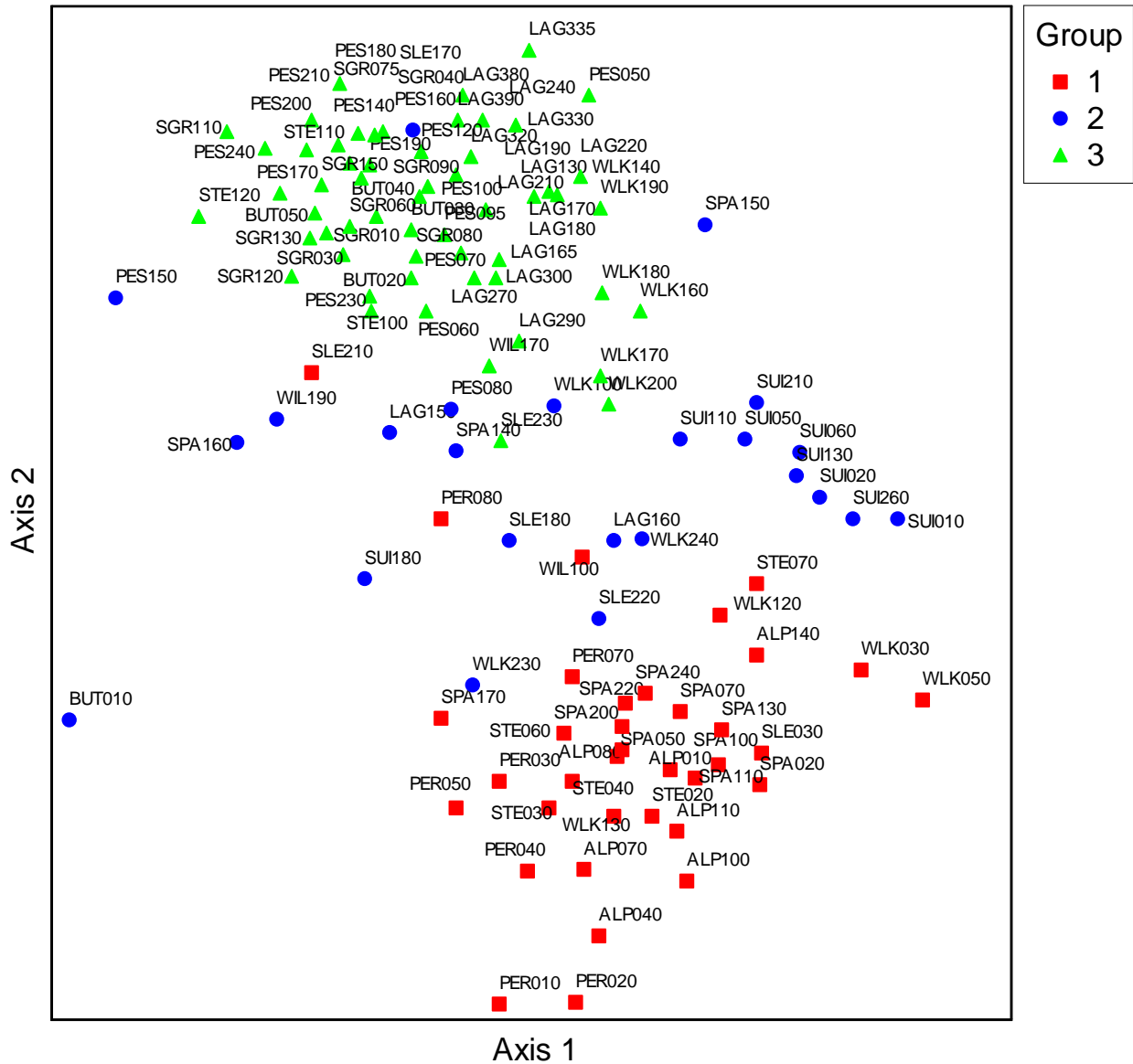


Figure 6-2. NMS ordination of bioassessment sites, labeled by assemblage group

The NMS (non-metric multidimensional scaling) ordination of sites by log-transformed taxa abundances produced a two dimensional solution. Axis 1 and Axis 2 explained 27 percent and 56 percent of original variability, respectively. Sites are coded by assemblage group membership, as determined from the cluster analysis (Figure 6-1).

6.1.1 Watershed Land Use

The assemblage groups described above generally reflect upstream land use. Sites in Assemblage Group 1 are dominated by urban land use, sites in Assemblage Group 2 primarily drain agricultural or grazing land, and Assemblage Group 3 sites include many open space and rural residential sites. This analysis suggests that watershed land use, and associated water quality and habitat conditions, may be an important factor affecting benthic assemblages in Bay Area streams.

6.1.1.1 Land Use Classes

The NMS ordination of log-transformed taxa abundance by site, grouped by land use class, shows clearly distinct clusters of sites within the same land use class (Figure 6-3). Most of the open space sites (red squares) are tightly clustered near the top of the figure. Variability within this class of sites reflects annual variability (as sites were sampled in different years) and geographic variability (due to factors such as flow intermittency, climate, and topography). The tight clustering of the open space sites suggests that, despite the large range in environmental conditions present at these sites, benthic assemblages are very similar relative to assemblages found in other land use classes. The two open space sites that are near the center of the figure, away from the other open space sites, are streams that dry up during the summer.

With the exception of several outliers, rural residential sites (yellow circles) cluster closely to open space sites, indicating that the benthic invertebrate assemblages are fairly similar. The cluster of rural residential sites is centered somewhat lower in the figure than the open space sites, however, suggesting that, on average, the rural residential sites may have slightly altered assemblages compared to the open space sites.

Grazing and agriculture sites are scattered throughout the ordination plot. Some of these sites are located near the open space and rural residential sites in the ordination, while others are more similar to the urban sites. The grazing and agriculture sites generally represent an intermediate level of disturbance between the open space and urban sites. Benthic assemblages at urban sites are consistently different from assemblages found at sites in all other land use classes, and are clustered together at the bottom of the figure.

Many biological metrics exhibited statistically significant differences (One-way ANOVA, paired comparisons using Student's *t*, $\alpha = 0.05$) in the number of EPT taxa between sites draining open space/parkland and every other land use type except rural residential (Figure 6-4). The agriculture, grazing, and mixed land use classes were not significantly different from one another, but the urban land use class was significantly different from all other land use classes.

While the mean number of EPT taxa at all other land use classes ranged from 14 to 22, on average only 2 EPT taxa were found at urban sites. Other biological metrics showed similar patterns with land use. Mean values of selected metrics for each land use class are shown in

Table 6-2. The metrics taxa richness, EPT richness, percent sensitive EPT, percent Coleoptera, and percent intolerant were generally highest in the open space and rural residential land use

types and lowest in the urban land use type. Conversely, the percent non-insect, percent Chironomidae, percent Oligochaeta, and tolerance value metrics were highest in the urban land use type and lowest in the open space land use type. The mean values of the non-insect taxa metric ranged between 5 and 7 and showed little variation between land use types.

Table 6-2. Biological metrics in different land use classes

The mean values of selected biological metric at sites from various land use classes are compared.

| | Agriculture | Grazing | Mixed | Open Space | Rural Res. | Urban |
|----------------------------------|-------------|---------|-------|------------|------------|-------|
| Number of Sites (n) | 9 | 10 | 38 | 13 | 22 | 25 |
| Richness Metrics | | | | | | |
| Taxa Richness | 27 | 28 | 32 | 46 | 40 | 13 |
| EPT Taxa | 14 | 14 | 15 | 22 | 20 | 2 |
| Non-Insect Taxa | 6 | 5 | 7 | 5 | 5 | 6 |
| Composition Metrics | | | | | | |
| % Sensitive EPT | 14 | 28 | 24 | 35 | 36 | 0 |
| % Non-insect | 7 | 9 | 9 | 4 | 5 | 26 |
| % Chironomidae | 37 | 30 | 28 | 14 | 20 | 56 |
| % Coleoptera | 4 | 3 | 7 | 14 | 8 | 0 |
| % Oligochaeta | 3 | 3 | 2 | 0 | 2 | 21 |
| Tolerance Metrics | | | | | | |
| Tolerance Value | 5 | 4.6 | 4.5 | 3.6 | 3.8 | 5.7 |
| % Intolerant | 10 | 19 | 21 | 32 | 32 | 0 |
| Functional Feeding Groups | | | | | | |
| % Collector-gatherer | 64 | 53 | 60 | 37 | 53 | 89 |
| % Collector-filterer | 17 | 12 | 10 | 5 | 8 | 8 |
| % Scraper | 10 | 9 | 15 | 28 | 18 | 1 |
| % Shredder | 4 | 14 | 7 | 13 | 7 | 0 |
| % Predator | 4 | 12 | 8 | 15 | 12 | 1 |

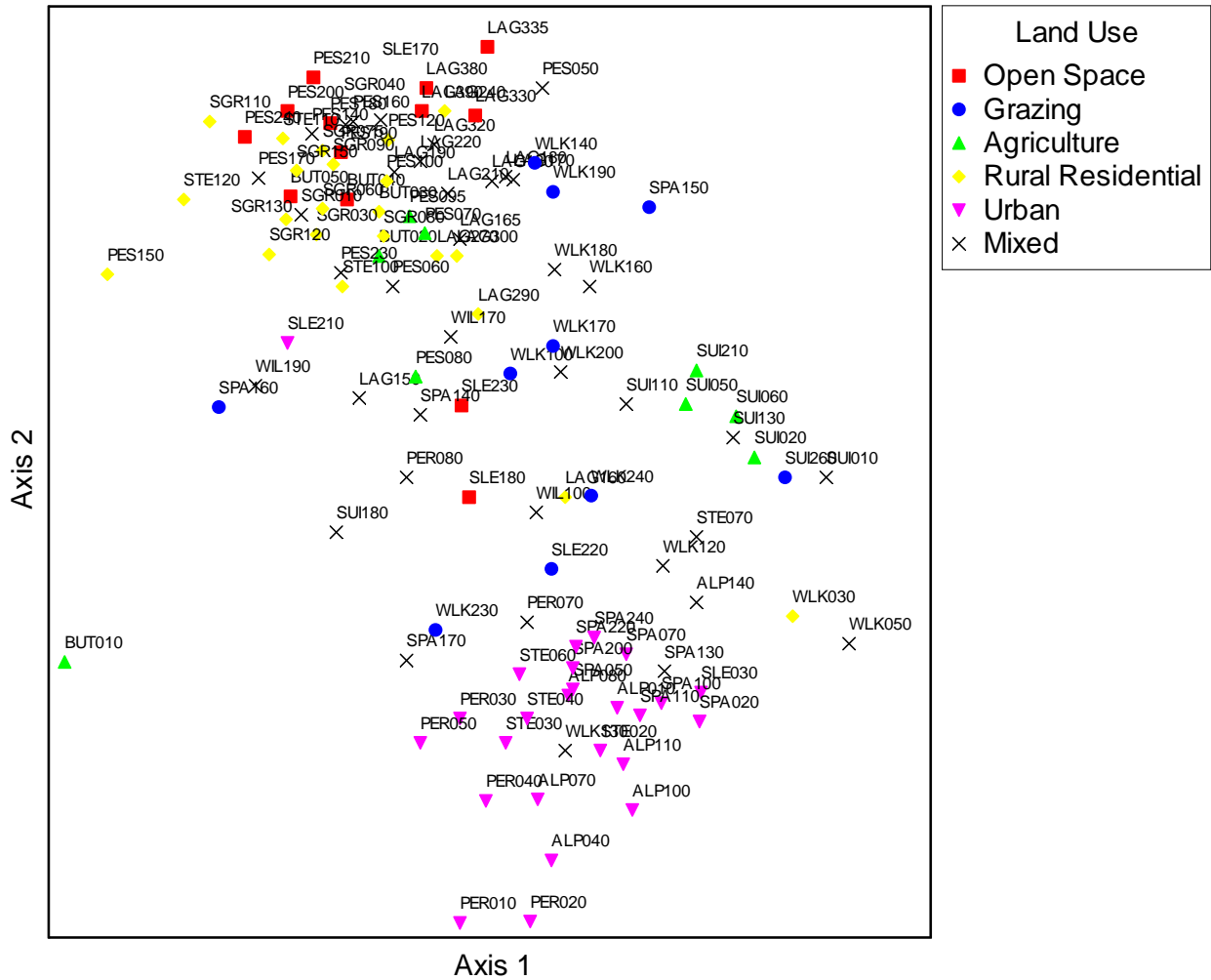


Figure 6-3. NMS ordination of bioassessment sites labeled by land use class

This NMS (non-metric multidimensional scaling) ordination is the same as in Figure 6-2, except sites are labeled by land use class.

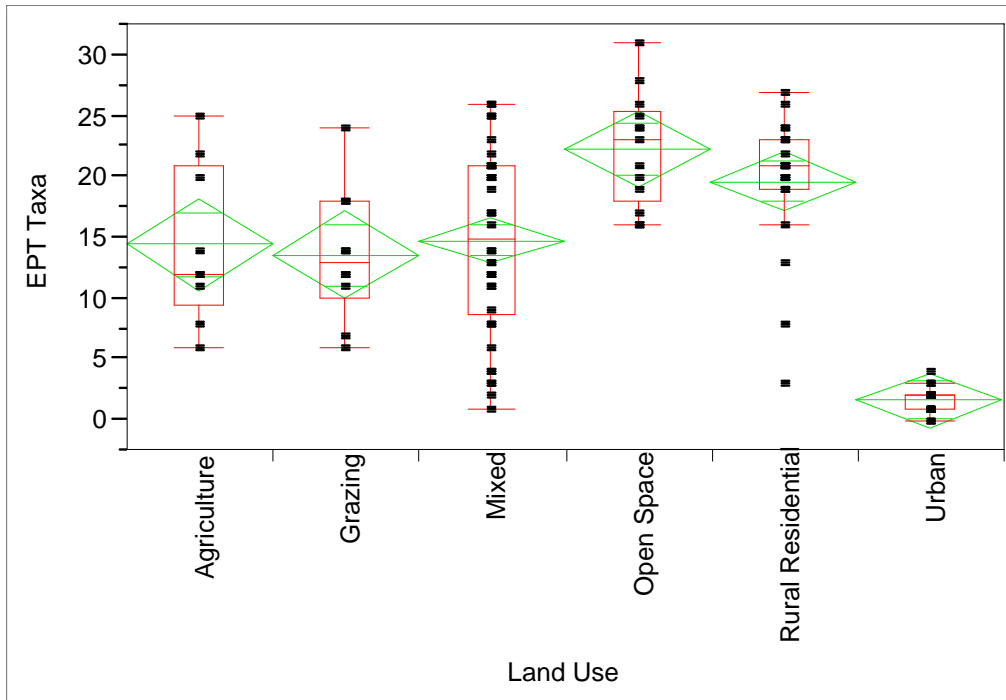


Figure 6-4. Boxplot of EPT taxa richness by land-use class

The box indicates the 75th percentile, 50th percentile (median), and 25th percentile, while the lines above and below the box indicate the 10th and 90th percentiles of the data. The diamond indicates the 95 percent confidence interval about the mean.

6.1.1.2 Channel and Riparian Conditions

To examine the relationships between local physical habitat conditions and benthic assemblages, correlations between P-HAB scores and taxa presence were examined using multivariate analyses. P-HAB scores are qualitative observations of reach-scale physical habitat quality and alteration. P-HAB scores were correlated with an NMS ordination of taxa presence at 87 sites that possessed complete P-HAB and benthic assemblage data (Figure 6-5).

Axis 1 of the ordination differentiates between the minimally disturbed sites (clustered on the right-hand side of the graph) and significantly disturbed sites in urban areas (on the left hand side of the graph). The bi-plot (the two red lines on Figure 6-5) shows the magnitude and direction of the correlation between two of the physical habitat variables and the ordination of sites. By convention, the origin of the lines of the bi-plot is the origin (mean score) of the ordination of sites. The bi-plot shows that the PHAB variables epifaunal substrate and channel alteration are positively correlated with Axis 1, and highest at the minimally disturbed sites (on the right hand side of the graph), meaning these sites tend to have the best epifaunal substrate and the least

channel alteration. Stream channels in urban areas are often highly altered, while channels at minimally disturbed sites are rarely modified.

Many PHAB variables, such as epifaunal substrate, channel alteration, bank vegetation, and riparian zone width were correlated ($r^2 > 0.35$) with Axis 1 of the ordination, while channel flow status and bank stability were least correlated with the ordination of taxa presence (Table 6-3). Other than 'channel flow status', PHAB variables were not strongly correlated with Axis 2, suggesting that this axis of the ordination does not represent physical and biological integrity as well as Axis 1.

In a study of streams with a range of human impacts in the Santa Clara Valley, Kearns *et al.* (2001) found that riffle frequency and riparian zone width were most highly correlated with biological condition, while bank vegetation and channel alteration showed weak correlation. In contrast, channel alteration was the most highly correlated PHAB variable with biological conditions in this SWAMP study (Table 6-3). Although channel alteration is highly correlated with biological integrity, identifying causal factors is difficult, as other factors (such as water quality) may have equal or more significant impacts on benthic assemblages.

Table 6-3. Correlation of P-HAB scores with macroinvertebrate ordination axes

| | Axis 1 | Axis 2 |
|-------------------------|--------|--------|
| Channel Alteration | 0.456 | 0.002 |
| Riparian Zone Width | 0.366 | 0.053 |
| Epifaunal Substrate | 0.49 | 0.041 |
| Embeddedness | 0.278 | 0.015 |
| Velocity/ Depth Regimes | 0.172 | 0.032 |
| Sediment Deposition | 0.261 | 0.017 |
| Channel Flow Status | 0.047 | 0.299 |
| Riffle Frequency | 0.343 | 0.004 |
| Bank Stability | 0.048 | 0.05 |
| Bank Vegetation | 0.372 | 0.037 |

The values listed represent the correlation (r^2) of P-HAB variables with axes from an NMS ordination (Figure 6-5) of taxa presence at 87 sites where P-HAB scores were available.

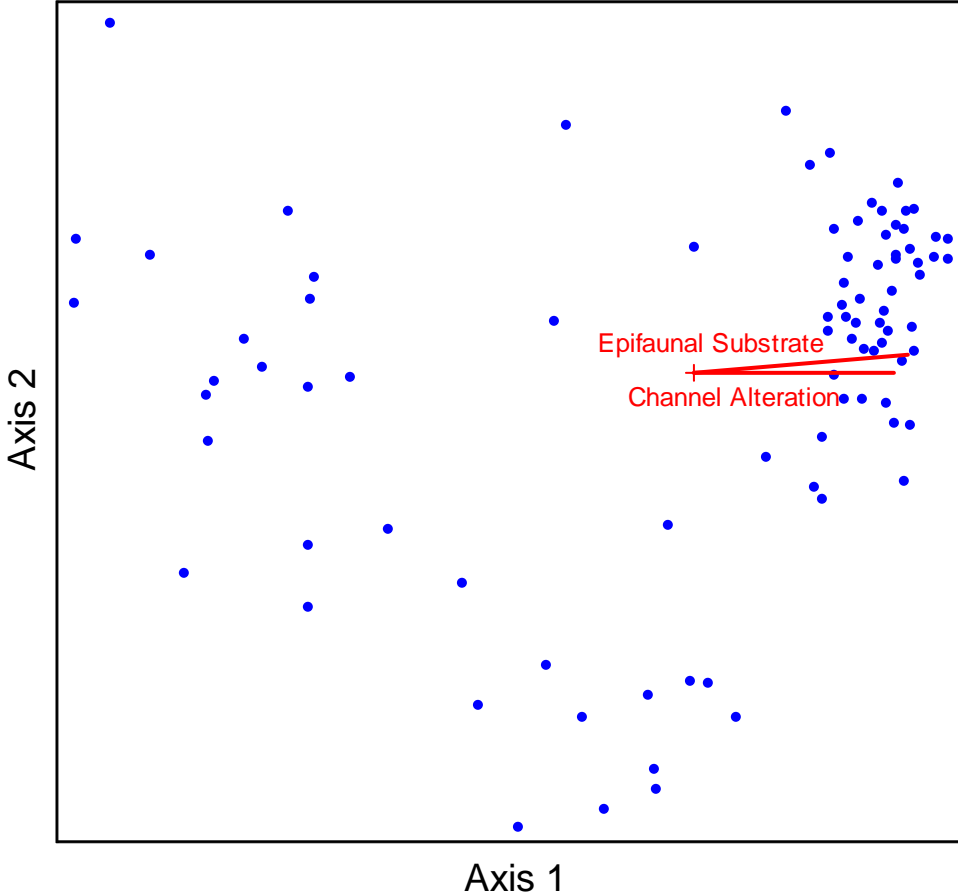


Figure 6-5. NMS ordination of bioassessment sites, with bi-plot of P-HAB variables

The NMS (non-metric multidimensional scaling) ordination of 87 sites by taxa presence produced a two dimensional solution. Axis 1 and 2 explained 74 percent and 13 percent of the variation in the original data, respectively. The sites clustered on the right-hand side of the graph are the minimally disturbed sites. The bi-plot (red lines) shows the magnitude and direction of the correlation between physical habitat variables and the ordination of sites. By convention, the origin of the lines of the bi-plot is the origin (mean score) of the ordination of sites. The bi-plot shows that the PHAB variables epifaunal substrate and channel alteration are highest at the minimally disturbed sites, meaning these sites tend to have the best epifaunal substrate and the least channel alteration. For clarity, the bi-plot only shows the P-HAB variables with correlation coefficients greater than 0.400. Other P-HAB variables were also highly correlated with axis 1 (see Table 6-3).

6.1.1.3 Water Operations/Impoundments

Sampling sites in several watersheds monitored by SWAMP were located immediately upstream and downstream of dams and reservoirs. In the Stevens Creek watershed, Stevens Creek Reservoir is located in the upper part of the watershed, draining mostly parkland. The reservoir serves as storage for groundwater recharge, in addition to providing recreational fishing and boating opportunities. Site STE100 was located above the reservoir in Stevens Creek County Park, while site STE070 was located directly downstream of the reservoir, also in Stevens Creek County Park. At the upstream site, 44 taxa were found; below the dam, taxa richness was only

17. Other common biological metrics followed this trend, with much lower values at the downstream site than the upstream site (Table 6-4). It is problematic to attribute downstream changes in biological condition directly to the effects of the reservoir, however, because a gravel quarry is located immediately adjacent to the reservoir. The dramatic decline in taxa of the order Ephemeroptera, generally believed to be intolerant of metals pollution, from 10 above the reservoir to 1 below the reservoir provides circumstantial evidence that runoff from the quarry may also have a significant effect on downstream biological conditions.

Four dams are located in the upper watershed of Lagunitas Creek, forming a sequence of four large reservoirs: Lake Lagunitas, Bon Tempe Lake, Alpine Lake, and Kent Lake. All four reservoirs, and their surrounding watershed lands, serve as water supply for southern Marin County. Four sites were located on undammed small tributaries to the reservoirs: LAG330, LAG335, LAG380, and LAG390. One site, LAG320, was located on Lagunitas Creek below the downstream dam, Shafter Dam. Although all 5 sites drain protected watershed lands, the upstream and downstream sites may not be directly comparable because of differences in stream size. Mean values of metrics from the upstream sites were very similar to values from the downstream sites (Table 6-4).

Other studies in the Bay Area have found that dams can have significant impacts on downstream biotic assemblages. Choy (2004) found that streams in the Santa Clara Valley below dams had higher levels of fine sediment and fewer EPT organisms than unregulated streams. Weissich (2005) observed that benthic invertebrate assemblages were much more variable over time in the same regulated streams in the Santa Clara Valley than in unregulated streams. Watersheds monitored by SWAMP contained dams, but upstream-downstream comparisons were less valid due to other confounding factors. A range of biological conditions are observed downstream of dams, suggesting that dam type and management play an important role in determining downstream biological effects.

Table 6-4. Biological metrics upstream and downstream of dams

Mean values of selected biological metrics are compared above and below dams.

| | Stevens Creek | | Lagunitas Creek | |
|----------------------------------|---------------|------|-----------------|------|
| | Up | Down | Up | Down |
| Richness Metrics | | | | |
| Taxa Richness | 44 | 17 | 42 | 41 |
| EPT Taxa | 20 | 4 | 19 | 20 |
| Composition Metrics | | | | |
| % EPT | 53 | 33 | 50 | 60 |
| % Sensitive EPT | 44 | 4 | 28 | 38 |
| % Chironomidae | 30 | 53 | 20 | 23 |
| % Coleoptera | 2 | 0 | 11 | 9 |
| Tolerance Metrics | | | | |
| % Intolerant | 40 | 4 | 26 | 41 |
| % Tolerant | 1 | 4 | 1 | 3 |
| Tolerance Value | 4 | 6 | 4 | 4 |
| Functional Feeding Groups | | | | |
| % Predator | 9 | 6 | 18 | 5 |

| | | | | |
|----------------------|----|----|----|----|
| % Collector-filterer | 3 | 3 | 2 | 4 |
| % Collector-gatherer | 73 | 84 | 38 | 47 |
| % Scraper | 7 | 0 | 22 | 27 |
| % Shredder | 8 | 4 | 20 | 16 |

6.2 Field Measurements of Basic Water Quality

Multi-parameter probes on YSI 6600 sondes measured pH, specific conductance, temperature, and dissolved oxygen (DO) every 15 minutes at sites in each watershed during deployments. These data allow characterization of the temporal variability in water quality conditions that affect aquatic life, including salmonid species. The data from continuous monitoring of basic water quality field measurements are presented in Appendix D.

6.2.1 Regional Trends

Region-wide, highest weekly average temperatures (HWATs) exceeded 17°C, a threshold for steelhead, at 25 of 114 station-season combinations. Most exceedances occurred in summer. Of the 13 coastal sites with HWATs, 12 exceeded the 14.8°C limit for coho salmon. Fifty percent of the 54 sites monitored had DO concentrations below 7 mg/L, the Basin Plan objective for coldwater habitat. The San Gregorio Creek watershed had the lowest rate of DO violations with 17 percent, while Arroyo Las Positas had the highest rate, with 75 percent of sites measuring below the Basin Plan thresholds. The great majority of high temperature and low DO values occurred during the summer deployments.

There were statistically significant relationships between physical habitat parameters and the in-stream measures of temperature and DO. Four key parameters were analyzed for their relationships to DO and temperature, and correlations with all four were statistically significant ($p < 0.05$). These four physical habitat parameters were indices of channel alteration, average bank vegetation, average riparian zone width, and average canopy cover. The four physical habitat parameters tended to co-vary: channel alteration correlated significantly with bank vegetation and riparian zone width ($p < 0.001$), and bank vegetation also correlated significantly with riparian zone width and canopy cover ($p < 0.001$ and 0.05 , respectively). The relationships of temperature and DO with channel alteration and canopy cover are shown below (Figure 6-6 and Figure 6-7), though similar trends occurred with average bank vegetation and average riparian zone width. The relationships appear important because the temperature and DO curves cross threshold lines. Sites with high indices for channel alteration (low alteration) and canopy cover tend to have acceptable temperature and DO values, while sites with more altered channels and less canopy cover tend to exceed salmonid temperature thresholds or violate Basin Plan DO objectives.

Values for continuously measured parameters were compared based on the land uses that characterized the areas draining to the measurement sites. The five sites draining open space land use were not sampled during the dry season, and of the three grazing sites, two have DO data for the dry season. Data are presented for the dry season monitoring, when the most guideline exceedances were measured. Rural residential areas produced the lowest dry season water temperatures, as indicated by 'highest weekly average temperature' values (HWAT), while agricultural and urban drainages had the highest water temperatures (Figure 6-8). Mixed land-use areas produced intermediate and more variable HWAT values. A similar pattern emerged from the dissolved oxygen data. Streams draining rural residential areas had higher minimum dissolved oxygen concentrations, while DO was lower in agricultural and urban drainages and intermediate in the mixed land use areas (Figure 6-9). Data from one site draining grazing land

had a high minimum dry season DO concentration. The site, WLK140, is in a canyon which is well-shaded and has a wide riparian zone. The other grazing site with summer DO measurements, LAG085 (Five Brooks), just barely fell below the 7 mg/L minimum.

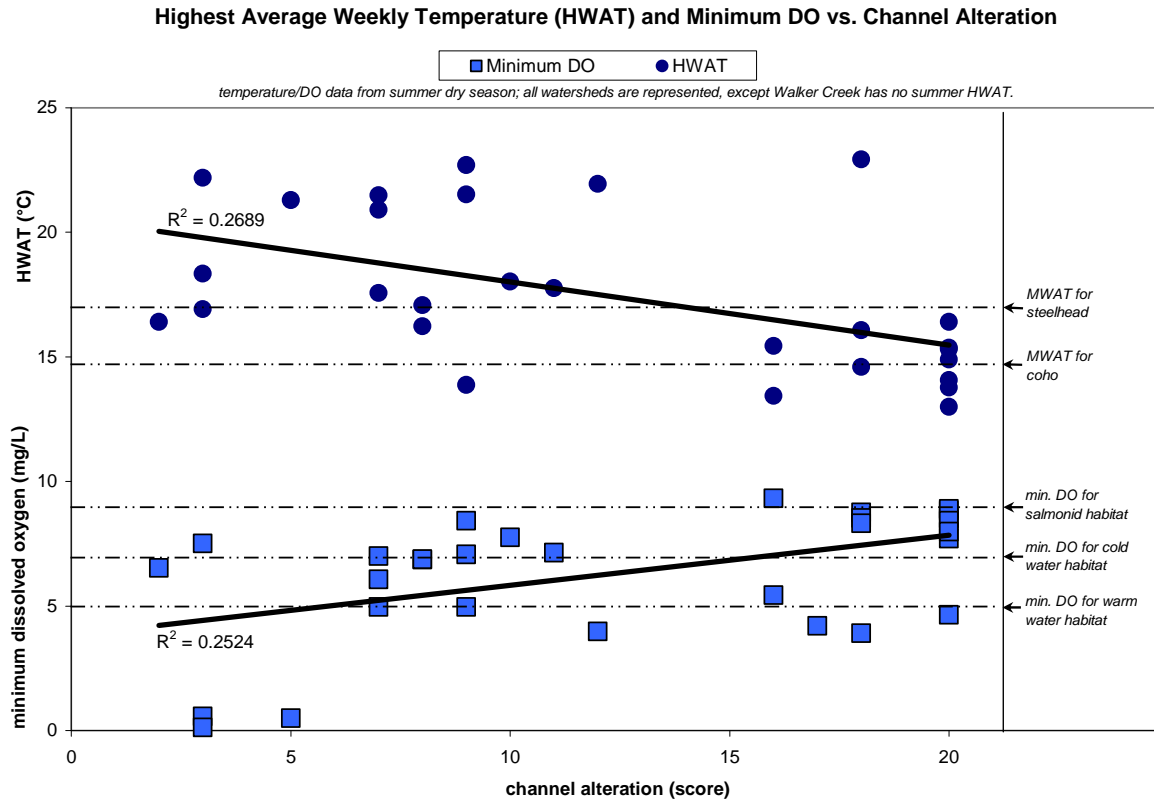


Figure 6-6. HWAT and minimum DO versus channel alteration

High channel alteration scores indicate channels that are least disturbed. Dashed lines indicate salmonid temperature and dissolved oxygen (9mg/L) thresholds and Basin Plan objectives for dissolved oxygen (7mg/L for cold water habitat, and 5mg/L for warm water habitat). Because sites on Walker Creek streams were monitored for less than seven days during the critical summer period, the HWAT for Walker is not represented; however the minimum DO from one Walker Creek site is included

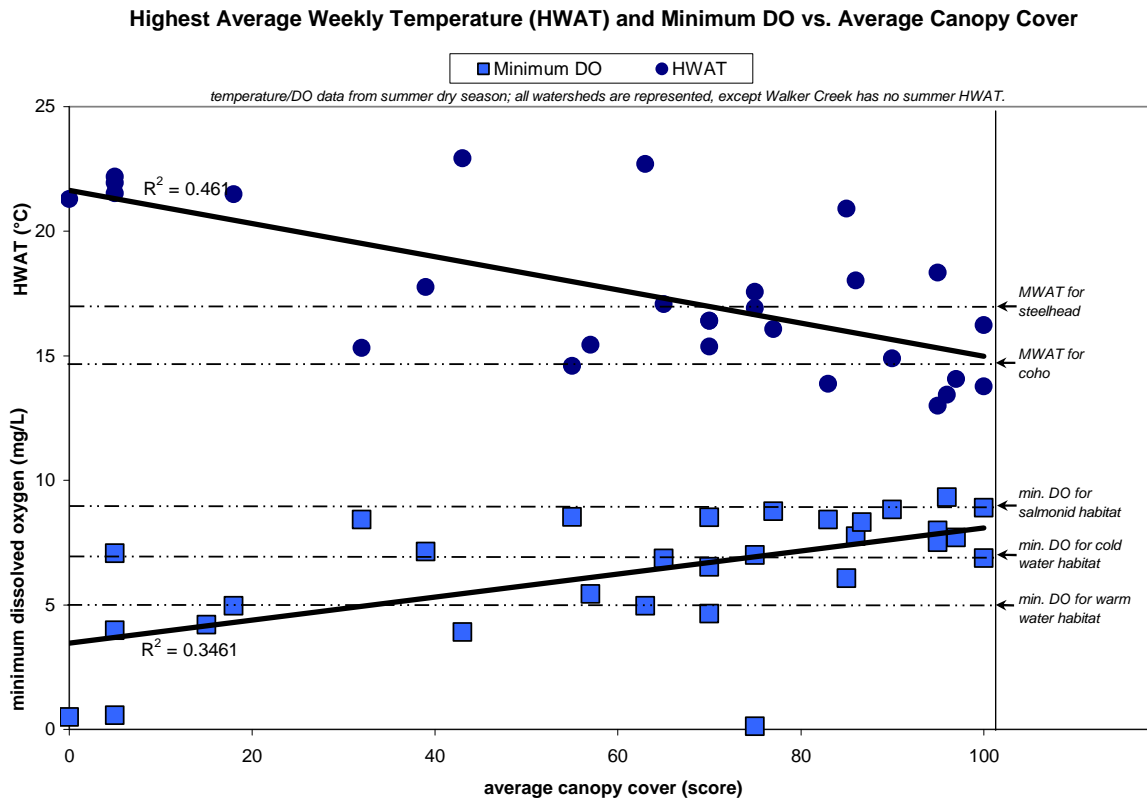


Figure 6-7. HWAT and minimum DO versus canopy cover

This graph represents relationships between average canopy-cover scores and continuously measured temperature and dissolved oxygen values. High average canopy cover scores indicate greater canopy cover. Dashed lines indicate salmonid temperature and dissolved oxygen (9mg/L) thresholds and Basin Plan objectives for dissolved oxygen (7mg/L for cold water habitat, and 5mg/L for warm water habitat).

Regional highest average weekly temperature (HWAT) vs. land use

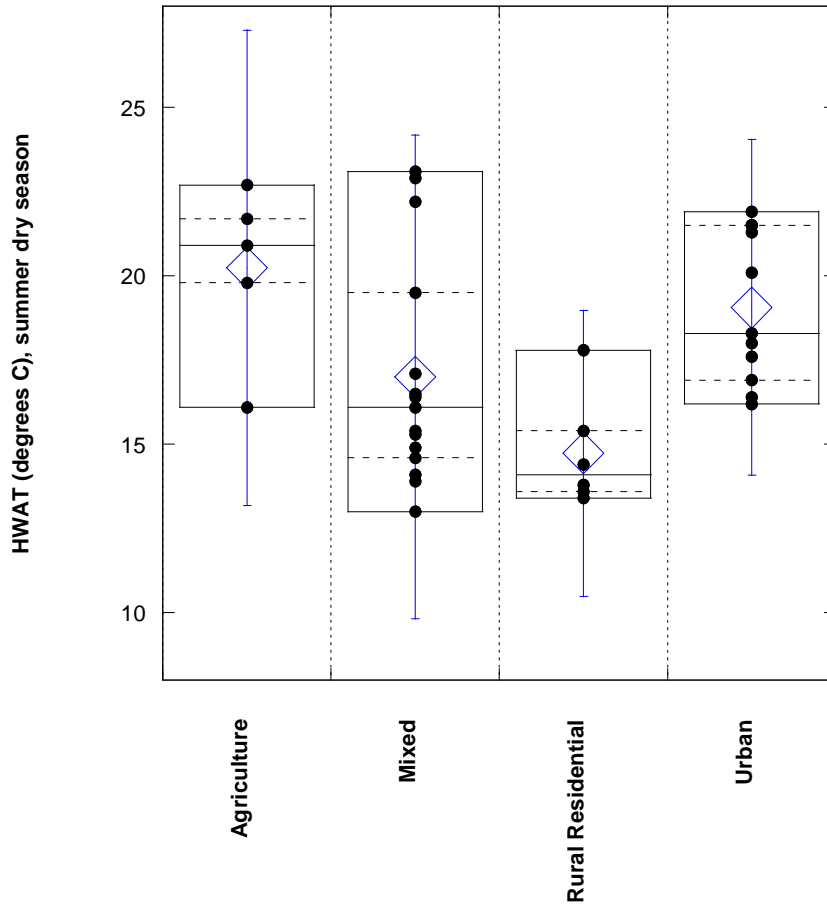


Figure 6-8. Regional HWAT versus land use

Highest average weekly temperature (HWAT) values are grouped by land-use category assigned to areas draining to the stream sampling site. Black dots represent all data points. Black boxes represent percentiles: 95, 75, 50, 25, 5. Center of the blue diamond represents the mean, with error bars representing the 95 percent confidence interval about the mean (two-tailed, $\alpha=0.05$).

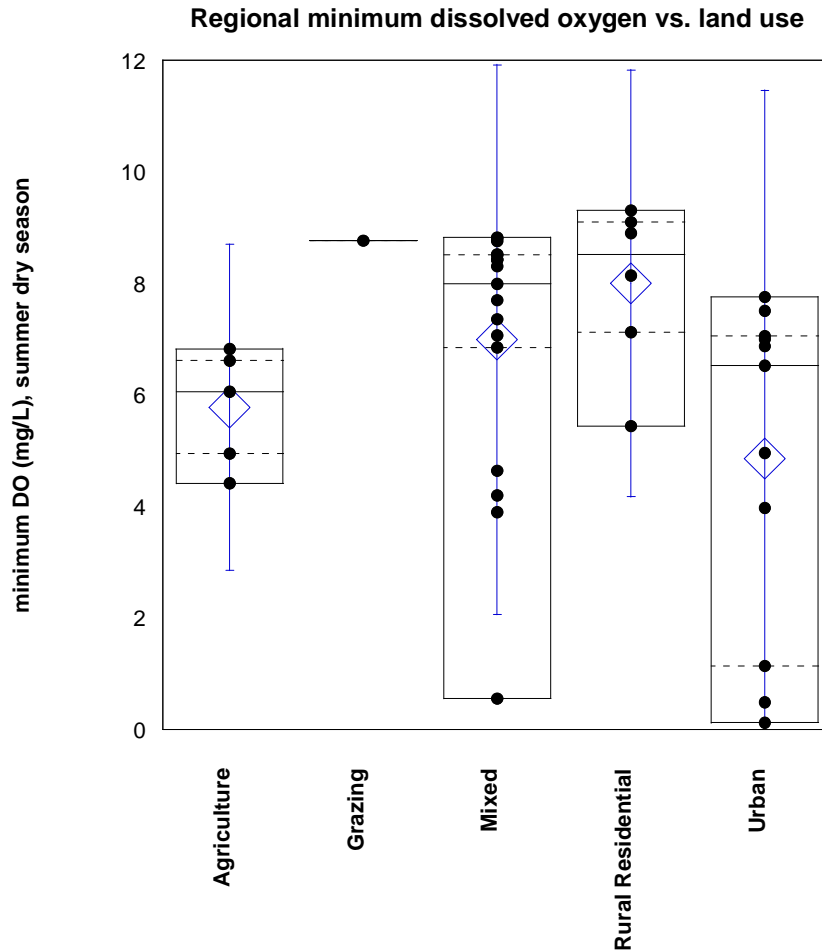


Figure 6-9. Regional minimum DO versus land use

Continuously monitored dissolved oxygen values are grouped by land-use category assigned to areas draining to the stream sampling site. Black dots represent all data points. Black boxes represent percentiles: 95, 75, 50, 25, and 5. Center of the blue diamond represents the mean, with error bars representing the 95 percent confidence interval about the mean (two-tailed, $\alpha=0.05$).

6.2.2 Temporal Patterns

For parameters with high daily or seasonal variability such as temperature and DO, periodic, continuous monitoring captures ranges and extremes that would be missed by discrete samples. The field parameters' cyclical temporal variability is driven by solar radiation, resulting in characteristic daily and seasonal patterns. Stochastic events, both natural and anthropogenic, are often discernible against these background patterns. A generalized trend from upstream to downstream and from minimally impaired to degraded conditions is reflected by increasing temperature, pH, and specific conductance, and by decreasing DO. Human impacts to streams may alter the natural patterns, typically by increasing temporal variation and decreasing spatial variation (Poole, Dunham *et al.*

2001). The patterns of the temperature and DO cycles characterize the water quality of a site and indicate whether photosynthesis or temperature dominates the DO cycle. The timing of the DO minima and maxima differ between productive and non-productive sites.

The variability of water quality field parameters is summarized in boxplots that describe the distribution of data points for each sampling period, with extremes marked by the whiskers, a central tendency represented by the interquartile range or middle 50 percent of the distribution (from the 25th to the 75th percentile), and a median. In section 7, the results are represented by watershed using boxplots for each parameter, excluding sampling events with rejected data. In sections 6.2.2.5, 6.2.2.6 and 6.2.2.7, a few time series graphs illustrate interpretations of the temporal patterns. The temperature and DO curves compare minimally-impaired conditions to more degraded conditions with characteristic temporal patterns which indicate whether photosynthesis or temperature dominates the DO cycle.

6.2.2.1 Temperature

Of the core field parameters measured—temperature, DO, pH, and specific conductance—temperature is fundamental, affecting each of the other parameters. It has the widest range of influences and effects, a complexity of spatial and temporal variability, the most consistent daily cycle, and is easiest to measure. Daily temperature fluctuations may be higher in smaller tributaries because of lower flow volume, or stable if the stream is fed primarily by groundwater. Stream reaches without substantial groundwater supply and with minimal riparian canopy, whether naturally or from altered land use, will experience greater daily temperature fluctuations as well as elevated temperatures. As with each parameter, the many possible confounding influences call for individual sites to be evaluated according to their specific conditions. Seasonal stream temperature variations are essentially a dampened curve of the mean monthly air temperature (Stevens *et al.* 1975).

The Basin Plan water quality objective states that any alteration of water temperature must be demonstrated neither to adversely affect beneficial uses, nor to increase natural stream temperature by more than 2.8 C°. The requirements for salmonids, which constitute the most limiting beneficial use for most streams, are complex: thresholds are specific to species, life history stages, stream reaches, and varying seasonal time periods. In lieu of the current narrative water quality objectives, this report uses three screening threshold levels to evaluate time series temperature data throughout the region: a salmonid survival limit as a daily maximum of 24°C, a steelhead Maximum Weekly Average Temperature (MWAT) of 17°C, and a coho salmon MWAT of 14.8°C (see Appendix C). These thresholds are used only as comparative tools, not as substitute water quality objectives.

6.2.2.2 Dissolved Oxygen

Dissolved oxygen (DO), which is critical for aquatic life, is variable and dependent on temperature and photosynthesis, as well as respiration, nitrification, chemical oxidation,

salinity, and atmospheric pressure. As with temperature, the DO requirements for salmonids and other aquatic life are complex and variable, both spatially and temporally.

The Basin Plan lists numeric water quality objectives for DO minima of 5 mg/L for warmwater habitats and 7 mg/L for coldwater habitats, with a 30-day average minimum of 80 percent saturation. The 9 mg/L threshold noted in the tables and maps of this report is an indication of good spawning habitat for salmonids, not a water quality objective.

Sites with high productivity tend to have low DO minima, whereas sites with moderate or low productivity tend to have higher DO minima and more stable overall water quality. Daily ranges of DO greater than 2 mg/L often indicate high productivity (excessive photosynthesis) with potential for nighttime hypoxia with associated fish kills. Four out of five sites in Arroyo Las Positas had DO levels below the minimum for warmwater habitat and maximum DO percent saturation levels above 120 percent, even up to 395 percent (ALP150 in the spring). Such supersaturation, together with high pH values, indicates high productivity associated with eutrophication.

Inspection of the DO curves yields additional understanding of ambient conditions at a specific site. The timing of the daily DO maxima tends to be distinctly earlier (before noon, just as the temperature rises) at temperature-dominated sites compared to photosynthesis-dominated ones, where the DO maxima tends to be at noon and up to a few hours later.

6.2.2.3 pH

The pH level provides an indirect measure of water quality by reflecting the status of chemical equilibrium reactions. Instream photosynthesis and respiration are the major drivers of the pH cycle through their effect on the concentration of carbon dioxide and the consequent equilibrium of the carbonate cycle (carbon dioxide ↔ carbonic acid ↔ bicarbonate ion ↔ carbonate ion). The buffering capacity of the carbonate system is the major regulator of the pH level. Typically, the pH cycle closely follows the DO cycle. The Basin Plan water quality objective for pH ranges between 6.5 and 8.5. None of the continuously monitored sites had pH below 6.5, but values above 8.5 occurred at 15 of 56 sites. Two sites had pH above 9 (WLK160 with 9.07, and ALP150 with 9.25). Most of the pH values above 8.5 occurred in the winter or spring seasons.

The Basin Plan Objective for pH states that “controllable water quality factors shall not cause changes greater than 0.5 units in normal ambient pH levels.” Normal ambient pH levels have not been established for Bay Area watersheds. Nevertheless, changes in pH greater than one unit and values above 8.5 suggest excessive photosynthesis and possible nutrient enrichment, especially when DO percent saturation is above 120 percent. All five sites in Arroyo Las Positas exceeded pH 8.5; four of them had DO saturation above 120 percent, and the fifth (ALP010) reached 119 percent.

6.2.2.4 Specific Conductance

Specific conductance provides an indirect measure of total dissolved solids (TDS) as an indicator of pollution or salinity levels. Electrical conductivity is a measure of how much

electricity is conducted through one centimeter of water at a specified temperature. Because electrical conductivity changes with temperature, comparisons across seasons and streams would be confounding. The use of specific conductance avoids this confusion by compensating all conductivity to 25°C. The YSI meters used in this study measure electrical conductivity and from that calculate specific conductance using associated temperature and a temperature coefficient of 0.0191 (YSI 1999). Freshwater specific conductance is measured in microSiemens per centimeter ($\mu\text{S}/\text{cm}$).

The specific conductance of a site tends to be characteristic and not to vary much daily; however, winter and spring rains will lower the values and increase the range compared to summertime, and values tend to increase with lower summer flow. The Basin Plan has no water quality objectives for specific conductance to protect aquatic life, but typically values above 850 $\mu\text{S}/\text{cm}$ are considered evidence of impact from land-use activity (Ketcham 2002). Sudden changes in specific conductance suggest a discharge, precipitation and runoff, or tidal influence (Figure 6-14). First flush after the dry season will typically show a spike in specific conductance, as accumulated residues are washed into streams before the diluting effect of freshwater lowers the levels.

6.2.2.5 Comparison of temporal patterns at two downstream sites

BUT010 (Bean Hollow) and PES060 (Community Church) are sites in comparable lower portions of adjacent watersheds (see map of Butano and Pescadero creeks, Figure 3-7). Only two and a half days of monitoring in early June, 2003, are compared in Figure 6-10. The monitoring probes were not post-calibrated for PES060, so the values were not used for official assessment in this report; nevertheless, the patterns are instructive. BUT010 has an impoverished invertebrate population and thus strikingly low bioassessment metrics compared to PES060. The pattern of temperature and DO timing, however, suggests that the water quality at BUT010 is better than PES060, perhaps because of full shade from riparian willows. Sedimentation with sand and fine sediment is likely the limiting factor affecting aquatic habitat for benthic macroinvertebrates and periphyton.

For the short time compared, the maximum temperature at BUT010 is equal to the minimum temperature at PES060 (16.32°C), and the average daily temperature range at PES060 (6.2 C°) is nearly 2 C° more than at BUT010 (4.3 C°). Although the sites are less than a mile from each other, PES060 has a two-hour delay in the average temperature minimum and about a one-and-a-half hour delay in the average temperature maximum in comparison to BUT010. The difference in riparian vegetative cover is most likely the predominant contributor to these differences.

The average daily DO range at PES060 (2.63 mg/L), more than six times the range at BUT010 (0.42 mg/L), suggests excessive photosynthesis. The average timing of DO maxima at PES060 is mid-afternoon (14:45), a typical pattern for a DO cycle dominated by photosynthesis, whereas the average time for maximum DO at BUT010 is before noon (11:15), a pattern more typical for a DO cycle dominated by temperature.

Inspection of the temperature and DO curves in Figure 6-10 provides the most straightforward clue to the difference between the sites: PES060 has a nearly flattened

DO curve at the bottom, rising very slowly in response to decreasing temperatures until after dawn when solar radiation powers photosynthesis and a steep rise in DO. In comparison, the DO curve for BUT010 rises as soon as it hits minimum sometime in the early evening (an average of 19:45 in this case), responding to decreasing temperatures. These patterns are characteristic of a site where photosynthesis dominates the DO pattern (PES060) and a site where temperature dominates the DO pattern (BUT010). Nutrient enrichment and lack of riparian cover are two prominent factors supporting the pattern of high productivity.

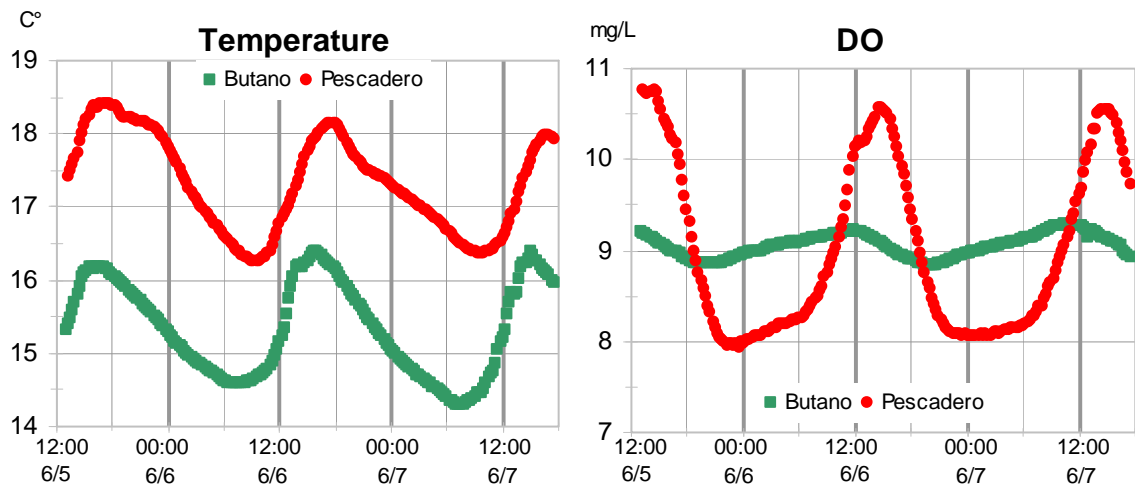


Figure 6-10. BUT010 and PES060 temperature and DO patterns

These are comparable downstream sites for Pescadero and Butano creeks (see map, Figure 3-7), monitored in June, 2003. The difference in DO patterns indicates photosynthesis-dominated timing at PES060 compared to the more shaded habitat at BUT010, where the effect of temperature on DO is evident.

6.2.2.6 Comparison of temporal patterns between upstream and downstream sites

A similar comparison of the temperature and DO patterns during the critical summer period illustrates a characteristic difference between an upstream, higher gradient, well-shaded site at LAG190 (Devils Gulch, a tributary with known coho salmon spawning habitat) and a downstream, lower gradient, less-shaded site at LAG050 (Bear Valley Road Bridge on Olema Creek, the most downstream station monitored for continuous field parameters in the Lagunitas Creek watershed).

For the 11.5-day period sampled in August, 2002, 82 percent of the DO maxima at downstream LAG050 occurred from 14:45 to 15:45, just before the temperature maxima, a timing characteristic of photosynthesis-dominated DO cycles. Upstream at LAG190, however, 64 percent of the maxima occurred from 08:30 to 09:45, just following the temperature minima, a timing characteristic of temperature-dominated DO cycles. At LAG050, 75 percent of the DO minima occurred from 21:45 to 23:30; upstream at LAG190, 92 percent of the DO minima occurred from 19:00 to 20:45. There is a consistent timing pattern of DO maxima before noon and minima in the early evening at

minimally impaired, well-shaded sites. Photosynthesis-driven sites tend to have DO maxima in mid-afternoon and minima later at night or early in the morning.

Graphically, the photosynthesis-dominated pattern more common at downstream sites is apparent as a DO curve that closely follows the temperature curve, but peaks slightly earlier and lingers at the bottom, rising slowly until dawn, when the increase is suddenly rapid (Figure 6-11).

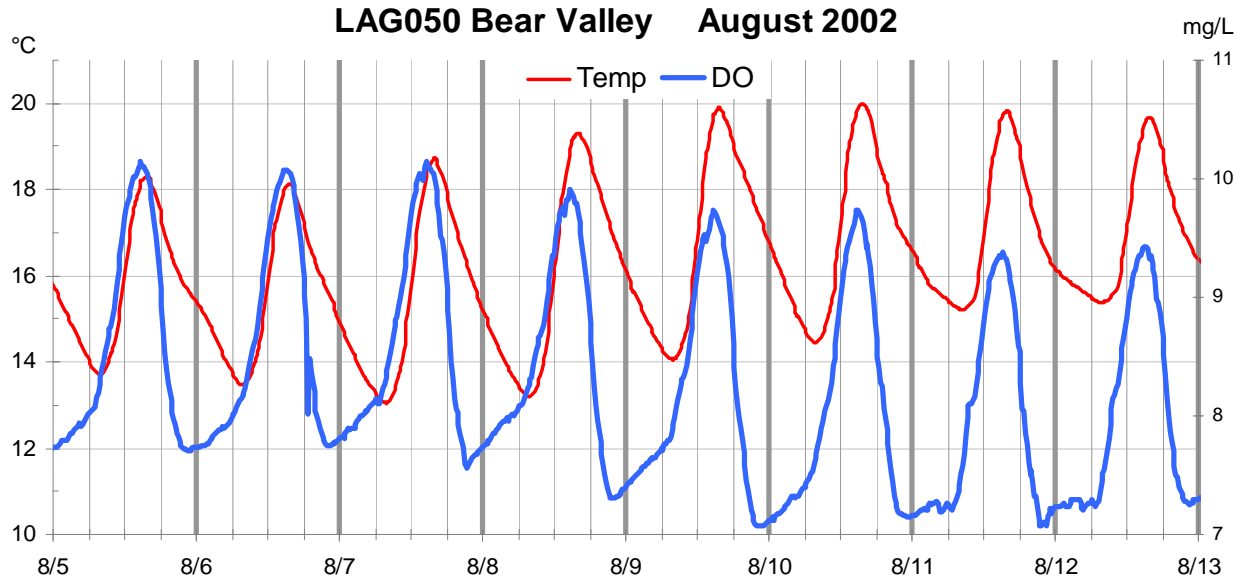


Figure 6-11. LAG050 temperature and DO patterns

By comparison, the temperature-dominated pattern of well-shaded sites more common at minimally disturbed sites upstream is apparent as a DO curve that is nearly the inverse of the temperature curve and recovers rapidly from its minimum (Figure 6-12).

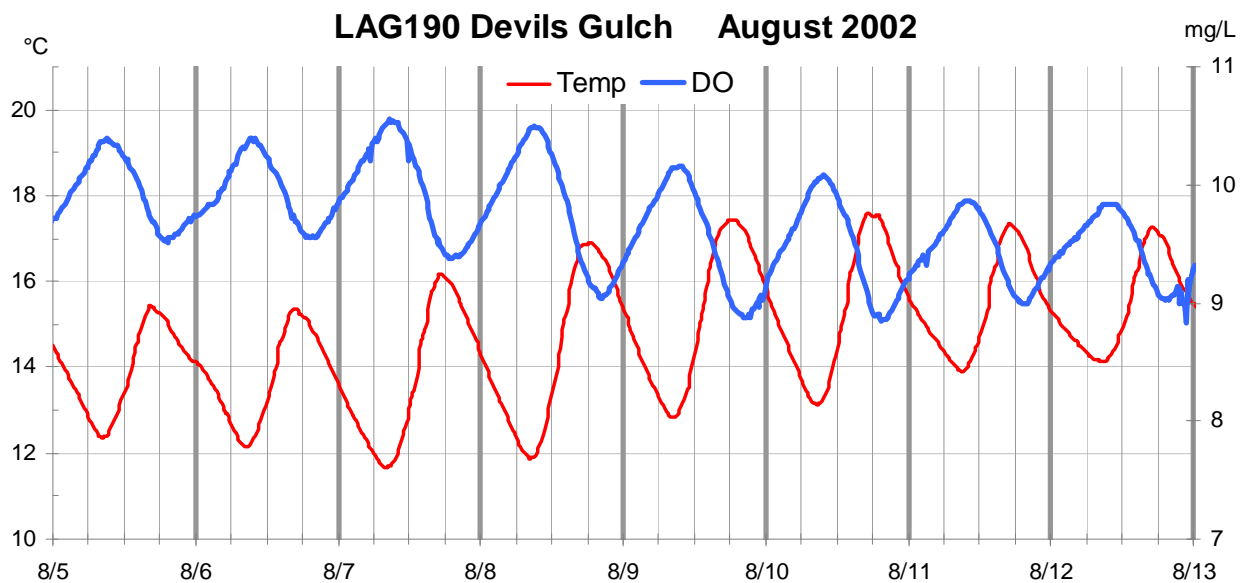


Figure 6-12. LAG190 temperature and DO patterns

6.2.2.7 Unusual temporal patterns

In the San Gregorio Creek watershed, SGR120 (Sky Londa, on La Honda Creek) had an unusual temperature and DO pattern from September 26 to October 4, 2002 (Figure 6-13). While the temperature range was not unusual, and a daily cycle is evident, the timing of the maxima raises concerns: six of nine maxima are within a half hour of midnight and the other three are also at uncharacteristic times (02:30, 09:45, and 20:15). The DO and pH ranges are inconsistently low compared with other sites in the watershed for the same time period, but the specific conductance levels are not (see boxplots in Figures 7-69, 7-70, and 7-71). The graph in Figure 6-13 reveals large and rapid declines and recoveries in the DO curve about the same time as the temperature maxima in the middle of the night, suggesting unnatural causes. The most dramatic event is a loss and recovery of more than 4 mg/L in less than two hours total, a rate unachievable by biological oxygen demand or photosynthesis, especially at night. A mechanical disturbance to the DO membrane would not be likely to recur so regularly.

Comparison of the specific conductance levels with DO in Figure 6-14 reveals distinctly coincident rapid changes. The patterns of the specific conductance measurements support the interpretation of a discharge, which was originally suggested by uncharacteristic timing of temperature maxima and DO minima. Evidence from more than one parameter highlights the value of multi-parameter continuous-monitoring interpretations of water quality. An incident such as this suspected discharge may occur just a few times for a brief period, but have a significant effect on aquatic life. By integrating over time, bioassessment metrics may reflect the impact of such events. Although the relationship is not necessarily causal, lower than expected bioassessment metrics were found at this site.

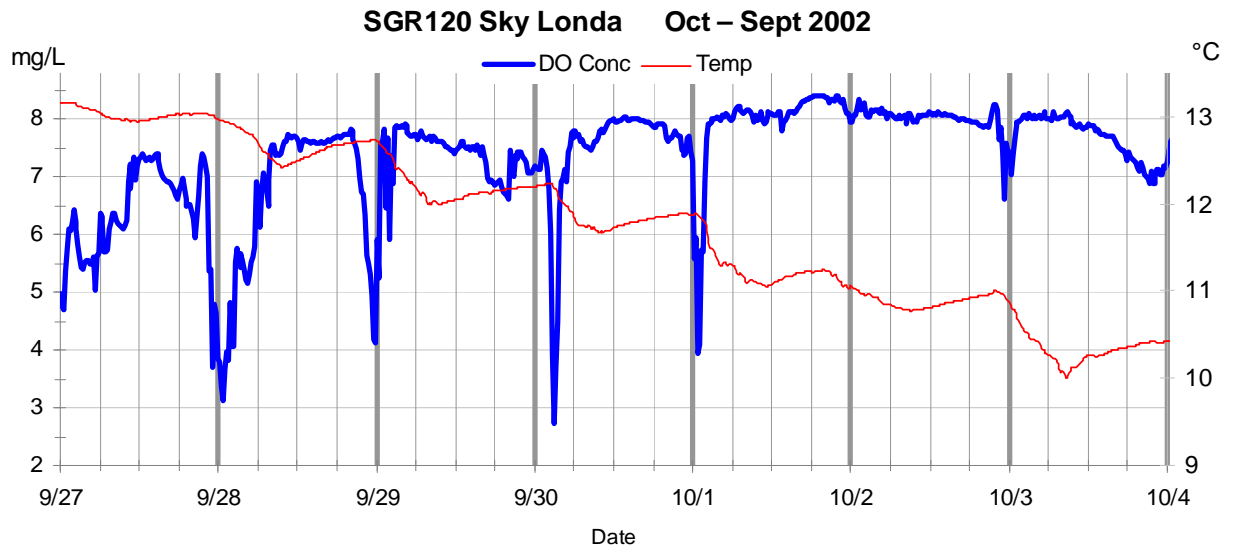


Figure 6-13. SGR120 temperature and DO patterns

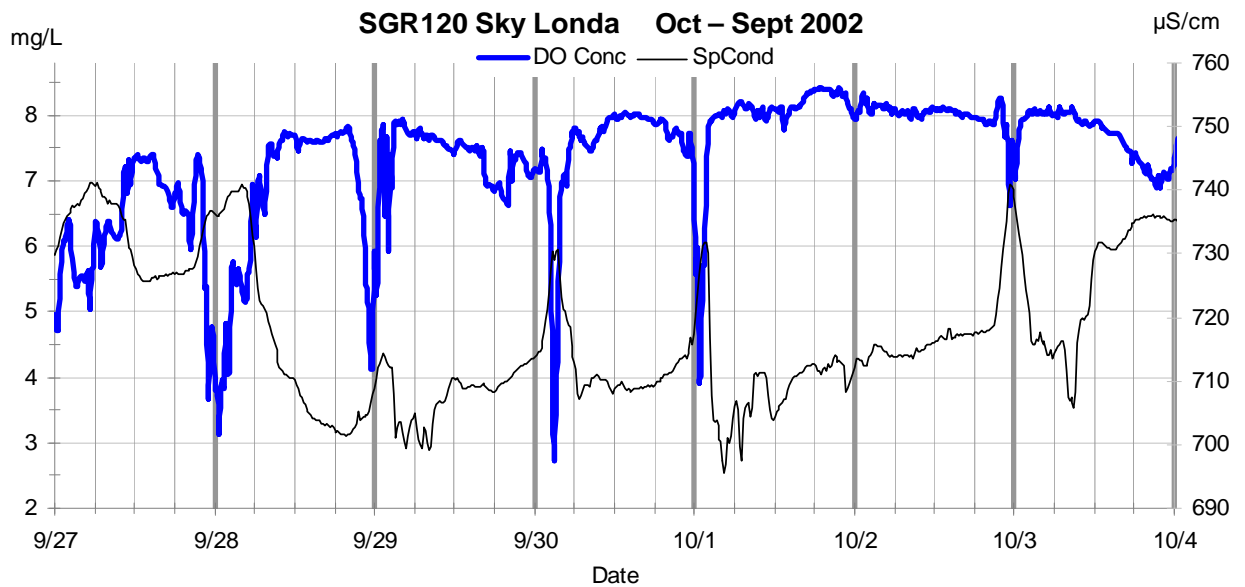


Figure 6-14. SGR120 specific conductance and DO patterns

6.3 Water, Sediment, and Clam Tissue Chemistry

6.3.1 Water Chemistry

Chemicals measured in regional water samples were compared to Basin Plan water quality objectives and/or guidelines established by other agencies to protect aquatic life beneficial uses.

6.3.1.1 Nutrients and Chlorophyll in Water

Nutrient concentrations were compared with U.S. EPA's reference guidelines for Aggregate Ecoregion III Ecoregion 6 (Southern and Central California Chaparral and Oak Woodlands) streams (U.S. EPA 2000). The guidelines are derived from 25th percentile values of stream monitoring data collected from 1990 through 1999, and are intended to represent reference conditions.

Approximately three-fourths of nitrate samples exceeded the U.S. EPA Ecoregion 6 reference guideline of 0.155 mg-N/L. For total phosphorus (as P), the reporting limit of 0.04 mg/L was above the Ecoregion 6 reference guideline of 0.030 mg/L. Ninety six samples exceeded the reporting limit, 41 samples were at the reporting limit of 0.04 mg/L and one sample was below the reporting limit. The minimum reporting limit for orthophosphorus (< 0.010 mg/L for most sample runs) is below the total phosphorus guideline value, and the majority of orthophosphorus values exceeded this guideline (Figure 6-15, Figure 6-16 and Figure 6-17). (U.S. EPA does not provide a reference guideline for orthophosphate. For this reason, measured orthophosphate concentrations will be compared to the total phosphorus guidelines throughout the remainder of this report.)

All samples in this study were below the Basin Plan objective for unionized ammonia. No samples collected upstream of municipal drinking water sources exceeded any standards for nitrite. The ten highest nitrate concentrations (all > 4 mg/L) were measured in the Arroyo Las Positas and San Leandro watersheds. Two sites, Arroyo Seco at Arroyo Las Positas (204ALP110) and San Leandro Creek at Empire Road (204SLE030) combined for the six highest concentrations measured (all > 6 mg/L), and the highest value for each site was measured in the spring (both > 7 mg/L).

The highest orthophosphate concentration (1.99 mg/L) was measured at Lauterwasser Creek in the San Pablo Creek watershed (206SPA200); all other measurements were less than 0.45 mg/L. Seven of the ten highest orthophosphate concentrations were measured in the San Pablo Creek watershed, with the other three from the San Leandro Creek (204SLE070; 0.36 mg/L), Permanente Creek (205PER010; 0.34 mg/L), and Stevens Creek (205STE020; 0.27 mg/L) watersheds (Appendix G; Figure 6-15, Figure 6-16, and Figure 6-17).

In the following graphs, where the analyte was detected but below the reporting limit, the reporting limit is represented.

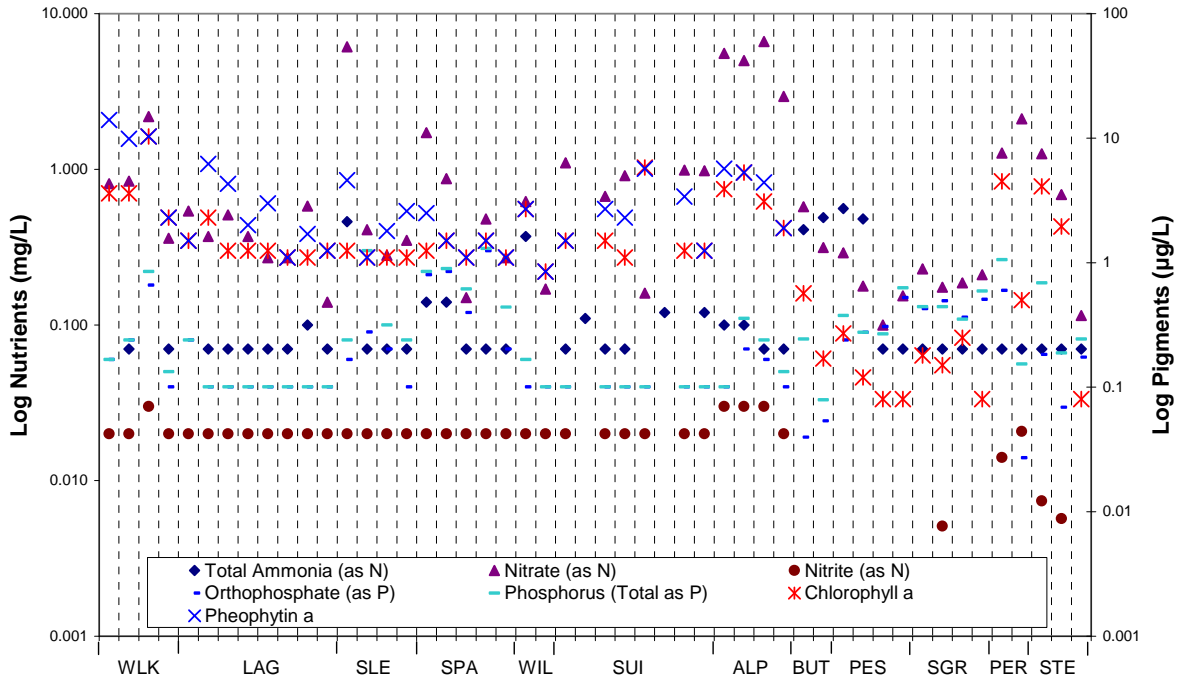


Figure 6-15. Wet season nutrients and chlorophyll

Wet season concentrations of nutrients and chlorophyll measured in water samples. Only watersheds are labeled below for readability. Sites are listed in the same order as in Appendix G. Analytes detected below the reporting limit are represented at the reporting limit.

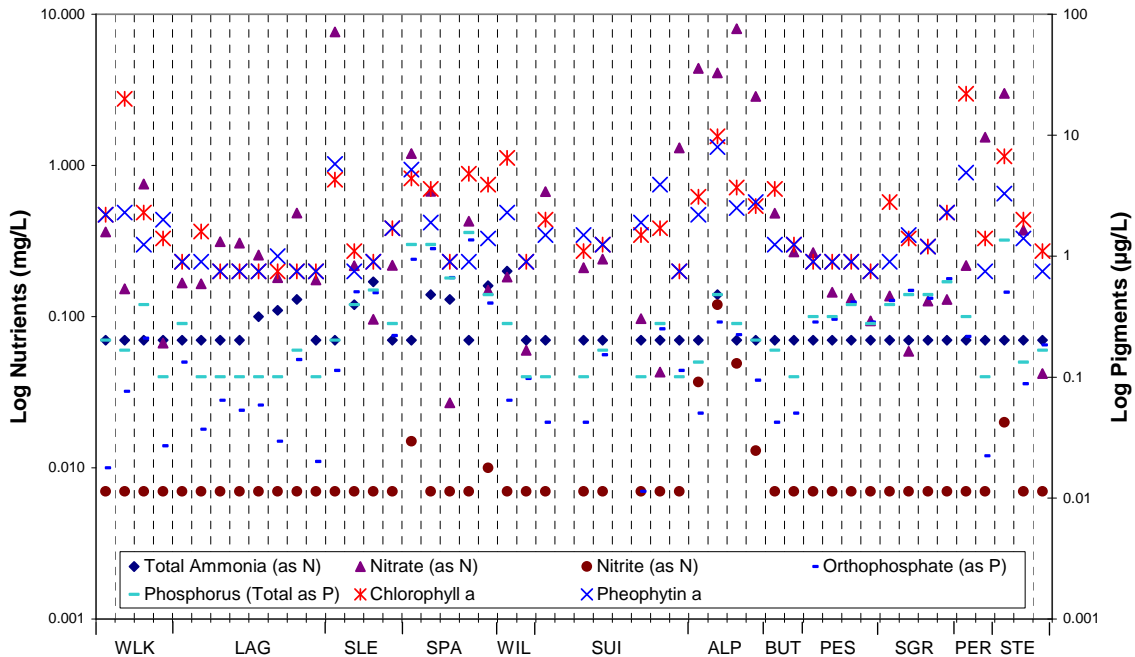


Figure 6-16. Spring season nutrients and chlorophyll

Spring season concentrations of nutrients and chlorophyll measured in water samples. Only watersheds are labeled below for readability. Sites are listed in the same order as in Appendix G. Analytes detected below the reporting limit are represented at the reporting limit.

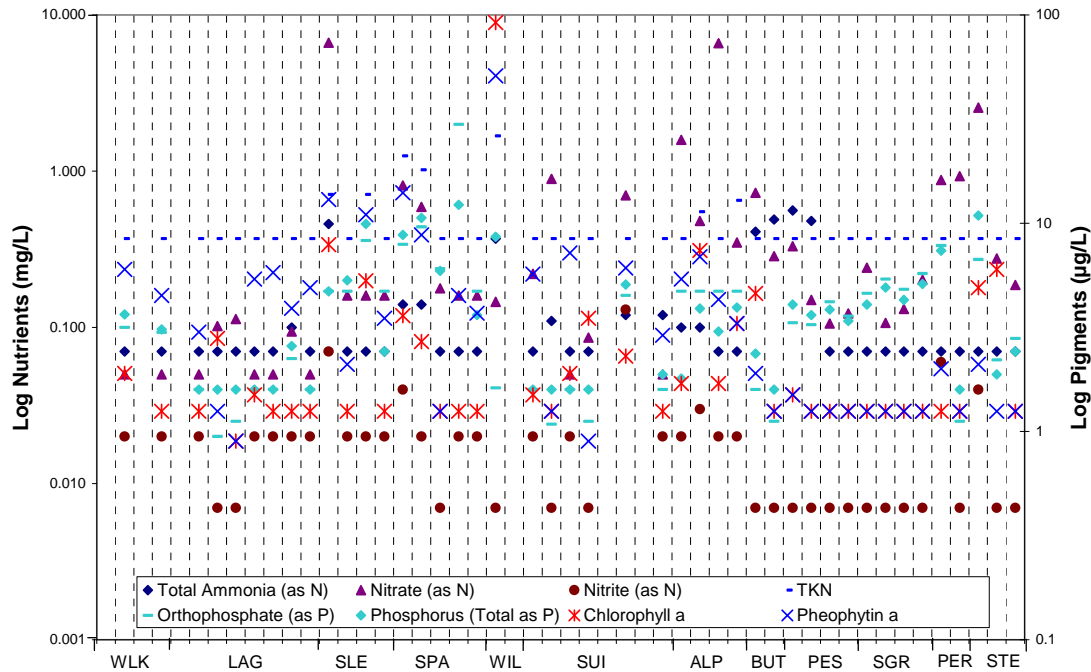


Figure 6-17. Dry season nutrients and chlorophyll

Dry season concentrations of nutrients and chlorophyll measured in water samples. Only watersheds are labeled below for readability. Sites are listed in the same order as in Appendix G. Analytes detected below the reporting limit are represented at the reporting limit. See Appendix G for detection limits.

There were no significant differences in nitrate values among the three seasons (ANOVA $p = 0.32$), with averages of 0.58 mg/L (dry season), 0.88 mg/L (spring season), and 1.04 mg/L (wet season). However, large seasonal differences were observed in some watersheds, as described later in this report.

There was a significant seasonal difference in orthophosphate values (ANOVA $p < 0.05$), with averages of 0.174 mg/L (dry season), 0.079 mg/L (spring season), and 0.078 mg/L (wet season). Total phosphorus results were similar, with dry season values significantly higher than spring and wet season values (ANOVA $p < 0.05$). Average total phosphorus concentrations were 0.153 mg/L (dry season), 0.101 mg/L (spring season), and 0.100 mg/L (wet season).

The ratio of nitrogen to phosphorus in aquatic systems determines which of these elements is most important in controlling eutrophication. Algae utilize nitrogen and phosphorus at a ratio of about 7:1 by mass (Redfield 1958). A ratio of these elements significantly narrower than 7:1 means that there is a greater supply of phosphorus than nitrogen, relative to algal needs, and that nitrogen is limiting growth. A wider ratio than 7:1 implies the opposite: phosphorus limits growth. A ratio close to 7:1 suggests that either or both elements may be limiting. Since the dissolved inorganic forms of nitrogen (nitrate + nitrite + ammonia) and phosphorus (orthophosphate) are the forms most available to algae, it is useful to compare these forms. The median ratio of inorganic

nitrogen to inorganic phosphorus for all sampled streams in the region was 4.5, indicating that the streams are predominantly phosphorus-limited.

Region-wide, there were few significant, or even apparent, relationships between any of the discrete measures of parameters associated with eutrophication: DO, pH, nitrate, orthophosphate, total phosphorus, chlorophyll a, and pheophytin. There was a significant correlation between DO and pH in the wet season ($p < 0.01$). The other two significant correlations were perhaps more interesting: conflicting relationships between concentrations of dissolved oxygen (percent saturation) and total phosphorus. Region-wide, the dry season samples showed decreasing DO with increasing total phosphorus ($p < 0.005$). The reverse was seen for the wet season samples ($p < 0.02$). However, the trend in the wet season was driven by very high DO values (> 120 percent saturation), with few low DO measurements. The dry season relationship, however, was driven more by DO values at or below 50 percent saturation, and may have been reflective of DO depletion through algal respiration as a result of eutrophication.

Several factors may account for the general lack of relationships among these variables. The most important is that primary productivity in shallow streams is usually dominated by benthic algae (periphyton) in contrast to planktonic algae, therefore measurement of water column chlorophyll does not accurately reflect the trophic status or potential impacts on these streams. Periphyton bioassessment methods specific to California are currently being developed. SWAMP will sample periphyton in the future.

An additional factor is that DO data collected in conjunction with water chemistry data were single measurements, sometimes taken in the morning and sometimes in the afternoon, as determined by sampling team logistics. Inconsistent timing of DO sampling can result in the data set being confounded by the strong diel DO swings observed in eutrophic streams. The relationship between DO and algal productivity in streams are also confounded by differences in stream velocity and depth (and therefore reaeration rate). Relationships between ambient nutrient concentrations and algal density can also be obscured by the rapid rate of nutrient uptake during algal growth.

Nitrate values region-wide over all seasons were highest and most variable in streams draining urban areas (average 1.56 mg/L), followed by mixed land use (0.93 mg/L), agriculture (0.68 mg/L), grazing (0.47 mg/L), rural residential (0.26 mg/L), and open space (0.17 mg/L; Figure 6-18). This difference was statistically significant (ANOVA on sqrt transformed data: $p = 0.01$). Differences between individual land use categories were not resolved significantly (Newman-Keuls for unequal sample sizes, $p > 0.20$). The mean nitrate level for urban streams was more than twice that for streams draining agricultural areas, and nearly ten times that of streams in open space areas.

Dry season orthophosphate and total phosphorus data also indicated higher average concentrations and greater variability in streams draining urban sites. There was little apparent difference between concentrations in streams draining areas characterized by other land uses (Figure 6-19 and Figure 6-20). Nitrates and phosphates are routinely and professionally applied as fertilizers in conventional agriculture, but are also applied more

variably for residential and urban landscaping. It appears that urban uses are either more intensive than agricultural uses, or that fertilizers are applied in ways that more directly lead to transport into aquatic habitats.

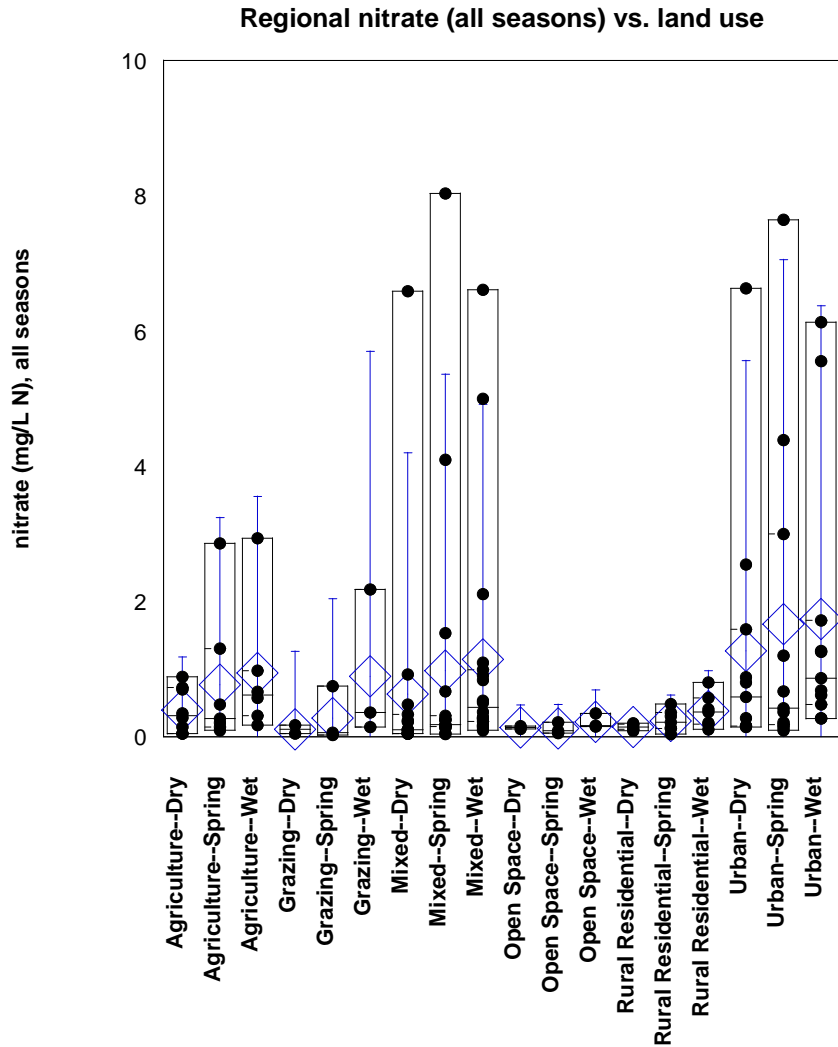


Figure 6-18. Nitrate (NO₃-N)

Nitrate concentrations are grouped by land use category and season. Black dots represent all data points. Black boxes represent percentiles: 95, 75, 50, 25, and 5. Center of blue diamond represents mean, with error bars representing the 95 percent confidence interval about the mean (two-tailed, $\alpha=0.05$).

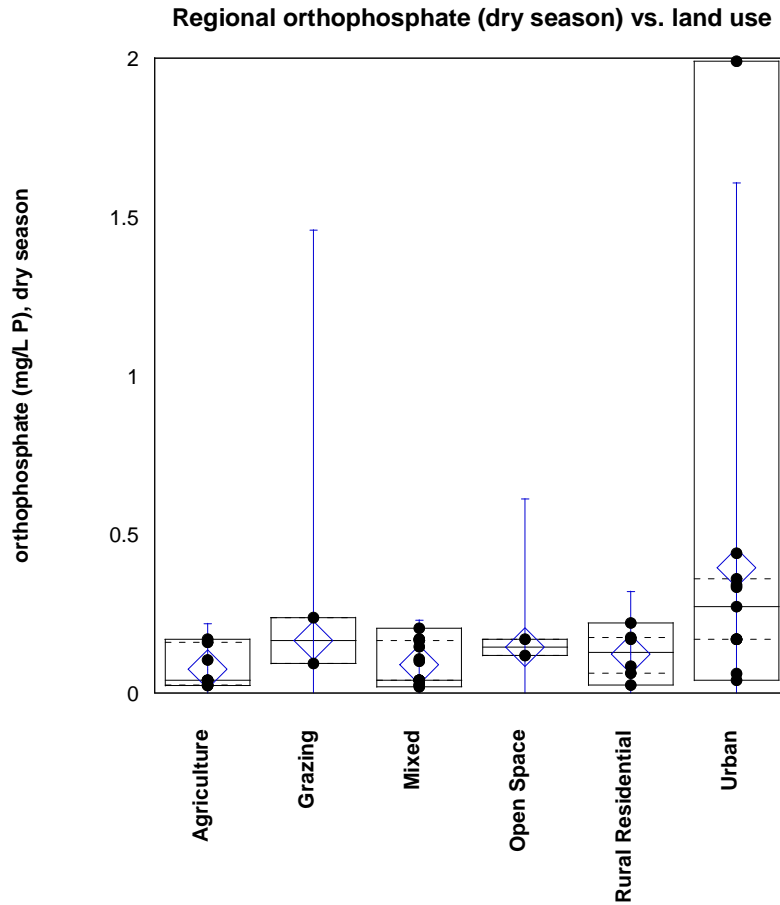


Figure 6-19. Orthophosphate (dry season)

Orthophosphate concentrations are grouped by land-use category during the dry season, when concentrations were highest. Black dots represent all data points. Black boxes represent percentiles: 95, 75, 50, 25, and 5. Center of blue diamond represents mean, with error bars representing the 95 percent confidence interval about the mean (two-tailed, $\alpha=0.05$).

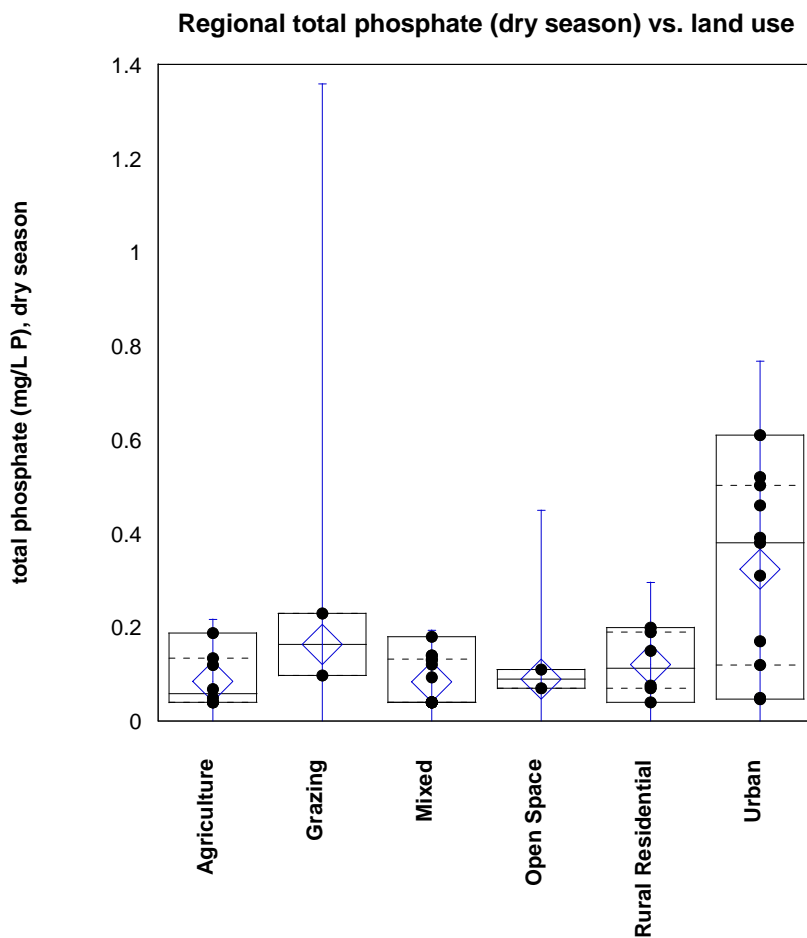


Figure 6-20. Total phosphorus (dry season)

Total phosphorus concentrations are grouped by land-use category during the dry season, when concentrations were highest. Black dots represent all data points. Black boxes represent percentiles: 95, 75, 50, 25, and 5. Center of blue diamond represents mean, with error bars representing the 95 percent confidence interval about the mean (two-tailed, $\alpha=0.05$).

6.3.1.2 Metals in Water

Only two trace metals—dissolved chromium and total selenium—were found at concentrations exceeding guideline values for aquatic life (see Table 4-2 and Appendix G). Two types of Basin Plan Objectives were used as guidelines for assessing metals: the chronic water quality objective (WQO) based on a 4-day average, and the acute WQO, based on a 1-hour average. Results from grab samples are appropriate to compare to an acute WQO; however, they may not be applicable to a chronic WQO. The chronic WQO is used here only as a screening level for further investigation. Dissolved chromium was measured in this study, but the objective is based on hexavalent chromium, which is usually a small fraction of the dissolved chromium (Hem 1985). Thus, applying the chromium VI objective to the measured dissolved chromium is a conservative estimate. Further investigation may be warranted in San Leandro Creek to measure hexavalent chromium and possibly establish a ratio between total and hexavalent chromium.

Over the two years of this study, metals were measured in 59 water samples; though mercury was measured in only 44 of these. The chronic WQO was exceeded in 2 samples (3 percent) for dissolved chromium, both in samples from San Leandro Creek site SLE030. The chronic WQO was exceeded in 5 samples (8 percent) for total selenium, all in samples from Arroyo Las Positas (sites ALP010 and ALP100) and Permanente Creek (site PER070). The acute WQO was exceeded in only one sample, for dissolved chromium at SLE030. There were no water sample exceedances of aquatic life thresholds for arsenic, cadmium, copper, lead, mercury, nickel, silver, or zinc. There are no Basin Plan aquatic life objectives for aluminum or manganese.

The spatial distribution of these measurements is discussed in more detail in the sections below dealing with land use and individual watersheds.

There were some seasonal differences between the spring and dry season samples in the region-wide measurements of trace metals in water. Paired t-tests ($n = 25$) indicated that total and dissolved arsenic ($p < 0.1$), total copper ($p < 0.1$), dissolved manganese ($p < 0.1$), total mercury ($p < 0.1$), and total and dissolved zinc ($p < 0.05$) all had higher concentrations in the dry season compared to the spring season. Total and dissolved selenium ($p < 0.05$) were higher in the spring season. One possible explanation for these general region-wide trends is that metals are conserved in some streams and are concentrated during low flows, while higher flows tend to dilute their ambient concentrations. The exception to this is during discrete rain-related runoff events when the highest concentrations would be expected. However, since this is an ambient monitoring program, discrete events were not sampled. Stream flow data are not available to allow determination of the seasonal trends in metal loading and transport.

6.3.1.3 Organic Chemicals in Water

Detectable concentrations of 43 synthetic organic chemicals were measured in regional water samples. Only 7 compounds of the number listed in Table 4-1 were detected in 5 or more samples: benzo(b)fluoranthene (5 samples, all dry season), benzo(k)fluoranthene (same 5 samples), perylene (5 samples), p',p'DDE (5), diazinon (20), disulfoton (11), and oxadiazon (15).

A limited number of pesticides were detected in a limited number of samples (Appendix G), but only two pesticides were measured above guideline values. The organophosphate pesticide chlorpyrifos was not detected by EPA analytical methods, but enzyme-linked immunosorbent assays (ELISAs) measured chlorpyrifos at concentrations above guidelines for aquatic life protection in water samples from four sites, all urban. Diazinon was detected in 20 samples by EPA methods and 8 samples by ELISA (Figure 6-9), and was found at concentrations above guidelines for aquatic life protection in water samples from three sites, all urban. Both pesticides were measured above guidelines at two of the sites. Diazinon was measured above guidelines by both EPA and ELISA methods in one sample, only by ELISA in a second, and only by EPA methods in the third. These two pesticides have been widely used in agricultural, residential and industrial applications.

Diazinon has previously been identified as causing toxicity to *Ceriodaphnia* in urban watersheds in the Bay Area. Urban creeks in the San Francisco Bay Region were listed on the 303(d) list for diazinon in 2002. As a result, a pesticide TMDL has been developed by the San Francisco Bay Regional Water Quality Control Board. SWAMP data were used in the TMDL to determine that toxicity due to diazinon has decreased in urban creeks since data were collected in the 1990s (SFBRWQCB 2005). The spatial distribution of these measurements is discussed in more detail in the sections below dealing with land use and individual watersheds. Oxadiazon was detected in 15 water samples from throughout the region. This herbicide was also found in clam tissues (see Section 6.3.3, below). Oxadiazon is used primarily for landscape maintenance, with this use accounting for about 50 percent of registered applications in California in 2004 (PAN 2004).

PCBs were detected in only two samples, both collected in the dry season from urban sites. Total PCBs did not exceed the standard (Table 4-2).

Polycyclic aromatic hydrocarbons (PAHs) were detected in a number of samples (Figure 6-10). Individual PAHs that were detected included benz(a)anthracene, benzo(a)pyrene, benzo(b)fluorathene, benzo(k)fluorathene, chrysene, and dibenz(a)anthracene. There are no regulatory aquatic life standards for PAHs, but in general, elevated PAHs were most prevalent at urban sites, and successively less frequent at mixed, rural residential, and agricultural land uses.

There were statistically significant seasonal trends in the concentrations of dissolved and total organic carbon in water samples region-wide (ANOVA, $p < 0.001$). At least 44 DOC/TOC measurements were taken in each of three seasons (dry, spring, wet). Mean DOC values were 6.6 mg/L (dry season), 5.3 mg/L (spring), and 1.6 mg/L (wet), while TOC means were 10.8 mg/L (dry season), 9.7 mg/L (spring), and 2.8 mg/L (wet). This pattern may be due to dilution and flushing of algae and accumulated organic debris during storms.

Seasonal trends in concentrations of organic compounds in water were difficult to quantify because of the relatively low number of detections region-wide. However, some patterns were apparent. Diazinon, the most frequently detected compound, was found in 9 dry season samples, 9 spring samples, and 2 wet season samples. PAHs were detected in 10 dry season samples, 2 spring samples, and no wet season samples. This pattern generally held for the other more commonly detected compounds (DDE, disulfoton, and oxadiazon), which, combined, had only one detection in the wet season (Appendix G). This may again be the result of dilution and/or lower rates of application in the winter for the currently used compounds.

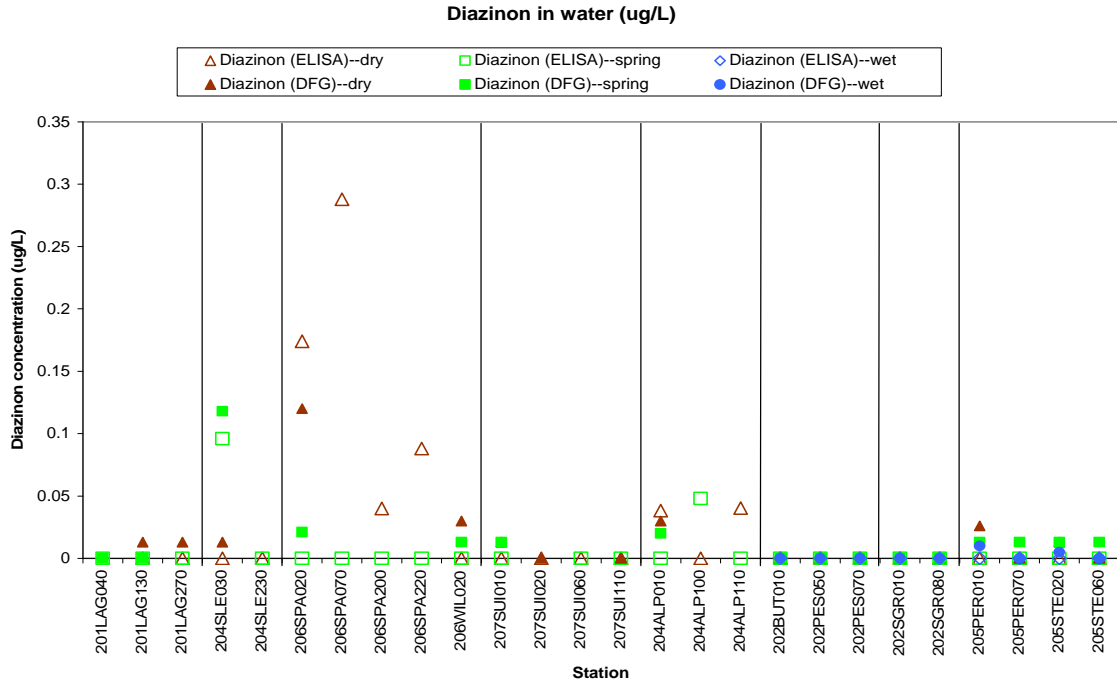


Figure 6-21. Diazinon in water

Concentrations of the organophosphate pesticide diazinon in water samples were measured by two techniques: gas chromatography (DFG Lab) and enzyme-linked immunosorbant assays (ELISA).

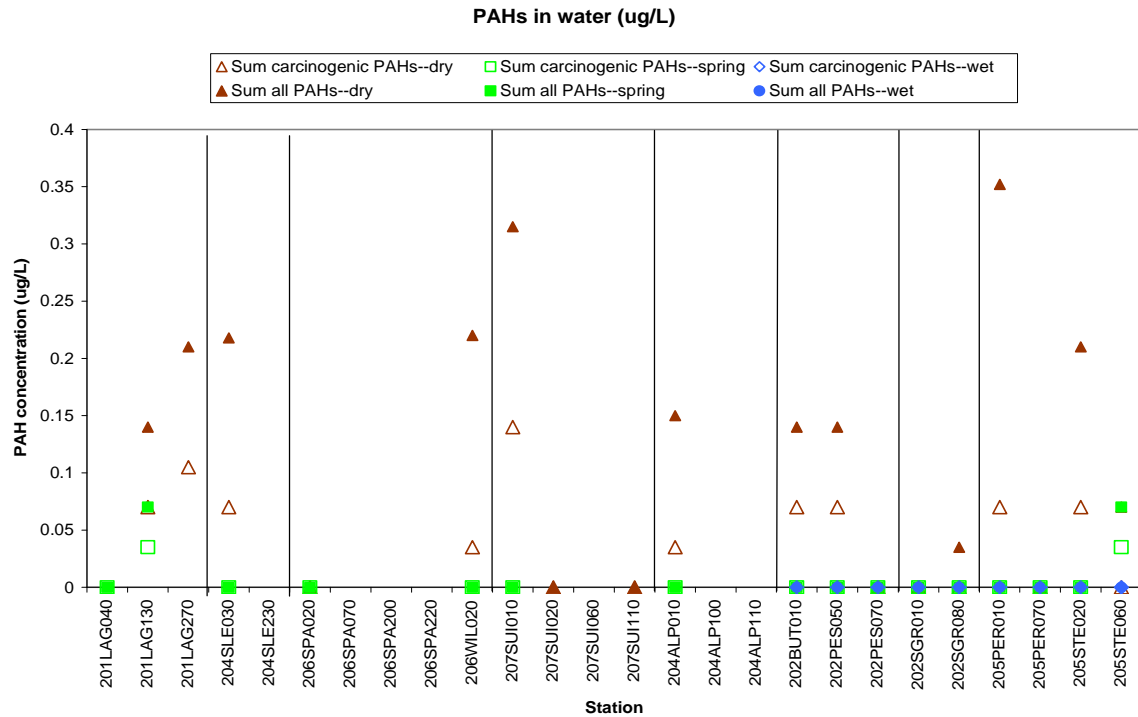


Figure 6-22. PAHs in water

Concentrations of polycyclic aromatic hydrocarbons (PAHs) were measured in water samples.

6.3.2 Sediment Chemistry

Sediments collected from 13 integrator sites at or near the mouths of all surveyed watersheds were analyzed for trace metal and trace organic chemicals (Figures 3-1 to 3-9). Concentrations were generally low relative to available sediment quality guidelines (Table 4-3). The two primary guidelines used for comparison were the threshold effects concentration (TEC), below which harmful biological effects are considered unlikely to be observed, and the probable effects concentration (PEC), above which harmful effects are considered likely to be observed (MacDonald *et al.* 2000). Two geologically abundant trace metals, chromium and nickel, had the largest number of guideline exceedances. Chromium exceeded the TEC in 11 of 13 samples, and the PEC in 5 of 13 samples, with no obvious relationship between concentration and land use. Nickel exceeded the TEC in 10 of 13 samples, and the PEC in 6 of 13 samples, also with no obvious relationship to land use. High concentrations of nickel and chromium are common due to the geology of this area. Lead exceeded the PEC in one sample (from the urban San Leandro site). No other chemicals were found at sediment concentrations exceeding their PEC values.

Of the 13 sites sampled, other chemicals found in exceedance of TEC values included the metals arsenic (3), cadmium (1), copper (4), mercury (5), and zinc (2); sum PCBs (2); and the pesticides DDD (2), DDE (3), DDT (2), and dieldrin (1). All but one of the pesticide exceedances were from two urban sites, near the mouths of San Leandro and Permanente Creeks. The sediment sample from San Leandro Creek was the only one in which amphipod survival was significantly reduced in the sediment toxicity tests (zero percent survival; other toxicity results discussed below by watershed). Permanente Creek was the only watershed with PAHs measured above a guideline value, with the TEC exceeded for benz(a)anthracene, benzo(a)pyrene, chrysene, dibenz(a)anthracene, fluoranthene, phenanthrene, pyrene, and sum PAHs.

Sediment quality guideline quotients (SQGQ) were calculated as a way of comparing the watershed integrator sites in terms of contamination from chemical mixtures (see Section 4.6.3). These SQGQ values can be compared to ranges evaluated by MacDonald *et al.* (2000), who found that 85 percent of sediment samples with mean PEC quotients greater than 0.5 were toxic to sediment-dwelling organisms. Similarly, 92 percent of sediment samples with mean PEC quotients greater than 1.0 were toxic to one or more species of aquatic organisms.

Creamery Gulch, (LAG270), in Lagunitas Creek watershed, had the highest SQGQ (0.58). However, this value was driven primarily by concentrations of trace metals, especially the geologically abundant chromium and nickel. No chemicals other than metals were found there at concentrations above TEC levels. San Leandro Creek had the second highest SQGQ (0.44). This value was also driven by metals, with chromium, nickel, and lead above PEC levels, but also DDD, DDE, and DDT concentrations above TECs. When sum DDE was included in the mean PEC quotient (as recommended by MacDonald *et al.* 2000) the SQGQ increased slightly to 0.47. Samples from all other sites

had SQGQ values less than 0.35 (with DDE included), indicating relatively low potential for sediment toxicity due to chemical mixtures.

Sediment grain size and total organic carbon (TOC) varied widely in samples from the integrator sites throughout the region. The San Leandro Creek sample had a very high TOC content (9.42 percent), and also had a relatively high percentage of fine grained sediment (75 percent). This sample had the second highest SQGQ value and was toxic to amphipods. TOC and fine grained sediments tend to retain contaminants, and numerous studies have documented correlations between these factors and chemical concentrations. The high TOC content of this sediment may also have contributed directly or indirectly to toxicity. The Creamery Gulch sample (LAG270), on the other hand, had low TOC (0.6 percent) and coarse sediment (3 percent fines), but still had high trace metal concentrations. This may indicate active metal loading into the stream, since the sediments there should not have had a high capacity for chemical retention. This loading consisted primarily of geologically abundant metals nickel and chromium, and may be more representative of increased erosion in the system rather than of anthropogenic sources of contaminants.

The San Pablo Creek integrator site sample had relatively high TOC (2.92 percent) and fine grain size (83 percent fines), but had a relatively low SQGQ, and generally low contaminant concentrations. Since this sediment would presumably have good chemical retention capacity, the low concentrations could be indicative of fairly low loading rates prior to the survey. There were no other distinct trends or patterns observed relative to sediment characteristics and contamination levels.

6.3.3 Clam Tissue Chemistry

Clams were deployed at the integrator sites near the mouths of all surveyed watersheds to help evaluate the bioavailability of specific chemicals and their potential to enter food webs. A number of chemicals were detected in clam tissues after one month deployments. However, concentrations of most chemicals were similar to those measured in tissues of control clams. Data are presented for a number of chemicals of concern previously identified in San Francisco Bay area studies.

Only two chemicals, copper and PCB aroclor 1248, were measured above the 85th percentile of the statewide clam tissue data base (Appendix I). Copper concentrations at watershed sites, however, generally fell within a narrow range about the control value (Figure 6-). Mercury tissue concentrations were highest in the Lagunitas watershed, perhaps reflecting geologic abundance, but these concentrations were all below the statewide 85th percentile (Figure 6-). Selenium tissue concentrations were similar to control values at all sites (Figure 6-). While selenium has been of concern in the Bay/Delta for many years, local watersheds are not assumed to be primary sources. These data tend to support that assumption.

Total PCB tissue concentrations were generally similar to control values, though tissues from clams deployed in Stevens and Permanente Creeks were about double the control

value (Figure 6-). Stevens and Permanente Creeks were among three sites that produced sediment samples with total PCBs above TEC guidelines. Total chlordanes have been associated with toxicity in San Francisco Bay sediments (Thompson *et al.* 1999), and total chlordanes were highest in tissue samples from Stevens, Permanente, and San Pablo Creeks (Figure 6-). Chlordanes were not detected in control clams. DDT compounds were among the highest in control clams (Figure 6-) so it is difficult to assign trends to the site data. Dieldrin and oxadiazon were the only other pesticides detected in clam tissues. The distribution of oxadiazon is striking for the elevated value in the Arroyo Las Positas watershed, about 10 times the control value (Figure 6-). Oxadiazon was detected at relatively low levels in sediments and water at that site (Appendices G and H), but was detected in 15 water samples throughout the region, none at concentrations above drinking water standards. This herbicide is traditionally used for turf and ornamental plant applications.

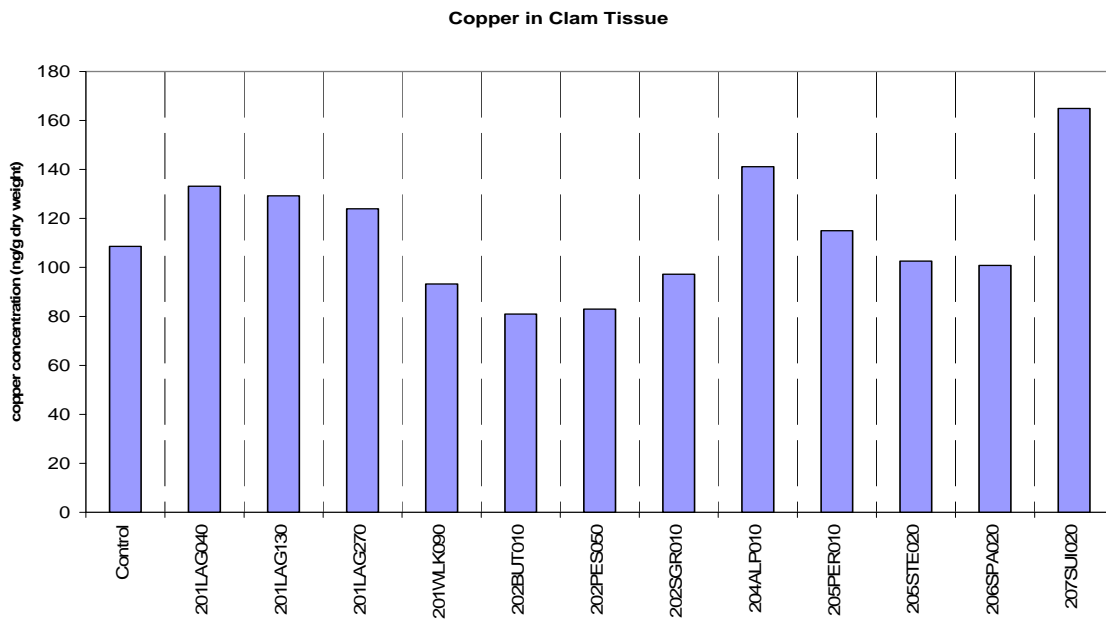


Figure 6-23. Copper in tissue

Concentrations of the trace metal copper were measured in tissues of clams deployed at integrator sites near the base of watersheds.

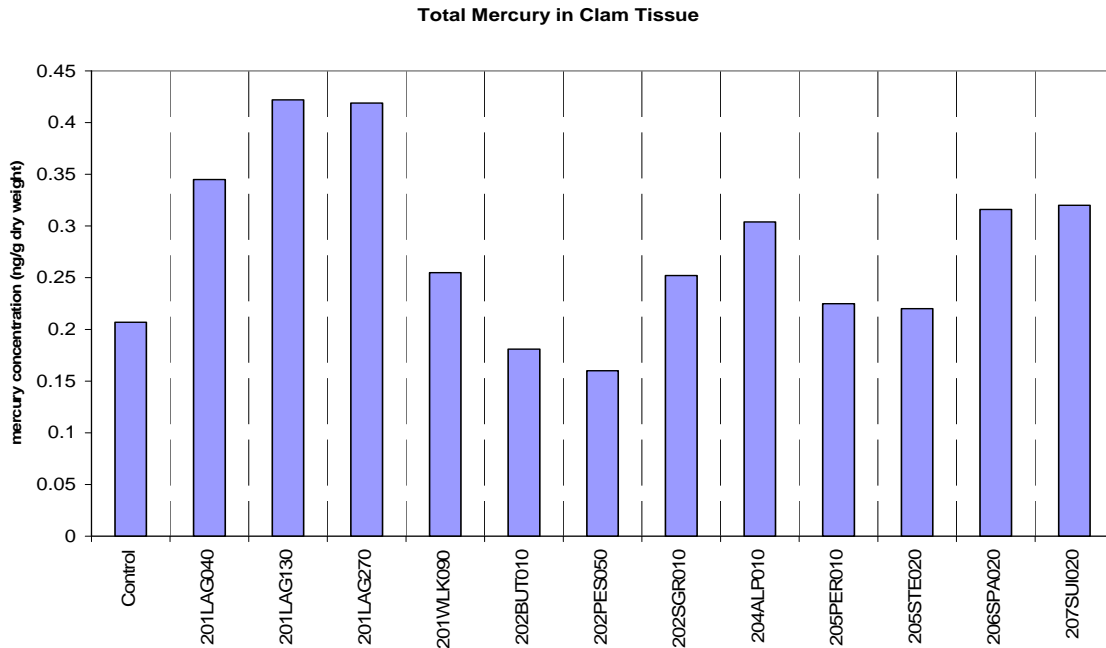


Figure 6-24. Mercury in tissue

Concentrations of the trace metal mercury were measured in tissues of clams deployed at integrator sites near the base of watersheds.

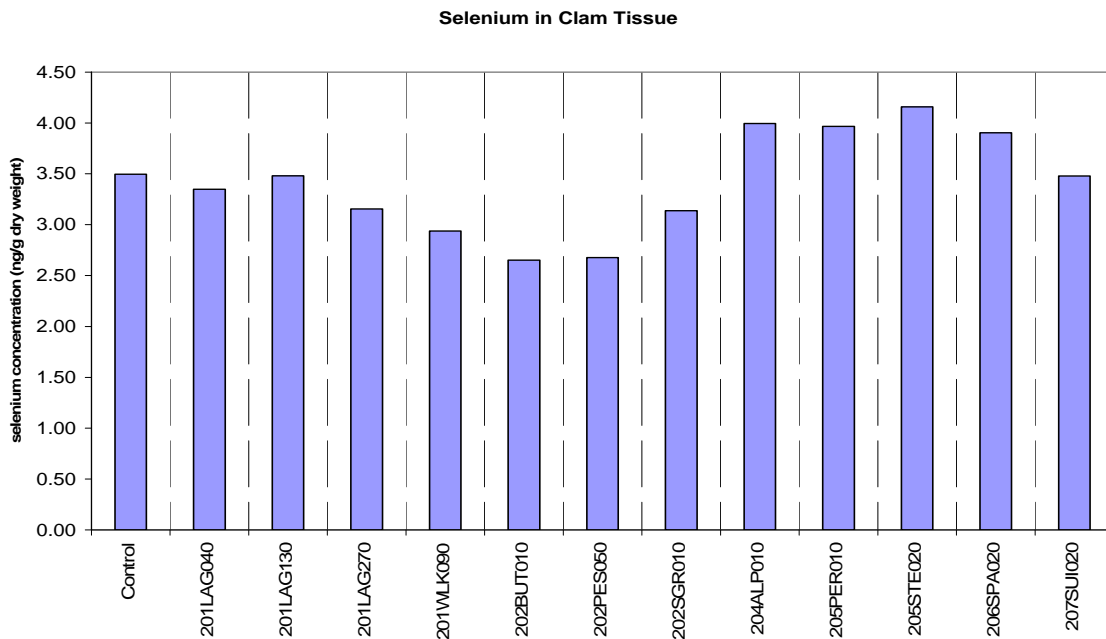


Figure 6-25. Selenium in tissue

Concentrations of the trace element selenium were measured in tissues of clams deployed at integrator sites near the base of watersheds.

PCBs in Clam Tissue; Congeners 201 - 209, plus Sum PCBs

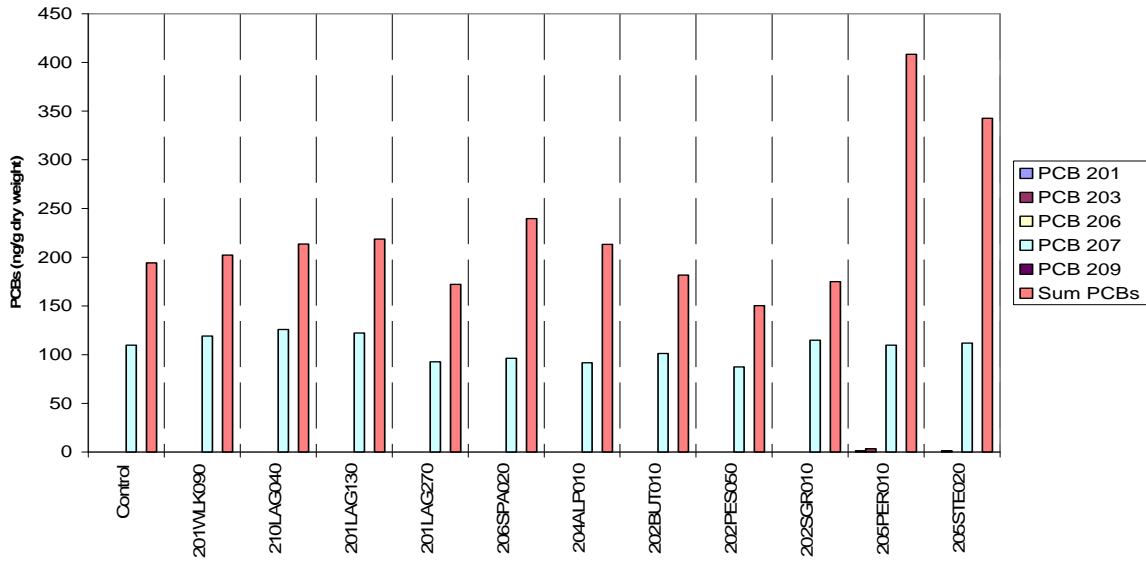


Figure 6-26. PCBs in tissue

Concentrations of the large PCB congeners and total PCBs were measured in tissues of clams deployed at integrator sites near the base of watersheds.

Chlordanes in Clam Tissue

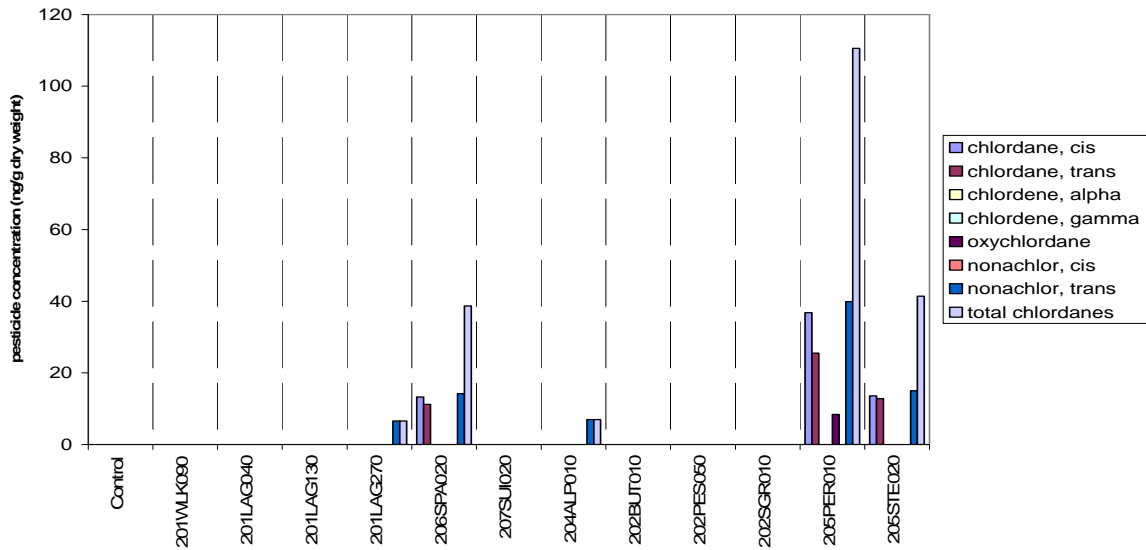


Figure 6-27. Chlordanes in tissue

Concentrations of chlordanes and total chlordane were measured in tissues of clams deployed at integrator sites near the base of watersheds.

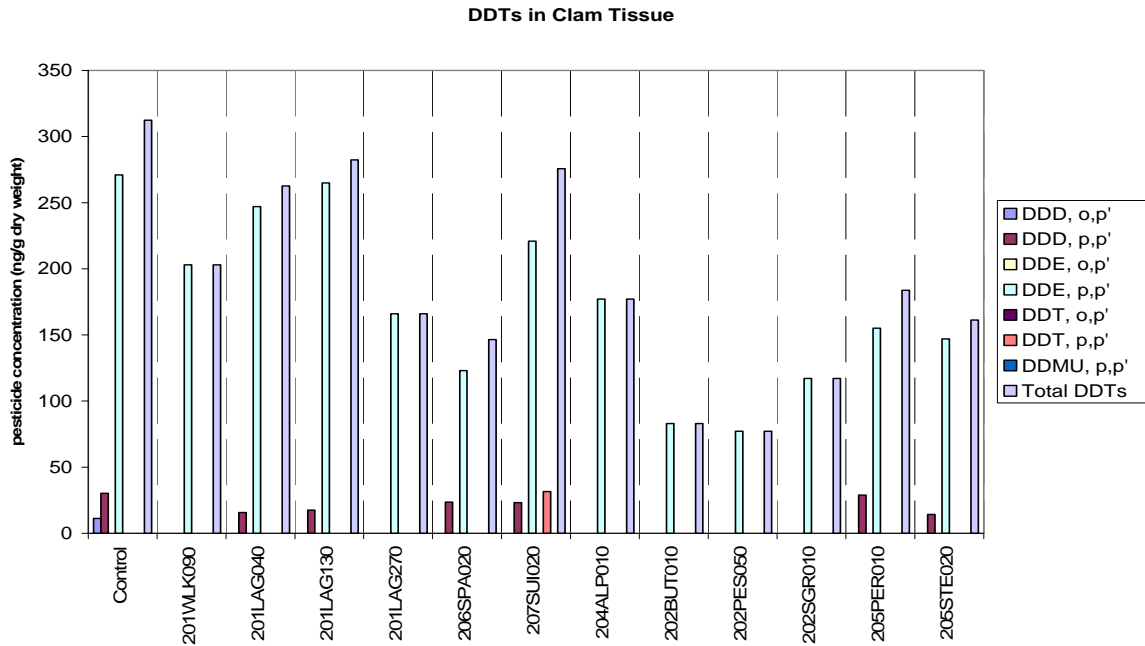


Figure 6-28. DDT in tissue

Concentrations of the pesticide DDT, DDT breakdown products, and total DDTs were measured in tissues of clams deployed at integrator sites near the base of watersheds.

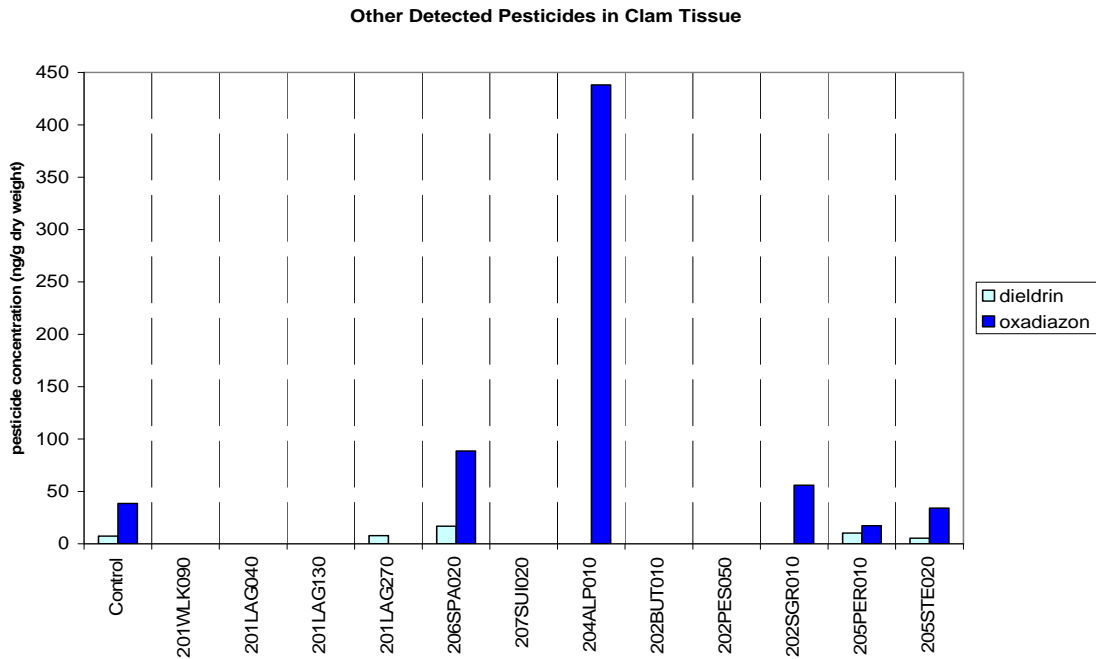


Figure 6-29. Other pesticides in tissue

Concentrations of other pesticides were detected in tissues of clams deployed at integrator sites near the base of watersheds.

6.4 Water and Sediment Toxicity

Three toxicity test species were exposed to grab water samples collected two to three times from 26 sites in the region, totaling 59 samples. Of these, significant reductions in survival, growth, or reproduction were observed in 19 samples from 10 sites. Algal growth was reduced in 15 samples, invertebrate reproduction or survival was adversely affected in three samples, and fish larval survival declined in one sample. All of the observed effects were relatively moderate, with the lowest observed responses (as a percent of the control) being 70 percent for *Ceriodaphnia* survival, 63 percent for *Ceriodaphnia* reproduction, 73 percent for fathead minnow survival, and 29 percent for algal growth.

Because toxicity to *Ceriodaphnia* has been linked to diazinon in many studies over the past 20 years, and because the herbicide oxadiazon was detected in a large number of samples during the present surveys, relationships were examined between these chemicals and water toxicity to *Ceriodaphnia* and *Selenastrum*, respectively. Diazinon was detected, either by ELISA or by GCMS, in 25 samples. However, the highest concentration measured, 0.288 µg/L by ELISA, was well below the *Ceriodaphnia* LC50 value of 0.32 (Bailey *et al.* 1997), and there was no apparent relationship with toxicity. Oxadiazon and toxicity to *Selenastrum* were measured synoptically in 43 samples, in which oxadiazon was detected in 15. Given the observed occurrence of this herbicide in water, sediment, and tissue samples, we investigated possible relationships with toxicity to the alga *Selenastrum*. The highest measured water concentration of oxadiazon in this study was 0.062 µg/L, about two orders of magnitude below published EC50 values for similar algae (4.1 to 126 µg/L; EPA ECOTOX data base, <http://cfpub.epa.gov/ecotox/>). Data from the present study showed no apparent relationship between measured oxadiazon concentration and observed toxicity to *Selenastrum* ($R^2 = 0.13$). While oxadiazon was commonly measured in water and sediment, and had among the highest measured concentrations for pesticides in clam tissues, it does not appear that there was any direct relationship between this herbicide and observed toxicity to the alga *Selenastrum*. See section 6.3.1.3 for regional trends of organic chemicals in water. Details are discussed by watershed in Section 7.

Sediment toxicity tests were conducted on samples collected during the dry season from the integrator sites near the mouth of each watershed. Observed adverse impacts were limited to moderate reductions in amphipod growth, with one exception. There was 100 percent amphipod mortality observed in a sediment sample collected near the mouth of San Leandro Creek, as discussed in section 6.3.2, below. Region-wide, 6 of 13 sites produced samples in which amphipod growth was significantly less than in controls, one site produced the sample with significant mortality, and no sediment toxicity was observed in samples from the other six sites (Appendix J). While toxicity was observed in samples with sediment chemical concentrations above guideline values, there was no significant or apparent linear relationship between the overall level of contamination and the degree of toxicity. Samples were collected at only one site per watershed during one season, so trends related to land use or seasonality cannot be evaluated.

6.5 Coliform Bacteria

The U.S. EPA *E. coli* limit of 126 MPN/100 mL applies to the geometric mean for all sites that could be used for recreation, which applies to all sites sampled; 235 MPN/100 mL applies only to designated beach sites. Four of the thirty sites sampled for coliforms in 2001 or 2002 are designated beaches that are used for water-contact recreation (Appendix K). One of these sites—Swimming Hole at Samuel P. Taylor State Park (LAG185)—had *E. coli* counts in excess of the Basin Plan objectives for water-contact recreation at designated beaches. Two other beaches—Memorial Park Swim 1 (PES134) and Swim 2 (PES135)—had total coliform counts in excess of Basin Plan objectives for water-contact recreation. The fourth beach, Turtle Pond (WLK162), did not exceed any coliform Basin Plan objectives for recreational use (Figure 6- through Figure 6-).

Twenty-six of the sites sampled for coliforms are not designated beaches; however, the beneficial use of REC-1 (body contact recreation) applies to these sites. Of these sites, 12 exceeded *E. coli* (and in some cases total and fecal coliform) Basin Plan objectives for water-contact recreation; 13 others exceeded at least one of the fecal or total coliform basin plan objectives for recreation. The Root Park site (SLE070), which is in a highly urbanized area, had the highest weekly *E. coli* count of any site (>16000 MPN/100 mL). The San Pablo City Park site (SPA060) had the highest *E. coli* geometric mean (872 MPN/100 mL); this site is also highly urbanized, and much trash and evidence of habitation was observed there during sampling. Another highly urbanized site, La Avenida (STE020), had the second highest *E. coli* geometric mean (796 MPN/100 mL) and the highest total coliform median value. Camp Cooley (STE090, misnamed for Cooley Picnic Area at Camp Costanoan) had extremely high total coliform counts on one day (30,000 MPN/100 mL). Although Camp Cooley has signs prohibiting swimming, swimmers were observed there at each sampling visit, and more than once, its *E. coli* counts exceeded the limit for designated beaches. One site, Cataract (LAG390), did not exceed any coliform Basin Plan objectives for recreational use. The site was included as a reference site, since no water contact is allowed and it is above a municipal drinking source, however it is accessible to hikers and wildlife, despite signs prohibiting swimming.

It is important to note that coliform detection limits varied with sampling date and site. Some sites had a maximum detection limit of 1600 MPN/100 mL, which limited the ability to accurately calculate mean and median values and interpret the maximum daily total coliform objective (10,000 MPN/100 mL).

With the exception of sites dominated by urban land use (see Appendix A), there were no obvious relationships between land use categories and coliform bacteria in recreational waters at the sites sampled. The Stevens/Permanente creeks watershed had consistently very high levels of total coliforms, even though land uses varied greatly between sites. Sites near grazing and horse-riding areas would be expected to have higher coliform counts; however, coliform counts at these sites were at the low end of the range across all sites. Of the seven major watersheds in which samples were collected for coliform bacteria measurement, only Walker Creek did not exceed the *E. coli* Basin Plan objective

(Figure 6-), however only one isolated swimming spot was sampled. Walker Creek watershed has been sampled more intensively for a pathogen TMDL (due to contamination of shellfish beds) and the results of that study are more indicative of general coliform bacteria conditions in the watershed. No water samples were collected in the Suisun or Arroyo Las Positas watersheds for coliform bacteria.

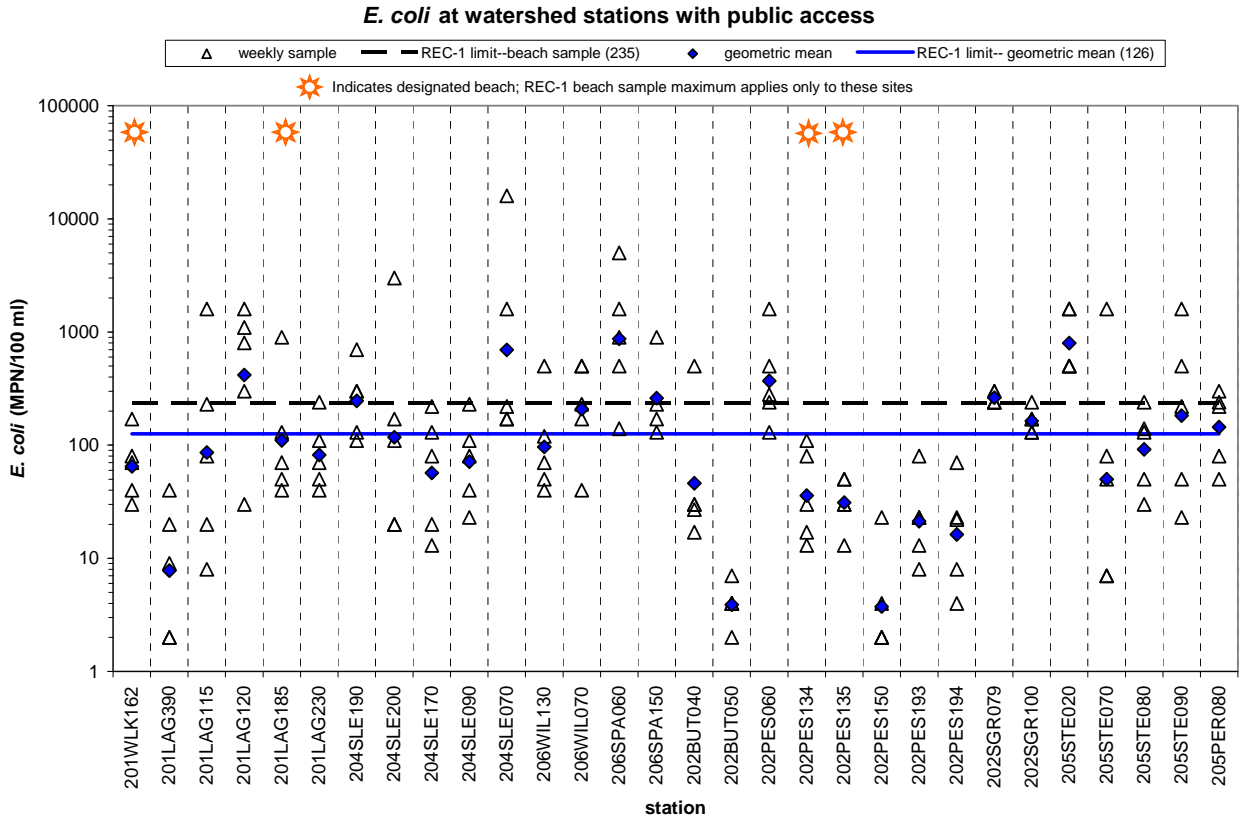


Figure 6-30. E. coli counts

Sites with blue diamonds (the geometric mean) above the blue line at 126 MPN/ 100 mL exceed the EPA steady state limit for E. coli. Sites with sun icons and triangles above the dashed line at 235 MPN/100 mL exceed the limit for designated beaches

Fecal coliforms at watershed stations with public access

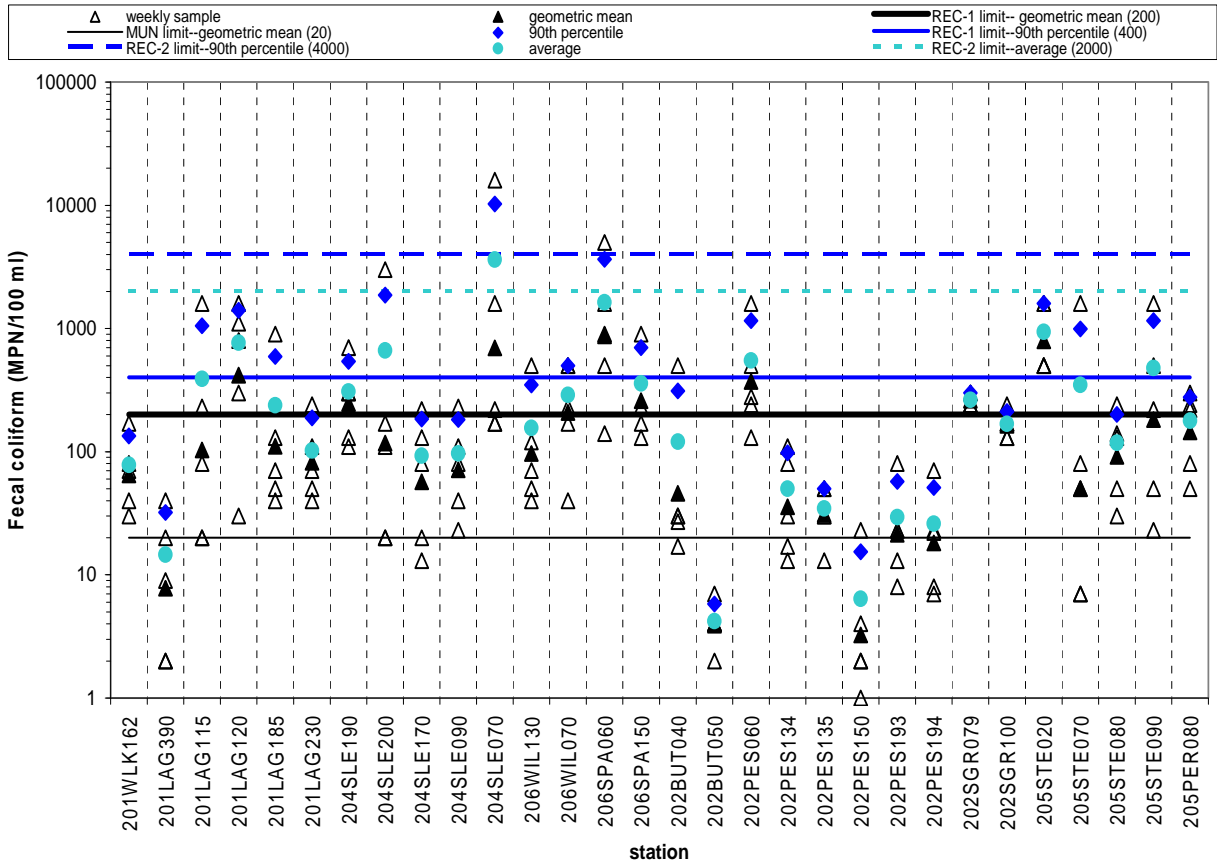


Figure 6-31. Fecal coliform bacteria counts

Counts of the most probable number (MPN) of fecal coliform bacteria measured in 100 milliliters of water samples collected from watershed sites.

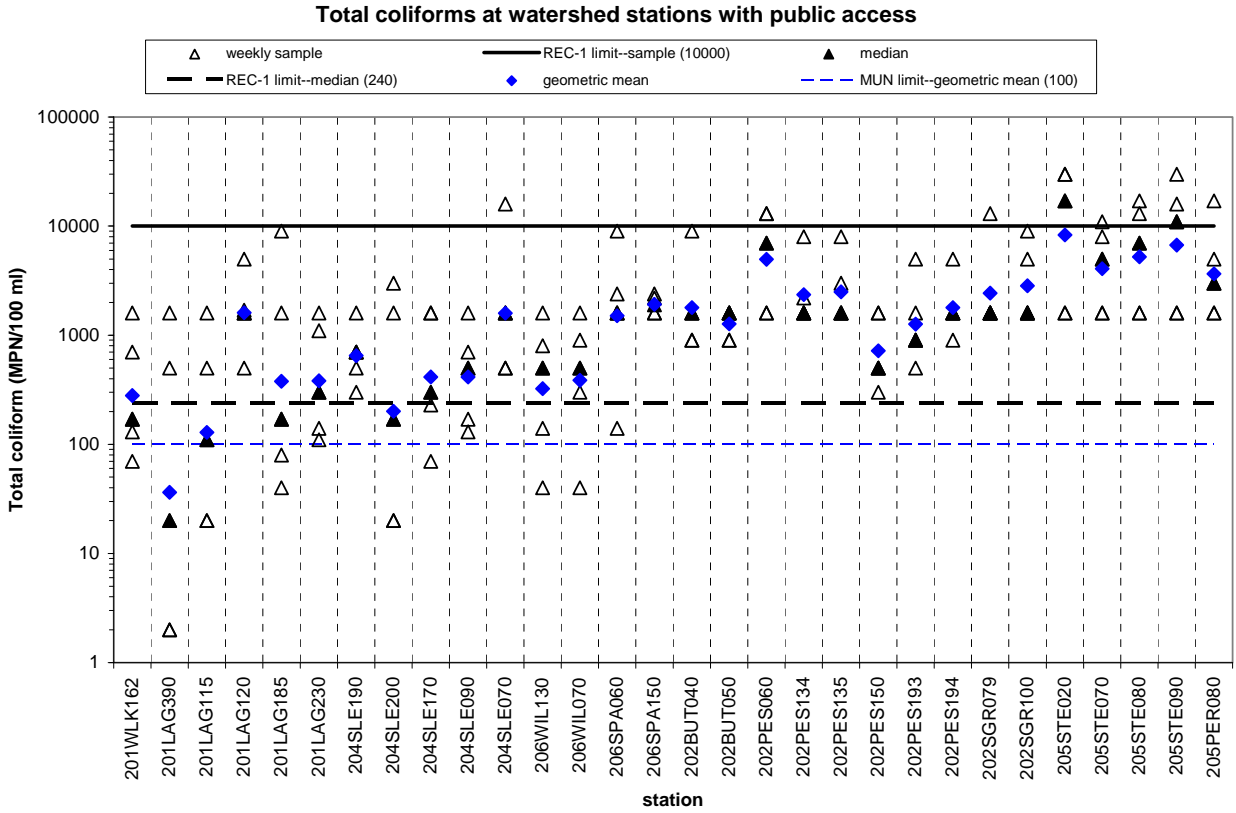


Figure 6-32. Total coliform bacteria counts

Counts of the most probable number (MPN) of total coliform bacteria measured in 100 mL water samples collected from watershed sites.

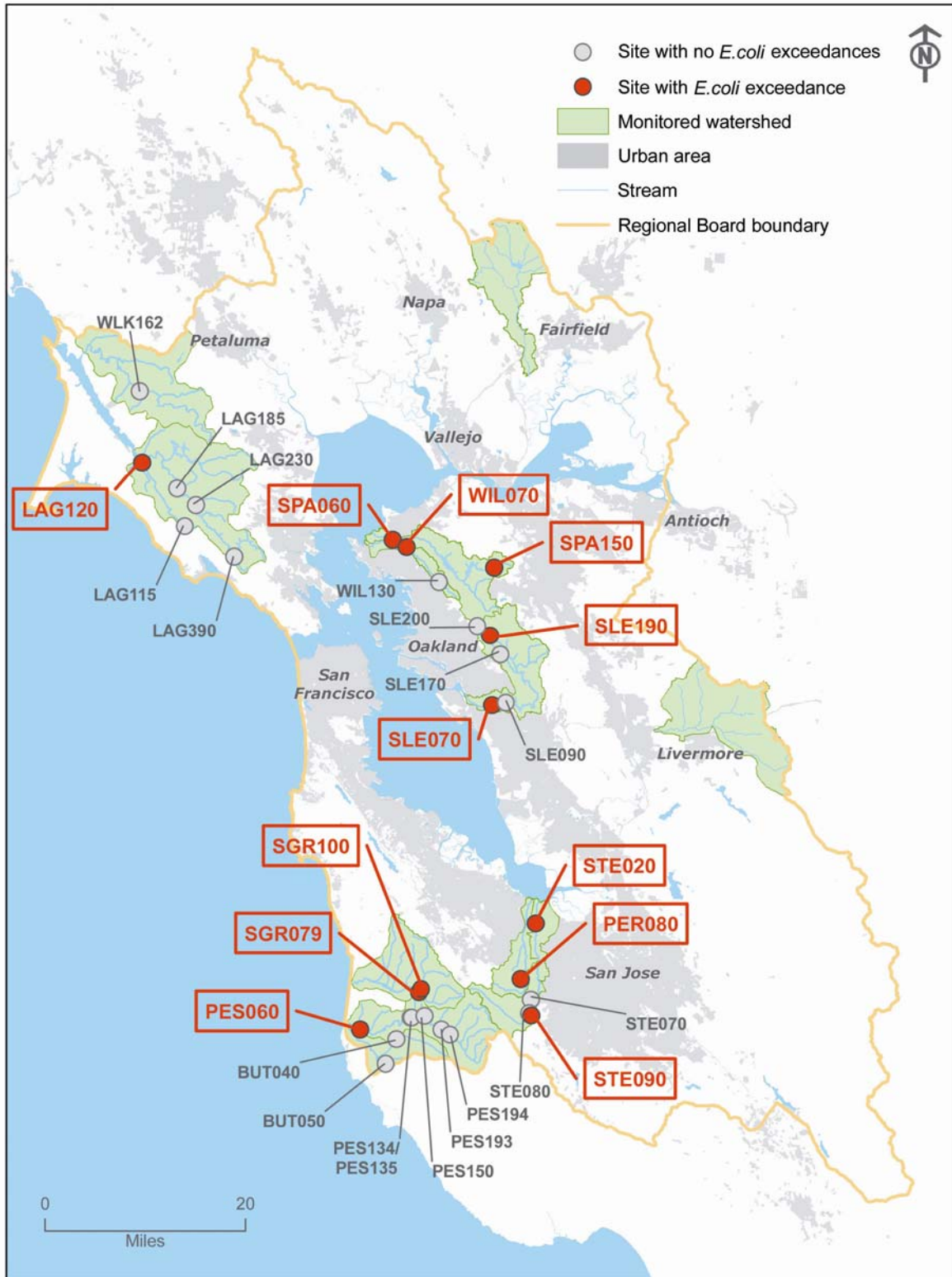


Figure 6-33. Map of *E. coli* bacteria exceedances

6.6 Regional Reference Conditions for Bay Area Streams

Expected background concentrations for certain pollutants (*i.e.* pesticides and other synthetic organics) in pristine, non-degraded systems are obvious: any detectable level of these substances is indicative of anthropogenic pollution. For most water quality indicators, however, it is not as simple to predict conditions in minimally disturbed systems. Biological systems are highly variable in both space and time. Likewise, the concentrations of organic and inorganic dissolved constituents in stream water depend greatly upon natural factors such as geology, hydrology, and vegetation. Defining reference conditions or the range of conditions expected in natural or minimally disturbed systems requires an understanding of the natural variability of a system. Careful descriptions of reference conditions are important in order to properly identify degraded conditions. Further detailed studies must be carried out in order to properly describe reference conditions. This section contains a brief description of approaches for defining reference conditions.

6.6.1 Benthic Macroinvertebrate Assemblages

6.6.1.1 Natural Variability

Detecting the effects of human disturbance on benthic assemblages requires setting appropriate expectations for biological conditions across a range of stream types. A necessary first step is to develop an understanding of the sources and ranges of natural variability in benthic assemblages. Here we consider (1) annual variability, generally associated with annual fluctuations in climate, and (2) geographic variability, due to spatial patterns in climate, topography, and geology.

6.6.1.1.1 Annual Variability

Although bioassessment samples in the Bay Area are all collected during the same April-May reference period, annual variation in climate can cause differences in benthic invertebrate assemblages from year to year. Annual rainfall is highly variable in Mediterranean climates, and changes in seasonal patterns (e.g. high rainfall totals in April and May) can greatly affect the timing of bed-scouring floods, summer baseflow, and water temperatures. These hydrologic factors, in turn, affect the survival, development, and timing of emergence of benthic assemblages. In order to compare data from sites sampled in different years we need to understand the year-to-year variability of benthic assemblages expected at a site. To date, biomonitoring programs in the Bay Area generally sample sites only once (SWAMP) or over two consecutive years (most county stormwater programs).

Four consecutive years of monitoring have been conducted in Sonoma Creek (by Sonoma Ecology Center) and Wildcat Creek and San Leandro Creek (Breux *et al.* 2005). These sites represent a range of conditions from minimally disturbed to heavily urbanized sites.

In general, annual variability of metrics at minimally disturbed sites draining open space is much lower than variability at urban sites (Table 6-5). Six metrics exhibited low (C.V. < 0.25) inter-annual variability in minimally disturbed streams, while only 3 metrics had low variability between years at urban sites. The taxa richness, Shannon Diversity, and tolerance value metrics generally exhibited the least annual variability, while composition metrics tended to be more variable. Percent predators had low annual variability only at minimally disturbed sites, while variability in percent collectors was low at urban sites because metric values exhibit a small range (90 to 99 percent) relative to the mean. These measures of annual variability should be taken as conservative estimates, as only four years of data, all from moderate rainfall years (2000-2003), are represented.

Taxa presence/absence data from the four-year study of Wildcat Creek and San Leandro Creek were examined in a multivariate ordination analysis (Figure 6-). Consecutive years' samples are generally grouped by site, but there is some overlap between sites. For most sites the year 2000 samples are more dissimilar than the years 2001-2003. Given that year 2000 samples for all sites have systematically lower values on Axis 3 than subsequent years' samples, regional climatic differences in 2000 may be responsible. Of the four years, rainfall in the 1999-2000 wet season was the highest (30 inches), including several late spring storms. Rainfall in 2000-2001, 2001-2002, and 2002-2003 was 20, 24, and 26 inches respectively. Rainfall just prior to sampling might also have resulted in taxonomic differences between years. In the two days before sampling began in 2000, 1.5 inches of rain fell in Oakland. In 2001 and 2002 no precipitation was recorded in the two weeks prior to sampling, except for 0.1 inches on the first day of sampling in 2002. In 2003, 0.9 inches of rain fell 5 days prior to sampling, and 0.6 inches of rain fell the day before sampling (all rainfall data from the Oakland North weather station).

These analyses suggest that year-to-year variability in taxa presence and some metric values can be considerable. Given that the available data come from years that were not extremely wet or dry, it is expected that the long-term annual variability in benthic assemblages is even more significant than is represented here. The data used in this report, in analyses of geographic variability and correlations with land use, come from multiple years (2001-2003). Thus, annual variability is reflected in these analyses, meaning that measured differences between site groups (*e.g.* urban versus open-space land use) are detectable, despite annual variability. A better understanding of the effects of climatic variability on benthic assemblages, however, will allow for more precise conclusions to be drawn about the effects of geographic variability and water quality. Weissich (2005) observed that benthic invertebrate assemblages exhibited high temporal variability throughout the year, even within the April-May reference period. One of the recommendations of this report is that several long-term bioassessment monitoring stations should be established in the Bay Area, in order to gain a better understanding of temporal variability and the role of climate.

Table 6-5. Annual variability for select metrics

| | Open Space (n = 4) | Other (n = 7) | Urban (n = 5) |
|------------------------|------------------------------|-------------------------|-------------------------|
| Taxa Richness | 0.08 | 0.23 | 0.23 |
| EPT Taxa | 0.1 | 0.25 | 0.77 |
| % EPT | 0.22 | 0.45 | 1.03 |
| % Sensitive EPT | 0.29 | 0.89 | 1.79 |
| Shannon Diversity | 0.06 | 0.17 | 0.26 |
| Tolerance Value | 0.13 | 0.14 | 0.11 |
| % Intolerant Organisms | 0.35 | 0.86 | 1.4 |
| % Tolerant Organisms | 0.92 | 0.73 | 0.5 |
| % Dominant Taxon | 0.3 | 0.35 | 0.35 |
| Percent Collectors | 0.46 | 0.47 | 0.16 |
| Percent Filterers | 0.46 | 0.65 | 0.87 |
| Percent Scrapers | 0.29 | 0.72 | 0.94 |
| Percent Predators | 0.12 | 0.56 | 1 |
| Percent Shredders | 0.47 | 1.2 | 1.64 |

The annual variability for select metrics is expressed as the average coefficient of variation (C.V.) for sites over four years. The number of sites in each site class is indicated by “n”. C.V.s less than 0.25 are in bold. “Other” sites include a mix of rural, agricultural, and mixed land use sites, and represent an intermediate level of disturbance. A subset of metrics is shown because not all metrics were calculated in the original datasets.

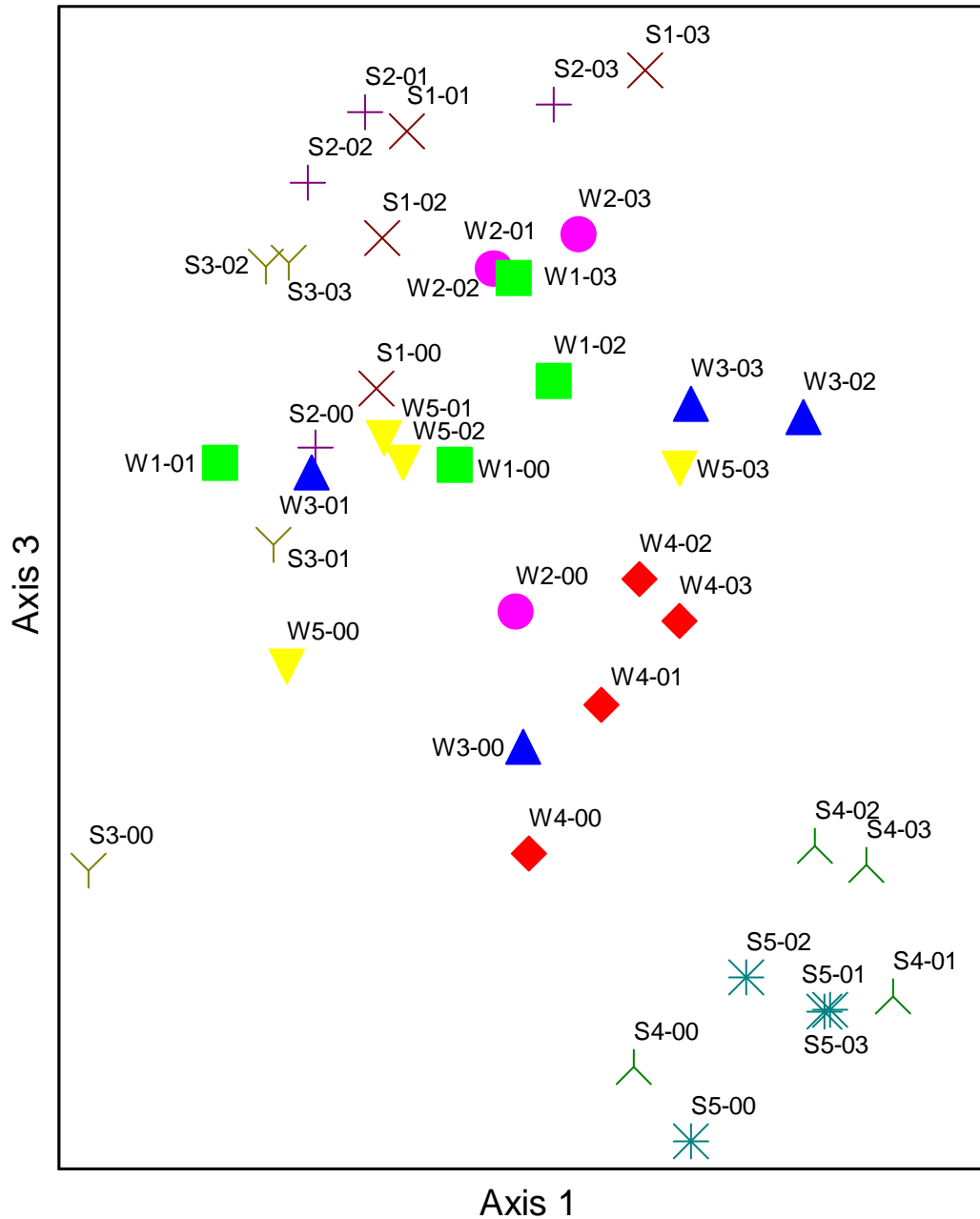


Figure 6-34. Taxa presence in Wildcat and San Leandro Creeks

NMS ordination of taxa presence at 5 sites in Wildcat Creek (W) and 5 sites in San Leandro Creek (S) with four consecutive years of data. Sites are numbered from downstream to upstream. The year of sampling is indicated by the last two digits of the year sampled (e.g., 2001 is -01). Each site is shown with a distinct symbol and color. Wildcat Creek sites are shown with filled symbols; San Leandro Creek sites are shown with unfilled symbols. Axis 3 represents 52 percent of the variation and Axis 1 represents 26 percent of the variation in the original data.

6.6.1.1.2 Geographic Variability

The pervasive, intense human land use in low elevation areas in the Bay Area presents a major difficulty in defining reference conditions. Minimally disturbed conditions are generally only present in the Bay Area in protected parklands in mountainous areas at high elevations. A longitudinal study of Santa Clara Valley streams by Fend et al. (2005) found a high degree of correlation between natural environmental variables (channel slope, elevation, temperature) and anthropogenic impacts (e.g., urban land use, reservoirs, reach habitat quality). Many studies have described natural variation in benthic invertebrate assemblages along longitudinal gradients (Alan 1995). Thus, teasing apart the covarying effects of natural factors and anthropogenic impacts is extremely difficult. Much more work is needed to define appropriate reference conditions for lowland streams in the Bay Area.

To examine the range of natural, undisturbed biological conditions across the Bay Area, biological metrics and community composition were examined from sites with minimal human influence. Minimal human influence was established using three criteria: (1) the absence of any urban, residential, grazing, or agricultural land use in the upstream watershed, (2) a P-HAB Channel Alteration score greater or equal than 17 (out of 20), and (3) P-HAB Riparian Zone Width scores for right and left bank greater or equal than 9 (out of 10). These two P-HAB scores were selected as criteria because they best represent local reach-scale habitat disturbance in the channel and riparian zone. Numerical criteria for these P-HAB scores were arbitrarily set to eliminate sites that scored lower than the upper range of the "optimal" condition category for each variable. The land use criteria limited the minimally disturbed sites to 13 sites classified as open space in the land use type classification (see above). The third criteria eliminated 1 of the sites, bringing the total number of minimally disturbed sites to 12.

Also included in this analysis were sites monitored by the Regional Board in 2003, as well as monitoring performed by other agencies within the Bay Area using the same field and laboratory protocols. The same criteria were used to identify sites with minimal human influence from these datasets. Four sites from the Regional Board's 2003 monitoring and 7 sites monitored by other agencies were identified, bringing the total number of minimally disturbed sites to 23 (Table 6-6). To ensure that the datasets were comparable, raw taxonomic data was combined together to resolve differences in taxonomic resolution prior to the re-calculation of metrics.

Table 6-6. Minimally disturbed sites used in the analysis of geographic variability

| Site ID | Stream | Watershed | Monitoring Agency |
|---------|-----------------------------|------------------------|--|
| BUT-050 | Little Butano Creek | Butano Creek | SWAMP |
| LAG-330 | Big Carson Creek | Lagunitas Creek | SWAMP |
| LAG-335 | Big Carson Creek | Lagunitas Creek | SWAMP |
| LAG-380 | Little Carson Creek | Lagunitas Creek | SWAMP |
| LAG-390 | Cataract Creek | Lagunitas Creek | SWAMP |
| PES-180 | Peters Creek | Pescadero Creek | SWAMP |
| PES-190 | Pescadero Creek | Pescadero Creek | SWAMP |
| PES-200 | Slate Creek | Pescadero Creek | SWAMP |
| PES-210 | Oil Creek | Pescadero Creek | SWAMP |
| PES-240 | Pescadero Creek | Pescadero Creek | SWAMP |
| SLE-180 | East Fork Redwood Creek | San Leandro Creek | SWAMP |
| SLE-230 | Kaiser Creek | San Leandro Creek | SWAMP |
| SPC-MF | Middle Fork San Pedro Creek | San Pedro Creek | San Mateo County Pollution Prevention Program |
| SPC-SF | South Fork San Pedro Creek | San Pedro Creek | San Mateo County Pollution Prevention Program |
| UP-6 | Upper Penitencia Creek | Upper Penitencia Creek | Santa Clara Valley Urban Runoff Program |
| UMR 2 | Upper Marsh Creek | Marsh Creek | Contra Costa County Clean Water Program (CCCCWP) |
| UMR 3 | Upper Marsh Creek | Marsh Creek | CCCCWP |
| SMA-180 | San Mateo Creek | San Mateo Creek | SWAMP |
| SMA-160 | San Mateo Creek | San Mateo Creek | SWAMP |
| MTD-120 | Mitchell Canyon Creek | Mt. Diablo Creek | SWAMP |
| MTD-140 | Donner Canyon Creek | Mt. Diablo Creek | SWAMP |
| NAP-Red | Redwood Creek | Napa River | Friends of the Napa River |
| NAP-Se | Segassia Creek | Napa River | Friends of the Napa River |

To determine if variation between benthic assemblages in minimally disturbed streams was ascribable to natural environmental factors, correlations between taxa presence and environmental variables were examined. At each minimally disturbed site, four environmental variables that play important roles in structuring stream communities (Allan 1995) were quantified at each minimally disturbed site: two climate/hydrology variables (flow intermittency, watershed annual precipitation) and two channel network variables (channel slope, drainage area). Flow intermittency was assessed during at least one late-summer (September) site visit, and was quantified as binary values (1 = intermittent, 0 = perennial), except one site with questionable intermittency that was coded as 0.5. Annual precipitation in the upstream drainage basin of each site was estimated from a precipitation contour map of the Bay Area by Rantz (1971). Channel slope and drainage area were estimated from digitized topographic maps (TOPO! v.2.0, Wildflower Productions, 1999). Note that flow intermittency and watershed annual precipitation generally reflect other geographic or landscape gradients: coastal watersheds have the highest precipitation levels and are almost always perennial, whereas

watersheds east of the Oakland-Berkeley Hills have the lowest precipitation levels and are mostly intermittent. Thus, these two factors are well-correlated with each other and reflect vegetation gradients and ecoregion boundaries. There is very little published information available on spatial patterns of flow intermittency in the Bay Area, however.

Intermittency and annual precipitation were highly correlated with the first axis of an NMS ordination of taxa presence at minimally disturbed sites, while stream slope and drainage area exhibited minimal correlation (Figure 6-, Table 6-7). Intermittency and annual precipitation are highly correlated with each other, however. When only perennial sites are examined (17 of 23 minimally disturbed sites), there is very low correlation between annual precipitation and the first ordination axis.

Many biological metrics at intermittent streams are significantly different (paired comparisons using student's *t*, $\alpha = 0.05$) from conditions at perennial streams (Table 6-7). In particular, many aquatic beetles (Order: Coleoptera) were absent from intermittent sites, probably due to their novel life history in which both the adult and larval stages of many taxa are aquatic. Likewise, caddisfly (Order: Trichoptera) richness was also significantly reduced at intermittent sites. In general, richness of most insect orders was lower at intermittent sites, reflected in the taxa richness and EPT richness metrics.

Watershed annual precipitation exhibited no relationship to any biological metrics. Watershed annual precipitation ranged from 400 to 1200 mm/yr (16 to 46 in/yr) at minimally disturbed sites, with a median value of 880 mm/yr (34 in/yr). Annual precipitation in the Bay Area ranges from 300 mm/yr (12 in/yr) east of Livermore to above 1300 mm/yr (50 in/yr) in the headwaters of the Napa River in the Mayacamas Mountains. Thus, it is probable that the range of watershed precipitation values in the set of minimally disturbed sites reflects the range of conditions in the Bay Area.

Channel slope varied from 1 to 15 percent among minimally disturbed sites, with most sites in the 2 to 6 percent range. Most streams in low elevation zones in the Bay Area, such as along the bayshore, have slopes less than 2%. Thus, the minimally disturbed streams discussed here may not reflect reference conditions for low gradient streams. Drainage area of the minimally disturbed streams ranged from 0.1 sq. km (0.1 sq. miles) to 40 sq. km (15 sq. miles), with most sites less than 10 sq. km. Test sites in the Bay Area tend to have larger drainage areas, on the order of 5-100 sq. km.

The biological expectations described in this analysis apply mainly to small, high-gradient streams in hilly or mountainous parts of the Bay Area. Minimally disturbed sites were exclusively gravel and cobble-bedded streams, and generally were located in mountainous terrain. In contrast, many streams in the Bay Area along the bay plain or in broad valleys have slopes less than one percent and are naturally sand-bedded streams. Whereas gravel-bedded streams share many common physical and ecological processes across their range of conditions, sand-bedded streams often have very different biological communities (Allan 1995). Additionally, streams with larger watersheds are expected to have different invertebrate communities than streams with small watersheds. Much more

research is needed in order to define appropriate reference conditions for larger, lower gradient streams in the Bay Area.

Table 6-7. Minimally disturbed intermittent and perennial sites' metric scores

| | Mean Intermittent n = 6 | Mean Perennial n = 17 | T-test Probability > t |
|----------------------------------|---------------------------------------|-------------------------------------|----------------------------------|
| Richness Metrics | | | |
| Taxa Richness | 32 | 46 | <0.0001 |
| EPT Taxa | 16 | 23 | 0.002 |
| Ephemeroptera Taxa | 7 | 9 | 0.0748 |
| Plecoptera Taxa | 5 | 6 | 0.6093 |
| Trichoptera Taxa | 4 | 9 | <0.0001 |
| Coleoptera Taxa | 2 | 7 | <0.0001 |
| Diptera Taxa | 8 | 10 | 0.1091 |
| Non-Insect Taxa | 5 | 6 | 0.2518 |
| Composition Metrics | | | |
| % Sensitive EPT | 33 | 38 | 0.2364 |
| % Non-insect | 3 | 5 | 0.315 |
| % Chironomidae | 22 | 13 | 0.0228 |
| % Coleoptera | 1 | 12 | 0.0071 |
| % Oligochaeta | 2 | 1 | 0.0435 |
| Tolerance Metrics | | | |
| Tolerance Value | 4 | 3.6 | 0.0263 |
| % Intolerant | 30 | 33 | 0.3342 |
| Functional Feeding Groups | | | |
| % Collector-gatherer | 68 | 40 | <0.0001 |
| % Collector-filterer | 2 | 5 | 0.0511 |
| % Scraper | 7 | 25 | 0.003 |
| % Shredder | 5 | 13 | 0.0228 |
| % Predator | 17 | 17 | 0.7533 |

Mean metric scores at minimally disturbed intermittent and perennial sites, and results of Student's t-test. T-test values <0.05 are considered significant and are bolded.

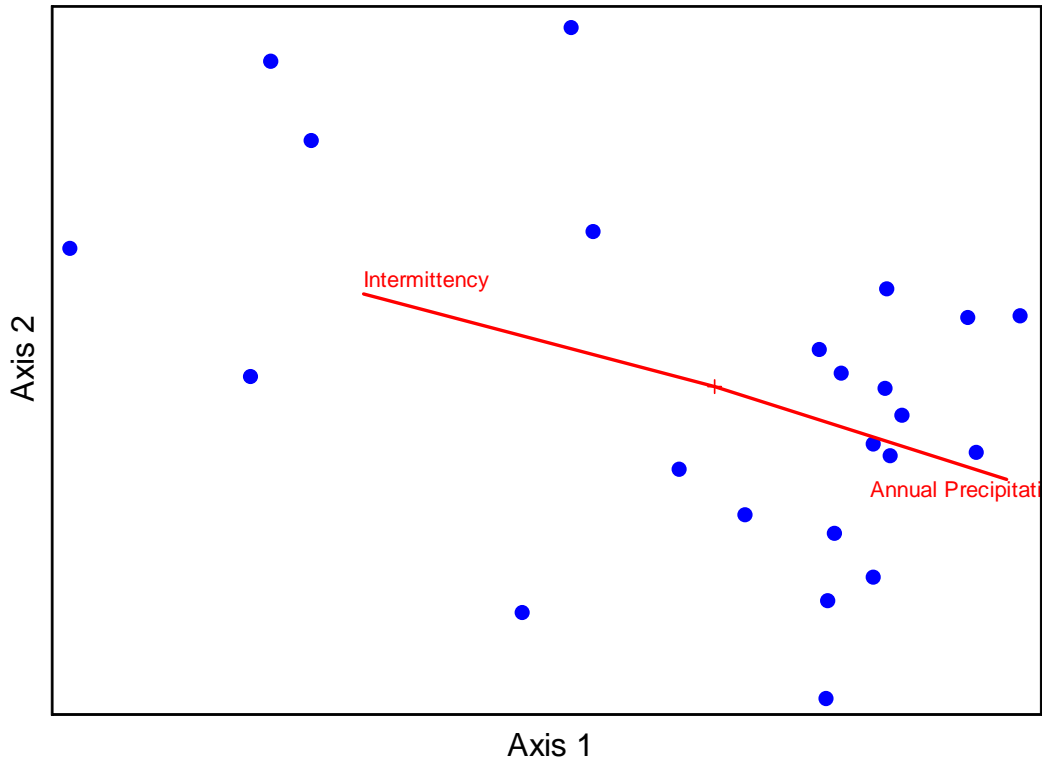


Figure 6-35. Correlation of environmental variables with taxa presence

Intermittency and annual precipitation are highly correlated with the first axis of this NMS ordination of BMI taxa presence at reference sites. Sites clustered on the right side of the graph have perennial flow, while sites on the left are intermittent.

6.6.1.2 Bioassessment Analytical Tools: Biological Metrics and Indicator Taxa

6.6.1.2.1 Biological Metrics

One major goal of bioassessment is to identify streams that have been negatively impacted by human activities relative to undisturbed conditions. An ideal analytical tool to accomplish this goal should be very sensitive to human disturbance, yet have low natural variability. Variability in metrics was examined within minimally disturbed perennial and intermittent sites.

Seven metrics from perennial sites and 8 metrics from intermittent sites had C.V.s less than 0.25 (Table 6-8). Metrics with low variability common to both site perennial and intermittent site classes were Diptera taxa, taxa richness, EPT taxa, percent EPT, and tolerance value.

Table 6-8. Coefficients of Variation from perennial and intermittent stream classes

| | Perennial C.V. | Intermittent C.V. |
|----------------------|-----------------------|--------------------------|
| Coleoptera Taxa | 0.23 | 0.49 |
| Diptera Taxa | 0.23 | 0.14 |
| Ephemeroptera Taxa | 0.24 | 0.35 |
| Plecoptera Taxa | 0.32 | 0.28 |
| Trichoptera Taxa | 0.27 | 0.22 |
| Non-Insect Taxa | 0.38 | 0.4 |
| Taxa Richness | 0.13 | 0.09 |
| EPT Taxa | 0.19 | 0.17 |
| % EPT | 0.19 | 0.16 |
| % Sensitive EPT | 0.27 | 0.23 |
| % Chironomidae | 0.55 | 0.45 |
| % Coleoptera | 0.75 | 1.82 |
| % Oligochaeta | 1.13 | 1.08 |
| % Non-insect | 0.8 | 0.41 |
| % Intolerant | 0.27 | 0.17 |
| % Tolerant | 1.27 | 0.53 |
| Tolerance Value | 0.1 | 0.08 |
| % Predator | 0.38 | 0.36 |
| % Collector-filterer | 0.71 | 1.05 |
| %Collector-gatherer | 0.29 | 0.14 |
| % Scraper | 0.5 | 1.05 |
| % Shredder | 0.57 | 0.53 |
| % Other | 1.47 | 0.94 |

Coefficients of Variation for metrics that were less than 0.25 are in bold. Metrics are not included when the mean for the metric was 0.

To examine the ability of 26 common metrics to discriminate between natural and altered conditions, metric scores at minimally disturbed sites were compared to values from two groups of sites representing moderate and high levels of alteration. Assemblage Groups 1 and 2, identified in the multivariate analysis of assemblage groups (see Assemblage Groups, above), are interpreted to represent high and moderate levels of human disturbance. Eleven out of 12 sites identified as minimally disturbed were classified into Assemblage Group 3, while Assemblage Groups 1 and 2 are dominated by streams draining urban and agricultural land uses, respectively. Disturbed sites in assemblage groups 1 and 2 were classified as perennial or intermittent streams and compared to the range of values from minimally disturbed sites with the same flow status.

The discrimination efficiency of a metric was defined as the percentage of disturbed sites with metric scores outside of the range of values (minimum to maximum) of minimally disturbed sites. In cases where metric values at disturbed sites were both greater and less than the maximum and minimum values at minimally disturbed sites, only the higher discrimination efficiency is reported. Discrimination efficiencies of 26 metrics ranged from 0 to 100 percent (Table 6-9). In general, metrics had higher discrimination efficiencies between minimally disturbed sites and Assemblage Group 1 (urban sites)

than Assemblage Group 2 (agricultural sites). Metrics with discrimination efficiencies both greater than 50 percent for Assemblage Group 2 and greater than 80 percent for Assemblage Group 1 are shown in bold in Table 6-9.

Table 6-9. Discrimination efficiencies of 26 metrics of inferred biological impairment

| | Response to Disturbance | Intermittent | | Perennial | |
|----------------------------------|-------------------------------|--------------|-----------|------------|------------|
| | | A.G. 2 | A.G. 1 | A.G. 2 | A.G. 1 |
| Richness Metrics | | | | | |
| Taxa Richness | - | 85 | 93 | 100 | 100 |
| EPT Taxa | - | 31 | 93 | 91 | 100 |
| Coleoptera Taxa | - | 31 | 53 | 91 | 100 |
| Diptera Taxa | - | 62 | 93 | 82 | 100 |
| Ephemeroptera Taxa | - | 38 | 100 | 73 | 100 |
| Plecoptera Taxa | - | 23 | 100 | 45 | 100 |
| Trichoptera Taxa | - | 31 | 80 | 18 | 95 |
| Non-Insect Taxa | + | 23 | 40 | 0 | 0 |
| Composition Metrics | | | | | |
| % EPT | - | 54 | 87 | 73 | 80 |
| % Sensitive EPT | - | 54 | 93 | 73 | 100 |
| % Chironomidae | + | 38 | 73 | 81 | 85 |
| % Coleoptera | - | 31 | 53 | 64 | 95 |
| % Oligochaeta | - | 23 | 53 | 36 | 80 |
| % Non-insect | + | 31 | 80 | 9 | 60 |
| Tolerance Metrics | | | | | |
| % Intolerant | - | 54 | 93 | 82 | 100 |
| % Tolerant | + | 54 | 67 | 9 | 0 |
| Tolerance Value | + | 54 | 93 | 90 | 100 |
| Functional Feeding Groups | | | | | |
| % Predator | - | 46 | 80 | 82 | 100 |
| % Collector-filterer | - | 38 | 53 | 36 | 15 |
| %Collector-gatherer | + | 46 | 60 | 64 | 95 |
| % Scraper | - | 46 | 73 | 73 | 100 |
| % Shredder | - | 54 | 87 | 64 | 90 |
| % Other | - | 62 | 53 | 0 | 0 |

Metrics in bold had discrimination efficiencies greater than 50 percent for Assemblage Group 2 and greater than 80 percent for Assemblage Group 1.

For intermittent streams, six metrics exhibited low natural variation (C.V.<0.25) and high discrimination efficiency includes: taxa richness, Diptera taxa, percent EPT, percent sensitive EPT, percent intolerant, and tolerance value. Seven metrics had low variation and high discrimination efficiency for perennial streams: taxa richness, Diptera taxa, Coleoptera taxa, Ephemeroptera taxa, EPT taxa richness, percent EPT, and tolerance value.

Redundancy between selected metrics across all sites was established by examining Pearson product-moment correlation coefficients, which measure the strength of the

linear relationship between two variables. Metrics with correlation coefficients $|r| \geq 0.90$ were considered to be redundant. Highly redundant metric pairs include: percent intolerant, tolerance value, and percent sensitive EPT; Ephemeroptera taxa, taxa richness, and EPT taxa; and Coleoptera taxa and taxa richness. Four metrics exhibited low redundancy at both perennial and intermittent sites: taxa richness, Diptera taxa richness, percent EPT, and tolerance value. These four metrics all exhibit low natural variability, high ability to discriminate disturbance, and relatively low redundancy. The range of values for these metrics from minimally disturbed streams is shown in Table 6-10.

This analysis suggests that certain biological metrics could be used to measure the response of stream biota to human disturbance in a subset of streams in the Bay Area. Further work is needed to (1) refine and develop this analytical approach and (2) carefully examine the range of physical settings that reference conditions can be developed for, given constraints on the availability of minimally disturbed streams in the Bay Area. This type of analytical tool could serve to further the utility of bioassessment in the Bay Area by establishing a set of standards that reflect minimally disturbed conditions with which to compare to biological conditions from test sites.

Table 6-10. Metrics for minimally disturbed streams in the San Francisco Bay Area

| | Intermittent | | | Perennial | | |
|------------------------|--------------|--------|---------|-----------|--------|---------|
| | Minimum | Median | Maximum | Minimum | Median | Maximum |
| Taxa Richness | 28 | 33 | 34 | 38 | 46 | 59 |
| Diptera Taxa | 7 | 8 | 10 | 7 | 9 | 14 |
| % EPT | 55 | 70 | 85 | 40 | 62 | 74 |
| Tolerance Value | 3.58 | 3.96 | 4.47 | 3.07 | 3.46 | 4.26 |

6.6.1.2 Indicator Taxa

Aquatic invertebrates exhibit a wide range of tolerances to natural habitat conditions and pollutants. Taxa that are commonly found in poor water quality conditions are considered to be tolerant of pollution. Benthic assemblages that are exclusively dominated by these tolerant taxa indicate low biological integrity. The taxa identified in this analysis could be useful taxa to focus on in future bioassessment projects in Bay Area streams.

Sites with poor water quality or habitat conditions at the time of invertebrate sampling (spring season) were identified using the following criteria:

1. Total pesticide concentrations (sum of 34 organochlorine and 44 organophosphate pesticides) in water greater than 0.05 ppb
2. Metals concentrations in water exceeding water quality standards
3. Nitrate concentrations greater than 4.0 mg/L, or ammonia or nitrite concentrations greater than 0.02 mg/L
4. Minimum dissolved oxygen concentrations below 1.0 mg/L
5. Concrete channels with little or no substrate or other habitat

Three sites exceeded the pesticide criteria: SLE030 (0.118 ppb), ALP010 (0.082 ppb), and SPA020 (0.063 ppb). One site exceeded a chronic standard for metals: SLE030 (30.6

ppb chromium). Five sites exceeded one or more of the criteria for nutrients: ALP010 (4.39 mg/L nitrate, 0.037 mg/L nitrite), ALP100 (4.10 mg/L nitrate, 0.120 mg/L nitrite), ALP110 (8.04 mg/L nitrate, 0.049 mg/L nitrite), SLE030 (7.65 mg/L nitrate), and STE020 (0.020 mg/L nitrite). One site failed to meet the minimum dissolved oxygen criteria: LAG160 (0.10 mg/L). At one site, PER020, the channel was a rectangular concrete channel with little available habitat. In summary, eight sites were identified as having poor water quality or habitat conditions out of all the BMI samples collected by SWAMP.

“Common tolerant taxa” are defined in this report as taxa that occur at 50 percent or more of sites identified as having poor water quality at the time of invertebrate sampling. Eight common tolerant taxa were identified:

1. Chironomidae, chironomid midges (Diptera)
2. Oligochaeta, segmented worms such as Naidids and Tubificids
3. *Simulium* sp., black fly larvae (Diptera: Simuliidae)
4. *Baetis* sp., a baetid mayfly (Ephemeroptera: Baetidae)
5. Ostracoda, seed shrimp
6. *Physa/Physella*, two genera of physid snails (Gastropoda)
7. Turbellaria, flatworms
8. Spermontidae, water mites, including *Sperchon* sp. (Acari)

Together, these 8 taxa made up over 95 percent of the total number of organisms at the 8 sites with poor water quality (Note: these taxa represent operational taxonomic units for the purposes of identification, and may in fact include tens or hundreds of individual species). Because these taxa are the dominant organisms present at sites with poor water quality, benthic assemblages with a high percentage of the total number of organisms belonging to these taxa may indicate poor biological integrity. At sites draining urban land uses, over 90 percent of organisms were common tolerant taxa (Table 6-11). Less than 50 percent of organisms at sites draining open space were common tolerant taxa, while sites in other land use classes exhibited a wide range of common tolerant taxa.

Table 6-11. Occurrence of tolerant taxa at sites in different land use classes

| | Common Tolerant Taxa | | |
|--------------------------|----------------------|--------|------|
| | Min. | Median | Max. |
| Open Space | 14% | 29% | 46% |
| Grazing | 16% | 62% | 95% |
| Agriculture | 47% | 66% | 89% |
| Rural Residential | 6% | 31% | 96% |
| Urban | 90% | 99% | 1% |

“Common intolerant taxa” are defined in this report as taxa that occur at greater than 50 percent of minimally disturbed sites and are absent from sites with poor water quality (identified above). These taxa represent organisms that are widespread throughout the Bay Area but are sensitive to water quality degradation. Because intermittency of flow was found to be the most important natural factor affecting taxon occurrence, separate

lists of intolerant taxa were developed for perennial and intermittent streams (Table 6-12 and Table 6-13).

Table 6-12. List of common intolerant taxa: perennial streams

-
1. *Rhyacophila* sp., a free-living caddisfly (Trichoptera: Rhyacophilidae)
 2. *Malenka* sp., a nemourid stonefly (Plecoptera: Nemouridae)
 3. *Dipheter hageni*, a baetid mayfly (Ephemeroptera: Baetidae)
 4. *Parthina* sp., a tube-case making caddisfly (Trichoptera: Odontoceridae)
 5. *Zaitzevia* sp. (larvae), a riffle beetle (Coleoptera: Elmidae)
 6. *Drunella* sp., an ephemereid mayfly (Ephemeroptera: Ephemereididae)
 7. *Ironodes* sp., a heptageniid mayfly (Ephemeroptera: Heptageniidae)
 8. *Paraleptophlebia* sp., a prong-gilled mayfly (Ephemeroptera: Leptophlebiidae)
 9. *Calineuria californica*, a golden stonefly (Plecoptera: Perlidae)
 10. *Neophylax* sp., a tube-case making caddisfly (Trichoptera: Uenoide)
 11. *Zaitzevia* sp. (adults), a riffle beetle (Coleoptera: Elmidae)
 12. *Lepidostoma* sp., a tube-case making caddisfly (Trichoptera: Lepidostoma)
 13. *Eubrianax edwardsi*, a water penny beetle (Coleoptera: Psephenidae)
 14. *Cinygmula* sp., a genus of heptageniid mayflies (Ephemeroptera: Heptageniidae)
 15. *Sweltsa* sp., a chloroperlid stonefly (Plecoptera: Chloroperlidae)
 16. *Maruina lanceolata*, a moth fly (Diptera: Psychodidae)
 17. *Limnophila* sp., a crane fly (Diptera: Tipulidae)
 18. *Serratella* sp., an ephemereid mayfly (Ephemeroptera: Ephemereididae)
 19. *Wormaldia* sp., a retreat-making caddisfly (Trichoptera: Philopotamidae)
 20. *Cordulegaster dorsalis*, the pacific spiketail dragonfly (Odonata: Cordulegasteridae)
 21. *Narpus* sp., a riffle beetle (Coleoptera: Elmidae)
 22. *Hesperoperla* sp., a golden stonefly (Plecoptera: Perlidae)
 23. *Micrasema* sp., a tube-case making caddisfly (Trichoptera: Brachycentridae)
 24. *Polycentropus* sp., a trumpet-net caddisfly (Trichoptera: Ploycentropodidae)
 25. *Torrenticola* sp., a water mite (Acari: Torrenticolidae)
-

Table 6-13. List of common intolerant taxa: intermittent streams

-
1. *Rhyacophila* sp., a free-living caddisfly (Trichoptera: Rhyacophilidae)
 2. *Drunella* sp., an ephemereid mayfly (Ephemeroptera: Ephemereididae)
 3. *Paraleptophlebia* sp., a prong-gilled mayfly (Ephemeroptera: Leptophlebiidae)
 4. *Sweltsa* sp., a chloroperlid stonefly (Plecoptera: Chloroperlidae)
 5. Capniidae, a family of stoneflies (Plecoptera)
 6. *Neohermes* sp., a fishfly (Megaloptera: Corydalidae)
 7. *Malenka* sp., a nemourid stonefly (Plecoptera Nemouridae)
 8. *Lepidostoma* sp., a tube-case making caddisfly (Trichoptera: Lepidostoma)
 9. *Dicranota* sp., a crane fly (Diptera: Tipulidae)
 10. Hydrobiidae, pebblesnails (Gastropoda)
 11. *Meringodixa* sp., a dixid fly (Diptera: Dixidae)
 12. *Baumannella alameda*, a perlodid stonefly (Plecoptera: Perlodidae)
 13. *Taenionema* sp., a stonefly (Plecoptera: Taeniopterigidae)
-

Tolerance values (from the CAMLnet STE) assigned to these taxa are generally low (less than 3), indicating pollution intolerance, with several exceptions. Hydrobiidae was found exclusively at sites with excellent water quality and habitat conditions, despite its tolerance value of 8. *Polycentropus* sp. was given a tolerance value of 6, but in the Bay Area all 11 sites where it was found are believed to exhibit excellent biological integrity. *Dipheter hageni* was assigned a tolerance value of 5, but was absent from urban sites and sites with poor water quality. *Paraleptophlebia* sp., *Ironodes* sp., *Zaitzevia* sp., and *Cinygmula* sp., which were assigned tolerance values of 4, were found at many sites with low to moderate levels of disturbance but not at sites with poor water quality.

Although most of the intolerant taxa from intermittent sites were also common at perennial sites, several taxa were found more frequently or exclusively at intermittent sites, including Capniidae, *Neohermes* sp., *Dicranota* sp., *Meringodixa* sp., *Baumannella alameda*, and *Taenionema* sp.

Although these 26 taxa are fairly widespread, occurring at 10 to 50 percent of sites, they often represent a small percentage of the total number of organisms collected at a site. Consequently, the number of “common intolerant taxa” found at a site, rather than the percentage of the total number of organisms made up of these taxa, may be an excellent indicator of biological integrity.

6.6.2 Water and Sediment Chemistry

The geology of the San Francisco Bay region is such that watersheds contain mineral deposits and ultramafic rocks that are naturally enriched in metals relative to the mean composition of the continental crust. Relatively high concentrations of chromium, nickel, and vanadium have eroded from the watersheds and deposited in San Francisco Bay sediments. Hornberger *et al.* (1998) found that concentrations of these metals have remained fairly constant since before major anthropogenic activities in the mid-1800s.

Chromium and nickel frequently exceed sediment quality guidelines throughout the region. However, these metals do not seem to be highly biologically available.

Although other metals are also naturally occurring, anthropogenic activities have increased their concentrations in sediments. Watersheds in the San Francisco Bay region contain some of the richest geologic deposits of mercury in the world. Mining activities that occurred primarily between 1850 and 1900 mobilized these deposits, thereby increasing concentrations in the San Francisco estuary and many of its watersheds. Guadalupe Creek, downstream from the second largest historical mercury mine in the world, has high concentrations of mercury that are contributing to fish contamination in San Francisco Bay.

6.6.3 Nutrients Reference Conditions

As noted in Section 6.3.1.1, nutrient concentrations in this report were compared with U.S. EPA's reference guidelines for Aggregate Ecoregion III, Ecoregion 6 (Southern and Central California Chaparral and Oak Woodlands) streams (U.S. EPA 2000a). The guidelines were derived from 25th percentile values of stream monitoring data collected from 1990 through 1999. U.S. EPA made the somewhat arbitrary assumption that the 25th percentile represents reference conditions, but encouraged states and tribes to establish more informed reference conditions and criteria.

In 2003, Tetra Tech, Inc. conducted a pilot study for U.S. EPA to, among other objectives, identify more meaningful reference conditions for nutrients in Ecoregion 6 streams. The study utilized data collected by a variety of sources from 1975 through 2002. Reference or "minimally impacted" streams were defined as those meeting the following criteria (U.S. EPA 2003):

- Beneficial uses met
- No field indicators of hydrological disturbance
- Land use predominantly natural
- Natural flow levels maintained (i.e. minimal flow diversions)
- Stable channel with well developed riparian zone and intact substrate

Median nutrient concentrations for these streams were reported as 0.02, 0.05, 0.25, 0.04, and 0.08 mg/L for NH₃-N, NO₃-N, TKN, ortho-P, and total P, respectively. These values are similar to, but somewhat higher than U.S. EPA reference values. Maximum concentrations reported for the minimally impacted streams were 3.25, 2.85, 1.20, 0.23, and 0.30 mg/L for the respective constituents.

In this study, only two sites designated as open space were sampled for nutrients: SLE230 in EBMUD protected-watershed land above Upper San Leandro Reservoir and PES190 on upper Pescadero Creek in Portola State Park. At both sites, results for NH₃-N and NO₃-N were above Tetra Tech's median concentration for reference streams in all seasons. In at least one season, concentrations for TKN, ortho-P, and total P were also above the median. However, all concentrations were notable below Tetra Tech's

maximum for minimally impacted streams. Nutrient data collected by SWAMP at reference sites add to our understanding of regional reference concentrations and can be used to support development of nutrient WQOs.

Table of Contents

7 Results by Watershed..... 7-1

7.1 Walker Creek Watershed 7-2

 7.1.1 Sites of Concern..... 7-2

 7.1.2 Water Quality in Relation to Land Use..... 7-3

 7.1.3 Macroinvertebrate Assemblages and Physical Habitat..... 7-3

 7.1.4 Basic Water Quality Field Measurements 7-12

 7.1.5 Water, Sediment, and Clam Tissue Chemistry 7-16

 7.1.6 Water and Sediment Toxicity 7-16

 7.1.7 Coliform Bacteria..... 7-16

7.2 Lagunitas Creek Watershed 7-17

 7.2.1 Sites of Concern..... 7-17

 7.2.2 Water Quality in Relation to Land Use..... 7-18

 7.2.3 Macroinvertebrate Assemblages and Physical Habitat..... 7-18

 7.2.4 Basic Water Quality Field Measurements 7-27

 7.2.5 Water, Sediment, and Clam Tissue Chemistry 7-31

 7.2.6 Water and Sediment Toxicity 7-32

 7.2.7 Coliform Bacteria..... 7-33

7.3 San Leandro Creek Watershed..... 7-34

 7.3.1 Sites of Concern..... 7-34

 7.3.2 Water Quality in Relation to Land Use..... 7-35

 7.3.3 Macroinvertebrate Assemblages and Physical Habitat..... 7-36

 7.3.4 Basic Water Quality Field Measurements 7-42

 7.3.5 Water, Sediment, and Clam Tissue Chemistry 7-46

 7.3.6 Water and Sediment Toxicity 7-47

 7.3.7 Coliform Bacteria..... 7-48

7.4 Wildcat/San Pablo Creek Watershed 7-49

 7.4.1 Sites of Concern..... 7-49

 7.4.2 Water Quality in Relation to Land Use..... 7-50

 7.4.3 Macroinvertebrate Assemblages and Physical Habitat..... 7-51

 7.4.4 Basic Water Quality Field Measurements 7-58

 7.4.5 Water, Sediment, and Clam Tissue Chemistry 7-62

 7.4.6 Water and Sediment Toxicity 7-63

 7.4.7 Coliform Bacteria..... 7-64

7.5 Suisun Creek Watershed 7-65

 7.5.1 Sites of Concern..... 7-65

 7.5.2 Water Quality in Relation to Land Use..... 7-66

 7.5.3 Macroinvertebrate Assemblages and Physical Habitat..... 7-66

 7.5.4 Basic Water Quality Field Measurements 7-73

 7.5.5 Water, Sediment, and Clam Tissue Chemistry 7-77

| | | |
|-------|---|-------|
| 7.5.6 | Water and Sediment Toxicity | 7-77 |
| 7.5.7 | Coliform Bacteria..... | 7-78 |
| 7.6 | Arroyo Las Positas Watershed..... | 7-79 |
| 7.6.1 | Sites of Concern..... | 7-79 |
| 7.6.2 | Water Quality in Relation to Land Use..... | 7-80 |
| 7.6.3 | Macroinvertebrate Assemblages and Physical Habitat..... | 7-80 |
| 7.6.4 | Basic Water Quality Field Measurements | 7-86 |
| 7.6.5 | Water, Sediment, and Clam Tissue Chemistry | 7-91 |
| 7.6.6 | Water and Sediment Toxicity | 7-91 |
| 7.6.7 | Coliform Bacteria..... | 7-92 |
| 7.7 | Pescadero Creek/Butano Creek Watershed | 7-93 |
| 7.7.1 | Sites of Concern..... | 7-93 |
| 7.7.2 | Water Quality in Relation to Land Use..... | 7-94 |
| 7.7.3 | Macroinvertebrate Assemblages and Physical Habitat..... | 7-94 |
| 7.7.4 | Basic Water Quality Field Measurements | 7-103 |
| 7.7.5 | Water, Sediment, and Clam Tissue Chemistry | 7-107 |
| 7.7.6 | Water and Sediment Toxicity | 7-107 |
| 7.7.7 | Coliform Bacteria..... | 7-108 |
| 7.8 | San Gregorio Creek Watershed | 7-109 |
| 7.8.1 | Sites of Concern..... | 7-109 |
| 7.8.2 | Water Quality in Relation to Land Use..... | 7-109 |
| 7.8.3 | Macroinvertebrate Assemblages and Physical Habitat..... | 7-110 |
| 7.8.4 | Basic Water Quality Field Measurements | 7-116 |
| 7.8.5 | Water, Sediment, and Clam Tissue Chemistry | 7-120 |
| 7.8.6 | Water and Sediment Toxicity | 7-120 |
| 7.8.7 | Coliform Bacteria..... | 7-121 |
| 7.9 | Stevens Creek/Permanente Creek Watershed..... | 7-122 |
| 7.9.1 | Sites of Concern..... | 7-122 |
| 7.9.2 | Water Quality in Relation to Land Use..... | 7-124 |
| 7.9.3 | Macroinvertebrate Assemblages and Physical Habitat..... | 7-124 |
| 7.9.4 | Basic Water Quality Field Measurements | 7-131 |
| 7.9.5 | Water, Sediment, and Clam Tissue Chemistry | 7-135 |
| 7.9.6 | Water and Sediment Toxicity | 7-135 |
| 7.9.7 | Coliform Bacteria..... | 7-136 |

List of Tables

| | |
|--|-------|
| Table 7-1. Summary of sites with exceedances in Walker Creek watershed | 7-2 |
| Table 7-2. Bioassessment sites in the Walker Creek watershed..... | 7-9 |
| Table 7-3. Biological metrics from Walker Creek watershed | 7-11 |
| Table 7-4. Summary of sites with exceedances in Lagunitas Creek watershed | 7-17 |
| Table 7-5. Bioassessment sites in the Lagunitas Creek watershed..... | 7-23 |
| Table 7-6. Biological metrics from the Lagunitas Creek watershed | 7-26 |
| Table 7-7. Summary of sites with exceedances in San Leandro Creek watershed..... | 7-34 |
| Table 7-8. Bioassessment sites in the San Leandro Creek watershed | 7-39 |
| Table 7-9. Biological metrics from the San Leandro Creek watershed..... | 7-41 |
| Table 7-10. Summary of sites with exceedances in Wildcat Creek watershed | 7-49 |
| Table 7-11. Summary of sites with exceedances in San Pablo Creek watershed..... | 7-50 |
| Table 7-12. Bioassessment sites in the Wildcat/San Pablo Creek watershed..... | 7-55 |
| Table 7-13. Biological metrics from the San Pablo Creek/Wildcat Creek watershed..... | 7-57 |
| Table 7-14. Summary of sites with exceedances in Suisun Creek watershed | 7-65 |
| Table 7-15. Bioassessment sites in the Suisun Creek watershed..... | 7-70 |
| Table 7-16. Biological metrics from the Suisun Creek watershed | 7-72 |
| Table 7-17. Summary of sites with exceedances in Arroyo Las Positas | 7-79 |
| Table 7-18. Bioassessment sites in the Arroyo Las Positas watershed..... | 7-83 |
| Table 7-19. Biological metrics from the Arroyo Las Positas watershed | 7-84 |
| Table 7-20. Percentage abundance of taxonomic groups in Arroyo Las Positas..... | 7-85 |
| Table 7-21. Summary of sites with exceedances in Pescadero Creek watershed..... | 7-93 |
| Table 7-22. Summary of sites with exceedances in Butano Creek watershed..... | 7-94 |
| Table 7-23. Bioassessment sites in the Pescadero Creek/Butano Creek watershed | 7-100 |
| Table 7-24. Biological metrics in the Pescadero Creek/Butano Creek watershed | 7-102 |
| Table 7-25. Summary of sites with exceedances in San Gregorio Creek watershed..... | 7-109 |
| Table 7-26. Bioassessment sites in the San Gregorio Creek watershed | 7-113 |
| Table 7-27. Biological metrics from San Gregorio Creek watershed..... | 7-115 |
| Table 7-28. Summary of sites with exceedances in Stevens Creek watershed..... | 7-122 |
| Table 7-29. Summary of sites with exceedances in Permanente Creek watershed | 7-123 |
| Table 7-30. Bioassessment sites in the Stevens Creek/Permanente Creek watershed..... | 7-128 |
| Table 7-31. Biological metrics from the Stevens Creek/Permanente Creek watershed | 7-130 |

List of Figures

| | |
|--|------|
| Figure 7-1. Map of Walker Creek benthic macroinvertebrate index scores | 7-8 |
| Figure 7-2. Cluster dendrogram of sites in the Walker Creek watershed..... | 7-9 |
| Figure 7-3. NMS ordination of sites in the Walker Creek watershed..... | 7-10 |
| Figure 7-4. Map of temperature and DO levels in Walker Creek watershed | 7-13 |
| Figure 7-5. Temperature monitoring in Walker Creek watershed..... | 7-14 |
| Figure 7-6. DO monitoring in Walker Creek watershed..... | 7-14 |
| Figure 7-7. pH monitoring in Walker Creek watershed | 7-15 |
| Figure 7-8. Specific conductance monitoring in Walker Creek watershed | 7-15 |
| | |
| Figure 7-9. Map of Lagunitas Creek benthic macroinvertebrate index scores | 7-22 |
| Figure 7-10. Cluster dendrogram of sites in the Lagunitas Creek watershed..... | 7-24 |
| Figure 7-11. NMS ordination of sites in the Lagunitas Creek watershed..... | 7-25 |
| Figure 7-12. Map of temperature and DO levels in Lagunitas Creek..... | 7-28 |
| Figure 7-13. Temperature monitoring in Lagunitas Creek watershed..... | 7-29 |
| Figure 7-14. DO monitoring in Lagunitas Creek watershed..... | 7-29 |
| Figure 7-15. pH monitoring in Lagunitas Creek watershed | 7-30 |
| Figure 7-16. Specific conductance monitoring in Lagunitas Creek watershed | 7-30 |
| Figure 7-17. Toxicity tests in Walker and Lagunitas creeks | 7-32 |
| | |
| Figure 7-18. Map of index scores for macroinvertebrates in San Leandro Creek..... | 7-38 |
| Figure 7-19. Cluster dendrogram of sites in the San Leandro Creek watershed | 7-39 |
| Figure 7-20. NMS ordination of sites in the San Leandro Creek watershed | 7-43 |
| Figure 7-21. Map of temperature and DO levels in San Leandro Creek | 7-43 |
| Figure 7-22. Temperature monitoring in San Leandro Creek..... | 7-44 |
| Figure 7-23. DO monitoring in San Leandro Creek | 7-44 |
| Figure 7-24. pH monitoring in San Leandro Creek | 7-45 |
| Figure 7-25. Specific conductance monitoring in San Leandro Creek..... | 7-45 |
| Figure 7-26. Toxicity tests in San Leandro Creek | 7-47 |
| | |
| Figure 7-27. Map of Wildcat and San Pablo creeks benthic macroinvertebrates index scores . | 7-54 |
| Figure 7-28. Cluster dendrogram sites in the Wildcat/San Pablo Creek watershed | 7-55 |
| Figure 7-29. NMS ordination of San Pablo Creek/Wildcat Creek sites | 7-56 |
| Figure 7-30. Map of temperature and DO levels in Wildcat and San Pablo creeks | 7-59 |
| Figure 7-31. Temperature monitoring in Wildcat and San Pablo creeks..... | 7-60 |
| Figure 7-32. DO monitoring in Wildcat and San Pablo creeks | 7-60 |
| Figure 7-33. pH monitoring in Wildcat and San Pablo creeks | 7-61 |
| Figure 7-34. Specific conductance monitoring in Wildcat and San Pablo creeks | 7-61 |
| Figure 7-35. Toxicity tests in Wildcat and San Pablo creeks | 7-64 |
| | |
| Figure 7-36. Map of Suisun Creek benthic macroinvertebrates index scores | 7-69 |
| Figure 7-37. Custer dendrogram of sites in the Suisun Creek watershed..... | 7-70 |
| Figure 7-38. NMS ordination of sites in the Suisun Creek watershed..... | 7-71 |
| Figure 7-39. Map of temperature and DO levels in Suisun Creek..... | 7-74 |
| Figure 7-40. Temperature monitoring in Suisun Creek | 7-75 |

| | |
|--|-------|
| Figure 7-41. DO monitoring in Suisun Creek..... | 7-75 |
| Figure 7-42. pH monitoring in Suisun Creek | 7-76 |
| Figure 7-43. Specific conductance monitoring in Suisun Creek | 7-76 |
| Figure 7-44. Toxicity tests in Suisun Creek | 7-78 |
| | |
| Figure 7-45. Map of Arroyo Las Positas benthic macroinvertebrate index scores..... | 7-82 |
| Figure 7-46. Cluster dendrogram of sites in the Arroyo Las Positas watershed..... | 7-83 |
| Figure 7-47. NMS ordination of sites in the Arroyo Las Positas watershed | 7-84 |
| Figure 7-48. Taxonomic composition in the Arroyo Las Positas watershed..... | 7-86 |
| Figure 7-49. Map of temperature and DO monitoring in Arroyo Las Positas..... | 7-88 |
| Figure 7-50. Temperature monitoring in Arroyo Las Positas..... | 7-89 |
| Figure 7-51. DO monitoring in Arroyo Las Positas | 7-89 |
| Figure 7-52. pH monitoring in Arroyo Las Positas | 7-90 |
| Figure 7-53. Specific conductance monitoring in Arroyo Las Positas | 7-90 |
| Figure 7-54. Toxicity tests in Arroyo Las Positas | 7-92 |
| | |
| Figure 7-55. Map of Pescadero and Butano creeks benthic macroinvertebrate index scores.... | 7-99 |
| Figure 7-56. Cluster dendrogram of sites in the Pescadero Creek/Butano Creek watershed .. | 7-100 |
| Figure 7-57. NMS ordination of sites in the Pescadero Creek/Butano Creek watershed | 7-101 |
| Figure 7-58. Riparian canopy cover and stream temperature in Pescadero Creek | 7-103 |
| Figure 7-59. Map of temperature and DO levels in Pescadero and Butano creeks | 7-104 |
| Figure 7-60. Temperature monitoring in Pescadero and Butano creeks..... | 7-105 |
| Figure 7-61. DO monitoring in Pescadero and Butano creeks | 7-105 |
| Figure 7-62. pH monitoring in Pescadero and Butano creeks | 7-106 |
| Figure 7-63. Specific conductance monitoring in Pescadero and Butano creeks | 7-106 |
| Figure 7-64. Toxicity tests from Pescadero and Butano creeks..... | 7-108 |
| | |
| Figure 7-65. Map of San Gregorio Creek benthic macroinvertebrate index scores | 7-112 |
| Figure 7-66. Cluster dendrogram of sites in the San Gregorio Creek watershed | 7-113 |
| Figure 7-67. NMS ordination of sites in the San Gregorio Creek watershed..... | 7-114 |
| Figure 7-68. Map of temperature and DO levels in San Gregorio Creek watershed..... | 7-117 |
| Figure 7-69. Temperature monitoring in San Gregorio Creek watershed | 7-118 |
| Figure 7-70. DO monitoring in San Gregorio Creek watershed..... | 7-118 |
| Figure 7-71. pH monitoring of sites in San Gregorio Creek watershed | 7-119 |
| Figure 7-72. Specific conductance monitoring in San Gregorio Creek watershed..... | 7-119 |
| Figure 7-73. Toxicity tests in San Gregorio Creek..... | 7-121 |
| | |
| Figure 7-74. Map of Stevens/Permanente creeks benthic macroinvertebrate index scores..... | 7-127 |
| Figure 7-75. Cluster dendrogram for the Stevens Creek/Permanente Creek watershed..... | 7-128 |
| Figure 7-76. NMS ordination of sites in the Stevens Creek/Permanente Creek watershed ... | 7-129 |
| Figure 7-77. Map of temperature and DO levels in Stevens and Permanente creeks..... | 7-132 |
| Figure 7-78. Temperature monitoring in Stevens and Permanente creeks | 7-133 |
| Figure 7-79. DO monitoring in Stevens and Permanente creeks..... | 7-133 |
| Figure 7-80. pH monitoring in Stevens and Permanente creeks..... | 7-134 |
| Figure 7-81. Specific conductance monitoring in Stevens and Permanente creeks | 7-134 |
| Figure 7-82. Toxicity tests in Stevens and Permanente creeks..... | 7-136 |

7 Results by Watershed

For a comprehensive understanding of the results presented in this section, review Section 6 for background on the study's analytical approach.

To more thoroughly understand the results for each watershed, refer to:

- Section 1: Map notes explaining the categories used in the maps
- Section 2: Watershed descriptions
- Section 3: Watershed maps and Table 3-4, which lists what was monitored where and general land use categories
- Section 4: Methods; Tables 4-2 and 4-3 for thresholds; Table 4-4 for biological metrics
- Section 6: Overview of regional results and context for watershed-specific results

Organization of this section

7.x.1: Brief summary of the sites of concern identified in this study. A table summarizes threshold exceedances for each watershed (See Table 4-2 and Table 4-3).

7.x.2: Summary description of water quality in relation to land use.

7.x.3: Analysis of bioassessment results based on findings related to benthic macroinvertebrate assemblages and physical habitat. Tables and figures follow the narrative:

- Map of index scores
- List of sampling sites
- Cluster dendrogram
- Ordination graph
- Table of bioassessment metrics

7.x.4: Summary of results for basic water quality field parameters from continuous monitoring followed by:

- Map of temperature and dissolved oxygen (DO) levels
- Boxplots summarizing temperature, DO, pH, and specific conductance

7.x.5: Summary of results for water, sediment, and tissue chemistry

7.x.6: Summary of results for water and sediment toxicity

7.x.7: Summary of results for coliform bacteria

Figures summarizing results in all watersheds for sections 7.x.5, 7.x.6, and 7.x.7 are presented and discussed in Section 6.

Refer to Appendices D through K for the original data.

7.1 Walker Creek Watershed

7.1.1 Sites of Concern

Exceedances noted in the summary table (Table 7-1) are discussed in the relevant sub-section for this watershed. A weight-of-evidence approach was used to evaluate sites of concern. Individual exceedances of parameters, particularly those without regulatory objectives, are considered in context. In general, nickel and chromium levels are geologically elevated in the Bay Area, particularly where there are serpentine deposits. Nutrient levels (nitrate and phosphorus) are above U.S. EPA guidelines at most sites.

Invertebrate communities were in poor condition in some sites, particularly Keys Creek at Tomales (WLK030), below Laguna Lake (WLK130) and the two most upstream sites, at Arroyo Sausal (WLK230) and the Cheese Factory (WLK240). These are discussed in Regional Trends, Section 6.1. The Walker Creek Ranch (WLK160) and Gambonini Mine (WLK180) sites had elevated water temperatures and either low DO or high pH in at least some measurements. Nickel and chromium are elevated in Walker Creek sediments due to high levels in geologic deposits. In general, mercury contamination from mining activities is the main chemical concern in this watershed, although the samples collected in this study were not highly elevated. Erosion and sedimentation caused by lack of riparian habitat protection and exacerbated by grazing operations appear to be primary issues affecting water quality in this watershed.

Table 7-1. Summary of sites with exceedances in Walker Creek watershed

| <i>Analyte</i> | <i>Standard</i> | <i># of sites</i> |
|--|-----------------|-------------------|
| FIELD MEASURES of 5 sites sampled | | |
| Temperature | MWAT, coho | 2 |
| Oxygen, dissolved | minimum, COLD | 1 |
| | minimum, WARM | 1 |
| | 3-month median | 1 |
| pH | range | 1 |
| CONVENTIONAL WATER QUALITY of 4 sites sampled | | |
| Nitrate as N | maximum | 4 |
| Phosphorus, total as P | maximum | 4 |
| SEDIMENT CHEMISTRY of 1 site sampled | | |
| <i>Metals</i> | | |
| Chromium | PEC | 1 |
| Mercury | TEC | 1 |
| Nickel | PEC | 1 |

7.1.2 Water Quality in Relation to Land Use

With the exception of the swimming beach at Turtle Pond, all sites drained areas used for grazing. Mining, agriculture, and residential uses were also common. There was generally insufficient data to analyze trends in any of the parameters relative to land use, with the exception of the bioassessment data discussed in Section 6.1 above. Benthic macroinvertebrate sample analyses from Chileno Creek provided a good example of improved water quality associated with a restoration effort that included grazing exclusion and riparian zone planting. Primary water quality stressors appear to be related to temperature and DO, parameters that vary region-wide with riparian habitat condition. Grazing practices that enhance riparian condition should translate into improved water quality.

7.1.3 Macroinvertebrate Assemblages and Physical Habitat

Three groups of similar sites were identified from the cluster dendrogram (Figure 7-2) and ordination graph (Figure 7-3).

- The Walker Creek/Salmon Creek group (WS) includes sites on Walker Creek above the confluence with Chileno Creek (WLK140, WLK160), Salmon Creek (WLK180, WLK190), and Verde Canyon Creek (WLK170). Sites in this group are the most closely related on the cluster dendrogram (Figure 7-2) and ordination graph (Figure 7-3), suggesting they are the most taxonomically similar.
- The Keys Creek (KC) group includes two sites on Keys Creek (WLK030, WLK050).
- The Upper Arroyo Sausal (UAS) group includes two sites on Arroyo Sausal above Soulajule Reservoir (WLK230, WLK240).

Sites that do not fall into these three groups, including those on Chileno Creek (WLK100, WLK120, WLK130) and Arroyo Sausal (WLK200), are discussed separately.

Walker Creek/Salmon Creek group

Metric values for four sites in this group (WLK140, WLK170, WLK180, and WLK190) are within or near the range of values from minimally disturbed perennial streams (Table 6-10). This suggests that benthic invertebrate assemblages in these locations were relatively undisturbed.

Metric values for the fifth site, WLK160, are well outside of this range. While invertebrate diversity was fairly high, the numerical dominance of pollution-tolerant taxa suggests that the invertebrate assemblage was moderately impacted from human-caused disturbances.

Abundant taxa found in the Walker Creek/Salmon Creek group (WS) of sites include:

| | |
|--------------------------------------|-----|
| Chironomidae | 24% |
| <i>Baetis</i> sp. (Baetidae) | 14% |
| <i>Gumaga</i> sp. (Sericostomatidae) | 11% |
| <i>Malenka</i> sp. (Nemouridae) | 7% |

| | |
|---------------------------------------|----|
| <i>Simulium</i> sp. (Simuliidae) | 6% |
| <i>Optioservus</i> sp. (Elmidae) | 5% |
| <i>Suwallia</i> sp. (Chloroperlidae) | 4% |
| <i>Dipheter hageni</i> (Baetidae) | 3% |
| <i>Agapetus</i> sp. (Glossosomatidae) | 3% |

Other rare intolerant taxa found at the sites include *Paracymus* sp. (Coleoptera: Hydrophilidae), *Protoptila* sp. (Glossosomatidae), and *Onocosmoecus* sp. (Limnephilidae). Two taxa from WLK140, *Acentrella* sp. (Baetidae) and *Culoptila* sp. (Glossosomatidae), have not been found elsewhere in the Bay Area.

Although these five sites were quite similar taxonomically, the relative abundance of taxa was quite different at WLK160 than at the other four sites.

- Chironomids composed less than 10 percent of organisms at WLK140 and WLK180, but nearly 50 percent at WLK160.
- *Simulium* sp. made up 20 percent of individuals at WLK160, but was much less abundant at other sites.

The percentage of common tolerant taxa ranged from 23 percent at WLK190 to 77 percent at WLK160. These compositional differences are reflected in the tolerance value metric, which ranged from 3.6 at WLK190 to 5.2 at WLK160. These differences suggest that the benthic assemblage at WLK160 was affected by water quality or habitat impacts, but the impacts were not severe enough to completely eliminate pollution-sensitive taxa.

Keys Creek group

The high percentages of common tolerant taxa observed at the Keys Creek sites are similar to conditions observed at sites in urban areas with very poor biological integrity.

Abundant taxa at two sites on Keys Creek (WLK030 and WLK050) included

| | |
|----------------------------------|------|
| Chironomidae | 39% |
| Oligochaeta | 19% |
| <i>Simulium</i> sp. (Simuliidae) | 16% |
| <i>Baetis</i> sp. (Baetidae) | 13%. |

Together, these common tolerant taxa made up 94 percent and 78 percent of all organisms at WLK030 and WLK050, respectively. Also abundant at WLK050 was *Hyaella* sp. (Amphipoda: Hyalellidae), a pollution tolerant organism (tolerance value = 8) that made up 15 percent of the sample.

Despite the numerical dominance of four tolerant taxa, diversity at these sites, especially WLK050, was higher than observed at urban sites. Less abundant (<1 percent) taxa found at both sites included a dytiscid beetle (*Agabus* sp.), an empid fly (*Clinocera* sp.), a relatively rare and intolerant hydroptilid caddisfly (*Oxyethira* sp.), the water mite Lebertiidae (Acari), and the

bivalve Sphaeriidae (Veneroidea). One common intolerant taxa was collected at each site, a single individual *Rhyacophila* sp. (Trichoptera: Rhyacophilidae) at WLK030 and one individual *Paraleptophlebia* sp. (Ephemeroptera: Leptophlebiidae) at WLK050. Several uncommon beetles, usually associated with lentic habitats, were found at WLK050: *Stictotarsus* sp. (Dytiscidae), *Tropisternus* sp. (Hydrophilidae), and *Cymbiodyta* sp. (Hydrophilidae). Although the taxa richness value at WLK050 (27) was very close to the range observed at intermittent minimally disturbed sites (28-34 taxa), all other metric values from both sites were very different from minimally disturbed conditions, suggesting that moderate anthropogenic impacts have altered the benthic communities at these sites.

Upper Arroyo Sausal group

The two sites in the Upper Arroyo Sausal (UAS) group (WLK230 and WLK240) were dominated by tolerant taxa, suggesting poor biological integrity:

| | |
|----------------------------------|-----|
| Chironomidae | 63% |
| <i>Simulium</i> sp. (Simuliidae) | 21% |
| <i>Baetis</i> sp. (Baetidae) | 4% |
| Hygrobatidae (Acari) | 4% |
| Oligochaeta | 3%. |

The percentages of common tolerant taxa at WLK230 and WLK240 were 87 percent and 95 percent, respectively. Other taxa found at both sites included *Laccobius* sp. (Coleoptera: Hydrophilidae), *Bezzia/ Palpomyia* (Diptera: Ceratopogonidae), and three intolerant taxa: the mayfly *Serratella* sp. (Ephemerelellidae) and the perlodid stoneflies *Isoperla* sp. and *Kogotus* sp. Several other sensitive taxa were found at one site including *Sweltsa* sp. (Plecoptera: Chloroperlidae), *Malenka* sp. (Plecoptera: Nemouridae), *Agapetus* sp. (Trichoptera: Glossosomatidae), and *Rhyacophila* sp. (Trichoptera: Rhyacophilidae).

While assemblages at WLK230 and WLK240 were numerically dominated by tolerant taxa, both sites contained several highly intolerant taxa, similar to the assemblages in the Keys Creek group. The presence of sensitive taxa indicates that toxic pollutant loads are not the primary cause of altered benthic assemblages. Rather, nutrient enrichment or habitat alteration may exclude taxa with specific habitat or food requirements, resulting in high relative abundances of common tolerant taxa. For example, lack of riparian vegetation and low inputs of leaf litter limits the abundance of shredders, invertebrates that feed on decomposing leaves and wood. Shredders, such as the caddisflies *Gumaga* sp. (Sericostrimatidae) and *Lepidostoma* sp. (Lepidostomatidae), the nemourid stonefly *Malenka* sp., and craneflies (Tipulidae), were abundant at sites in the Walker Creek/Salmon Creek group (WS) but were virtually absent from sites in the Keys Creek (KC) and Upper Arroyo Sausal (UAS) groups.

Chileno Creek sites

Sites on Chileno Creek, the largest tributary of Walker Creek, had very dissimilar benthic assemblages. The most upstream site on Chileno Creek, WLK130, is located just downstream of shallow, eutrophic Laguna Lake, and is surrounded by grazing land. The benthic assemblage was completely dominated by four taxa: Chironomidae, 58 percent; *Hyaella* sp. (Amphipoda:

Hyaellidae), 30 percent; *Simulium* sp., 7 percent; and Oligochaeta, 5 percent. Only two other taxa were found: three individual *Baetis* sp. mayflies and one individual planorbid snail. The low taxa richness (just six taxa) and taxonomic composition of WLK130 is similar to the most highly disturbed sites in urban streams in the Bay Area (such as WIL050, Wildcat Creek; and PER020, Permanente Creek).

The site downstream of WLK130 on Chileno Creek, WLK120, has undergone physical habitat restoration by the landowners. Cows have been excluded from the stream and riparian vegetation has been restored, resulting in much improved habitat conditions that are reflected in the benthic invertebrate assemblage. Biological metrics are substantially better at WLK120 than at WLK130, such as taxa richness (25 taxa vs. 6), percent EPT (37 percent vs. 0), and tolerance value (5.1 vs. 6.5). The percentage of common tolerant taxa dropped from 70 percent (100 percent if *Hyaella* sp. is included) at WLK130 to 43 percent (67 percent if *Hyaella* sp. is included) at WLK120. Numerically dominant taxa at WLK120 included:

| | |
|---|-----|
| Chironomidae | 24% |
| <i>Hyaella</i> sp. (Hyaellidae) | 24% |
| <i>Isoperla</i> sp. (Perlodidae) | 16% |
| <i>Ephemerella</i> sp. (Ephemerellidae) | 12% |
| Oligochaeta | 10% |
| <i>Baetis</i> sp. (Baetidae) | 9% |

In addition to the pollution intolerant taxa *Isoperla* sp. and *Ephemerella* sp., other sensitive taxa collected at WLK120 but not found at WLK130 include the caddisflies *Gumaga* sp. (Sericostomatidae) and *Lepidostoma* sp. (Lepidostomatidae) and the mayfly *Paraleptophlebia* sp. (Leptophlebiidae). Numerous molluscs were found at WLK120, including the bivalve Sphaeriidae as well as physid (*Physa/Physella*) and planorbid (*Gyraulus* sp. and *Menetus* sp.) snails.

WLK100 is located further downstream on Chileno Creek, just above the confluence with Walker Creek. The benthic assemblage at WLK100 was also dominated by a mix of tolerant and intolerant taxa:

| | |
|---|-----|
| <i>Simulium</i> sp. (<i>Simulium</i> sp.) | 37% |
| <i>Baetis</i> sp. (Baetidae) | 29% |
| Chironomide | 15% |
| <i>Serratella</i> sp. (Ephemerellidae) | 4% |
| <i>Malenka</i> sp. (Nemouridae) | 4% |
| Oligochaeta | 3% |
| <i>Wormaldia</i> sp. (Philopotamidae) | 3% |
| <i>Paraleptophlebia</i> sp. (Leptophlebiidae) | 2% |

Other sensitive taxa found at WLK100 include the riffle beetle *Optioservus* sp.; the mayflies *Dipheter hageni*, *Ephemerella* sp., and *Ironodes* sp.; the stoneflies *Suwallia* sp., *Isoperla* sp., and *Kogotus* sp.; and the caddisflies *Hydroptila* sp., *Rhyacophila* sp., and *Gumaga* sp. Biological metrics such as taxa richness, percent EPT, and tolerance value were similar between WLK100 and WLK120, while there were far more EPT taxa at WLK100 (14) than at WLK120 (6). Metric values were slightly outside of the range of values observed at minimally disturbed intermittent sites, suggesting that biological integrity does not fully recover from upstream disturbances in the Chileno Creek watershed.

Arroyo Sausal below Soulajule Reservoir

The benthic assemblage at WLK200 included many intolerant taxa, suggesting that biological integrity in Arroyo Sausal was higher below the reservoir than at upstream sites (WLK230 and WLK240).

Dominant taxa at WLK200 included a mix of tolerant and intolerant taxa:

| | |
|---|-----|
| Chironomidae | 33% |
| <i>Malenka</i> sp. (Nemouridae) | 12% |
| <i>Baetis</i> sp. (Baetidae) | 11% |
| <i>Dipheter hageni</i> (Baetidae) | 6% |
| Planorbidae | 6% |
| <i>Menetus</i> sp. (Planorbidae) | 5% |
| <i>Neophylax</i> sp. (Uenoidae) | 5% |
| <i>Gumaga</i> sp. (Sericostomatidae) | 3% |
| <i>Hydropsyche</i> sp. (Hydropsychidae) | 3% |
| Oligochaeta | 3% |

Many sensitive or rare caddisflies were present, including *Amiocentrus aspilus* (Brachycentridae), *Heteroplectron* (Calamoceratidae), *Glossosoma* sp. (Glossosomatidae), *Lepidostoma* sp. (Lepidostomatidae), and *Rhyacophila* sp. (Rhyacophilidae). Despite the relatively high EPT taxa richness (14) and total taxa richness (31), no beetles were found at WLK200. Releases from Soulajule Reservoir are believed to result in perennial flow downstream in Arroyo Sausal. Several taxa that require perennial flow are present, so it is unclear what environmental factors are causing the absence of Coleoptera.

Biological metric values at WLK200 are generally within the lower range of values from minimally disturbed sites, except percent EPT is lower than the range observed at intermittent streams and taxa richness is lower than perennial streams. This indicates that Arroyo Sausal below Soulajule Reservoir has fairly high biological integrity, but the invertebrate assemblage shows some evidence of anthropogenic disturbance.

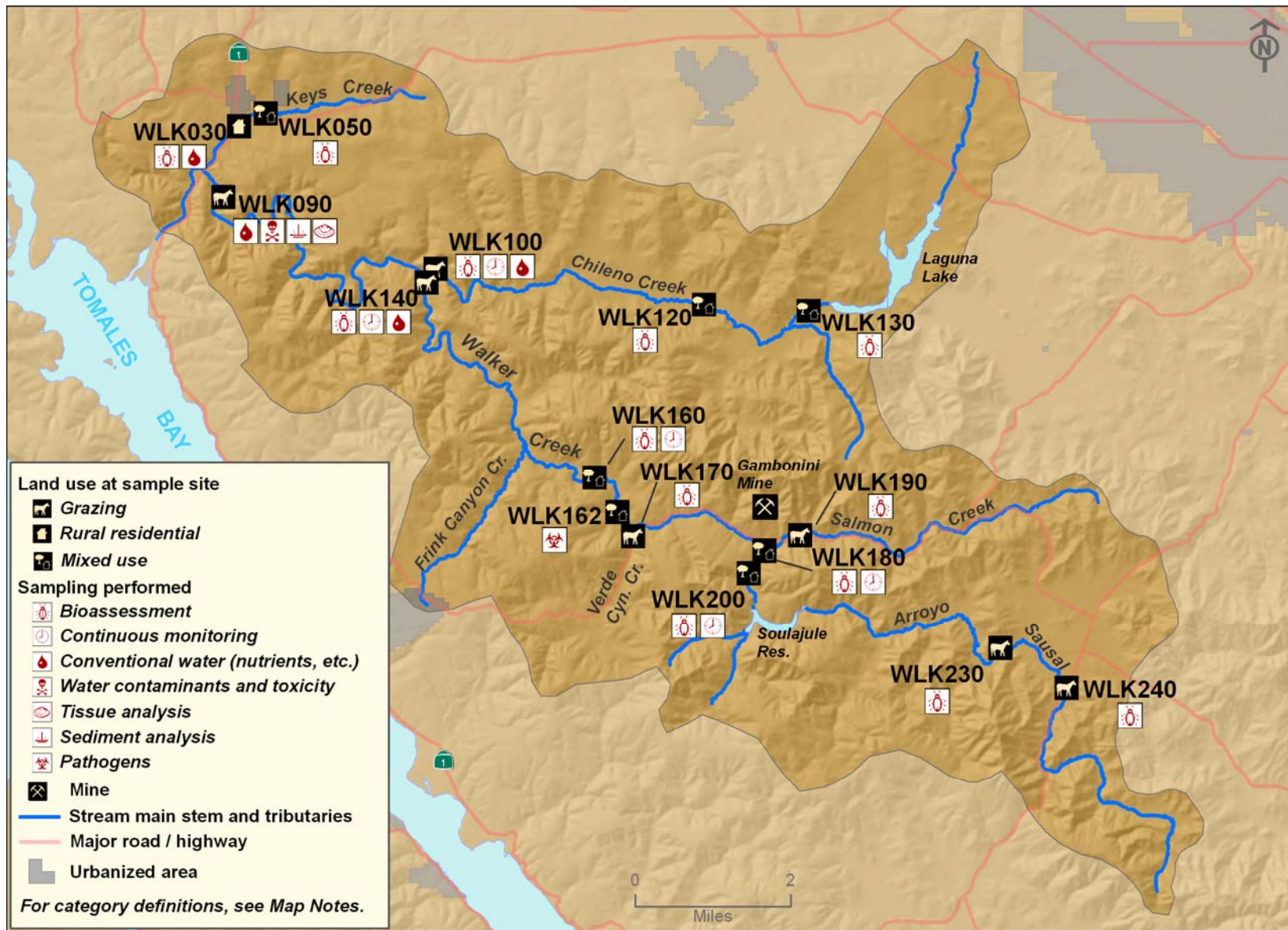


Figure 7-1. Map of Walker Creek benthic macroinvertebrate index scores

Table 7-2. Bioassessment sites in the Walker Creek watershed

| Site ID | Site Name | Stream |
|---------|--------------------|--------------------|
| WLK030 | Keys at Tomales | Keys Creek |
| WLK050 | Keys at Irvin Road | Keys Creek |
| WLK090 | Walker Creek | Walker Creek |
| WLK100 | Chileno Canyon | Chileno Creek |
| WLK120 | Chileno Valley | Chileno Creek |
| WLK130 | Laguna Lake | Chileno Creek |
| WLK140 | Walker Canyon | Walker Creek |
| WLK160 | Walker Creek Ranch | Walker Creek |
| WLK170 | Verde Canyon | Verde Canyon Creek |
| WLK180 | Gambonini Mine | Salmon Creek |
| WLK190 | Salmon Creek | Salmon Creek |
| WLK200 | Soulajule | Arroyo Sausal |
| WLK230 | Arroyo Sausal | Arroyo Sausal |
| WLK240 | Cheese Factory | Arroyo Sausal |

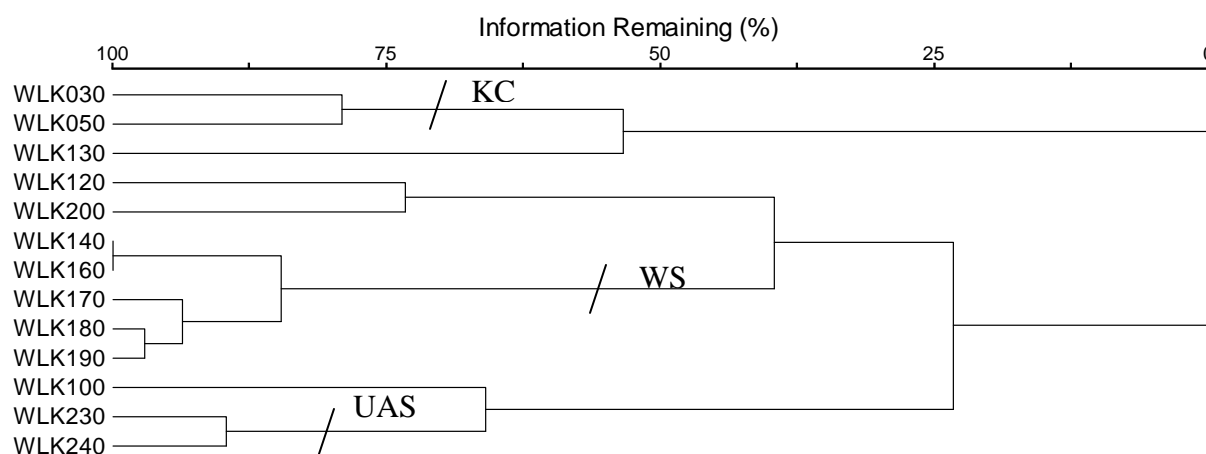


Figure 7-2. Cluster dendrogram of sites in the Walker Creek watershed

This cluster dendrogram groups sites based on similarity of macroinvertebrate taxa present in the Walker Creek watershed. Sites that are closely joined together have many similar taxa present, while groups with longer chain lengths are more dissimilar. The Information Remaining axis at the top signifies how similar the sites are with respect to taxa presence. Abbreviations: KC, Keys Creek; WS, Walker Creek/Salmon Creek; UAS, Upper Arroyo Sausal.

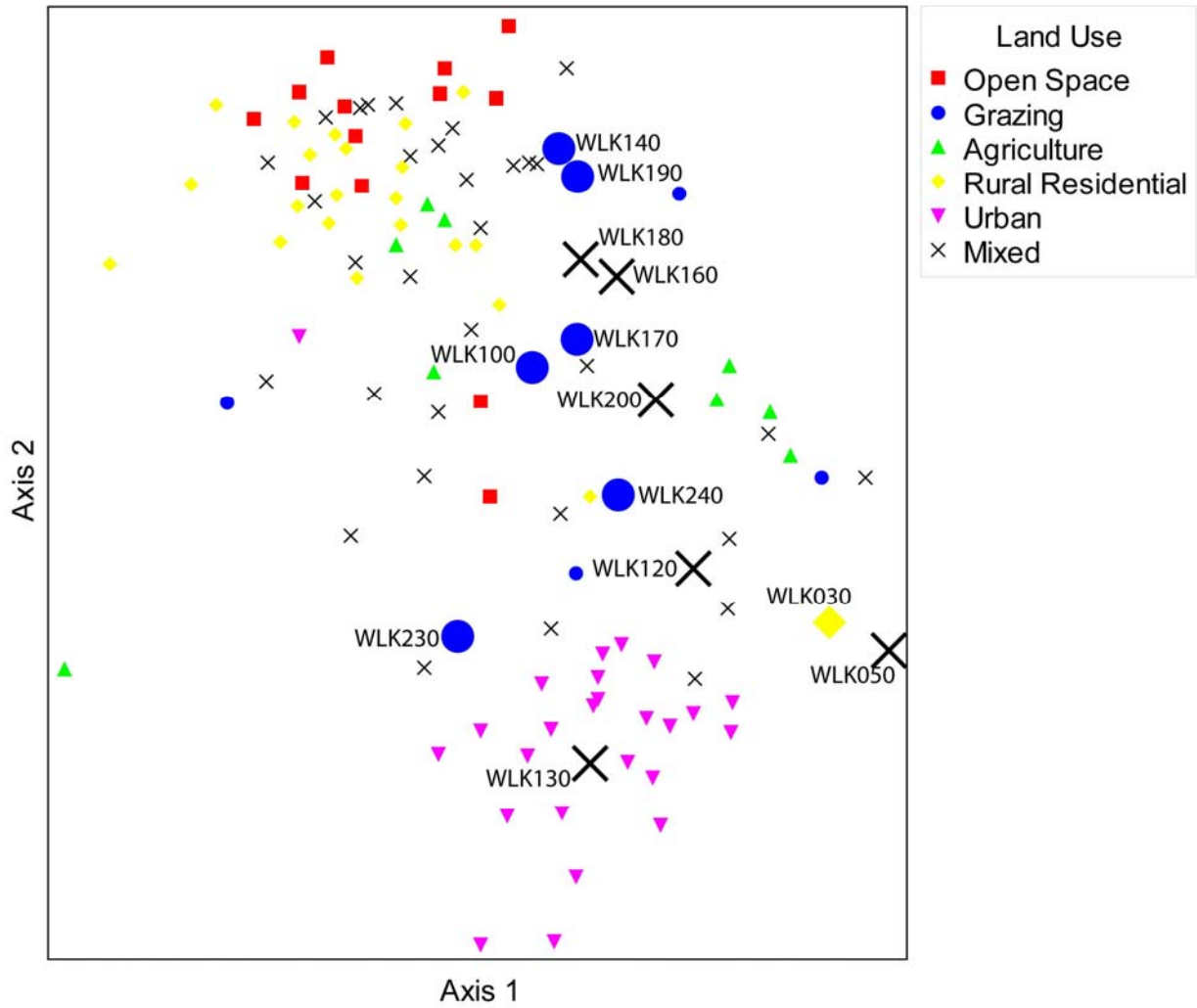


Figure 7-3. NMS ordination of sites in the Walker Creek watershed

The NMS ordination of sites by macroinvertebrate taxa abundance. Sites in the Walker Creek watershed are labeled and symbols are enlarged. For an explanation of ordination graphs, refer to Section 6.1.

Table 7-3. Biological metrics from Walker Creek watershed

| | WLK 030 | WLK 050 | WLK 100 | WLK 120 | WLK 130 | WLK 140 | WLK 160 | WLK 170 | WLK 180 | WLK 190 | WLK 200 | WLK 230 | WLK 240 |
|----------------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|
| Coleoptera Taxa | 1 | 9 | 2 | 1 | 0 | 6 | 4 | 3 | 3 | 4 | 0 | 2 | 1 |
| Diptera Taxa | 4 | 4 | 4 | 5 | 2 | 5 | 3 | 5 | 4 | 7 | 7 | 4 | 4 |
| Ephemeroptera Taxa | 1 | 2 | 6 | 3 | 1 | 7 | 5 | 5 | 6 | 5 | 3 | 2 | 3 |
| Plecoptera Taxa | 0 | 0 | 4 | 1 | 0 | 5 | 3 | 4 | 4 | 4 | 2 | 3 | 3 |
| Trichoptera Taxa | 2 | 1 | 4 | 2 | 0 | 12 | 12 | 9 | 10 | 9 | 9 | 1 | 1 |
| Non-Insect Taxa | 7 | 10 | 2 | 13 | 3 | 6 | 9 | 10 | 12 | 7 | 10 | 4 | 3 |
| EPT Taxa | 3 | 3 | 14 | 6 | 1 | 24 | 20 | 18 | 20 | 18 | 14 | 6 | 7 |
| Taxa Richness | 15 | 27 | 22 | 25 | 6 | 41 | 36 | 36 | 39 | 36 | 31 | 16 | 15 |
| % EPT | 5 | 23 | 45 | 37 | 0 | 70 | 23 | 52 | 74 | 59 | 47 | 4 | 9 |
| % Sensitive EPT | 0 | 1 | 13 | 28 | 0 | 47 | 13 | 26 | 54 | 53 | 32 | 1 | 2 |
| % Chironomidae | 44 | 31 | 15 | 24 | 58 | 9 | 49 | 33 | 9 | 18 | 32 | 63 | 64 |
| % Coleoptera | 1 | 3 | 0 | 0 | 0 | 16 | 5 | 0 | 3 | 7 | 0 | 1 | 0 |
| % Oligochaeta | 21 | 17 | 3 | 10 | 5 | 0 | 0 | 4 | 1 | 0 | 3 | 4 | 3 |
| % Non-insect | 25 | 35 | 3 | 37 | 35 | 2 | 2 | 9 | 7 | 10 | 18 | 12 | 3 |
| % Intolerant | 0 | 0 | 10 | 28 | 0 | 25 | 7 | 21 | 31 | 24 | 17 | 1 | 2 |
| % Tolerant | 4 | 18 | 0 | 26 | 30 | 1 | 1 | 3 | 2 | 2 | 4 | 7 | 0 |
| Tolerance Value | 5.8 | 5.9 | 5.1 | 5.1 | 6.5 | 3.7 | 5.2 | 4.7 | 3.6 | 3.6 | 4.8 | 6.0 | 5.8 |
| % Predator | 3 | 4 | 2 | 18 | 0 | 14 | 4 | 14 | 4 | 25 | 5 | 11 | 2 |
| % Collector-filterer | 26 | 8 | 40 | 1 | 7 | 13 | 21 | 2 | 11 | 1 | 6 | 18 | 23 |
| %Collector-gatherer | 71 | 86 | 53 | 79 | 93 | 41 | 58 | 65 | 37 | 26 | 58 | 70 | 74 |
| % Scraper | 0 | 0 | 1 | 2 | 0 | 25 | 7 | 2 | 15 | 13 | 14 | 1 | 1 |
| % Shredder | 0 | 0 | 5 | 0 | 0 | 8 | 6 | 17 | 33 | 35 | 17 | 0 | 0 |
| % Other | 0 | 2 | 0 | 0 | 0 | 0 | 3 | 1 | 0 | 0 | 0 | 1 | 0 |
| % Common Tolerant | 94 | 78 | 83 | 43 | 70 | 32 | 77 | 64 | 33 | 23 | 50 | 87 | 95 |
| Common Intol. Taxa | 1 | 1 | 7 | 2 | 0 | 12 | 7 | 7 | 9 | 11 | 6 | 3 | 3 |

7.1.4 Basic Water Quality Field Measurements

Continuously measured variables were within acceptable ranges (Appendix D; Figure 7-7 and Figure 7-8), with three exceptions:

- Late spring temperatures (May 23-30, 2002) measured at the Walker Creek Ranch (WLK160) and Gambonini Mine (WLK180) sites produced Highest Weekly Average Temperature (HWAT) values above 14.8°C, the threshold for coho salmon. Temperatures measured in the watershed during the summer (August 16-21, 2002) did not exceed threshold values (Appendix D; and Figure 7-5).
- The Gambonini Mine site (WLK180) had low DO values during the late spring and summer deployments (Figure 7-6).
- The maximum pH value from the Walker Creek Ranch site in late spring was 9.07, well above the Basin Plan maximum for all uses.

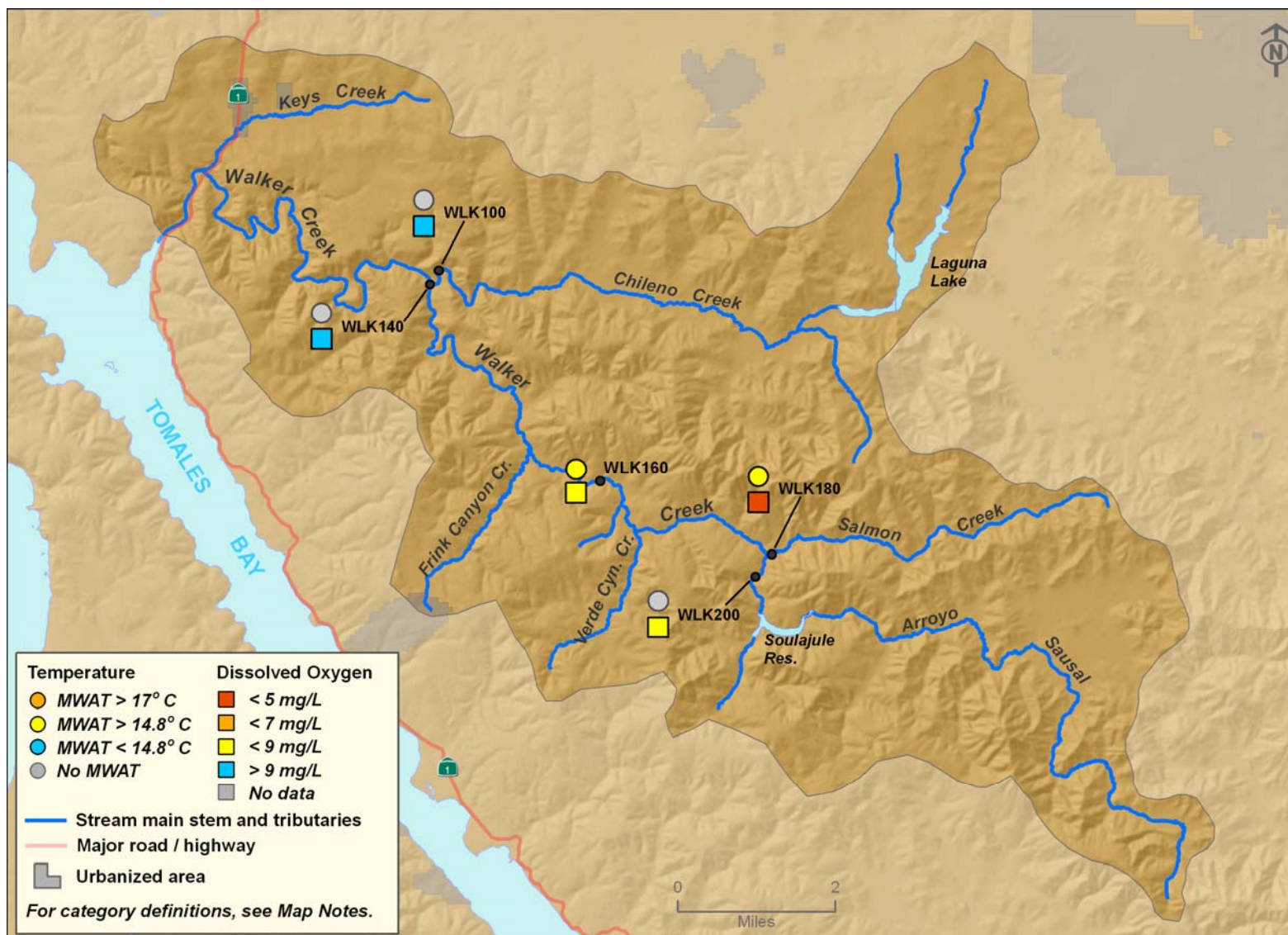


Figure 7-4. Map of temperature and DO levels in Walker Creek watershed

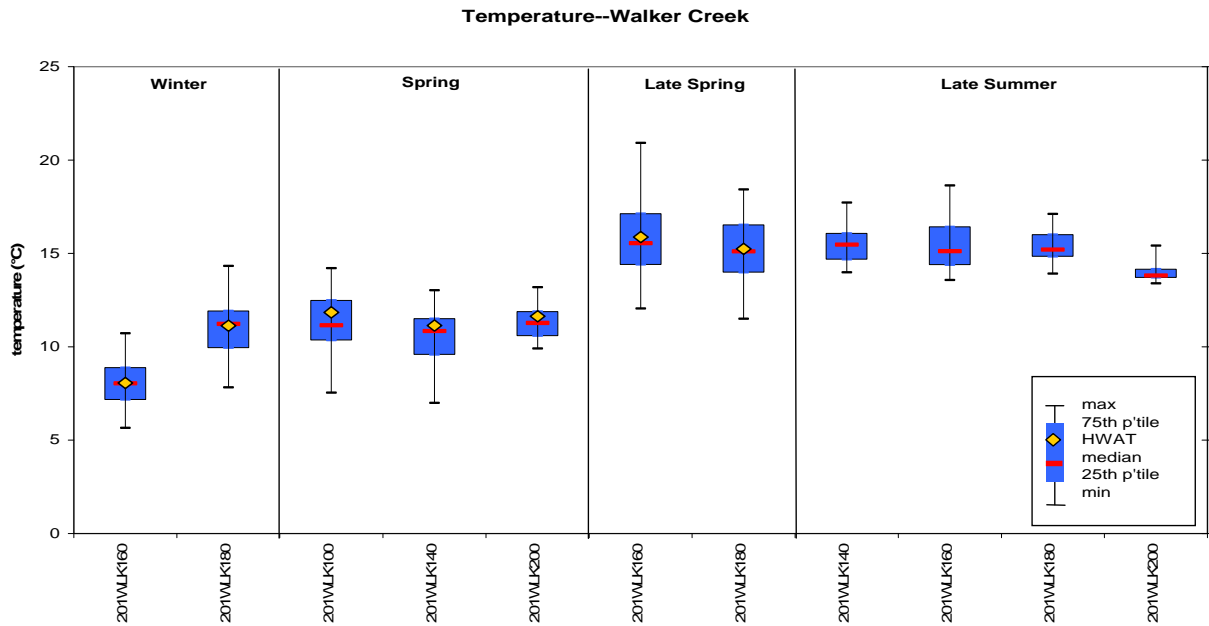


Figure 7-5. Temperature monitoring in Walker Creek watershed

The sites monitored in late summer were sampled for less than a week, so do not have HWAT values.

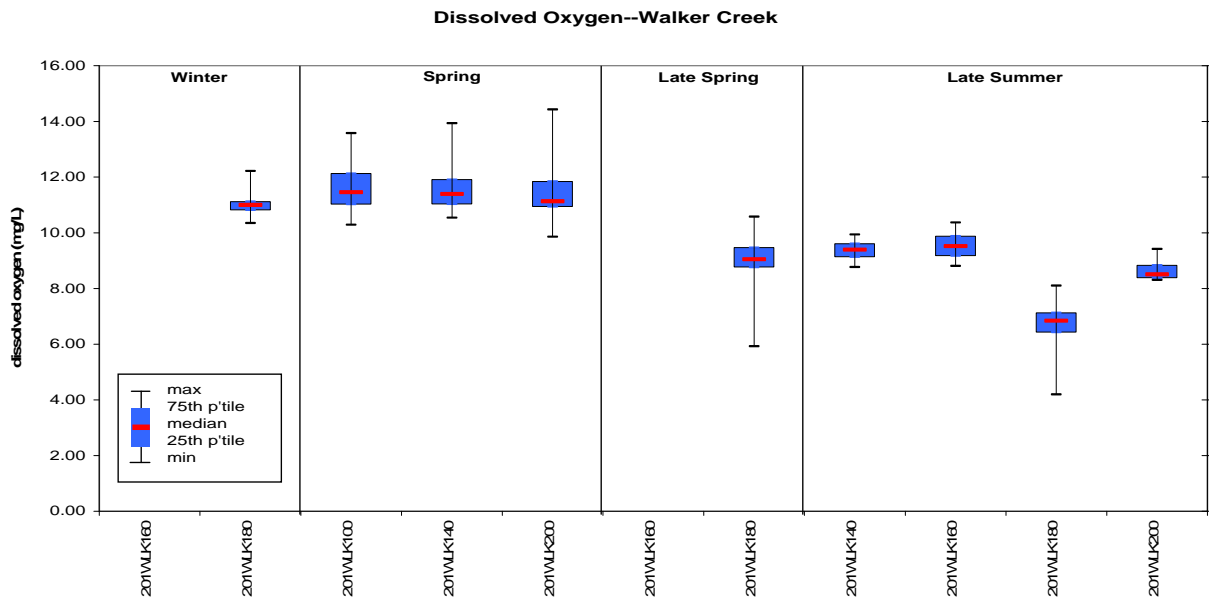


Figure 7-6. DO monitoring in Walker Creek watershed

Data were rejected for WLK160 in winter and late spring.

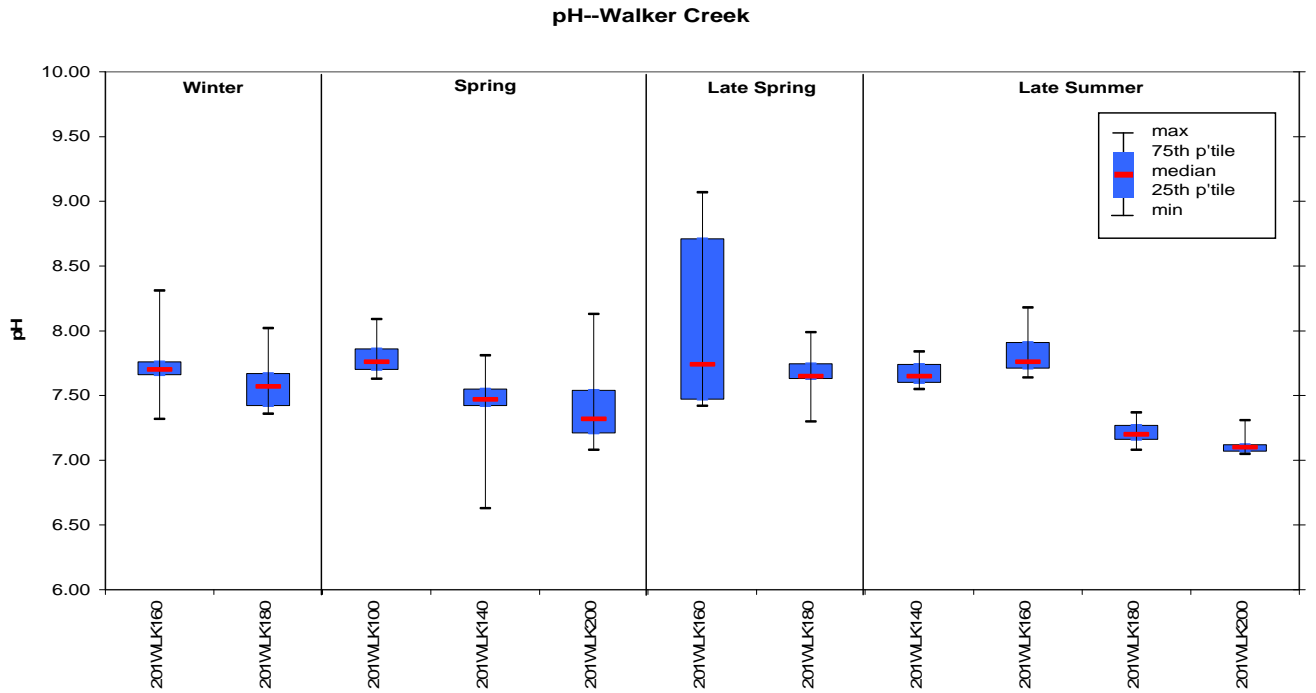


Figure 7-7. pH monitoring in Walker Creek watershed

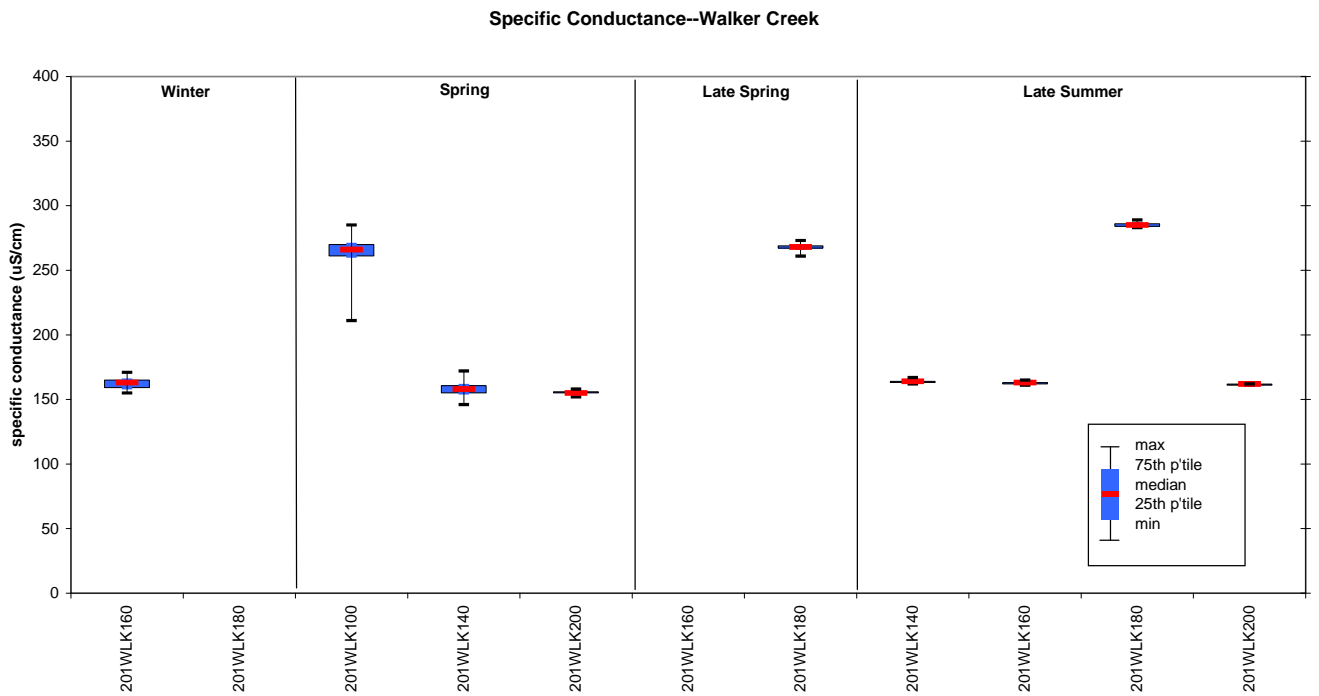


Figure 7-8. Specific conductance monitoring in Walker Creek watershed

Data were rejected for WLK180 in the winter and WLK160 in the late spring.

7.1.5 Water, Sediment, and Clam Tissue Chemistry

Nitrate concentrations in the water column were generally fairly low. While no dry season concentrations exceeded the U.S. EPA ecoregional reference guideline of 0.155 mg /L, most wet season concentrations exceeded the guideline. The highest nitrate concentration for this watershed was the wet season sample from Chileno Canyon (2.18 mg/L).

The highest orthophosphate concentration measured was 0.18 mg/L, and orthophosphorus values exceeded the U.S. EPA total phosphorus guideline for aquatic life of 0.030 mg/L at all sites in at least one season. Total phosphorus values exceeded the U.S. EPA guideline at most sites, with a maximum of 0.2 mg/L (Appendix G).

No samples were analyzed for water chemistry in this watershed. Sediment was collected from the Walker Creek site (WLK090) in October 2001. Three geologically abundant trace metals were measured at concentrations above guideline values: chromium and nickel (above the probable effects concentration), and mercury (above the threshold effects concentration; Appendix H). Although elevated levels of nickel and chromium are common due to serpentine deposits in the area, elevated mercury concentrations reflect mineral mobilization from past mining activities.

Clams deployed at the Walker Creek site yielded tissue concentrations similar to control values for all detected analytes (Figures 6-21 through 6-27). Even mercury concentrations were elevated only slightly relative to controls and were lower than in many other watersheds. This is of interest because of the historic mercury mining upstream and the concern for fish contamination in Tomales Bay receiving waters. It has been shown, however, that clams do not concentrate high levels of mercury even in Tomales Bay (Gunther 1999; Gassel *et al.* 2004).

7.1.6 Water and Sediment Toxicity

No water samples from the Walker Creek watershed were analyzed for toxicity. The sediment sample collected at WLK 090 was not toxic to test amphipods. For the graph of toxicity tests in Walker Creek, see Lagunitas Creek section 7.2.4 (Figure 7-17).

7.1.7 Coliform Bacteria

The single sampling site at Turtle Pond (WLK 162) is isolated from Walker Creek proper, and is not representative of overall conditions in the watershed. The Tomales Bay Pathogen TMDL reported numerous exceedances in Walker Creek during the rainy season (SFBRWQCB 2005). Turtle Pond is a designated swimming beach used heavily for water-contact recreation during camping season, which took place before—not during—the period sampled. Samples from this site did not exceed any coliform Basin Plan objectives for recreational use (Figures 6-28 through 6-30).

7.2 Lagunitas Creek Watershed

7.2.1 Sites of Concern

Exceedances noted in the summary table (Table 7-4) are discussed in the relevant sub-section for this watershed. A weight-of-evidence approach was used to evaluate sites of concern. Individual exceedances of parameters, particularly those without regulatory objectives, are considered in context. In general, nickel and chromium levels are geologically elevated in the Bay Area, particularly where there are serpentine deposits. Nutrient levels (nitrate and phosphorus) are above U.S. EPA guidelines at most sites, but may still be relatively low.

- Upper San Geronimo Creek sites had poorer benthic invertebrate metric scores than sites downstream.
- The Creamery Gulch site (LAG 270) had elevated water PAHs and metals, although none exceeded thresholds. Sediments from this site had the highest Sediment Quality Guideline Quotient found anywhere in the region—mainly due to metals, some of which were likely of geologic origin.
- Clam tissues at Creamery Gulch (LAG270), Olema Creek (LAG040), and Gallagher’s Ranch (LAG130) had the highest mercury concentrations measured in this study.
- Lagunitas Creek at Hwy 1 (Green Bridge) had the highest bacterial counts in the watershed.
- Nicasio (LAG160) had significantly degraded biological integrity in its benthic community. Continuous monitoring results showed unusual timing patterns and DO levels below survivability for most organisms during the spring (not a typically stressful season even for intermittent streams).

Table 7-4. Summary of sites with exceedances in Lagunitas Creek watershed

| <i>Analyte</i> | <i>Standard</i> | <i># of sites</i> |
|--|------------------|-------------------|
| FIELD MEASURES of 6 sites sampled | | |
| Temperature | MWAT, coho | 5 |
| | minimum, COLD | 2 |
| Oxygen, dissolved | minimum, WARM | 1 |
| | 3-month median | 4 |
| CONVENTIONAL WATER QUALITY of 8 sites sampled | | |
| Nitrate as N | maximum | 8 |
| Phosphorus, total as P | maximum | 3 |
| SEDIMENT CHEMISTRY of 2 sites sampled | | |
| <i>Metals</i> | | |
| Arsenic | TEC | 1 |
| Chromium | TEC | 2 |
| Copper | TEC | 1 |
| Mercury | TEC | 1 |
| Nickel | PEC | 2 |
| COLIFORMS of 5 sites sampled | | |
| <i>E. coli</i> | steady state | 1 |
| | designated beach | 1 |
| Fecal coliform | log mean | 1 |
| | 90th percentile | 3 |
| Total coliform | median | 2 |

7.2.2 Water Quality in Relation to Land Use

None of the sampling sites in the watershed drain land areas characterized as urban; most sites for which multiple indicators were measured were mixed use or rural residential. There was generally insufficient data to analyze trends in any of the parameters relative to land use, with the exception of the bioassessment data discussed in Section 6.1.1. The presence of elevated levels of PAHs and copper, usually associated with automobile traffic (combustion and brake pads) is unexpected in these rural residential areas. The Creamery Gulch site (LAG270) may have been influenced by the upstream treatment plant as well as local sources.

7.2.3 Macroinvertebrate Assemblages and Physical Habitat

Benthic assemblages in Lagunitas Creek generally indicate excellent water quality conditions, but conditions in several tributaries are less than optimal. It should be noted that benthic invertebrates are valuable indicators of gross levels of pollution, but they respond to physical habitat conditions at different spatial scales than fish. Habitat conditions required by populations of coho salmon and steelhead in Lagunitas Creek, such as deep pools and woody debris, cannot be evaluated using benthic invertebrate assemblages. Although invertebrate assemblages suggest that water quality conditions in Lagunitas Creek are generally excellent, habitat conditions for salmonids or other wildlife could be less than optimal.

From the cluster dendrogram (Figure 7-10) and NMS ordination (Figure 7-11) of taxa presence, five groups of sites are apparent:

- The mainstem of Lagunitas Creek, including the Cheda Creek and Devils Gulch tributaries (MSL)
- The San Geronimo Creek watershed, including Woodacre Creek (SG)
- Tributaries in the upper Lagunitas Creek watershed (ULT)
- Nicasio Creek and Halleck Creek (NH).
- Olema Creek (OL)

Mainstem Lagunitas Creek

Sites on the mainstem of Lagunitas Creek (LAG130, LAG165, LAG170, LAG210, LAG220, LAG320), from near the mouth above the town of Point Reyes Station (LAG130) to downstream of Shafter Dam (LAG320), exhibited similar benthic assemblages. All six sites were closely clustered on the NMS ordination graph (Figure 7-11) and the cluster dendrogram (Figure 7-10).

Biological metric values were very similar among the six sites (Table 7-4), and generally fall between the minimum and median values observed at perennial minimally disturbed sites (Table 7-4). This is somewhat surprising, given the significant differences in physical habitat conditions between downstream and upstream sites: sites near the mouth (LAG130, LAG165) have high amounts of fine sediment (40-50%), while sites further upstream have little fine sediment. Biological assemblages suggest that water quality conditions in the mainstem of Lagunitas Creek are generally good. Benthic assemblages from the mainstem of Lagunitas Creek closely resemble

minimally disturbed conditions, and, with San Gregorio Creek, represent the best biological conditions yet sampled from large streams in the Bay Area. Less than 50 percent of organisms at all six sites belonged to common tolerant taxa (Table 7-4), which is within the range of conditions at sites draining open space (Table 6-12).

Taken together, these six sites were numerically dominated by the following taxa:

| | |
|----------------------------------|-----|
| Chironomid midges | 20% |
| <i>Optioservus</i> sp. (Elmidae) | 11% |
| <i>Simulium</i> sp. (Simuliidae) | 8% |
| <i>Malenka</i> sp. (Nemouridae) | 8% |
| <i>Baetis</i> sp. | 5% |

Many common intolerant taxa were present at all six sites including *Rhyacophila* sp., *Sweltsa* sp., *Zaitzevia* sp., *Neophylax* sp., and *Calinueria californica*. Also notable was collection of a single endangered California Freshwater Shrimp, *Syncaris pacifica* (Decapoda), from LAG220. The California Freshwater Shrimp, restricted to streams in Marin, Sonoma, and Napa counties, was not expected in benthic samples from riffles because it prefers pools and bank-side habitat.

The Devils Gulch (LAG190) and Cheda Creek (LAG180) tributaries were taxonomically similar to sites on the Lagunitas Creek mainstem. On the cluster dendrogram (Figure 7-10), these two sites are most closely related to a mainstem Lagunitas Creek site (LAG165) and are grouped together with the mainstem sites (MSL, Mainstem Lagunitas Creek and tributaries). Taxa richness at LAG180 and LAG190 was 45 and 47 taxa, respectively, the highest values in the Lagunitas Creek watershed (Table 7-4). Cheda Creek had the greatest number of Coleoptera taxa (8) and non-insect taxa (10) in the entire watershed, as a result of the presence of many genera of riffle beetles (Elmidae) and water mites (Acari). The greatest number of mayfly (10) and caddisfly (10) taxa in the watershed were found at Devils Gulch, including less common taxa such as *Ecclisomyia* sp. (Trichoptera: Limnephilidae) and *Cinygma* (Ephemeroptera: Heptageniidae). Low percentages of common tolerant organisms and high numbers of common intolerant taxa (Table 7-6) indicate that Devils Gulch and Cheda Creek exhibit excellent biological integrity.

San Geronimo Creek watershed

Sites in the San Geronimo Creek watershed (LAG240, LAG270, LAG290, LAG300) clustered together both on the NMS ordination graph (Figure 6-33) and on the cluster dendrogram (Figure 7-11). This is the most densely developed portion of the Lagunitas Creek basin. While benthic assemblages at these sites exhibited some signs of reduced biological integrity, conditions were significantly better than in other urbanized watersheds in the Bay Area.

Unlike many other watersheds in the Bay Area, where biological conditions generally decline from upstream to downstream, San Geronimo Creek showed recovery in the downstream direction. The farthest upstream site on San Geronimo Creek, at the water treatment plant (LAG290), appears the most degraded, with low taxa richness and low EPT composition metrics relative to the rest of the watershed (Figure 7-11). Woodacre Creek (LAG300), a tributary to San

Geronimo Creek, had similarly low EPT composition metrics, but taxa richness values were within the range of minimally disturbed conditions (Figure 7-11). Assemblages at both sites were numerically dominated by chironomid midges. Functional differences are also evident, including low representation of predator, scraper, and shredder feeding groups at LAG290 and LAG300.

Richness, composition, tolerance, and functional metrics at the downstream sites on San Geronimo Creek (LAG240 and LAG270) more closely resemble values from sites on the mainstem of Lagunitas Creek (Table 7-4). Taxonomically, LAG240 and LAG270 are notable for high abundances of the crane fly *Linophila* sp. (Diptera: Tipulidae) and the presence of the stonefly nymph *Paraperla* sp. (Plecoptera: Chloroperlidae), which has not been found elsewhere in the Bay Area.

Tributaries in the upper Lagunitas Creek watershed

Big Carson Creek (LAG330 and LAG335) and Little Carson Creek (LAG380) are small tributaries that flow into Kent Lake, the lowest of four reservoirs in the upper Lagunitas Creek watershed. Cataract Creek (LAG390) is a steep tributary that empties into Alpine Lake, the next highest reservoir. These four sites exhibited metric values within the range of minimally disturbed conditions, and very high numbers of common intolerant taxa (17-20). More Plecoptera taxa, including the rare intolerant taxa *Despaxia augusta* (Leuctridae) and *Soliperla* sp. (Peltoperlidae), were found at Cataract Creek than at other sites in the watershed. The greatest number of Diptera taxa in the watershed was found at LAG335, including the uncommon blepharicerid *Agathon* sp. and the tipulid *Erioptera* sp.

Nicasio Creek and Halleck Creek

Benthic assemblages in Nicasio Creek (LAG160) and Halleck Creek (LAG150) were taxonomically distinct from other sites in the watershed, as shown in the ordination graph (Figure 7-11) and cluster dendrogram (Figure 7-10). The dissimilarity may be due in part to the fact that both streams usually go dry during the summer.

Biological metrics from Halleck Creek (Table 7-6) are very similar to mean metric values from minimally disturbed intermittent streams (Table 6-10). Taxa with life cycles greater than one-year were virtually absent from Halleck Creek, with the possible exception of the chloroperlid stonefly *Suwallia* sp. Additionally, only three individual beetles were found (two larvae and one adult of *Optioservus* sp.), suggesting that the periodic disturbance of drying may preclude taxa that require stable habitat conditions for more than one year. There is anecdotal evidence that the channel bed at the Halleck Creek sampling site is annually excavated in an effort to reduce flooding from channel aggradation. This management practice is likely partly responsible for the unusual benthic assemblage at this site.

When compared to conditions at minimally disturbed intermittent streams (Table 6-10) the benthic assemblage of Nicasio Creek appears significantly degraded. The sample was numerically dominated by chironomid midges (80 percent of individuals), and only 8 EPT taxa were present. The benthic assemblage at this site, although better than conditions found in heavily urbanized areas, was very different from other sites in the watershed and indicative of poor water quality.

Olema Creek

Sites on Olema Creek (LAG040, LAG050, LAG060, LAG075, LAG085, LAG100; described as OL1, OL2, OL3, OL4, OL5, OL6 in Lee and Ketcham (2001) are more similar to each other taxonomically than to other sites in the Lagunitas Creek watershed, although they more closely resemble sites in the San Geronimo Creek watershed than other sites in the basin. Metric values were not calculated for Olema Creek sites because (unlike other sites in the Lagunitas Creek watershed) samples were collected from one riffle only. Common tolerant taxa made up 46 to 79 percent of organisms at the six sites, higher than the range of conditions at minimally disturbed sites. Olema Creek sites are located between the San Geronimo Creek sites and Halleck Creek on the ordination graph (Figure 7-11), suggesting a low to moderate level of degradation.

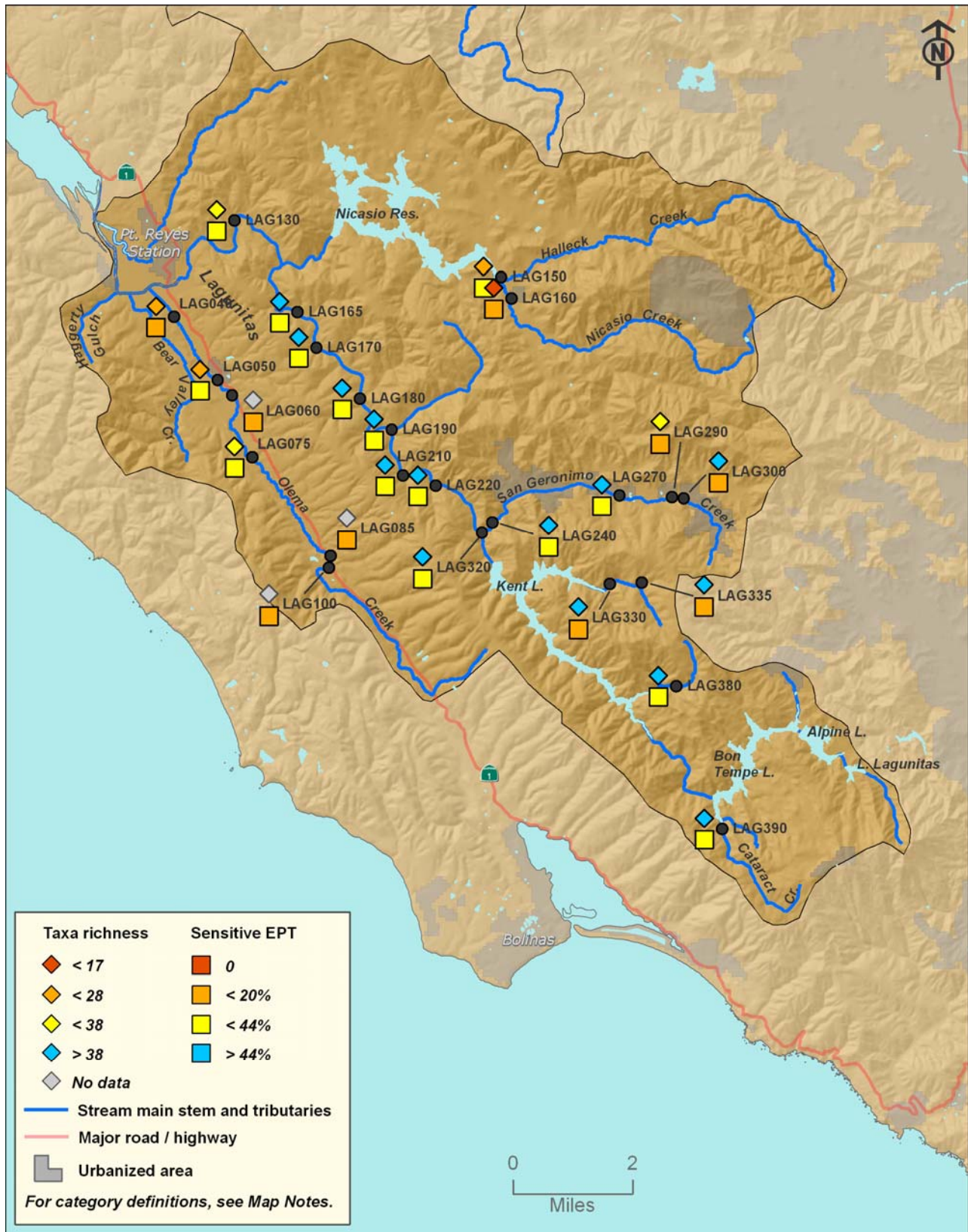


Figure 7-9. Map of Lagunitas Creek benthic macroinvertebrate index scores

Table 7-5. Bioassessment sites in the Lagunitas Creek watershed

| Site Code | Site Name | Stream |
|------------------|-------------------------|----------------------------|
| LAG040 | Olema Low | Olema Creek |
| LAG050 | Bear Valley Road Bridge | Olema Creek |
| LAG060 | Vedanta | Olema Creek |
| LAG075 | Truttman | Olema Creek |
| LAG085 | Five Brooks | Olema Creek |
| LAG100 | Blue Line | Olema Creek |
| LAG130 | Gallagher's Ranch | Lagunitas Creek |
| LAG150 | Halleck | Halleck Creek |
| LAG160 | Nicasio | Nicasio Creek |
| LAG165 | Below Tocaloma | Lagunitas Creek |
| LAG170 | Tocaloma Bridge | Lagunitas Creek |
| LAG180 | Cheda | Cheda Creek |
| LAG190 | Devils Gulch | Devils Gulch |
| LAG210 | Taylor Park | Lagunitas Creek |
| LAG220 | Irving Bridge | Lagunitas Creek |
| LAG240 | White Horse Bridge | San Geronimo Creek |
| LAG270 | Creamery Gulch | San Geronimo Creek |
| LAG290 | Water Treatment Plant | San Geronimo Creek |
| LAG300 | Woodacre Creek | Woodacre Creek |
| LAG320 | Shafter Bridge | Lagunitas Creek |
| LAG330 | Big Carson 1 | Big Carson Creek |
| LAG335 | Big Carson 2 | Big Carson Creek Tributary |
| LAG380 | Little Carson | Little Carson Creek |
| LAG390 | Cataract | Cataract Creek |

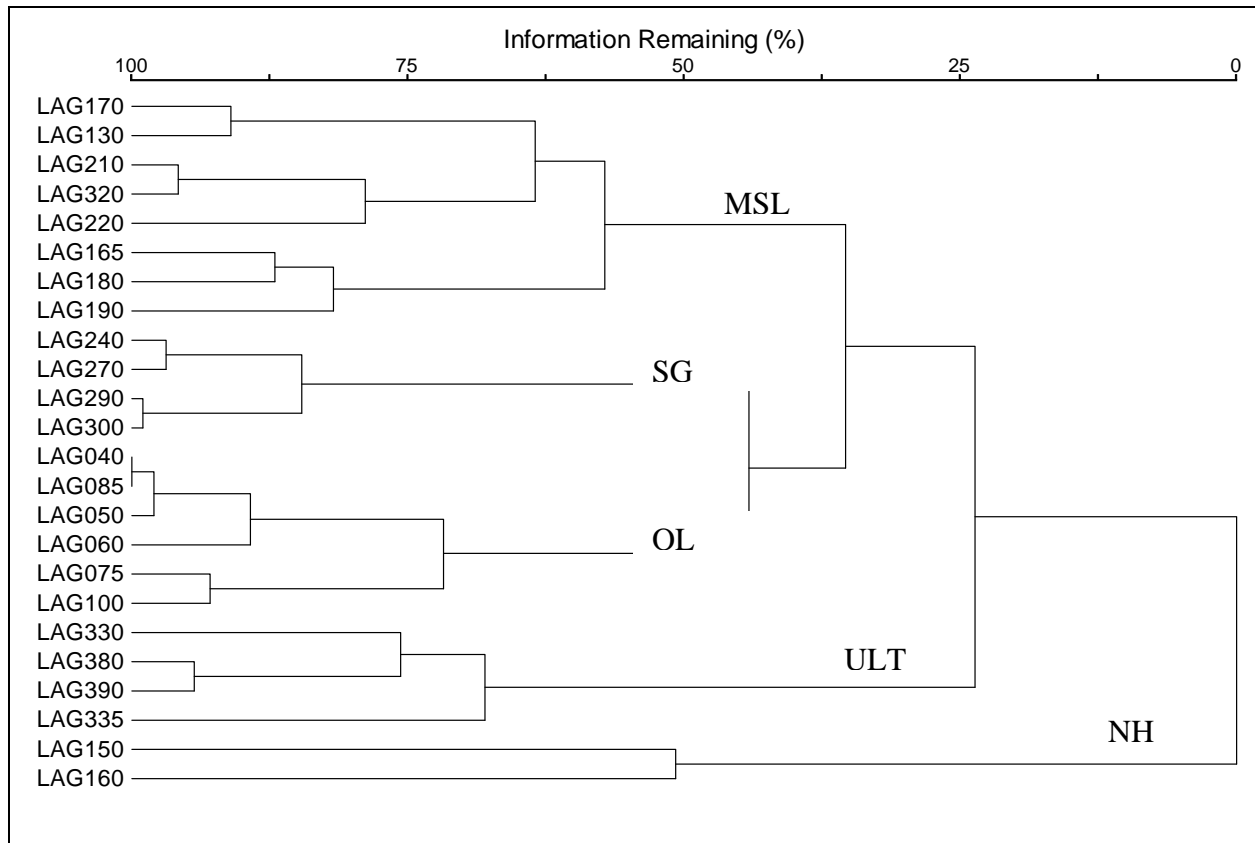


Figure 7-10. Cluster dendrogram of sites in the Lagunitas Creek watershed

This cluster dendrogram groups sites based on similarity of macroinvertebrate taxa present in the Lagunitas Creek watershed. Sites that are closely joined together have many similar taxa present, while groups with longer chain lengths are more dissimilar. The Information Remaining axis at the top signifies how similar the sites are with respect to taxa presence. Abbreviations: MSL, Mainstem Lagunitas Creek and tributaries; SG, San Geronimo Creek watershed; OL, Olema Creek watershed; ULT, Upper Lagunitas Creek tributaries; NH, Nicasio Creek and Halleck Creek.

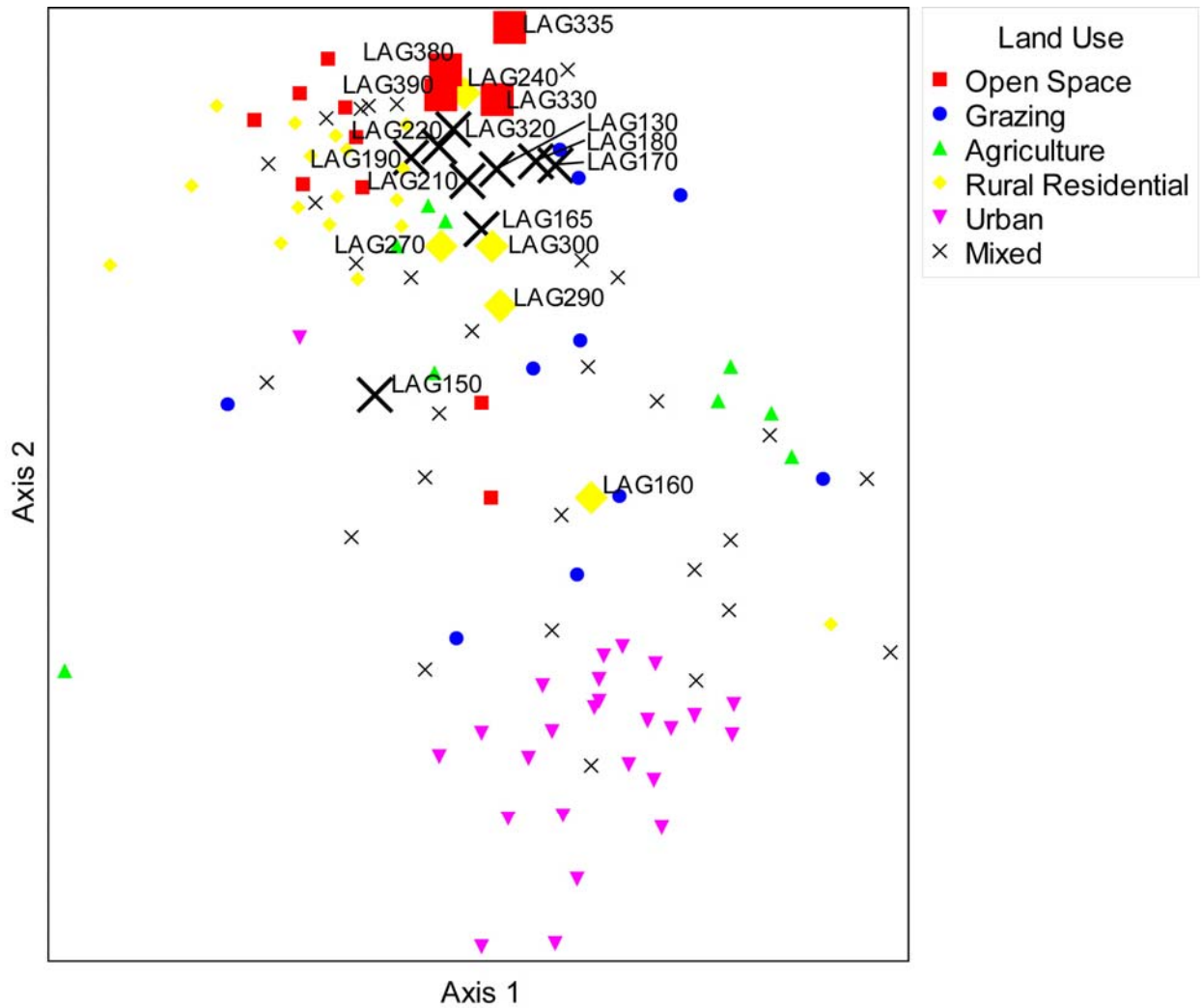


Figure 7-11. NMS ordination of sites in the Lagunitas Creek watershed

The NMS ordination of sites by macroinvertebrate taxa abundance. Sites in the Lagunitas Creek watershed are labeled and symbols are enlarged. For an explanation of ordination graphs, refer to Section 6.1.

Table 7-6. Biological metrics from the Lagunitas Creek watershed

The biological metrics are from benthic macroinvertebrate assemblages in the Lagunitas Creek watershed. Data from Olema Creek are available in a separate report (see Ketcham 2001).

| LAG | 130 | 150 | 160 | 165 | 170 | 180 | 190 | 210 | 220 | 240 | 270 | 290 | 300 | 320 | 330 | 335 | 380 | 390 |
|----------------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|
| Coleoptera Taxa | 6 | 2 | 0 | 8 | 4 | 8 | 7 | 5 | 7 | 6 | 7 | 7 | 6 | 5 | 7 | 4 | 7 | 6 |
| Diptera Taxa | 3 | 4 | 3 | 5 | 7 | 6 | 11 | 8 | 7 | 7 | 7 | 4 | 7 | 9 | 8 | 13 | 10 | 9 |
| Ephemeroptera Taxa | 8 | 8 | 4 | 8 | 8 | 9 | 10 | 8 | 8 | 7 | 7 | 5 | 8 | 8 | 7 | 7 | 6 | 8 |
| Plecoptera Taxa | 5 | 5 | 3 | 6 | 6 | 4 | 5 | 6 | 4 | 6 | 6 | 4 | 5 | 6 | 3 | 2 | 5 | 7 |
| Trichoptera Taxa | 8 | 3 | 1 | 8 | 7 | 8 | 10 | 7 | 7 | 6 | 7 | 4 | 7 | 6 | 6 | 8 | 9 | 8 |
| Non-Insect Taxa | 5 | 4 | 3 | 5 | 6 | 10 | 4 | 5 | 9 | 6 | 5 | 7 | 8 | 7 | 7 | 5 | 7 | 7 |
| EPT Taxa | 21 | 16 | 8 | 22 | 21 | 21 | 25 | 21 | 19 | 19 | 20 | 13 | 20 | 20 | 16 | 17 | 20 | 23 |
| Taxa Richness | 35 | 26 | 14 | 40 | 38 | 45 | 47 | 39 | 42 | 38 | 39 | 31 | 41 | 41 | 38 | 40 | 44 | 45 |
| % EPT | 43 | 58 | 17 | 51 | 34 | 43 | 51 | 47 | 49 | 49 | 41 | 23 | 21 | 60 | 60 | 40 | 43 | 58 |
| % Sensitive EPT | 21 | 44 | 10 | 37 | 21 | 32 | 36 | 24 | 34 | 32 | 28 | 10 | 11 | 38 | 34 | 14 | 30 | 35 |
| % Chironomidae | 16 | 37 | 80 | 20 | 20 | 19 | 30 | 22 | 25 | 17 | 38 | 53 | 58 | 23 | 20 | 25 | 23 | 11 |
| % Coleoptera | 30 | 0 | 0 | 16 | 22 | 28 | 11 | 10 | 18 | 14 | 9 | 10 | 12 | 9 | 7 | 13 | 18 | 5 |
| % Oligochaeta | 0 | 1 | 0 | 4 | 0 | 0 | 2 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 0 |
| % Non-insect | 9 | 3 | 0 | 5 | 2 | 4 | 5 | 3 | 4 | 1 | 2 | 5 | 2 | 5 | 5 | 4 | 7 | 9 |
| % Intolerant | 17 | 44 | 10 | 33 | 18 | 26 | 33 | 25 | 24 | 29 | 25 | 8 | 6 | 41 | 33 | 19 | 23 | 31 |
| % Tolerant | 1 | 1 | 0 | 0 | 1 | 1 | 0 | 1 | 1 | 0 | 0 | 0 | 1 | 3 | 1 | 0 | 1 | 3 |
| Tolerance Value | 4.0 | 3.6 | 5.5 | 3.9 | 4.4 | 4.0 | 3.8 | 4.2 | 3.9 | 3.5 | 4.2 | 5.1 | 5.2 | 3.6 | 3.6 | 4.3 | 3.9 | 3.5 |
| % Predator | 2 | 28 | 9 | 14 | 10 | 10 | 14 | 7 | 12 | 20 | 12 | 5 | 5 | 5 | 16 | 21 | 18 | 16 |
| % Collector-filterer | 8 | 0 | 1 | 10 | 22 | 5 | 5 | 19 | 3 | 1 | 4 | 9 | 4 | 4 | 2 | 0 | 2 | 5 |
| %Collector-gatherer | 35 | 61 | 90 | 41 | 34 | 34 | 56 | 40 | 43 | 43 | 63 | 71 | 73 | 47 | 45 | 43 | 38 | 25 |
| % Scraper | 44 | 10 | 0 | 26 | 30 | 33 | 16 | 28 | 35 | 17 | 10 | 13 | 13 | 27 | 12 | 21 | 30 | 24 |
| % Shredder | 10 | 1 | 0 | 9 | 3 | 18 | 7 | 5 | 7 | 18 | 11 | 2 | 4 | 16 | 25 | 13 | 12 | 30 |
| % Other | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 0 | 0 |

7.2.4 Basic Water Quality Field Measurements

Patterns in the basic water quality data at continuously monitored sites in the Lagunitas Creek watershed are in agreement with the assemblage groups identified using bioassessment data. In the MSL bioassessment group, the Lagunitas Creek mainstem site (LAG165) and Devils Gulch (LAG190) both have narrow ranges for temperature, DO, pH, and specific conductance, indicating stable habitats. Sites in the OL bioassessment group (LAG050 and LAG085) have slightly greater ranges of temperature, DO, pH, and specific conductance in the summer dry season. The lower Olema Creek site (LAG050) also showed greater ranges in the spring season for each parameter. The site in the SG bioassessment group (LAG270), which showed lower biological integrity than the mainstem group, had low DO, and both a wide range and high values of specific conductance in the summer (Figure 7-14 and Figure 7-16).

Similar to the pattern seen in the invertebrate assemblages, the Nicasio site (LAG160) exhibited unusually stressful conditions in May, particularly DO levels that ranged from 0.1 to 8.76 mg/L. While the temperature range at LAG160 in the spring was not unusual, the timing of the maximum temperature (at night) and the DO pattern were anomalous, with no obvious explanation. In the winter, however, DO levels were high with full saturation (Figure 7-14). While this reach of stream is dry in summer, the exceedingly low DO in spring and the poor benthic invertebrate assemblage are uncharacteristic of properly functioning, healthy intermittent streams.

Stream temperatures exceeded some thresholds at all continuously monitored sites (Figure 7-12, Figure 7-13, and Appendix D). Highest Weekly Average Temperature (HWAT) values exceeded the Maximum Weekly Average Temperature (MWAT) threshold for coho salmon (14.8°C) during the spring wet season at Bear Valley Road Bridge (LAG050), and during the summer dry season at all five sites (LAG050, LAG085, LAG165, LAG190, and LAG270). Physical habitat scores were available for three of these sites (LAG165, LAG190, and LAG270), but varied little among them. Scores for channel alteration, bank vegetation, riparian zone width, and canopy cover ranged from the 57th to the 100th percentile for all categories (Appendix E).

Dissolved oxygen concentrations fell below Basin Plan objectives and/or salmonid thresholds at all continuously monitored sites during at least one season (Figure 7-12, Figure 7-14, and Appendix D). The measured pH at Bear Valley Road Bridge (LAG050) during the spring wet season varied from a minimum pH value of 6.60 to a maximum of 7.71, a notable variation of more than one pH unit (Figure 7-15). Such a wide variation usually indicates high photosynthetic activity, often with a high biological oxygen demand. This was the lowest pH value measured in the watershed.

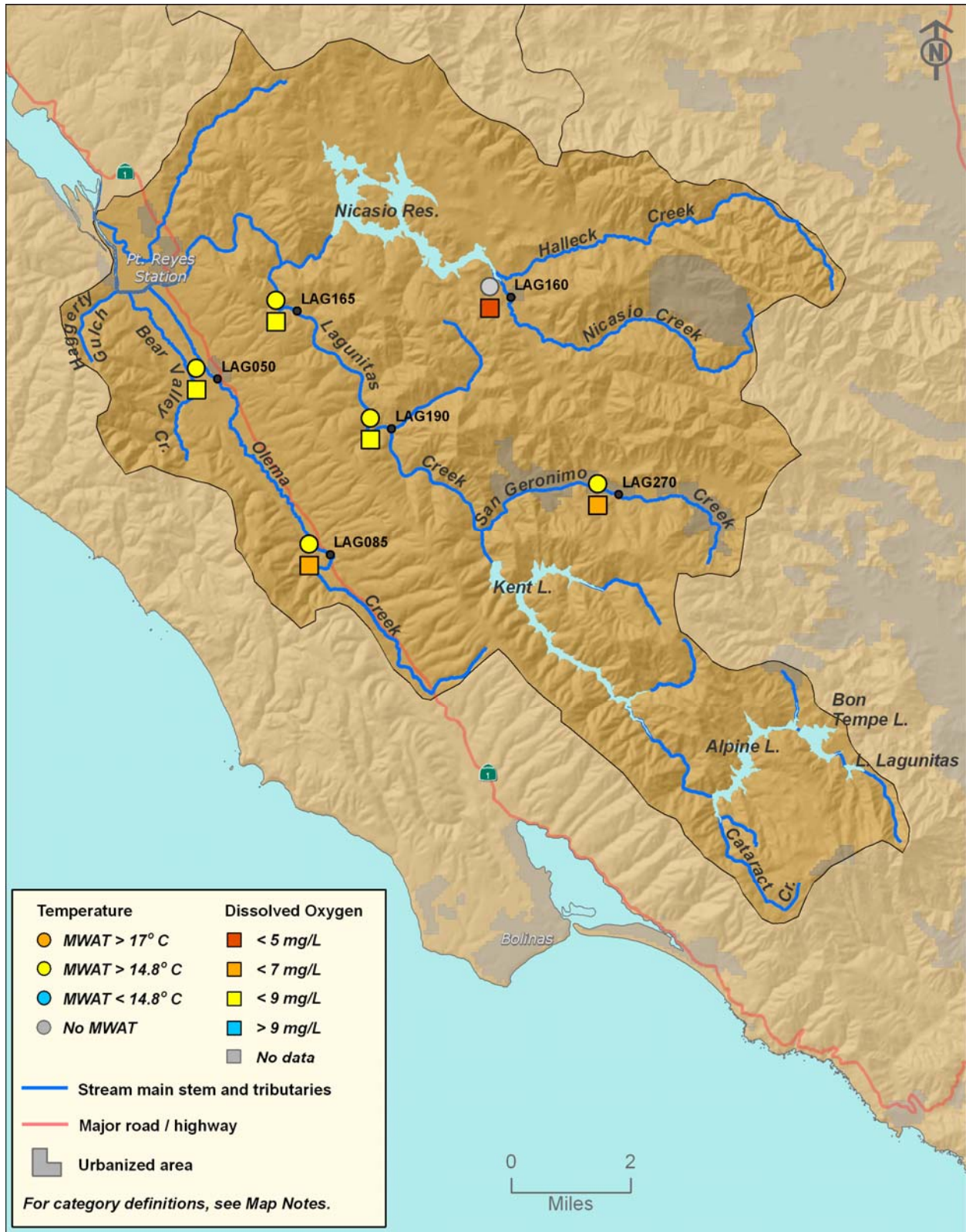


Figure 7-12. Map of temperature and DO levels in Lagunitas Creek

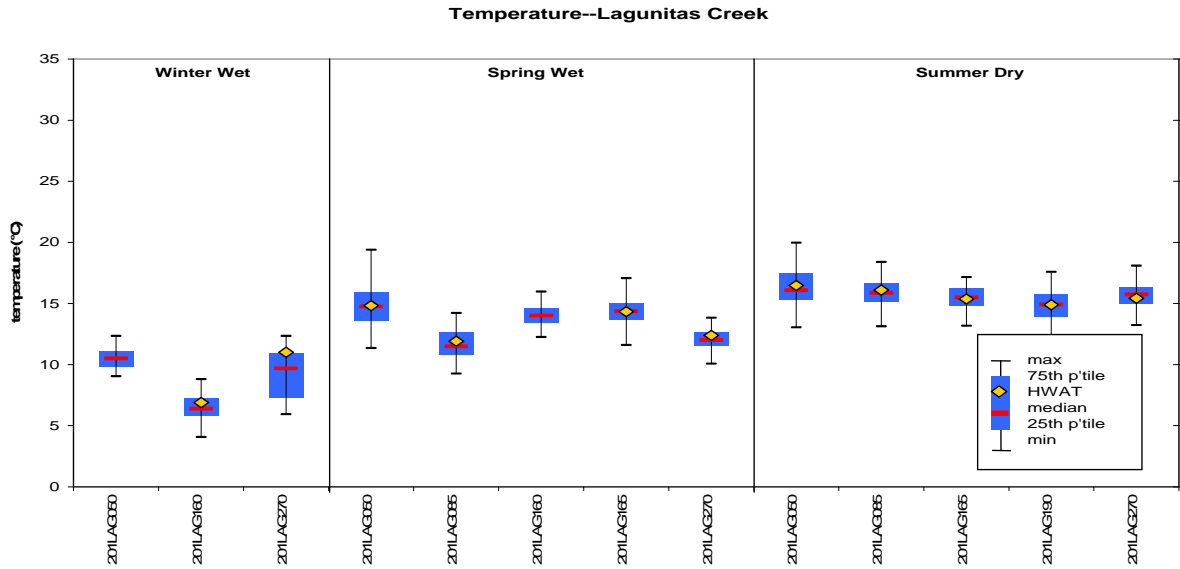


Figure 7-13. Temperature monitoring in Lagunitas Creek watershed

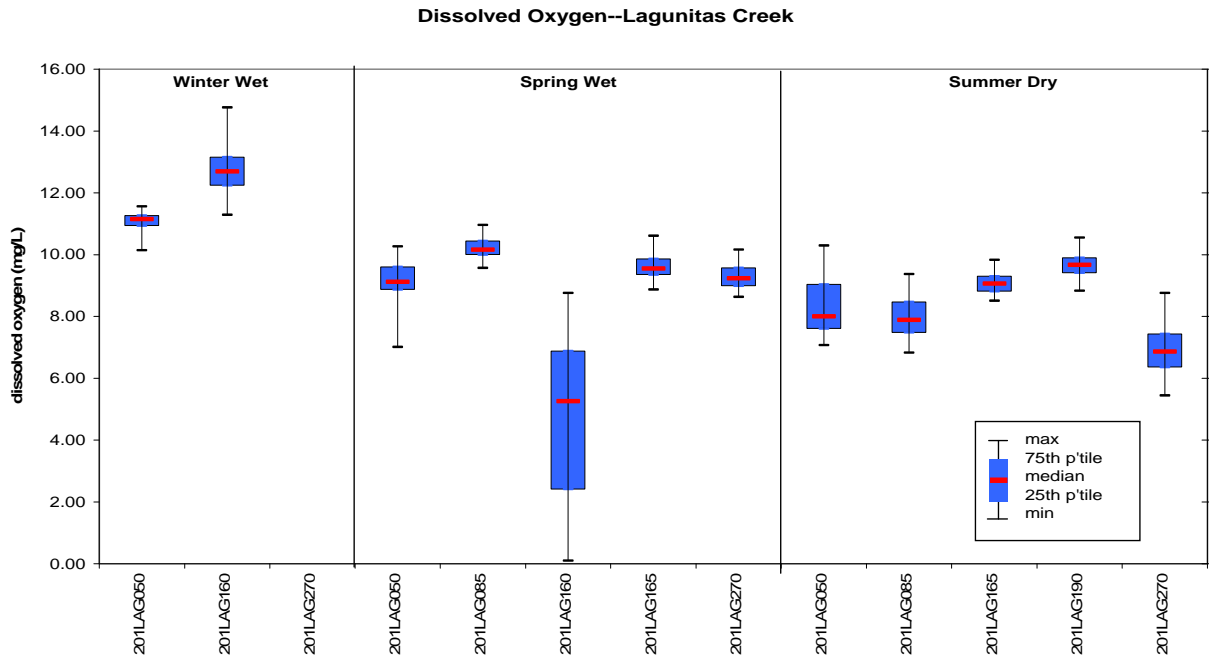


Figure 7-14. DO monitoring in Lagunitas Creek watershed

Data were rejected for LAG270 in the winter.

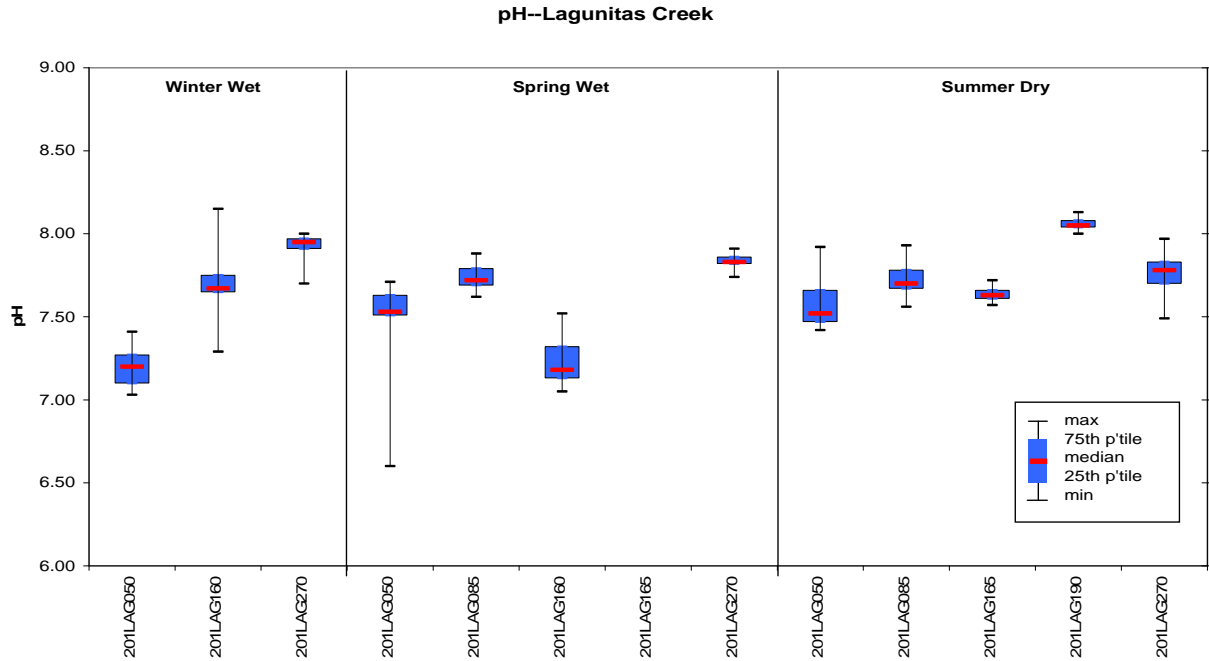


Figure 7-15. pH monitoring in Lagunitas Creek watershed

Data were rejected for LAG165 in the spring.

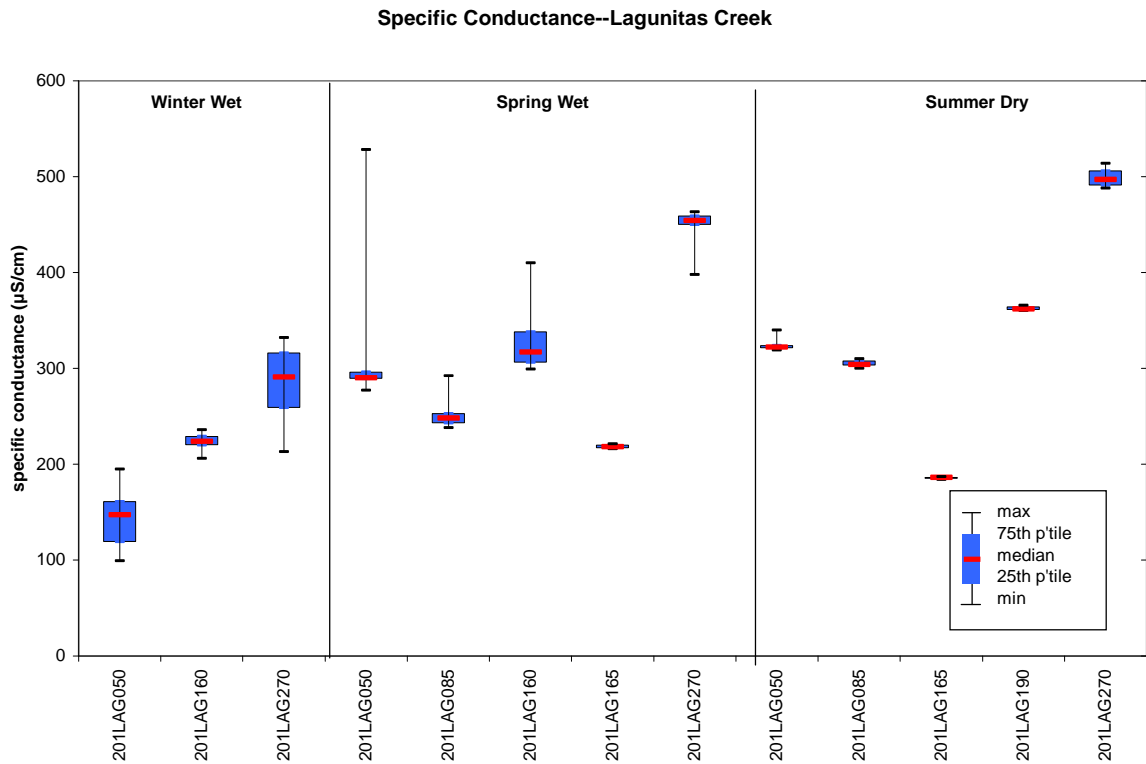


Figure 7-16. Specific conductance monitoring in Lagunitas Creek watershed

7.2.5 Water, Sediment, and Clam Tissue Chemistry

Water samples from the Lagunitas Creek watershed all had relatively low nitrate concentrations, with all dry season values falling below the U.S. EPA reference guideline. Wet season concentrations were significantly higher than those measured in the dry season; all wet season values exceeded the U.S. EPA guideline (Figures 6-15 through 6-17; Appendix G). There was little variation among sites.

Chlorophyll-a values were also generally low, with only two measurements above the 1.78 µg/L guideline value (Appendix G). Those two measurements above guideline values came from samples collected in the dry season at Halleck Creek (LAG 150) and the wet season at Gallagher's Ranch (LAG130). Total phosphorus measurements were all low, 0.09 mg/L or less, with many values at the minimum reporting limit of 0.04 mg/L.

Orthophosphate values were among the lowest in the region; however only four measurements were below the 0.030 mg/L U.S. EPA region-specific guideline for total phosphorus.

Metal and organic contaminants were measured at three sites: Lower Olema Creek (LAG040), Gallagher's Ranch (LAG130), and Creamery Gulch (LAG270). Land use at these sites was characterized as mixed at the first two and rural residential at the third (Table 3-4). Basin Plan water quality objectives (WQOs) for protection of aquatic life were not exceeded in any water samples from this watershed.

Sediments were collected for analysis from two sites: Green Bridge on Lagunitas Creek at Hwy 1 (LAG120, mixed land use), and Creamery Gulch on San Geronimo Creek (LAG270, rural residential), both during the dry season. No organic chemicals were measured at concentrations above sediment quality guidelines (Table 4-3).

However, geologically abundant chromium and nickel were measured above the probable effects concentration (PEC) in sediments from both sites, and arsenic, copper, and mercury were measured above the threshold effects concentrations (TECs) in sediments from San Geronimo Creek (LAG270). These elevated metals concentrations gave the San Geronimo Creek site a sediment quality guideline quotient (SQGQ) value 0.58, the highest of any site measured in the region during this study. In comparison, the SQGQ value for Lagunitas Creek at Hwy 1 was 0.18. Generally, SQGQ values greater than 0.5 tend to be associated with acute toxicity to infaunal invertebrates (MacDonald *et al.* 2000). However, the bioavailability and toxicity of these geologically available metals is uncertain in this setting. These concentrations may be more representative of increased erosion in the watershed than of anthropogenic sources of toxicants.

Tissues of clams deployed at lower Olema Creek (LAG040), Gallagher's Ranch (LAG130), and Creamery Gulch (LAG270) had the highest mercury concentrations measured anywhere in the region (Figure 6-22). Copper concentrations were higher than in controls (Figure 6-21), but selenium and organic chemicals were similar to control values. These tissue results reflect the sediment chemistry measurements of elevated metals and lower concentrations of organics.

7.2.6 Water and Sediment Toxicity

Toxicity tests were conducted on water samples from lower Olema Creek (LAG040), Gallagher's Ranch (LAG130), and Creamery Gulch (LAG270); and on sediment samples from Lagunitas Creek at Hwy 1 (LAG120) and Creamery Gulch on San Geronimo Creek (LAG270). No significant toxicity was observed in any of these samples (Figure 7-17). This is worth noting because of the elevated trace metal concentrations in San Geronimo Creek sediments, and indicates that these metals were not highly bioavailable.

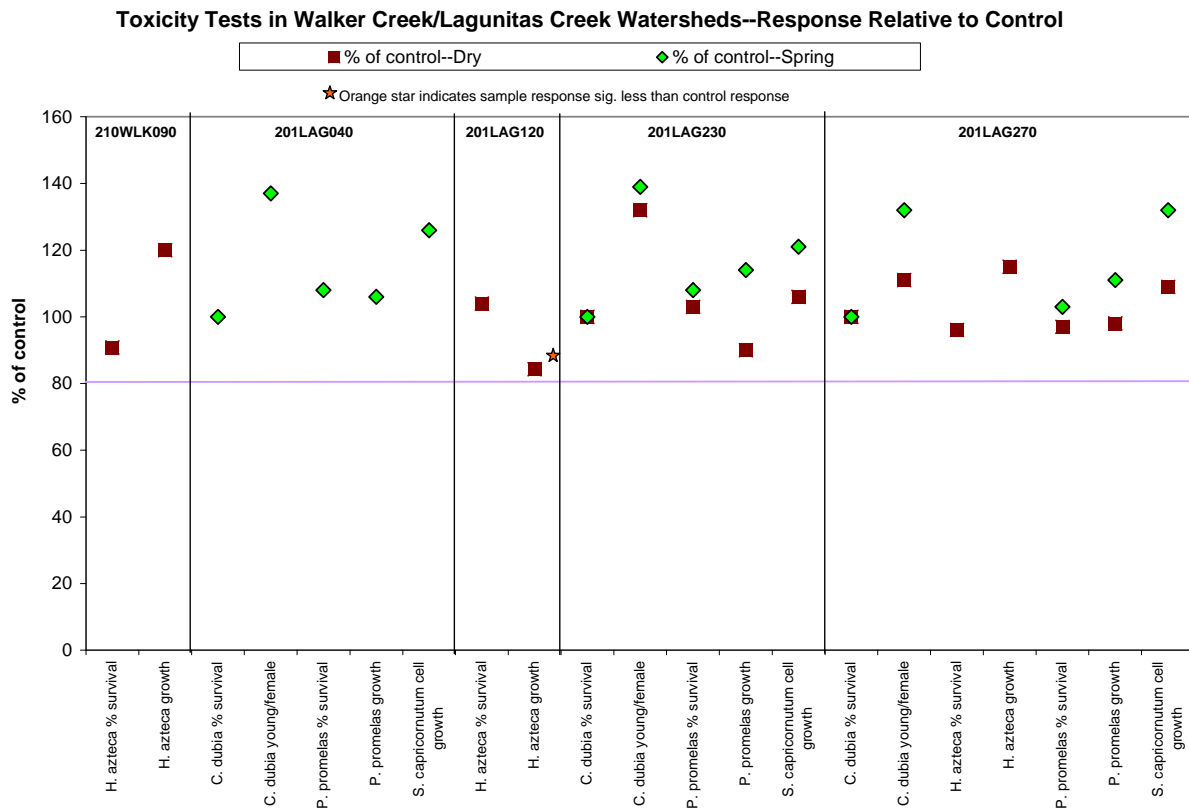


Figure 7-17. Toxicity tests in Walker and Lagunitas creeks

Results of toxicity tests are shown with three species in water samples and one species in sediment samples from Walker and Lagunitas Creeks. See text for description of species, endpoints, and seasons. Stars indicate statistically significant differences from test controls. Samples with stars that were also below 80 percent of the control were considered toxic for this assessment.

7.2.7 Coliform Bacteria

Bacterial counts were made in water samples from five sites in the watershed (Figures 6-28 through 6-30).

- Hagmaier Pond (LAG115) had acceptable levels of *E. coli* and total coliforms, but exceeded the Basin Plan objective for 90th percentile fecal coliform.
- Lagunitas Creek at Hwy 1 (Green Bridge, LAG120) exceeded Basin Plan objectives for all bacterial types.
- Inkwells (LAG230) exceeded only the objective for total coliforms.
- The Swimming Hole at Samuel P. Taylor State Park (LAG185) exceeded only the *E. coli* Basin Plan objective for water-contact recreation.
- Cataract (LAG390), a site above a municipal water source in the Lagunitas Creek headwaters, did not exceed any Basin Plan coliform objectives for recreational use.

7.3 San Leandro Creek Watershed

7.3.1 Sites of Concern

Exceedances noted in the summary table (Table 7-7) are discussed in the relevant sub-section for this watershed. A weight-of-evidence approach was used to evaluate sites of concern. Individual exceedances of parameters, particularly those without regulatory objectives, are considered in context. In general, nickel and chromium levels are geologically elevated in the Bay Area, particularly where there are serpentine deposits. Nutrient levels (nitrate and phosphorus) are above U.S. EPA guidelines at most sites, but may still be relatively low.

The most contaminated site measured in the watershed was the most downstream site, Lower San Leandro Creek at Empire Road (SLE030). Measured concentrations of numerous chemicals exceeded guideline values, water was toxic to invertebrate reproduction, and sediment was highly toxic to amphipods. This urban site had the highest nutrient and chlorophyll values in the watershed as well.

Coliform indicator concentrations exceeded at least one Basin Plan objective at all recreational sites. Contaminants appear to be a concern in this watershed along with conventional parameters such as temperature and DO.

Ongoing deployment of temperature loggers at key sites would help clarify conditions for salmonids, and continued monitoring of sediment contamination is recommended to help identify chemicals of concern in the watershed and possible sources to the Bay.

Table 7-7. Summary of sites with exceedances in San Leandro Creek watershed

| <i>Analyte</i> | <i>Standard</i> | <i># of sites</i> |
|--|-----------------|-------------------|
| FIELD MEASURES of 7 sites sampled | | |
| Temperature | MWAT, coho | 1 |
| Oxygen, dissolved | minimum, COLD | 5 |
| | 3-month median | 5 |
| pH | range | 1 |
| CONVENTIONAL WATER QUALITY of 4 sites sampled | | |
| Nitrate as N | maximum | 4 |
| Phosphorus, total as P | maximum | 4 |
| WATER METALS of 2 sites sampled | | |
| Chromium VI, dissolved | acute | 1 |
| | chronic | 1 |
| WATER ORGANICS of 2 sites sampled | | |
| Diazinon | acute | 1 |
| WATER TOXICITY of 2 sites sampled | | |
| <i>Ceriodaphnia</i> | reproduction | 1 |
| SEDIMENT CHEMISTRY of 1 site sampled | | |

| Analyte | Standard | # of sites |
|--|-----------------|-------------------|
| <i>Metals</i> | | |
| Arsenic | TEC | 1 |
| Cadmium | TEC | 1 |
| Chromium | PEC | 1 |
| Copper | TEC | 1 |
| Lead | PEC | 1 |
| Mercury | TEC | 1 |
| Nickel | PEC | 1 |
| Zinc | TEC | 1 |
| <i>Organics</i> | | |
| Chlordane | TEC | 1 |
| DDD (sum op + pp) | TEC | 1 |
| DDE (sum op + pp) | TEC | 1 |
| DDT (total) | TEC | 1 |
| SEDIMENT TOXICITY of 1 site sampled | | |
| <i>Hyalella</i> | survival | 1 |
| COLIFORMS of 5 sites sampled | | |
| <i>E. coli</i> | steady state | 2 |
| Fecal coliform | log mean | 2 |
| | 90th percentile | 3 |
| Total coliform | median | 5 |
| | maximum | 1 |

7.3.2 Water Quality in Relation to Land Use

The most urban site at the downstream end of the watershed had high contaminant and nutrient concentrations, but since contaminants were measured at few sites, it is not possible to determine statistical relationships with land use. Samples from the upstream site on Moraga Creek (SLE210), which drains an urban area, had relatively high nutrient and algal concentrations. Benthic invertebrate assemblages at sites in urban areas were severely degraded relative to conditions at sites draining open space parkland. Relationships between bioassessment data and land use are further discussed in Section 6.1.1 above.

7.3.3 Macroinvertebrate Assemblages and Physical Habitat

Three benthic assemblage groups were identified in the San Leandro Creek watershed from the cluster dendrogram (Figure 7-19):

- The Middle Mainstem (MM) group included two sites located on the mainstem of San Leandro Creek in the heavily urbanized City of San Leandro (SLE050 and SLE070).
- The Undeveloped Tributaries (UD) group included five sites on small tributaries draining watersheds with relatively low intensity human land use (SLE180, SLE190, SLE200, SLE220, and SLE230).
- The third assemblage group included sites from throughout the watershed.
 - SLE030 is located at the bottom of the watershed near the maximum extent of tidal influence.
 - SLE090 is on the mainstem of San Leandro Creek below Chabot Dam.
 - SLE170 is on the West Fork of Redwood Creek, a small tributary to Upper San Leandro Reservoir, in Redwood Regional Park.
 - SLE210 is located on Moraga Creek, a small tributary to Upper San Leandro Reservoir, in suburban Moraga..

These sites all possess an intermediate to high level of human land use in their upstream watersheds.

Middle Mainstem group

Benthic assemblages at sites in the Middle Mainstem Group (MM) were dominated by the following taxa:

| | |
|--------------|-----|
| Oligochaeta | 82% |
| Chironomidae | 12% |

Other taxa present at both sites include the mothfly *Psychoda* sp., Ostracoda, the water mite family Libertiidae, and the pond snail *Fossaria* sp. No specimens from the Ephemeroptera, Plecoptera, Trichoptera, or Coleoptera insect orders were collected at either site.

The assemblage composition suggests extremely poor water quality. Mothflies feed on microorganisms associated with raw sewage, while Oligochaetes and Chironomids are often dominant when poor conditions inhibit other insects such as the mayflies, stoneflies, and caddisflies.

Undeveloped Tributaries group

The Undeveloped Tributaries (UD) group contained a diverse assemblage of invertebrates, with taxa richness values at sites ranging from 27 to 39. The benthic assemblages at these five sites reflect excellent water quality conditions. Differences in taxonomic composition most likely reflect minor environmental differences among the sites. The numerically dominant taxa at these sites were:

| | |
|---|-----|
| Chironomidae | 30% |
| <i>Optioservus</i> sp. (Elmidae) | 8% |
| <i>Baetis</i> sp. (Baetidae) | 7% |
| <i>Parthina</i> sp. (Odontoceridae) | 6% |
| <i>Simulium</i> sp. (Simuliidae) | 3% |
| <i>Malenka</i> sp. (Nemouridae) | 2% |
| <i>Paraleptophlebia</i> sp. (Leptophlebiidae) | 2% |

Many other EPT taxa were abundant and found at a majority of sites, including *Calineuria californica*, *Dipheter hageni*, and *Lepidostoma* sp.

Other sites throughout the watershed

Sites in the third assemblage group were dominated by four common tolerant taxa that are usually the most abundant invertebrates in highly disturbed streams:

| | |
|----------------------------------|-----|
| Chironomidae | 58% |
| <i>Simulium</i> sp. (Simuliidae) | 16% |
| <i>Baetis</i> sp. (Baetidae) | 6% |
| Oligochaeta | 4% |

Few other EPT taxa were found, other than several caddisflies and stoneflies at SLE170.

The four sites in this group reflect a moderate to high level of disturbance, although not as high as the Middle Mainstem (MM) sites. Given the range of land uses that drain these sites, these results indicate that a small amount of intensive land use can drastically alter benthic communities. For example, the West Fork of Redwood Creek (SLE170) drains a forested watershed partly in a regional park. Human land use in the watershed is relatively minor, but includes roads, horse stables, and several residences. Moraga Creek (SLE210) drains a suburban residential neighborhood. SLE090 is located on San Leandro Creek just below the dam at Lake Chabot. Benthic assemblages at these three sites are similar to those found at SLE030, the site on the mainstem of San Leandro Creek below the heavily urbanized City of San Leandro.

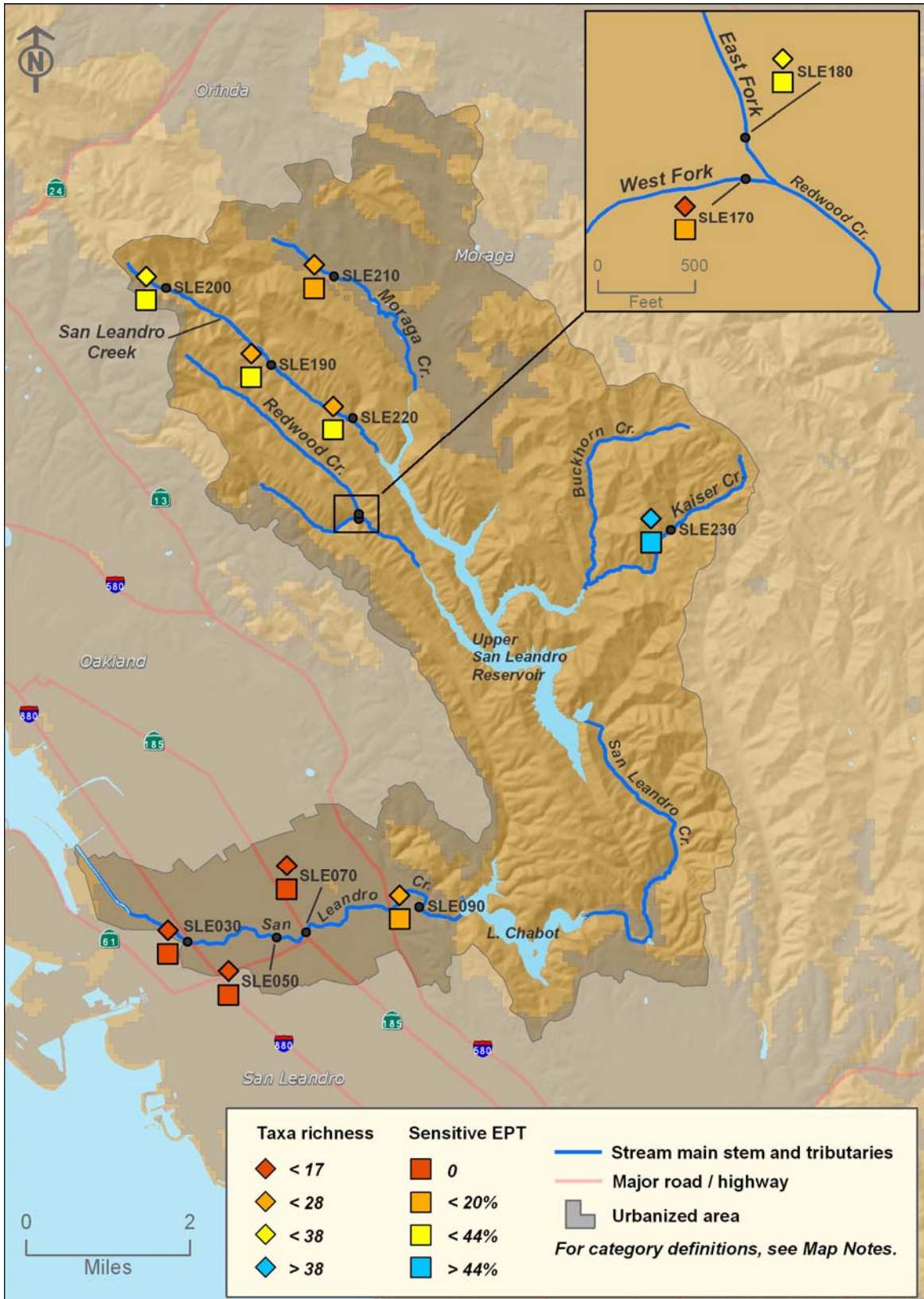


Figure 7-18. Map of index scores for macroinvertebrates in San Leandro Creek

Table 7-8. Bioassessment sites in the San Leandro Creek watershed

| Site Code | Site Name | Stream |
|-----------|-------------------|-------------------------|
| SLE030 | Empire Road | Lower San Leandro Creek |
| SLE050 | San Leandro BART | Lower San Leandro Creek |
| SLE070 | Root Park | Lower San Leandro Creek |
| SLE090 | Chabot Park | Lower San Leandro Creek |
| SLE170 | West Fork Redwood | West Fork Redwood Creek |
| SLE180 | East Fork Redwood | East Fork Redwood Creek |
| SLE190 | Canyon School | Upper San Leandro Creek |
| SLE200 | Huckleberry | Upper San Leandro Creek |
| SLE210 | Moraga | Moraga Creek |
| SLE220 | Indian | Indian Creek |
| SLE230 | Kaiser | Kaiser Creek |

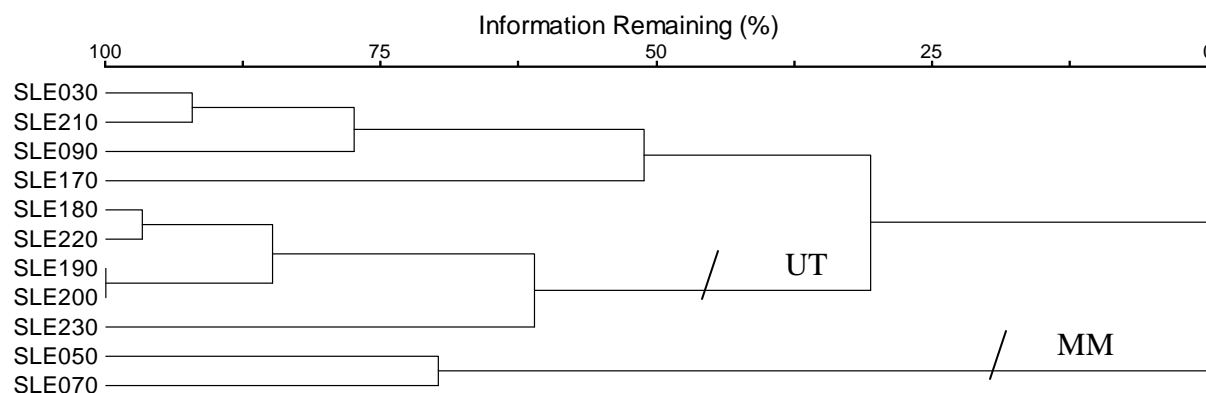


Figure 7-19. Cluster dendrogram of sites in the San Leandro Creek watershed

The cluster dendrogram groups sites based on similarity of macroinvertebrate taxa present. Sites that are closely joined together have many similar taxa present, while groups with longer chain lengths are more dissimilar. The Information Remaining axis at top signifies how similar the sites are with respect to taxa presence. Abbreviations: UT, Undeveloped Tributaries; MM, Middle Mainstem.

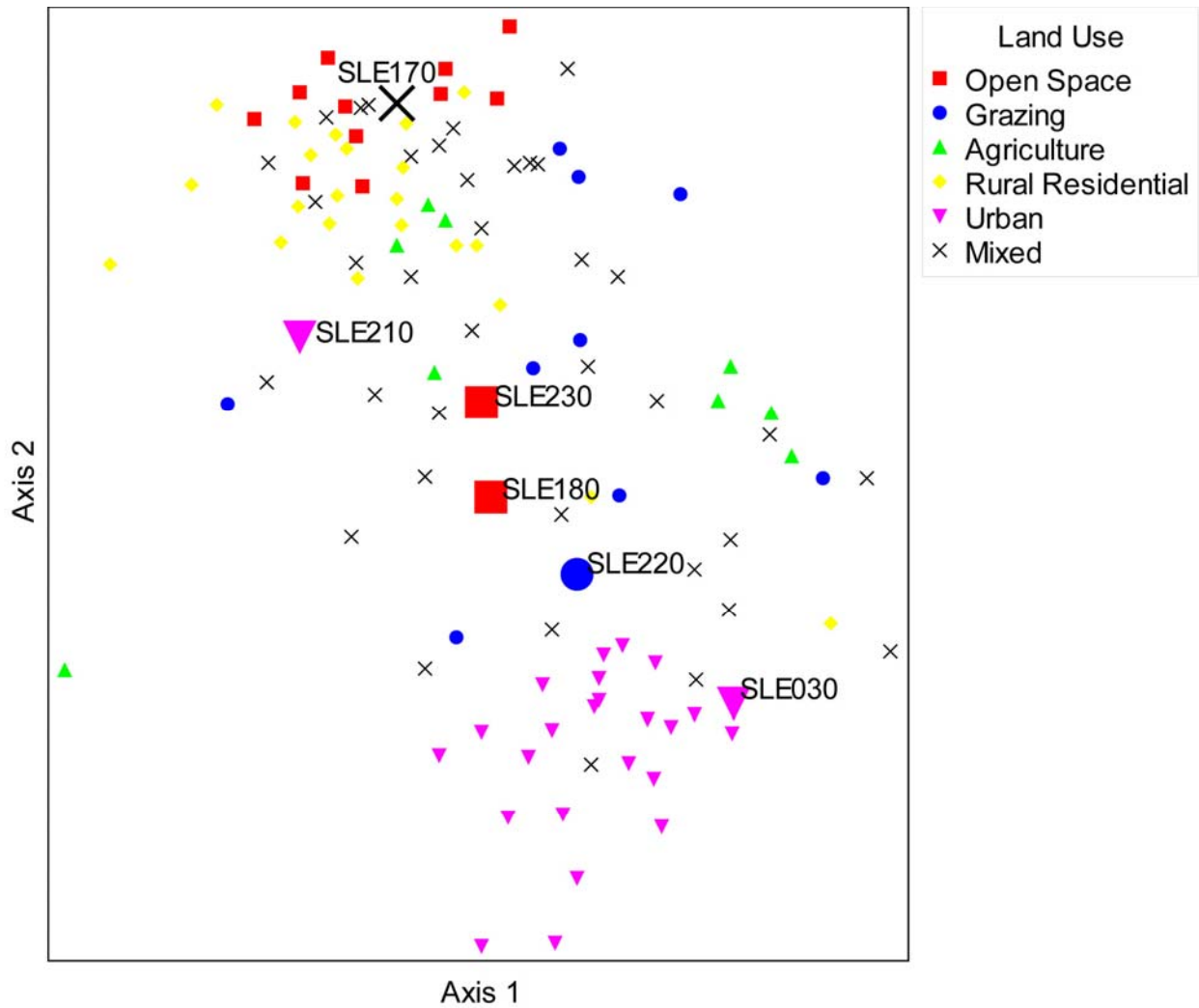


Figure 7-20. NMS ordination of sites in the San Leandro Creek watershed

The NMS ordination of sites by macroinvertebrate taxa abundance. Sites in the San Leandro Creek watershed are labeled and symbols are enlarged. Note that several sites are not included in the ordination because the sites were sampled as part of a different study (see Breaux et al. 2005). For an explanation of ordination graphs, refer to Section 6.1.

Table 7-9. Biological metrics from the San Leandro Creek watershed

| | SLE030 | SLE050 | SLE070 | SLE090 | SLE170 | SLE180 | SLE190 | SLE200 | SLE210 | SLE220 | SLE230 |
|----------------------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| Coleoptera Taxa | 0 | 0 | 0 | 1 | 2 | 2 | 6 | 4 | 2 | 2 | 7 |
| Diptera Taxa | 2 | 5 | 7 | 8 | 3 | 10 | 7 | 11 | 5 | 7 | 7 |
| Ephemeroptera Taxa | 1 | 0 | 0 | 1 | 2 | 5 | 6 | 8 | 1 | 5 | 6 |
| Odonata Taxa | 0 | 0 | 0 | 1 | 0 | 1 | 1 | 1 | 1 | 0 | 0 |
| Plecoptera Taxa | 0 | 0 | 0 | 0 | 2 | 7 | 5 | 3 | 1 | 3 | 4 |
| Trichoptera Taxa | 0 | 0 | 0 | 2 | 4 | 4 | 6 | 5 | 0 | 4 | 9 |
| Non-Insect Taxa | 9 | 6 | 6 | 13 | 3 | 6 | 3 | 5 | 8 | 6 | 5 |
| Taxa Richness | 12 | 12 | 13 | 26 | 16 | 35 | 35 | 37 | 18 | 27 | 39 |
| EPT Taxa | 1 | 0 | 0 | 3 | 8 | 16 | 17 | 16 | 2 | 12 | 19 |
| % EPT | 0 | 0 | 0 | 2 | 12 | 55 | 44 | 46 | 14 | 47 | 66 |
| % Sensitive EPT | 0 | 0 | 0 | 0 | 3 | 28 | 42 | 37 | 0 | 27 | 48 |
| % Chironomidae | 48 | 35 | 7 | 57 | 67 | 30 | 22 | 46 | 60 | 28 | 22 |
| % Coleoptera | 0 | 0 | 0 | 0 | 6 | 4 | 29 | 3 | 1 | 1 | 9 |
| % Oligochaeta | 1 | 52 | 88 | 9 | 2 | 0 | 1 | 0 | 9 | 3 | 0 |
| % Non-insect | 31 | 61 | 90 | 21 | 2 | 4 | 2 | 1 | 12 | 6 | 2 |
| % Intolerant | 0 | 0 | 0 | 0 | 3 | 27 | 29 | 35 | 0 | 26 | 35 |
| % Tolerant | 8 | 90 | 89 | 13 | 0 | 1 | 1 | 1 | 1 | 1 | 1 |
| Tolerance Value | 5.8 | 7.8 | 7.7 | 6.0 | 5.6 | 4.1 | 3.6 | 3.6 | 5.7 | 4.2 | 3.4 |
| % Predator | 9 | 4 | 4 | 9 | 2 | 23 | 17 | 14 | 1 | 18 | 7 |
| % Collector-filterer | 24 | 0 | 0 | 55 | 11 | 1 | 12 | 16 | 14 | 13 | 3 |
| %Collector-gatherer | 60 | 95 | 94 | 30 | 79 | 55 | 26 | 38 | 84 | 57 | 54 |
| % Scraper | 7 | 1 | 2 | 7 | 6 | 14 | 31 | 9 | 1 | 5 | 15 |
| % Shredder | 0 | 0 | 0 | 1 | 2 | 7 | 14 | 23 | 0 | 7 | 18 |
| % Other | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |

7.3.4 Basic Water Quality Field Measurements

Stream temperature data were not available to calculate HWAT values in the urbanized sites in lower San Leandro Creek because deployments were less than seven days. In the upper creek above the reservoir, only Moraga Creek at Moraga (SLE210) exceeded a threshold value—the MWAT threshold for coho salmon (14.8°C) during the summer dry season (see Figure 7-, Figure 7-, and Appendix D). Even in the absence of HWAT values, there appeared to be the potential for temperature stress at other sites, such as Lower San Leandro Creek at Root Park (SLE070) and Chabot Park (SLE090), where the median dry season measured temperatures exceeded 15°C.

Dissolved oxygen concentrations failed to meet Basin Plan objectives (7 mg/L) on Lower San Leandro Creek at San Leandro BART (SLE050); Root Park (SLE070); Chabot Park (SLE090); West Fork Redwood Creek (SLE170), and Moraga Creek (SLE210), mostly during the dry season. Upper San Leandro Creek at Canyon School (SLE190) had DO values below 9 mg/L during two dry season measurements (Figure 7-, Figure 7-, and Appendix D). The Moraga Creek site (SLE210) had maximum pH measured above the Basin Plan objective during the spring wet season, with a maximum pH value of 8.66 (Figure 7-).

Dry season physical habitat scores were available for West Fork Redwood Creek (SLE170) and Moraga Creek (SLE210). Both had high canopy cover scores and relatively low scores for channel modification and riparian zone width.

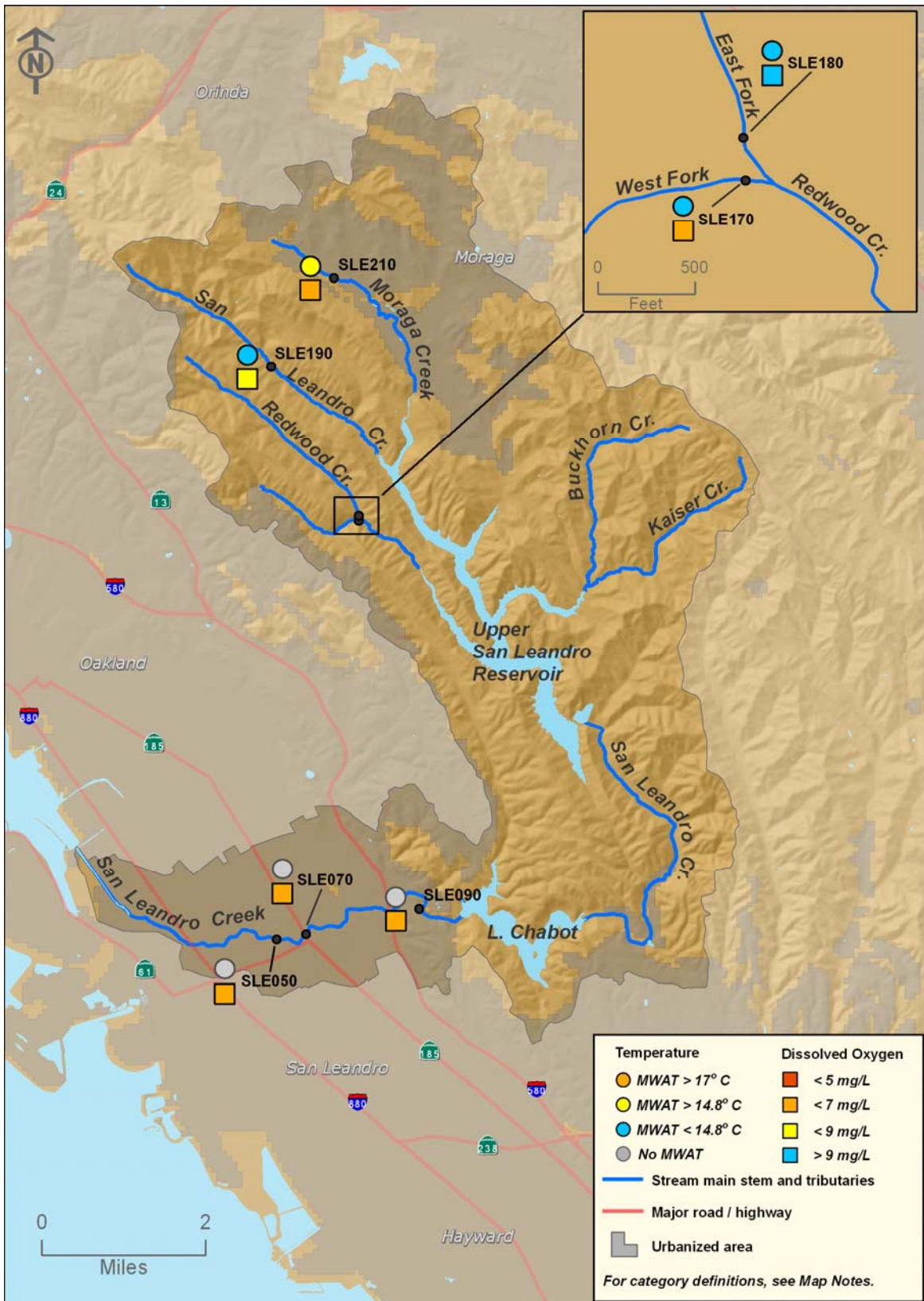


Figure 7-21. Map of temperature and DO levels in San Leandro Creek

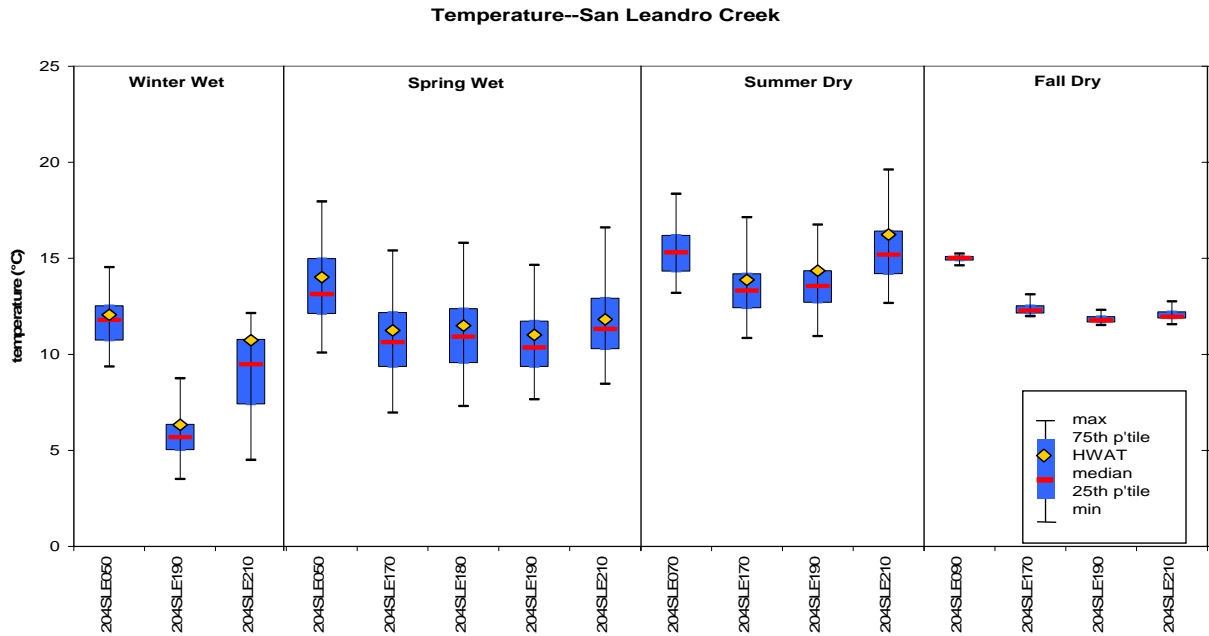


Figure 7-22. Temperature monitoring in San Leandro Creek

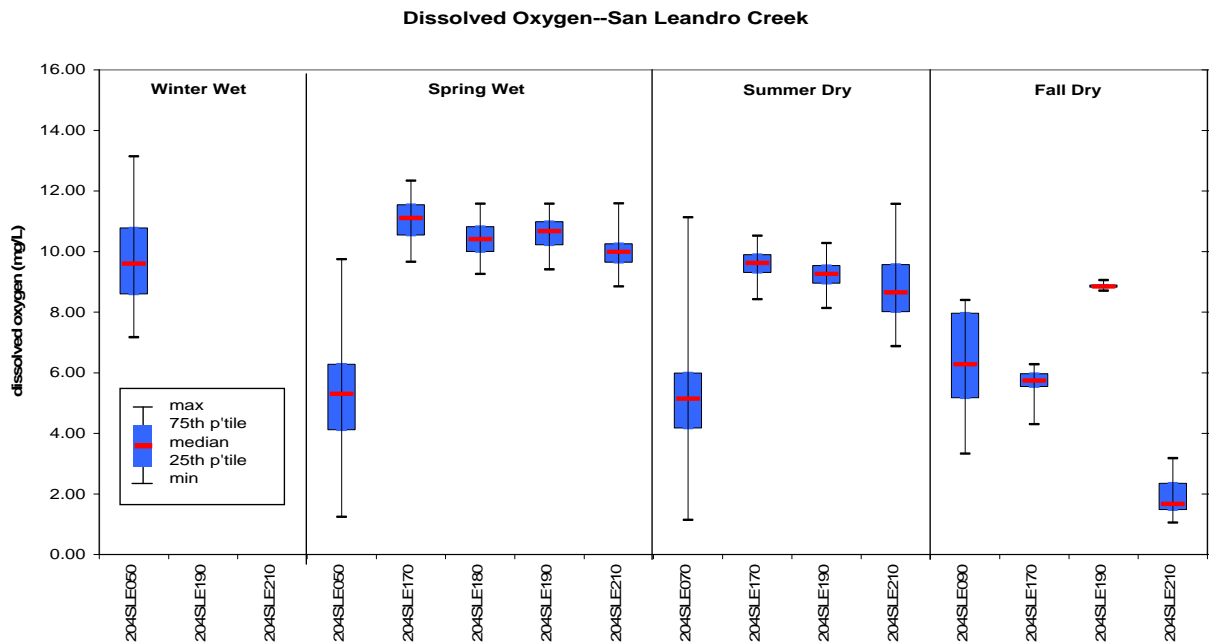


Figure 7-23. DO monitoring in San Leandro Creek

Data were rejected for SLE190 and SLE210 in the winter season.

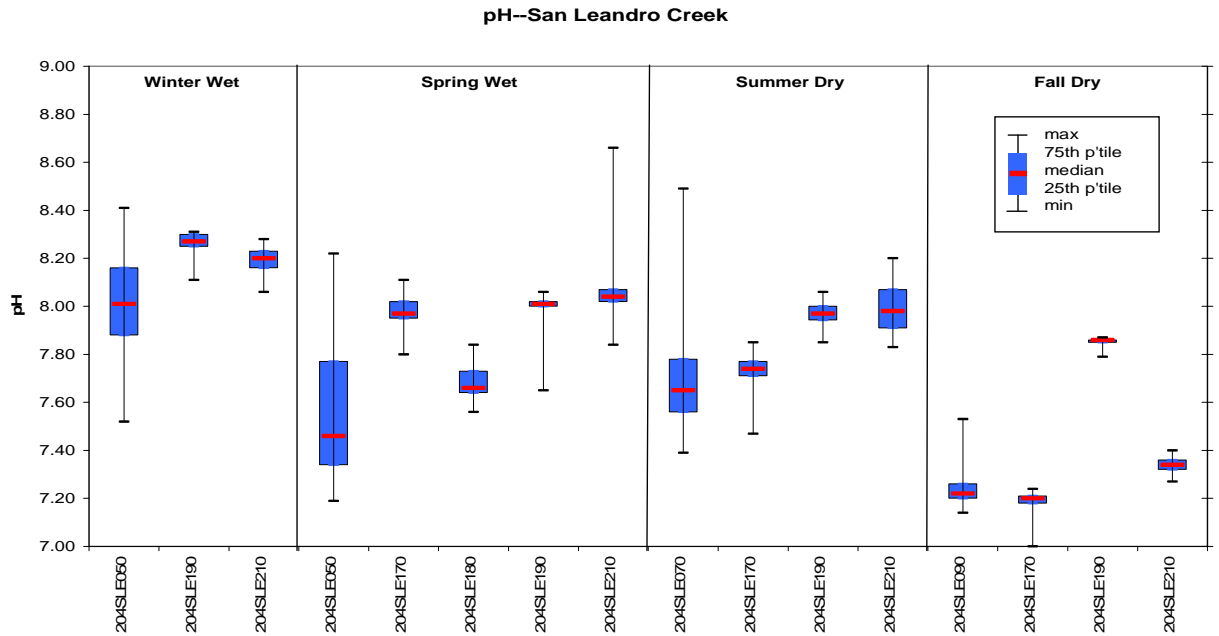


Figure 7-24. pH monitoring in San Leandro Creek

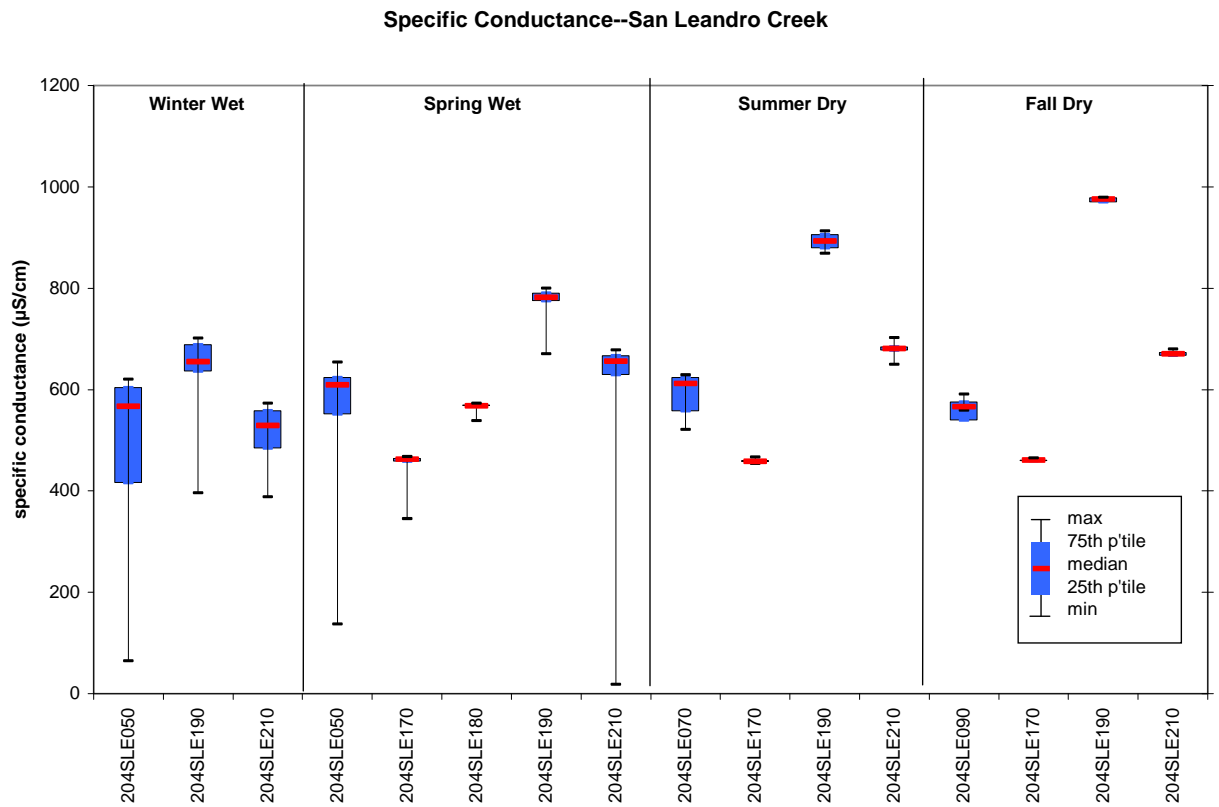


Figure 7-25. Specific conductance monitoring in San Leandro Creek

7.3.5 Water, Sediment, and Clam Tissue Chemistry

Water samples from the upper San Leandro watershed had moderately low nitrate concentrations, but most values exceeded 0.16 mg/L, the U.S. EPA reference guideline. Lower San Leandro Creek at Empire Road (SLE030) had very high nitrate concentrations between 6 and 8 mg/L (Figures 6-15 to 6-17; Appendix G).

Chlorophyll-a values were relatively high at SLE030 as well, with the dry season concentration of 7.9 µg/L being among the highest measured region-wide. The Moraga Creek site (SLE210) also had a relatively high dry season chlorophyll-a measurement of 5.3 µg/L. All others were below the 1.78 µg/L guideline value (Appendix G). (Care should be taken in interpreting water column chlorophyll values in these streams, since benthic algae are often dominant but not represented in water column samples.)

Total phosphorus measurements were highest at the Moraga Creek site (SLE210), with a dry season concentration of 0.46 mg/L. As with all samples in this study, values in the San Leandro Creek watershed were above the 0.030 mg/L U.S. EPA reference guideline value.

Orthophosphate values were also highest at the Moraga Creek site (SLE210), with a dry season concentration of 0.36 mg/L. All orthophosphate measurements in the watershed exceeded the 0.030 mg/L U.S. EPA guideline for total phosphorus.

Trace metal and organic contaminants were measured in water samples from Lower San Leandro Creek at Empire Road (SLE030), and metals were measured at Kaiser Creek (SLE230). Basin Plan objectives for aquatic life protection were exceeded for chromium and diazinon at SLE030. The chronic WQO was exceeded for chromium in the dry season, and both the acute and chronic WQOs were exceeded in the spring sample. Diazinon was measured above the WQO in the spring.

Sediments were collected for analysis from the most downstream site in the watershed, Lower San Leandro Creek at Empire Road (SLE030). Geologically abundant chromium and nickel were measured above the Probable Effects Concentration in sediments from this site, as was lead. The sediment lead concentration of 130 mg/kg was about 10 times higher than for any other sediment sample measured in this study. Sediment arsenic, cadmium, copper, lead, mercury, zinc, sum DDD, sum DDE and sum all DDTs were measured above the Threshold Effects Concentrations. This Lower San Leandro Creek sediment had the second highest sediment quality guideline quotient (SQGQ) value measured in the region during this study at 0.44 (0.47 with DDE included). These SQGQ values approach the 0.5 value evaluated by MacDonald *et al.*, 2000, but are not likely to be associated with acute sediment toxicity due to contaminant mixtures.

Clam deployments were unsuccessful in this watershed, and tissue data are unavailable.

7.3.6 Water and Sediment Toxicity

Three-species toxicity tests were conducted on water samples from Lower San Leandro Creek at Empire Road (SLE030), and Kaiser Creek (SLE230); amphipod tests were conducted on a sediment sample from SLE030. Diazinon has been commonly linked to lethal effects on *C. dubia*. There was a significant adverse effect on *C. dubia* reproduction in a dry season water sample from the Lower San Leandro Creek site (Figure 7-). Diazinon was measured in this sample below the lethal concentration. Other compounds were measured in this sample as well, and toxicity may be the result of a mixture of chemicals. No toxicity was observed in any of the other water samples.

The sediment sample from SLE030 was highly toxic, with 100 percent mortality in all replicates. It is important to note the very high total organic carbon concentration in this sample (9.4 percent), which may have negatively affected the test organisms directly, and which may be indicative of high contaminant loading rates. Of the chemicals measured for which probable effects concentration (PEC) guidelines are available, only lead exceeded the PEC value, while the legacy pesticides DDE, DDD, and DDT exceeded the threshold effects concentrations (TECs). The SQGQ for this sample was 0.44 (0.47 with DDE included), which is just below the 0.5 value evaluated by MacDonald *et al.* (2000) for association with acute toxicity.

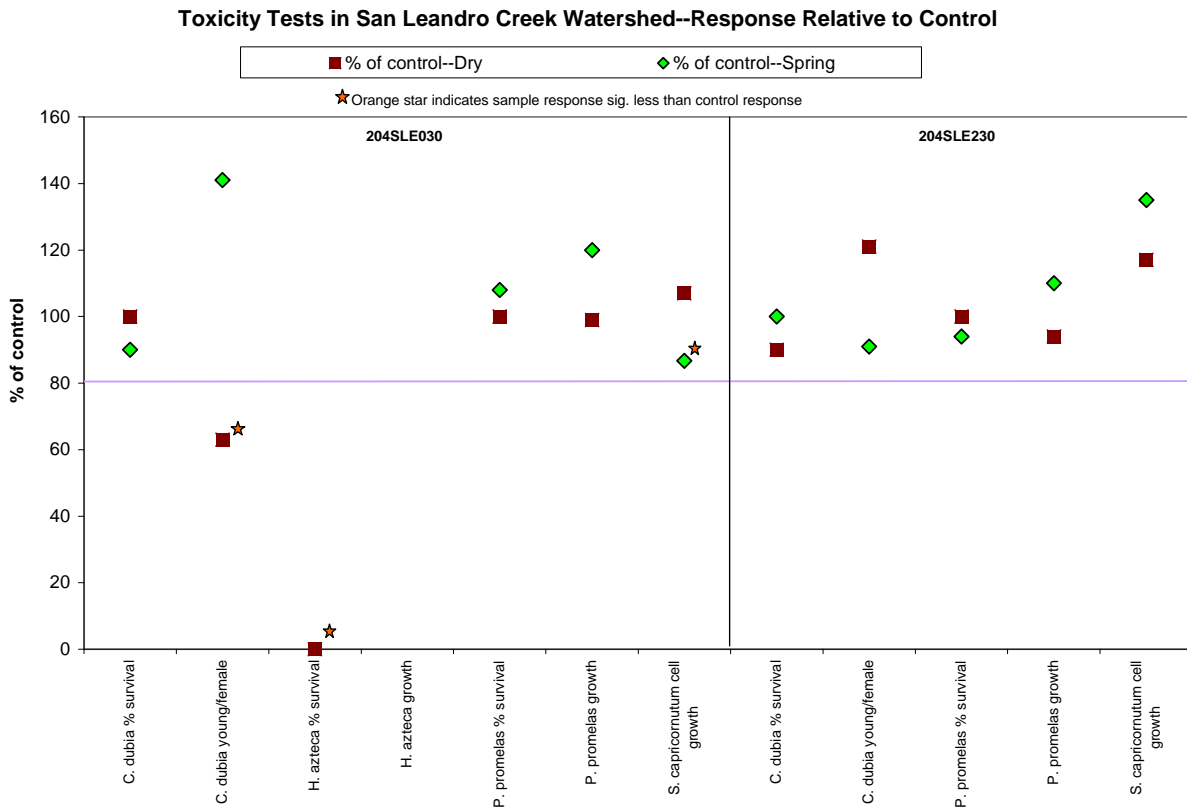


Figure 7-26. Toxicity tests in San Leandro Creek

Results of toxicity tests are shown, with three species in water samples and one species in sediment samples from San Leandro Creek. See text for description of species and endpoints. Stars indicate statistically significant differences from test controls. Samples with stars that were also below 80 percent of the control were considered toxic for this assessment.

7.3.7 Coliform Bacteria

Bacterial counts were made in water samples from five sites in the watershed (Figures 6-28 through 6-30). Basin Plan objectives for all three indicators (*E. coli*, fecal coliforms, and total coliforms) were exceeded at both the Lower San Leandro Creek site at Root Park (SLE070) and the Upper San Leandro Creek site at Canyon School (SLE190); the counts were considerably higher at the lower creek site. Basin Plan objectives for total coliforms only were exceeded at both the Lower San Leandro Creek site at Chabot Park (SLE090) and the W. Fork Redwood Creek site (SLE170), where the concentrations were comparable. Basin Plan objectives for fecal coliforms only were exceeded at Upper San Leandro Creek at Huckleberry (SLE200).

7.4 Wildcat/San Pablo Creek Watershed

7.4.1 Sites of Concern

Exceedances noted in the summary tables (Table 7-10 and Table 7-11) are discussed in the relevant sub-section for this watershed. A weight-of-evidence approach was used to evaluate sites of concern. Individual exceedances of parameters, particularly those without regulatory objectives, are considered in context. In general, nickel and chromium levels are geologically elevated in the Bay Area, particularly where there are serpentine deposits. Nutrient levels (nitrate and phosphorus) are above U.S. EPA guidelines at most sites, but may still be relatively low.

Benthic invertebrate assemblages generally reflected poor conditions throughout the watershed. Urban sites in the lower watershed had moderate concentrations of contaminants. Measurements from the Lauterwasser Creek site (SPA200) exceeded the short-term maximum salmonid temperature of 24°C, had dissolved oxygen concentrations below the salmonid threshold during the dry season, and had the highest orthophosphate concentration in the region.

Table 7-10. Summary of sites with exceedances in Wildcat Creek watershed

| <i>Analyte</i> | <i>Standard</i> | <i># of sites</i> |
|--|-----------------|-------------------|
| FIELD MEASURES of 3 sites sampled | | |
| Temperature | MWAT, coho | 2 |
| Oxygen, dissolved | minimum, WARM | 2 |
| | 3-month median | 3 |
| pH | range | 1 |
| CONVENTIONAL WATER QUALITY of 2 sites sampled | | |
| Nitrate as N | maximum | 2 |
| Phosphorus, total as P | maximum | 1 |
| WATER METALS of 1 site sampled | | |
| Mercury, total | chronic | 1 |
| WATER ORGANICS of 1 site sampled | | |
| Chlorpyrifos | chronic | 1 |
| SEDIMENT CHEMISTRY of 1 site sampled | | |
| <i>Metals</i> | | |
| Chromium | PEC | 1 |
| Nickel | PEC | 1 |
| <i>Organics</i> | | |
| Chlordance | TEC | 1 |
| PCB (total) | TEC | 1 |
| SEDIMENT TOXICITY of 1 site sampled | | |
| Hyaella | growth | 1 |
| COLIFORMS of 2 sites sampled | | |
| <i>E. coli</i> | steady state | 1 |
| Fecal coliform | log mean | 1 |

| | | |
|----------------|-----------------|-------------------|
| | 90th percentile | 1 |
| Analyte | Standard | # of sites |
| Total coliform | median | 2 |
| | max | 1 |

Table 7-11. Summary of sites with exceedances in San Pablo Creek watershed

| Analyte | Standard | # of sites |
|--|-------------------|-------------------|
| FIELD MEASURES of 5 sites sampled | | |
| | maximum, salmonid | 1 |
| Temperature | MWAT, steelhead | 2 |
| | MWAT, coho | 4 |
| Oxygen, dissolved | minimum, WARM | 1 |
| | 3-month median | 2 |
| pH | range | 1 |
| CONVENTIONAL WATER QUALITY of 5 sites sampled | | |
| Nitrate as N | maximum | 5 |
| Phosphorus, total as P | maximum | 5 |
| WATER ORGANICS of 4 sites sampled | | |
| Chlorpyrifos | chronic | 2 |
| Diazinon | acute | 2 |
| WATER TOXICITY of 4 sites sampled | | |
| <i>Selenastrum</i> | growth | 2 |
| SEDIMENT CHEMISTRY of 1 site sampled | | |
| <i>Metals</i> | | |
| Chromium | TEC | 1 |
| Mercury | TEC | 1 |
| Nickel | TEC | 1 |
| <i>Organics</i> | | |
| Chlordane | TEC | 1 |
| COLIFORMS of 2 sites sampled | | |
| <i>E. coli</i> | steady state | 2 |
| Fecal coliform | log mean | 2 |
| | 90th percentile | 2 |
| Total coliform | median | 2 |
| | max | 1 |

7.4.2 Water Quality in Relation to Land Use

High manganese and highly variable nickel concentrations in water samples from the Wildcat Creek site at Richmond Parkway (WIL020) may be the result of industrial activities in this urban area, or they may result from mobilization of geologic sources caused by activities such as grading for housing development. Concentrations of the herbicide oxadiazon found in clam tissues at 3rd St. Bridge on San Pablo Creek (SPA020), where toxicity to algae was observed, may indicate excessive herbicide use possibly in residential, nursery, or golf course applications. High orthophosphate concentrations in the Lauterwasser Creek site (SPA200) likely result from residential fertilizer applications in the Lafayette/Orinda suburban areas.

7.4.3 Macroinvertebrate Assemblages and Physical Habitat

Four benthic assemblage groups were identified in the San Pablo Creek/Wildcat Creek planning watershed from the cluster dendrogram (Figure 7-):

- The Urban Mainstem (UMS) group included six sites located near the mouths of San Pablo Creek (SPA020, SPA050, SPA070) and Wildcat Creek (WIL030, WIL050, WIL060), in the Richmond-San Pablo urban area. One site on San Pablo Creek in Orinda (SPA220) had a similar benthic assemblage and was classified into this group as well.
- The San Pablo Creek Tributaries (SPT) group consisted of sites on small tributary streams in the San Pablo Creek watershed including Appian Creek (SPA100), Wilkie Creek (SPA110), Castro Creek (SPA130), Lauterwasser Creek (SPA200), and the San Pablo Creek headwaters (SPA240).
- The Wildcat Canyon (WC) group included a site at the mouth of Wildcat Canyon at Alvarado Park (WIL070); a site downstream of Jewel Lake in Tilden Park (WIL100); and a site at the Lone Oak Picnic area in Tilden Park (WIL130).
- The Bear Creek and Wildcat Headwaters (BCWH) group consisted of three sites on Bear Creek (SPA140, SPA150, SPA160), a tributary to San Pablo Creek that drains into Briones Reservoir, and two sites in upper Wildcat Creek (WIL170, WIL190).

One site (SPA170) was dissimilar and did not fit into the four assemblage groups.

Urban Mainstem group

Benthic assemblages at sites in the Urban Mainstem group (UMG) are indicative of highly disturbed conditions common in urban streams in the Bay Area. Assemblages at these sites were dominated by common tolerant taxa. Oligochaetes (35 percent of the total number of organisms) and Chironomidae (34 percent) were both very abundant at every site. *Baetis* sp. (18 percent) was very abundant at most sites except WIL040 and WIL050. Similarly, *Simulium* sp. (9 percent) was abundant at four sites but not at three others (SPA020, WIL040, WIL060). No other taxa made up greater than one percent of benthic organisms.

Other taxa found at a majority of sites include the relatively tolerant caddisfly *Hydroptila* sp. (Hydroptilidae), the water mite family Sperchontidae (Acari), and the tolerant snail taxa Physa/Physella (Gastropoda: Physidae).

Several moderately intolerant taxa, including the stonefly *Kogotus* sp. (Perlodidae) and the caddisfly *Lepidostoma* sp. (Lepidostomatidae) were found at two sites, WIL060 and SPA220. *Lepidostoma* is commonly found in less impacted urban creeks in the Bay Area.

Many common moderately intolerant taxa were completely absent from all sites, including the caddisfly family Hydropsychidae, the mayfly families Heptageniidae and Leptophlebiidae, and the stonefly *Malenka* sp. (Nemouridae).

San Pablo Creek Tributaries

The San Pablo Creek Tributary (SPT) sites were dominated by the same common intolerant taxa, but differences in taxa abundance suggest significant ecological differences between the tributary streams. Chironomid midges were the dominant taxa at the lower San Pablo Creek tributaries, but present at lower abundances at the upper sites. The most common taxon at the upper tributaries was the mayfly *Baetis* sp., which was present at very low numbers at the lower tributaries:

| | Chironomidae | <i>Baetis</i> sp. (Baetidae) |
|-------------------|--------------|------------------------------|
| Lower tributaries | | |
| SPA100 | 52% | <1% |
| SPA 110 | 59% | <1% |
| SPA130 | 64% | <1% |
| Upper tributaries | | |
| SPA200 | 20% | 48% |
| SPA240 | 30% | 48% |

Other EPT taxa were largely absent; no other mayflies or stoneflies were found at any site. The moderately tolerant caddisfly *Hydroptila* sp. was present in low numbers at SPA110 and SPA200.

Wildcat Canyon group

Benthic macroinvertebrate assemblages at sites in the Wildcat Canyon Group suggest some upstream water quality impacts at WIL130 and WIL100, but improved biological conditions at the lowest site in this group, WIL070. WIL130 is located in Tilden Park downstream of a golf course and botanic garden. WIL100 is located below the outlet of Jewel Lake, a small dammed pond. WIL070 is located approximately four miles downstream near the mouth of the canyon at the urban boundary.

The assemblage at WIL130 was dominated by the tolerant taxa Oligochaeta (48 percent), Chironomidae (32 percent), and *Baetis* sp. (15 percent). *Baetis* and chironomids were also abundant at WIL100 and WIL070, but Oligochaetes were present at those sites only in low abundances (4 percent and 8 percent, respectively). In contrast, the assemblage at WIL100 was dominated by the blackfly *Simulium* sp., a filter-feeding organism that is often observed at high abundances below reservoirs (41%).

The benthic assemblage at WIL070 was in better condition than at upstream sites, based on the higher taxa richness (26) and greater numbers of pollution intolerant taxa such as the caddisflies *Rhyacophila* sp., *Arctopsyche* sp., and *Agapetus* sp. The rare beetle *Elodes* sp. (Scirtidae) was also found at WIL070.

Bear Creek and Wildcat Headwaters group

Sites in the Bear Creek and Wildcat Headwaters (BCWH) group had the highest biological integrity in the San Pablo Creek/Wildcat Creek watershed. Taxonomic richness was high (>25) at

all sites, and the assemblages were generally dominated by EPT taxa. Metric values were similar or slightly lower than values from intermittent minimally disturbed sites. The flow status of the Bear Creek sites and Wildcat Creek headwater sites is unknown, although there is anecdotal evidence that the Bear Creek sites may dry up and the Wildcat Creek sites are perennial. There were fewer more-sensitive EPT organisms at SPA140 and WIL170 (<18 percent) than at the other sites (>47 percent), suggesting that there may be slightly more anthropogenic disturbance at these two sites, possibly due to impacts from the Tilden Park golf course and botanical garden (WIL170) and forest management activities (SPA140). Overall, however, the sites in this group exhibit low levels of water quality impacts.

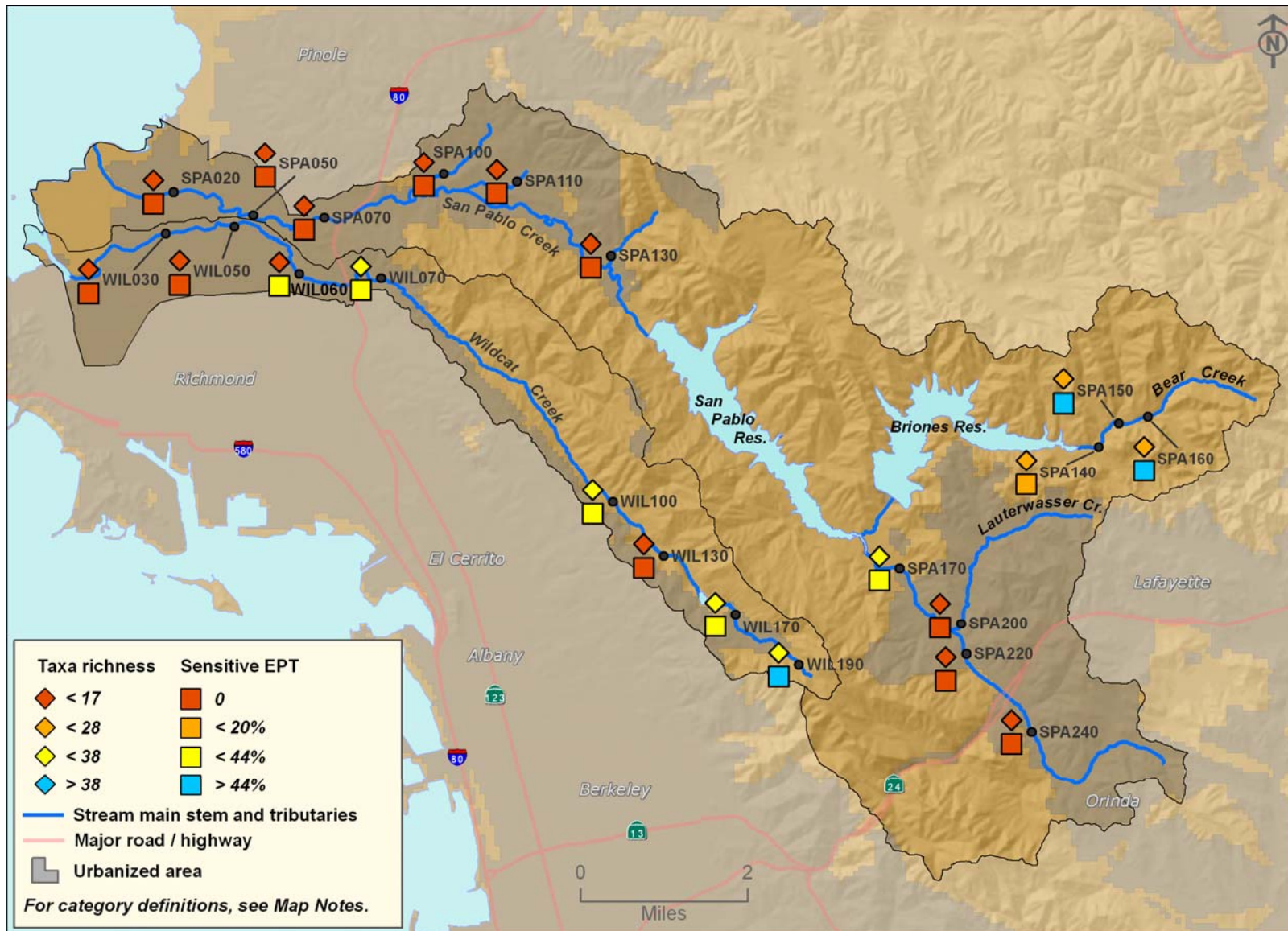


Figure 7-27. Map of Wildcat and San Pablo creeks benthic macroinvertebrates index scores

Table 7-12. Bioassessment sites in the Wildcat/San Pablo Creek watershed

| Site ID | Site Name | Stream |
|---------|---------------------------------|-----------------------|
| SPA020 | 3 rd St Bridge | San Pablo Creek |
| SPA050 | 20 th St. at Road 20 | San Pablo Creek |
| SPA070 | Cemetery Bridge | San Pablo Creek |
| SPA100 | Appian Way | Appian Way Creek |
| SPA110 | De Anza School | Wilkie Creek |
| SPA130 | Castro Ranch | Castro Creek |
| SPA140 | Xmas Tree Farm | South Fork Bear Creek |
| SPA150 | Briones 1 | Bear Creek |
| SPA160 | Briones 2 | Bear Creek |
| SPA170 | Bear Cr. Road | San Pablo Creek |
| SPA200 | Lauterwasser | Lauterwasser Creek |
| SPA220 | Orinda Village | San Pablo Creek |
| SPA240 | Camino Encinas | San Pablo Creek |
| WIL030 | 3rd St. | Wildcat Creek |
| WIL050 | Davis Park | Wildcat Creek |
| WIL060 | Vale Road | Wildcat Creek |
| WIL070 | Alvarado Park | Wildcat Creek |
| WIL100 | Jewel Lake Outlet | Wildcat Creek |
| WIL130 | Lone Oak | Wildcat Creek |
| WIL170 | Botanic Garden | Wildcat Creek |
| WIL190 | Possum Picnic Area | Wildcat Creek |

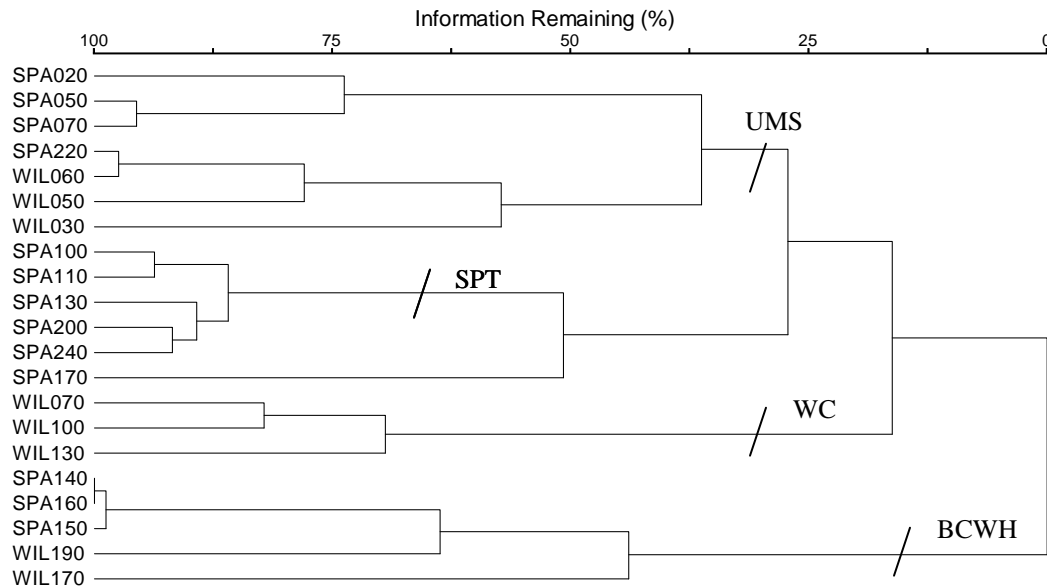


Figure 7-28. Cluster dendrogram sites in the Wildcat/San Pablo Creek watershed

The cluster dendrogram groups sites based on similarity of macroinvertebrate taxa present in the Wildcat Creek/San Pablo Creek watershed. Sites that are closely joined together have many similar taxa present, while groups with longer chain lengths are more dissimilar. The Information Remaining axis at top signifies how similar the sites are with respect to taxa presence. Abbreviations: UMS, Upper Mainstem; SPT, San Pablo Creek Tributaries; WC, Wildcat Canyon; BCWH, Bear Creek Headwaters.

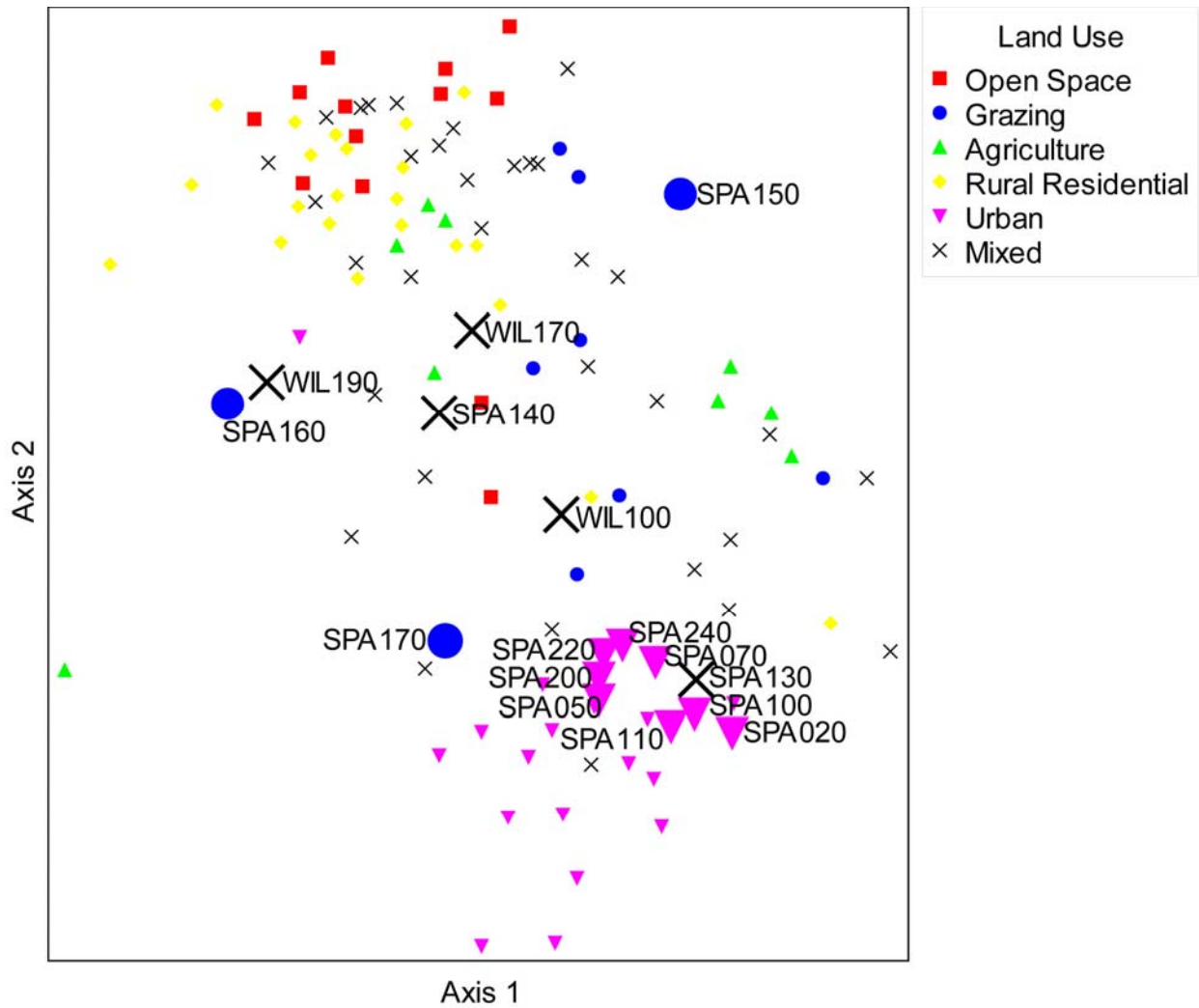


Figure 7-29. NMS ordination of San Pablo Creek/Wildcat Creek sites

The NMS ordination of sites by macroinvertebrate taxa abundance. Sites in the Wildcat/San Pablo watershed are labeled and symbols are enlarged. Note that several sites are not included in the ordination because the sites were sampled as part of a different study (see Breaux et al. 2005). For an explanation of ordination graphs, refer to Section 6.1.

Table 7-13. Biological metrics from the San Pablo Creek/Wildcat Creek watershed

| | SPA 020 | SPA 050 | SPA 070 | SPA 100 | SPA 110 | SPA 130 | SPA 140 | SPA 150 | SPA 160 | SPA 170 | SPA 200 | SPA 220 | SPA 240 | WIL 030 | WIL 050 | WIL 060 | WIL 070 | WIL 100 | WIL 130 | WIL 170 | WIL 190 |
|----------------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|
| Coleoptera Taxa | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 0 | 2 | 1 | 0 | 0 | 2 | 0 | 1 | 3 | 2 | 2 | 2 | 1 |
| Diptera Taxa | 2 | 2 | 5 | 5 | 5 | 7 | 9 | 6 | 8 | 6 | 3 | 4 | 4 | 4 | 2 | 3 | 4 | 3 | 2 | 5 | 9 |
| Ephemeroptera Taxa | 1 | 1 | 1 | 1 | 1 | 1 | 4 | 4 | 5 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 4 | 2 | 1 | 6 | 5 |
| Plecoptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 4 | 5 | 1 | 0 | 1 | 0 | 0 | 0 | 1 | 3 | 1 | 1 | 1 | 3 |
| Trichoptera Taxa | 1 | 1 | 1 | 0 | 1 | 1 | 4 | 4 | 4 | 1 | 1 | 2 | 0 | 0 | 0 | 2 | 5 | 5 | 1 | 4 | 5 |
| Non-Insect Taxa | 6 | 4 | 4 | 6 | 6 | 5 | 3 | 3 | 4 | 7 | 7 | 7 | 7 | 7 | 3 | 7 | 5 | 6 | 6 | 8 | 3 |
| Taxa Richness | 10 | 9 | 13 | 15 | 14 | 18 | 26 | 26 | 28 | 19 | 16 | 15 | 13 | 15 | 6 | 15 | 24 | 19 | 14 | 26 | 26 |
| EPT Taxa | 2 | 2 | 2 | 1 | 2 | 2 | 12 | 12 | 14 | 3 | 2 | 4 | 1 | 1 | 1 | 4 | 12 | 8 | 3 | 11 | 13 |
| % EPT | 18 | 32 | 12 | 1 | 1 | 0 | 39 | 78 | 79 | 56 | 48 | 46 | 48 | 2 | 1 | 17 | 47 | 23 | 16 | 34 | 84 |
| % Sensitive EPT | 0 | 0 | 0 | 0 | 0 | 0 | 18 | 51 | 47 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 22 | 3 | 0 | 15 | 61 |
| % Chironomidae | 32 | 25 | 45 | 52 | 59 | 64 | 26 | 12 | 11 | 36 | 20 | 21 | 30 | 57 | 43 | 18 | 37 | 28 | 32 | 41 | 10 |
| % Coleoptera | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 1 | 1 | 2 | 0 |
| % Oligochaeta | 40 | 36 | 9 | 39 | 36 | 3 | 0 | 0 | 2 | 3 | 3 | 22 | 8 | 34 | 42 | 62 | 8 | 4 | 48 | 0 | 0 |
| % Non-insect | 51 | 36 | 11 | 42 | 38 | 4 | 3 | 5 | 8 | 5 | 6 | 23 | 16 | 39 | 42 | 64 | 9 | 7 | 50 | 2 | 2 |
| % Intolerant | 0 | 0 | 0 | 0 | 0 | 0 | 14 | 40 | 33 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 21 | 3 | 0 | 7 | 46 |
| % Tolerant | 0 | 0 | 2 | 1 | 1 | 1 | 3 | 5 | 4 | 1 | 2 | 1 | 1 | 46 | 42 | 63 | 8 | 1 | 48 | 2 | 0 |
| Tolerance Value | 5.4 | 5.3 | 5.8 | 5.5 | 5.6 | 6.0 | 5.0 | 3.6 | 3.8 | 5.4 | 5.5 | 5.3 | 5.3 | 6.6 | 6.5 | 6.9 | 4.7 | 5.6 | 6.5 | 5.2 | 3.1 |
| % Predator | 0 | 0 | 2 | 2 | 1 | 3 | 4 | 9 | 7 | 4 | 1 | 1 | 1 | 7 | 2 | 1 | 10 | 1 | 1 | 4 | 20 |
| % Collector-filterer | 0 | 6 | 31 | 5 | 1 | 28 | 31 | 3 | 4 | 0 | 25 | 9 | 6 | 2 | 15 | 2 | 14 | 42 | 8 | 20 | 5 |
| %Collector-gatherer | 99 | 93 | 67 | 93 | 97 | 67 | 34 | 28 | 32 | 96 | 72 | 89 | 92 | 88 | 82 | 95 | 51 | 53 | 89 | 64 | 48 |
| % Scraper | 0 | 0 | 0 | 0 | 0 | 0 | 17 | 20 | 26 | 0 | 1 | 0 | 0 | 4 | 0 | 1 | 16 | 0 | 2 | 8 | 0 |
| % Shredder | 0 | 0 | 0 | 0 | 0 | 0 | 13 | 39 | 31 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | 3 | 0 | 4 | 28 |
| % Other | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |

7.4.4 Basic Water Quality Field Measurements

Stream temperatures exceeded some thresholds at a number of continuously monitored sites (Figure 7-, Figure 7-, and Appendix D). During the spring wet season (May 9-21, 2002) HWAT values exceeded the MWAT threshold for coho salmon (14.8°C) on Wildcat Creek at Richmond Parkway (WIL020) and on San Pablo Creek at Road 20 (SPA050). This threshold was also exceeded in the summer dry season at 5 stations. On Wildcat Creek at Jewel Lake Outlet (WIL100), from May 31-June 18, 2002, the temperature exceeded 14.8 for a total of 94.5 hours. The maximum measured temperature at San Pablo Creek at Road 20 (SPA050) exceeded the short-term maximum salmonid temperature of 24°C. This site, along with sites WIL100, SPA070, SPA200, and SPA220 had HWAT values ranging from 16.4 to 18.3 °C.

Temperature measurements in San Pablo Creek did not follow the regional pattern of increasing temperature with decreasing riparian habitat. Instead, the lower temperature values were recorded at sites below the San Pablo Dam and reservoir, and the higher temperatures were recorded at sites above the reservoir, which are on the other side of the coastal hills and thus in a warmer climate. Other relationships between physical habitat scores and continuously monitored variables were difficult to determine because of the low sample size.

During the spring and summer season deployments, DO concentrations failed to meet Basin Plan objectives (7 mg/L) on Wildcat Creek at Richmond Parkway (WIL020), Alvarado Park (WIL070), Jewel Lake Outlet (WIL100), and on San Pablo Creek at Road 20 (SPA050). San Pablo Creek at Cemetery Bridge (SPA070), Orinda Village (SPA220), Camino Encinas (SPA240), and Lauterwasser Creek (SPA200) all had DO values below the salmonid threshold during the dry season deployment (Figure 7-, Figure 7-, and Appendix D).

The Basin Plan maximum pH objective was exceeded by measurements of 8.56 at both Wildcat Creek at Richmond Parkway (WIL020) and San Pablo Creek at Orinda Village (SPA220; Figure 7-).

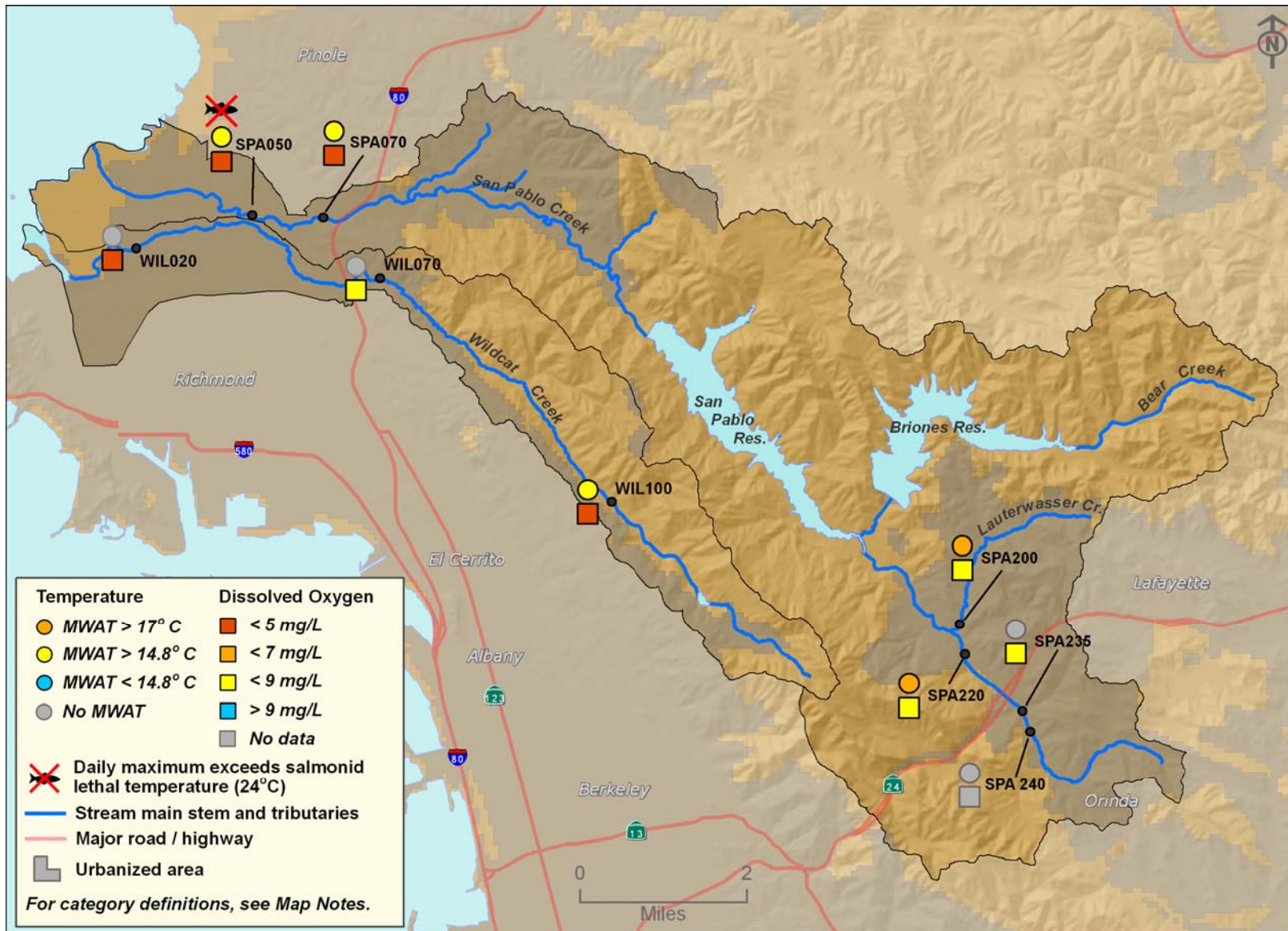


Figure 7-30. Map of temperature and DO levels in Wildcat and San Pablo creeks

Temperature--Wildcat/San Pablo Creeks

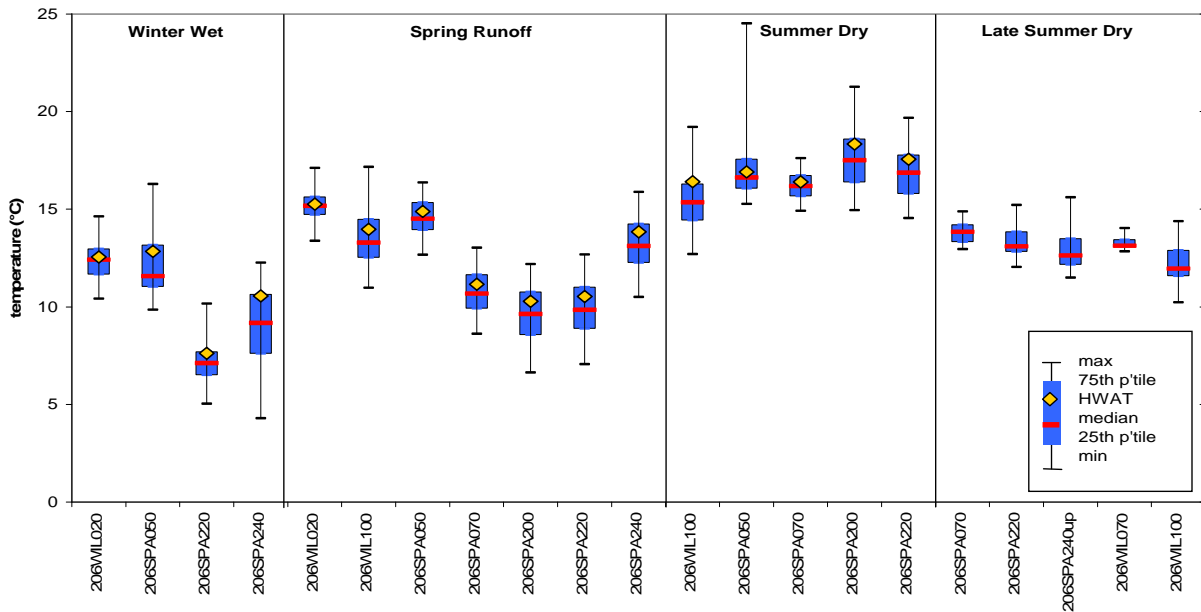


Figure 7-31. Temperature monitoring in Wildcat and San Pablo creeks

Dissolved Oxygen--Wildcat/San Pablo Creeks

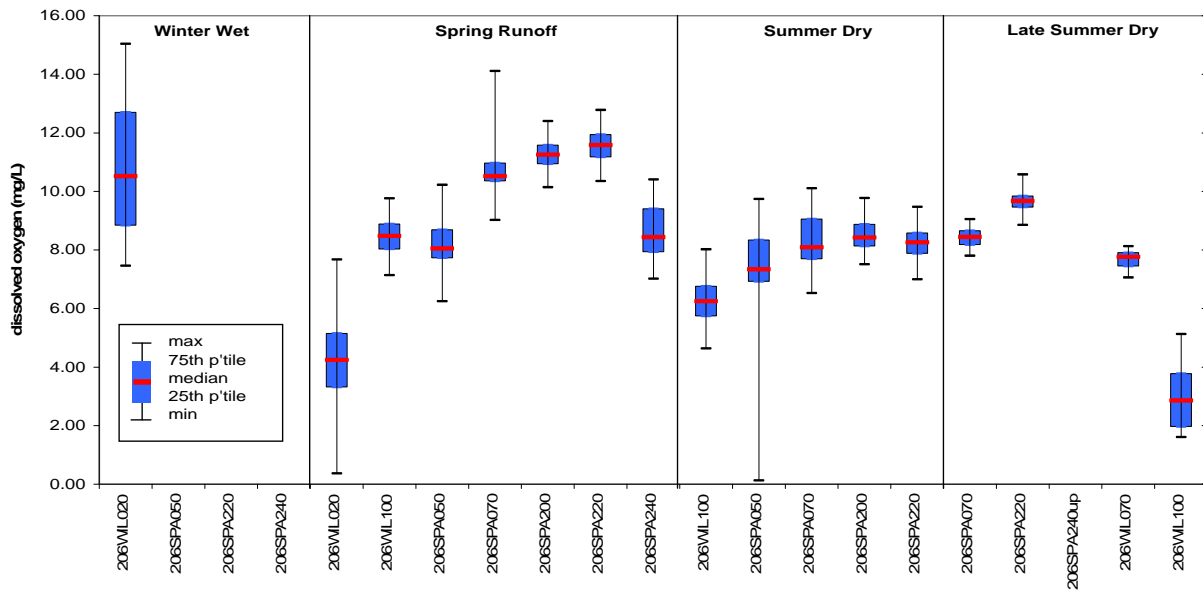


Figure 7-32. DO monitoring in Wildcat and San Pablo creeks

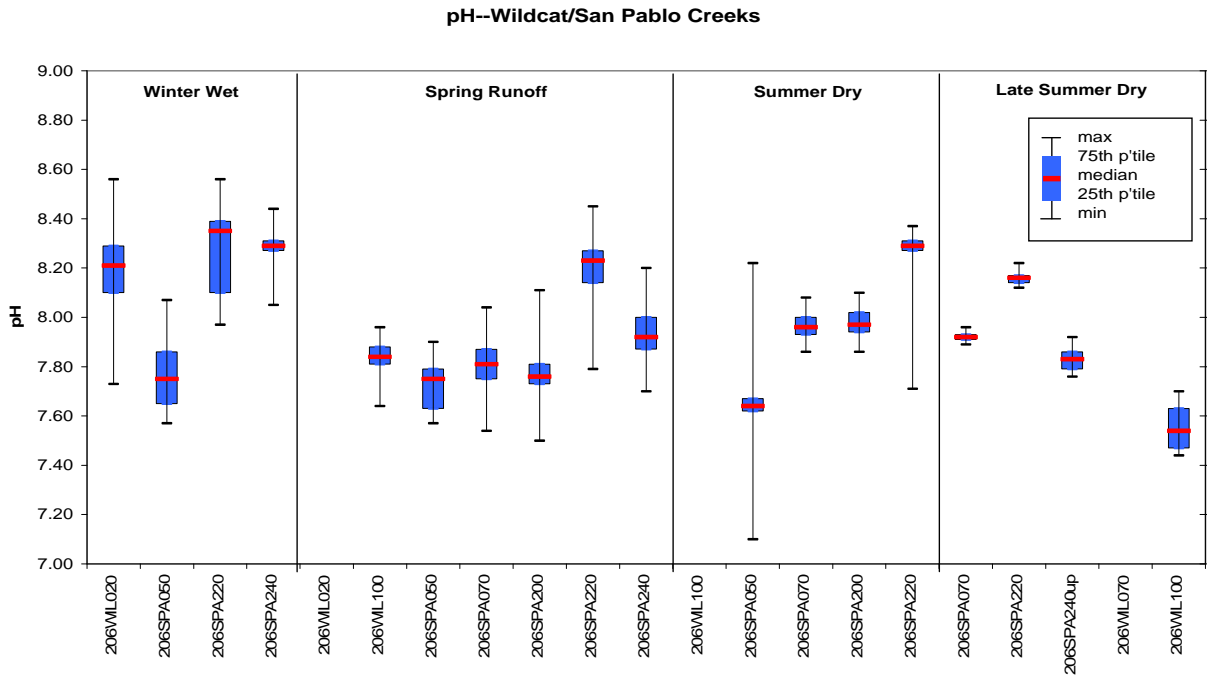


Figure 7-33. pH monitoring in Wildcat and San Pablo creeks

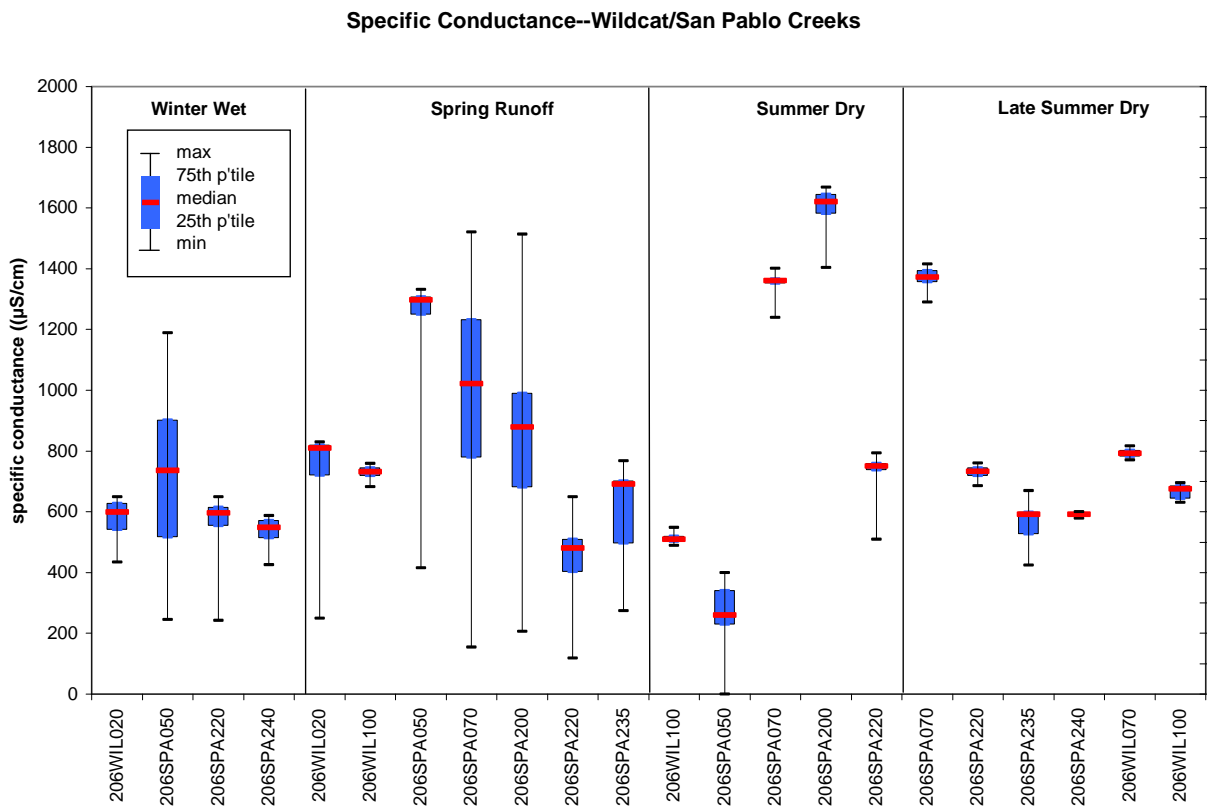


Figure 7-34. Specific conductance monitoring in Wildcat and San Pablo creeks

7.4.5 Water, Sediment, and Clam Tissue Chemistry

Water samples from the San Pablo and Wildcat Creek watersheds had generally low to moderate nitrate concentrations, with only three samples falling below the U.S. EPA reference guideline. The San Pablo Creek at 3rd St Bridge site (SPA020) had spring and wet season nitrate values of 1.2 and 1.7 mg/L (Figures 6-15 to 6-17; Appendix G). There were no apparent seasonal trends and only modest variation among sites.

Chlorophyll-a values were high relative to other watersheds in the region, with most spring and dry season samples measuring above the 1.78 µg/l guideline value (Appendix G). There was one extremely high chlorophyll-a value: 92 µg/L was recorded in a dry season sample from Wildcat Creek at the Richmond Parkway site (WIL020). This value is about four times greater than the second highest regional value (20 µg/L in Walker Creek). Nitrate and orthophosphate concentrations were low in this sample, and total phosphorus was not exceptionally high.

Total phosphorus measurements were the highest in the region, with an average of 0.24 mg/L. This was about twice the average concentration in the watershed with the next highest measurements, Stevens/Permanente. All samples from San Pablo/Wildcat were substantially above the 0.030 U.S. EPA guideline value.

Orthophosphate values were also the highest in the region, averaging 0.27 mg/L. The highest orthophosphate concentration was 1.99 mg/L at Lauterwasser Creek (SPA200), the highest value measured in the region. All of the orthophosphate guideline exceedances came from San Pablo Creek sites. Phosphorus measurements from the two Wildcat Creek sites were below the 0.030 mg/L total phosphorus guideline value.

Trace metals in water were measured at five sites: Wildcat Creek at Richmond Parkway (WIL020); San Pablo Creek at 3rd St Bridge (SPA020); Cemetery Bridge (SPA070); Orinda Village (SPA220); and Lauterwasser Creek (SPA200). Values were much higher in the dry season than in spring (Appendix G). No Basin Plan objectives for aquatic life protection were exceeded for any metals in samples from this watershed. Total manganese was very high at all sites, though there are no aquatic life water quality objectives (WQOs) for this metal. Concentrations ranged from 109 to 1356 µg/L, with the high value from a Wildcat Creek sample (WIL020). (Note that the next highest concentration measured in the region was 67 µg/L in Butano Creek.) Nickel concentrations in water samples were higher in this watershed than elsewhere in the region. The mercury concentration in water at Richmond Parkway (WIL020) for the dry season was the highest in this study (23.1 ng/L). The results for mercury, nickel, and manganese may be indicative of urban sources and transport mechanisms such as atmospheric deposition, traffic from industrial sites, illicit discharge, sheet flow, wind dust blow and spills.

Organic contaminants were measured at five sites. The ELISA measurement of the spring season sample from Wildcat Creek at Richmond Parkway (WIL020) exceeded the Central Valley Regional Water Quality Control Board's aquatic life objective for chlorpyrifos (0.057 µg/L). The dry season measurement from San Pablo Creek at 3rd St Bridge (SPA020) exceeded both the Central Valley Water Board's objective for chlorpyrifos (0.079 µg/L by ELISA) and the San Francisco Bay Basin Plan objective for diazinon (0.174 µg/L by ELISA and 0.120 µg/L by

GCMS). Similarly, WQOs for chlorpyrifos (0.071 µg/L) and diazinon (0.288 µg/L) were exceeded in ELISA measurements of the dry season sample from Cemetery Bridge (SPA070). Land use in areas draining to all of these sites was characterized as urban.

Sediments were collected for analysis from two sites, WIL020 and SPA020. No organic compounds exceeded available sediment quality guidelines at either site. WIL020 had sediment concentrations of chromium above the TEC and nickel above the PEC guidelines, while SPA020 had chromium, mercury, and nickel above the TEC guidelines. The sediment quality guideline quotient values for these sites were relatively low, at 0.14 and 0.11, respectively.

Tissues of clams deployed at lower San Pablo Creek at 3rd St Bridge (SPA020) had about average concentrations of copper, mercury, selenium, and PCBs, compared to the rest of the region (Figures 6-22 through 6-24). However, concentrations were higher than average for the pesticides chlordane, DDT, dieldrin, and oxadiazon (Figures 6-25 through 6-27).

7.4.6 Water and Sediment Toxicity

Toxicity tests were conducted on water samples from Wildcat Creek at Richmond Parkway (WIL020); San Pablo Creek at 3rd St Bridge (SPA020), Cemetery Bridge (SPA070), Orinda Village (SPA220); and Lauterwasser Creek (SPA200); and on sediment samples from WIL020 and SPA 020. There were significant decreases in algal population growth in the spring sample from SPA070, and in both spring and dry season samples from SPA020. SPA020 had elevated clam tissue concentrations of oxadiazon. No significant toxicity was observed in any other water samples (Figure 7-).

Amphipod growth in the sediment sample from WIL020 was significantly less than in controls. While growth integrates a number of physiological processes, the relationship of the amphipod growth endpoint to specific stressors is not well established. Concentrations of contaminants for which guideline values have been established were relatively low in this sediment sample.

Toxicity Tests in Wildcat/San Pablo Creeks Watershed--Response Relative to Control

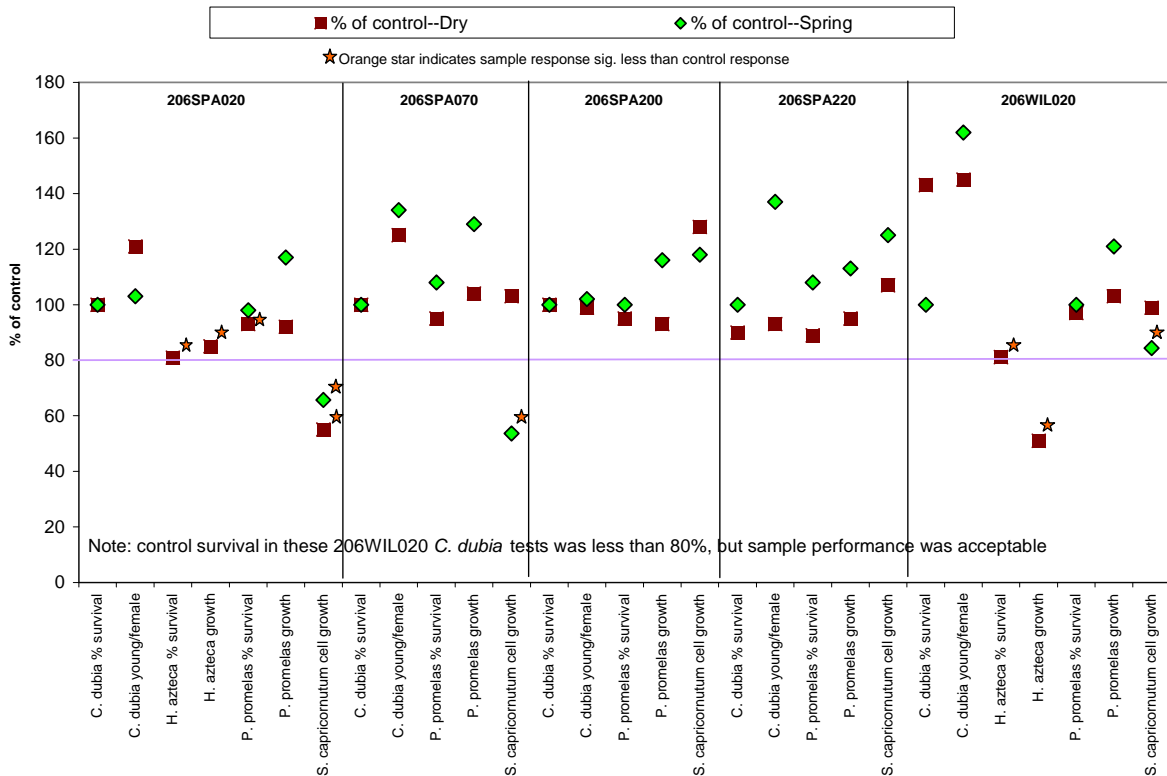


Figure 7-35. Toxicity tests in Wildcat and San Pablo creeks

Results of toxicity tests are shown with three species in water samples and one species in sediment samples from Wildcat and San Pablo Creeks. See text for description of species and endpoints. Stars indicate statistically significant differences from test controls. Samples with stars that were also below 80 percent of the control were considered toxic for this assessment.

7.4.7 Coliform Bacteria

Bacterial counts were recorded in water samples from four sites in the watershed (Figures 6-28 through 6-30). Basin Plan objectives for all three indicators (*E. coli*, fecal coliforms, and total coliforms) were exceeded at three sites: Wildcat Creek at Alvarado Park (WIL070), San Pablo Creek at San Pablo City Park (SPA060), and Bear Creek at Briones 1 (SPA150). Basin Plan objectives for total coliforms only were exceeded at Wildcat Creek at Lone Oak (WIL130).

7.5 Suisun Creek Watershed

7.5.1 Sites of Concern

Exceedances noted in the summary table (Table 7-14) are discussed in the relevant sub-section for this watershed. A weight-of-evidence approach was used to evaluate sites of concern. Individual exceedances of parameters, particularly those without regulatory objectives, are considered in context. In general, nickel and chromium levels are geologically elevated in the Bay Area, particularly where there are serpentine deposits. Nutrient levels (nitrate and phosphorus) are above U.S. EPA guidelines at most sites, but may still be relatively low.

Bioassessment scores indicated fair condition for benthic invertebrate communities throughout the watershed. Temperature and dissolved oxygen measurements exceeded thresholds at most sites, and pH was elevated at over half of the continuously monitored sites. There was no toxicity in this watershed and no individual sites stood out as being of specific concern. It is interesting to note, however, that the highest concentration of copper in tissue reported in this study was at Rockville (SUI020) on Suisun Creek, upstream from Suisun Slough. Consistent toxicity attributable to copper has been found in Grizzly Bay near the mouth of Suisun Slough (Phillips *et al.* 2003).

Table 7-14. Summary of sites with exceedances in Suisun Creek watershed

| <i>Analyte</i> | <i>Standard</i> | <i># of sites</i> |
|--|-------------------|-------------------|
| FIELD MEASURES of 7 sites sampled | | |
| Temperature | maximum, salmonid | 4 |
| | MWAT, steelhead | 6 |
| Oxygen, dissolved | minimum, COLD | 5 |
| | minimum, WARM | 3 |
| | 3-month median | 2 |
| pH | range | 4 |
| CONVENTIONAL WATER QUALITY of 8 sites sampled | | |
| Nitrate as N | maximum | 7 |
| Phosphorus, total as P | maximum | 8 |
| SEDIMENT CHEMISTRY of 1 site sampled | | |
| <i>Metals</i> | | |
| Arsenic | TEC | 1 |
| Chromium | TEC | 1 |
| Copper | TEC | 1 |
| Nickel | TEC | 1 |
| TISSUE CHEMISTRY of 1 site sampled | | |
| Copper | EDL 85 | 1 |
| PCB 1248 aroclor | EDL 85 | 1 |

7.5.2 Water Quality in Relation to Land Use

Land use throughout the watershed is a combination of grazing, agriculture, and mixed use. As there was minimal variability among sites in terms of most measured parameters, relationships with specific land uses could not be resolved definitively. Because degraded stream habitat conditions exist in much of the watershed, improved riparian management is recommended to reduce the potential for temperature and dissolved oxygen stress on aquatic populations.

7.5.3 Macroinvertebrate Assemblages and Physical Habitat

Two major benthic assemblages were identified in the Suisun Creek watershed based on the cluster dendrogram (Figure 6-58).

- The Lower Suisun Creek (LS) group includes sites on the mainstem of Suisun Creek below Lake Curry (SUI010, SUI020, SUI050, SUI060, SUI130).
- The Wooden Valley (WV) group includes sites on Wooden Valley Creek (SUI110, SUI210) and White Creek (SUI180), an intermittent tributary to Wooden Valley Creek.

Although the site on Suisun Creek above Lake Curry (SUI260) is located near other Suisun Creek sites on the ordination plot (Figure 6-59), it is the most dissimilar site on the cluster dendrogram (Figure 6-58) and is discussed separately.

Lower Suisun Creek group

Biological metrics were fairly consistent among the five sites, but metric values were outside the range observed at minimally disturbed sites. Sites in the Lower Suisun Creek group exhibit a moderate level of biological integrity. Benthic assemblages have retained many features characteristic of natural systems, but taxonomic diversity has been reduced relative to minimally disturbed streams.

Common tolerant taxa made up 81–89 percent of organisms at these sites, a relatively high proportion relative to minimally disturbed sites but less than sites in urban areas. While common tolerant taxa were the four most abundant taxa, intolerant taxa such as the mayfly family Ephemerellidae and the caddisfly *Agapetus* sp. were fairly abundant:

| | |
|---------------------------------------|-----|
| Chironomidae | 39% |
| <i>Baetis</i> sp. (Baetidae) | 26% |
| <i>Simulium</i> sp. (Simuliidae) | 12% |
| Oligochaeta | 4% |
| Ephemerellidae | 3% |
| <i>Agapetus</i> sp. (Glossosomatidae) | 2% |

Many other intolerant taxa were present at low abundances, including *Serratella* sp. (Ephemerellidae), the mayfly family Heptageniidae, *Malenka* sp. (Plecoptera: Nemouridae), *Isoperla* sp. (Plecoptera: Perlodidae), *Wormaldia* sp. (Trichoptera: Philopotamidae), *Tinodes* sp. (Diptera: Psychomyiidae), and *Gumaga* sp. (Trichoptera: Sericostomatidae).

Rare taxa found at these sites include the uncommon baetid *Fallceon quilleri*, the riffle beetle *Dubiraphia* sp., and the amphipod *Corophium* sp. (Corophiidae).

Wooden Valley group

Biological metrics from the site on White Creek (SUI180), such as percent EPT and tolerance value, were within the range of values from minimally disturbed intermittent streams, while the taxa richness metric was slightly outside of the range. Taxa richness, percent EPT, and Diptera taxa richness metric values at SUI110 were close to the lower end of values from minimally disturbed intermittent sites, but the tolerance value metric was more dissimilar. Metrics at SUI210 were significantly different from minimally disturbed conditions. Overall, metric scores suggest that the biological integrity of White Creek was high, but possibly slightly impacted, while Wooden Valley Creek was moderately impacted by human disturbances.

The most abundant taxa in this group of sites were:

| | |
|---|-----|
| Chironomidae | 25% |
| <i>Baetis</i> sp. (Baetidae) | 21% |
| <i>Simulium</i> sp. (Simuliidae) | 19% |
| <i>Ephemerella</i> sp. (Ephemerellidae) | 9% |
| <i>Paraleptophlebia</i> sp. (Leptophlebiidae) | 7% |
| Planorbidae | 3% |

Other common intolerant taxa included *Drunella* sp., *Serratella* sp., *Malenka* sp., and *Wormaldia* sp. Stream flow at two of the sites (SUI110 and SUI180) is known to be intermittent, while the status of the third site is uncertain. Noticeably absent from all three sites were long-lived taxa, perlid stoneflies, and taxa with fully aquatic life cycles, such as riffle beetles (Elmidae).

Upper Suisun Creek site

The taxa richness and Diptera taxa richness values were within the range of minimally disturbed conditions for intermittent streams, but because of the abundance of non-EPT tolerant taxa, such as chironomids and worms, the percent EPT and tolerance value metrics were considerably different from minimally disturbed conditions. Low physical habitat scores for epifaunal substrate (5), embeddedness (5), and sediment deposition (5) suggest that fine sediment may be partly responsible for the high relative abundances of chironomids and worms

Abundant taxa in the benthic assemblage from upper Suisun Creek (SUI260) included:

| | |
|--|-----|
| Chironomidae | 51% |
| Oligochaeta | 15% |
| Torrenticolidae (Acari) | 8% |
| <i>Serratella</i> sp. (Ephemerellidae) | 7% |
| <i>Simulium</i> sp. (Simuliidae) | 4% |
| Hygrobatidae (Acari) | 4% |

Four genera of Baetidae were present, including *Centroptilum/ Procloeon*, *Baetis* sp., *Dipheter hageni*, and *Fallceon quilleri*. Several intolerant taxa were found, including the stonefly *Isoperla* sp. (Perlodidae), the caddisflies *Agapetus* sp. (Glossosomatidae), *Glossosoma* sp. (Glossosomatidae), and *Wormaldia* sp. (Philopotamidae), and the tipulids *Hexatoma* sp. and *Molophilus* sp.

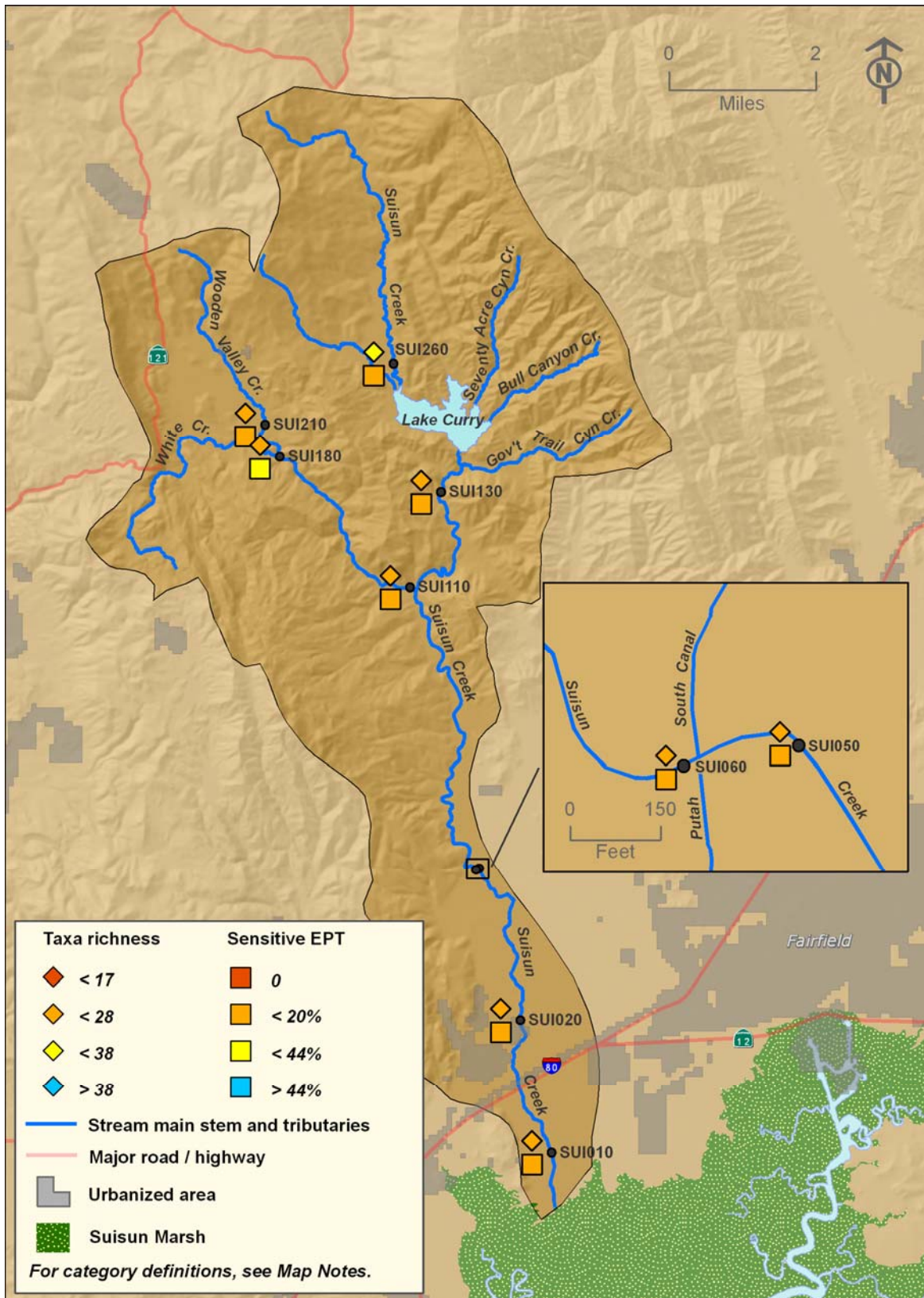


Figure 7-36. Map of Suisun Creek benthic macroinvertebrates index scores

Table 7-15. Bioassessment sites in the Suisun Creek watershed

| Site ID | Site Name | Stream |
|---------|--------------------------------|---------------------|
| SUI010 | Cordelia | Suisun Creek |
| SUI020 | Rockville | Suisun Creek |
| SUI050 | Putah South Canal - Downstream | Suisun Creek |
| SUI060 | Putah South Canal - Upstream | Suisun Creek |
| SUI110 | Wooden Valley | Wooden Valley Creek |
| SUI130 | Lake Curry Road | Suisun Creek |
| SUI180 | White Creek | White Creek |
| SUI210 | East Wooden Valley | Wooden Valley Creek |
| SUI260 | Upper Suisun | Suisun Creek |

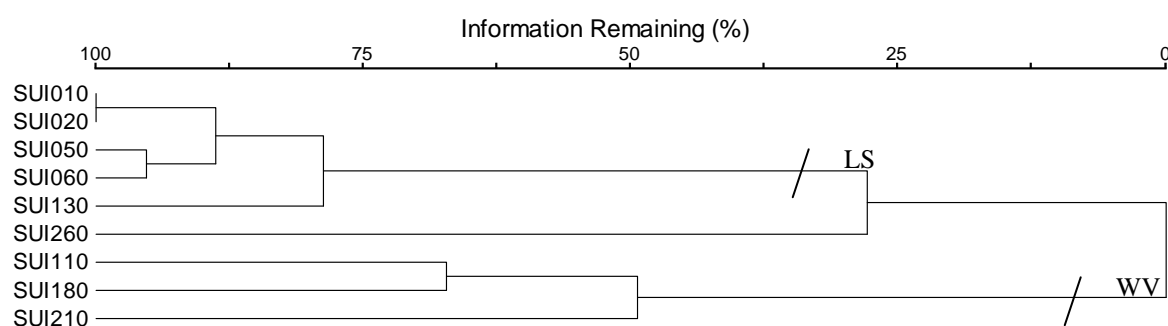


Figure 7-37. Cluster dendrogram of sites in the Suisun Creek watershed

This cluster dendrogram groups sites based on similarity of macroinvertebrate taxa present in the Suisun Creek watershed. Sites that are closely joined together have many similar taxa present, while groups with longer chain lengths are more dissimilar. The Information Remaining axis at top signifies how similar the sites are with respect to taxa presence. Abbreviations: LS, Lower Suisun; WV, Wooden Valley.

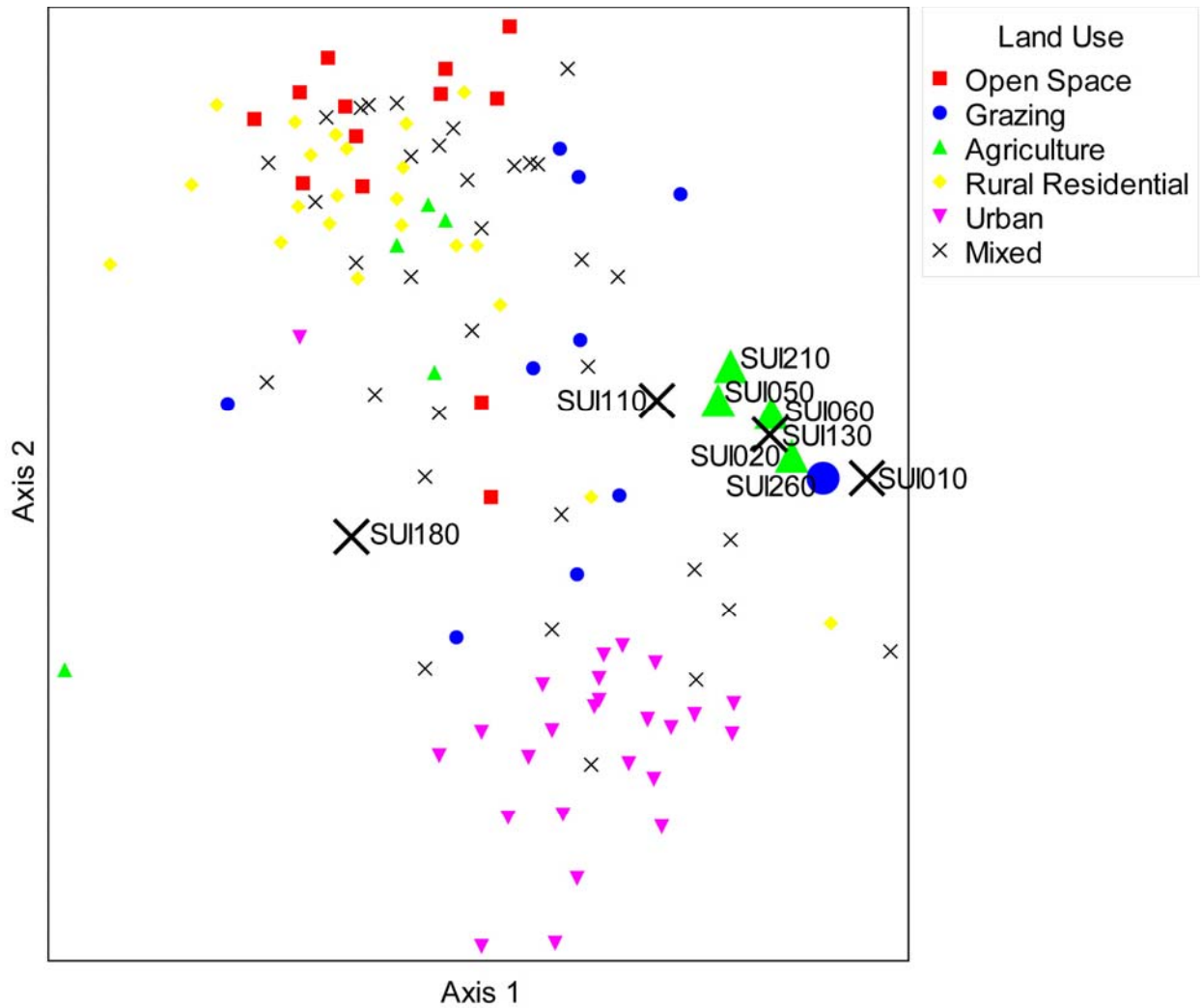


Figure 7-38. NMS ordination of sites in the Suisun Creek watershed

The NMS ordination of sites by macroinvertebrate taxa abundance. Sites in the Suisun Creek watershed are labeled and symbols are enlarged. For an explanation of ordination graphs, refer to Section 6.1.

Table 7-16. Biological metrics from the Suisun Creek watershed

| | SUI010 | SUI020 | SUI050 | SUI060 | SUI110 | SUI130 | SUI180 | SUI210 | SUI260 |
|------------------------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| Coleoptera Taxa | 2 | 1 | 1 | 0 | 1 | 1 | 0 | 1 | 1 |
| Diptera Taxa | 2 | 3 | 2 | 2 | 6 | 5 | 3 | 4 | 9 |
| Ephemeroptera Taxa | 4 | 5 | 6 | 4 | 7 | 3 | 8 | 3 | 6 |
| Plecoptera Taxa | 1 | 1 | 2 | 1 | 1 | 2 | 2 | 2 | 1 |
| Trichoptera Taxa | 6 | 6 | 6 | 6 | 5 | 9 | 2 | 3 | 4 |
| Non-Insect Taxa | 9 | 9 | 5 | 7 | 7 | 6 | 10 | 5 | 7 |
| EPT Taxa | 11 | 12 | 14 | 11 | 13 | 14 | 12 | 8 | 11 |
| Taxa Richness | 24 | 25 | 22 | 20 | 27 | 26 | 25 | 19 | 28 |
| % EPT | 25 | 36 | 35 | 43 | 57 | 52 | 61 | 25 | 11 |
| % Sensitive EPT | 3 | 9 | 6 | 8 | 11 | 11 | 37 | 4 | 8 |
| % Chironomidae | 52 | 42 | 37 | 37 | 18 | 23 | 23 | 34 | 51 |
| % Coleoptera | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| % Oligochaeta | 11 | 5 | 2 | 0 | 1 | 3 | 3 | 0 | 15 |
| % Non-insect | 18 | 18 | 6 | 4 | 4 | 6 | 15 | 1 | 30 |
| % Intolerant | 1 | 7 | 5 | 7 | 11 | 11 | 36 | 3 | 9 |
| % Tolerant | 5 | 3 | 3 | 3 | 2 | 2 | 2 | 1 | 8 |
| Tolerance Value | 5.6 | 5.1 | 5.4 | 5.2 | 5.0 | 4.9 | 3.9 | 5.4 | 5.5 |
| % Predator | 5 | 3 | 3 | 3 | 1 | 3 | 5 | 4 | 16 |
| % Collector-filterer | 10 | 7 | 25 | 19 | 22 | 20 | 1 | 39 | 4 |
| %Collector-gatherer | 83 | 84 | 67 | 72 | 73 | 75 | 80 | 53 | 78 |
| % Scraper | 0 | 4 | 3 | 5 | 4 | 1 | 12 | 2 | 1 |
| % Shredder | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 |
| % Other | 2 | 1 | 2 | 1 | 0 | 1 | 2 | 0 | 0 |
| % Common Tolerant | 88 | 78 | 89 | 86 | 83 | 81 | 42 | 74 | 74 |
| Common Intolerant Taxa | 2 | 2 | 3 | 4 | 5 | 3 | 6 | 4 | 3 |

7.5.4 Basic Water Quality Field Measurements

Stream temperatures exceeded some thresholds at all continuously monitored sites (see Figure 7-, Figure 7-, and Appendix D). HWAT values exceeded the MWAT threshold for coho salmon (14.8°C) during the spring season at the following sites:

- Suisun Creek at Rockville (SUI020)
- Putah South Canal - Downstream (SUI050)
- Capp Confluence (SUI090)
- Middle Suisun (SUI125)
- Wooden Valley Creek at Wooden Valley (SUI110)
- Above White Creek (SUI185)
- White Creek (SUI180)

All of these sites except SUI180, which wasn't monitored in the summer dry period, had HWAT values above the higher MWAT threshold for steelhead (17°C) in the dry season. Four sites (SUI020, SUI110, SUI050, and SUI185) had recorded maximum temperatures above 24°C, the maximum temperature for salmonid survival.

Minimum dissolved oxygen concentrations in spring fell below 9 mg/L at all of the above sites except SUI185 (see Figure 7-, Figure 7-, and Appendix D). The median percent saturation fell below 80 percent in dry season measurements at SUI020 and SUI050. Dry season measurements failed to meet the coldwater Basin Plan objectives for DO at SUI020, SUI110, SUI050, SUI090, and SUI185.

The Basin Plan objective for maximum pH (8.5) was exceeded at SUI050, SUI110, SUI125, and SUI185 (Figure 7-).

Physical habitat scores were available for three of these sites (SUI020, SUI050, and SUI110). The two downstream sites had poor scores for channel alteration and riparian zone width, but moderate to good scores for canopy cover and bank vegetation. The upstream site (SUI110) had the opposite: better scores for channel alteration and riparian zone width, but poor scores for canopy cover and bank vegetation. All three had similar HWAT values and DO violations.

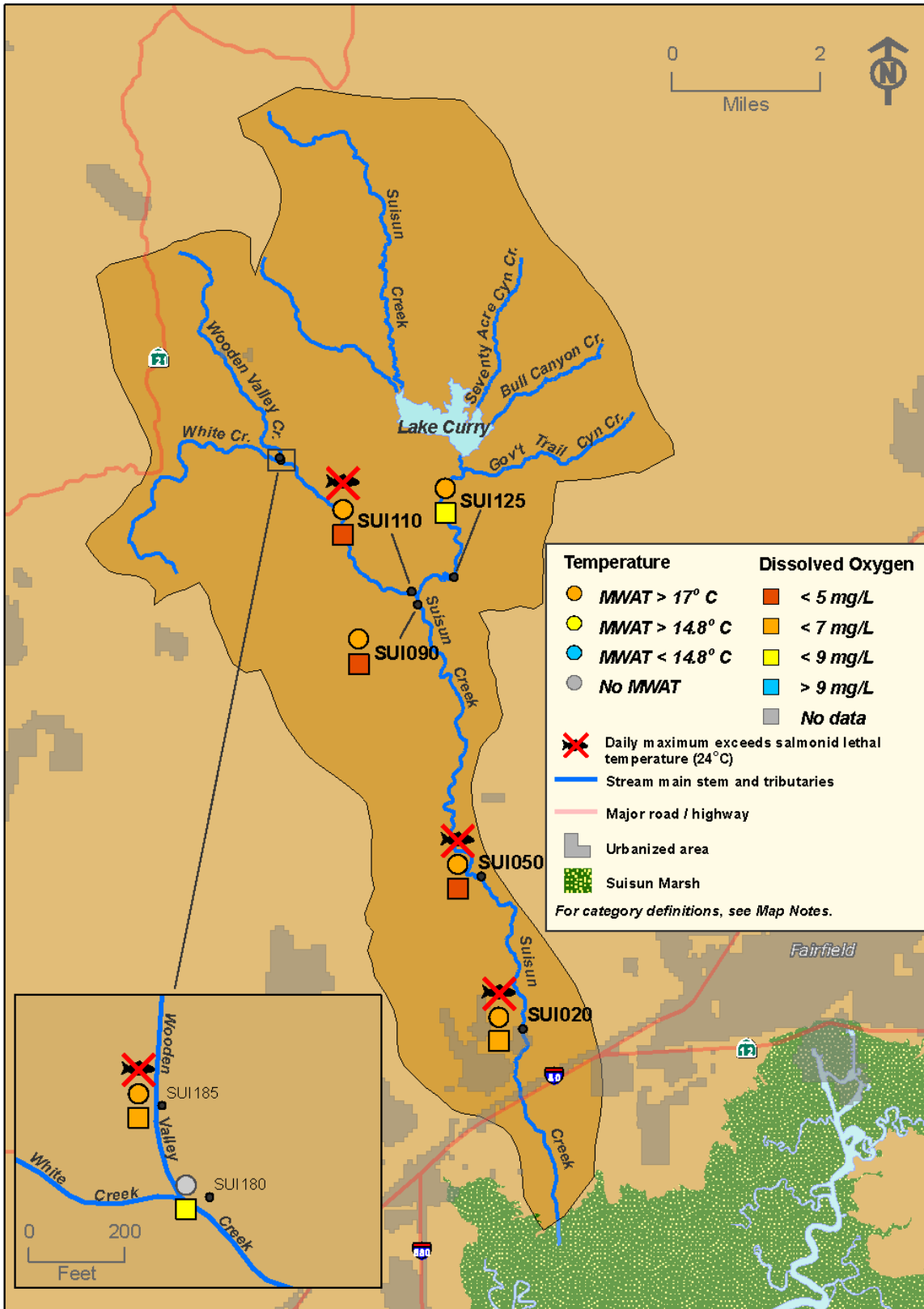


Figure 7-39. Map of temperature and DO levels in Suisun Creek

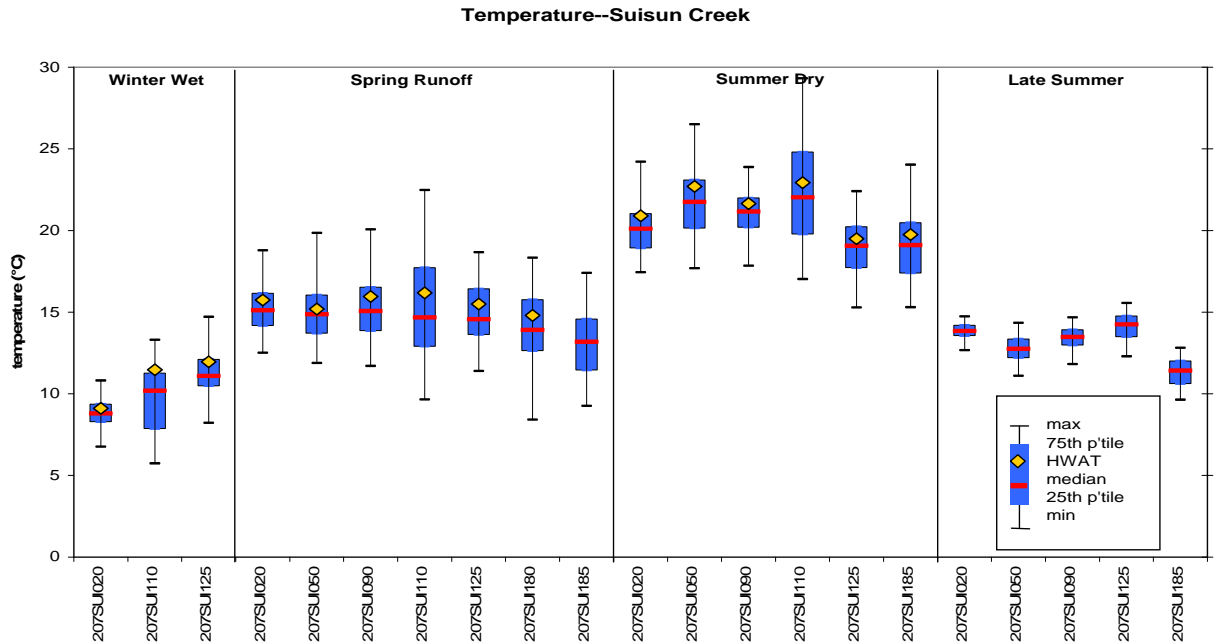


Figure 7-40. Temperature monitoring in Suisun Creek

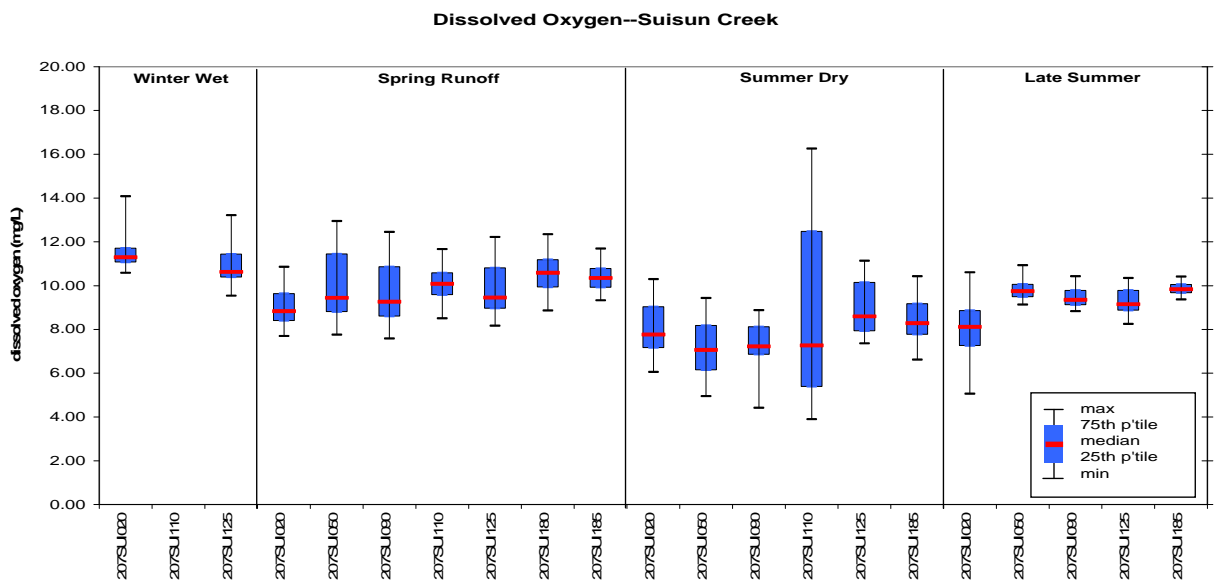


Figure 7-41. DO monitoring in Suisun Creek

Data were rejected for SUI110 in the winter.

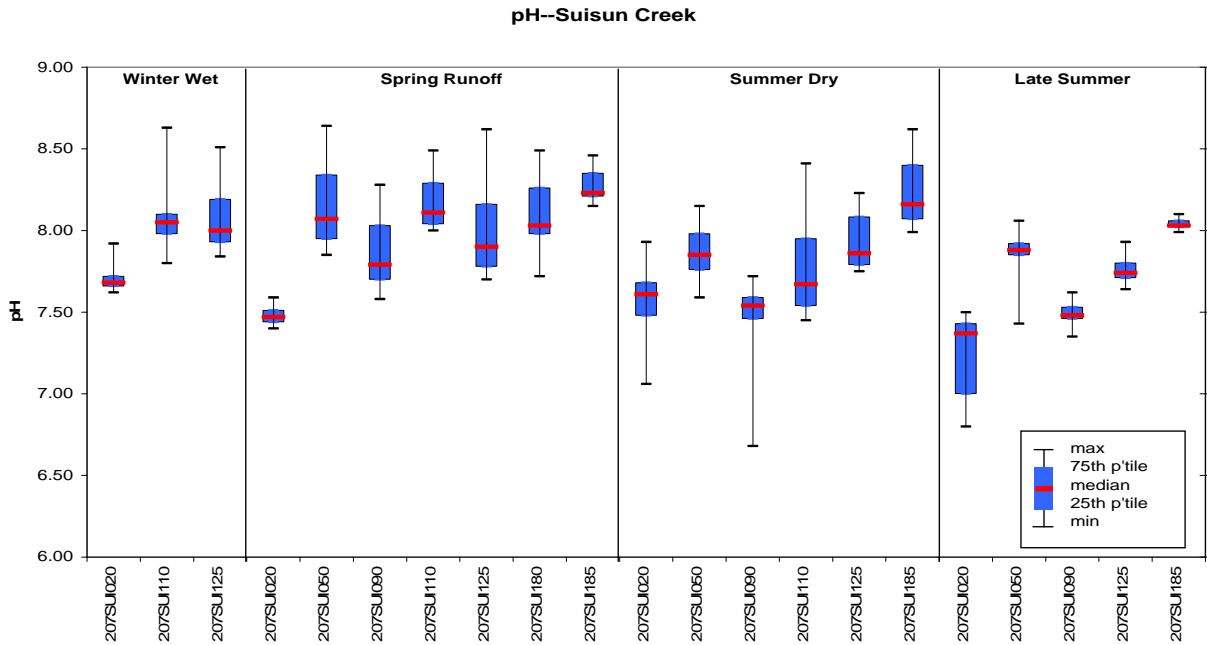


Figure 7-42. pH monitoring in Suisun Creek

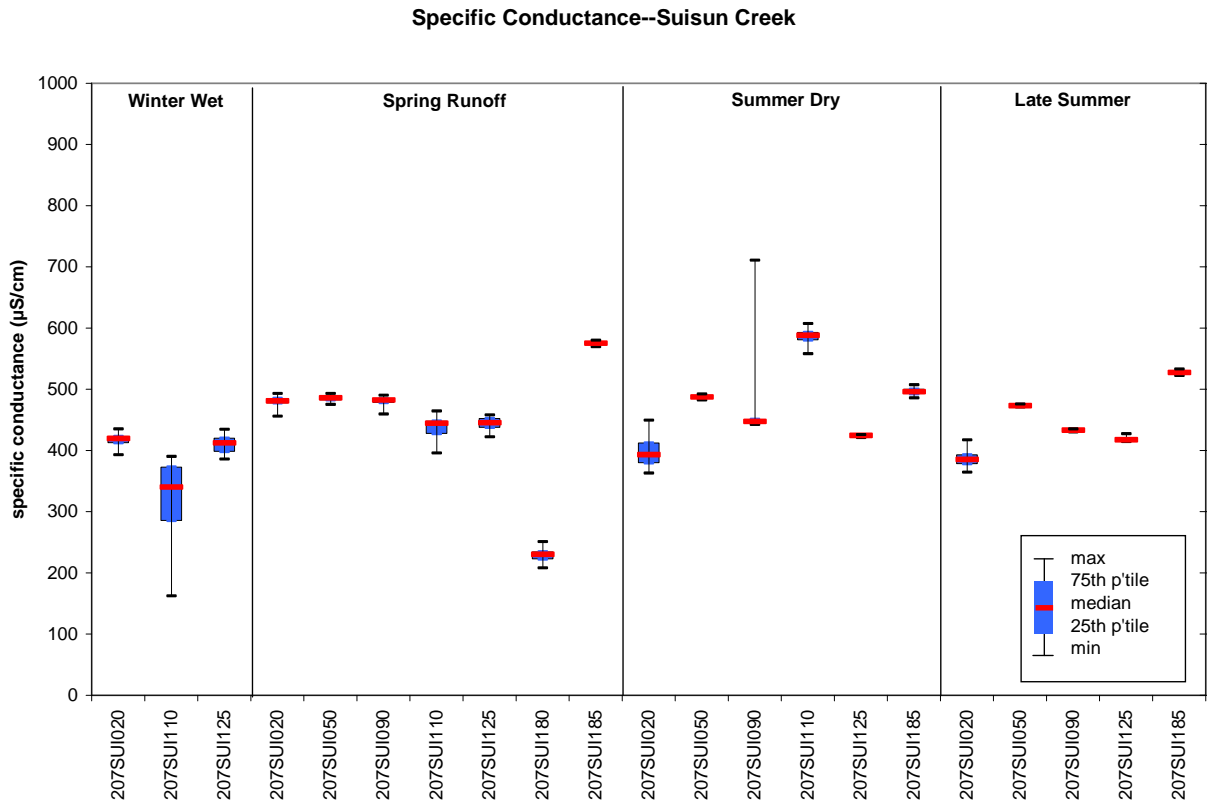


Figure 7-43. Specific conductance monitoring in Suisun Creek

7.5.5 Water, Sediment, and Clam Tissue Chemistry

Water samples from the Suisun Creek watershed had relatively low nitrate concentrations, although most exceeded 0.16 mg/L, the U.S. EPA reference guideline (Figure 6-15, Figure 6-16, Figure 6-17, and Appendix G).

Chlorophyll-a values were also generally low, only exceeding the 1.78 µg/L guideline value in the wet season sample from SUI125, a spring sample from SUI010, and dry season samples from SUI 060, SUI110, and SUI130 (Appendix G).

Most total phosphorus measurements were equal to the minimum reporting limit, with a few high values such as the dry season SUI130 value of 0.188 mg/L. Most orthophosphate values modestly exceeded the U.S. EPA total phosphorus guideline.

No Basin Plan objectives for metals or organics in water were exceeded in samples from this watershed. Trace metals were measured in water sampled at four sites.

Sediments were collected for analysis from only one site, the Suisun Creek site at Rockville (SUI020). No probable effects concentrations (PECs) were exceeded in this sample, but the lower threshold effects concentrations (TECs) were exceeded for arsenic, chromium, copper, and nickel. The sediment quality guideline quotient (SQGQ) value was low at 0.11 (Appendix H).

Tissues of clams deployed at the Suisun Creek site at Rockville (SUI020) had values near the regional average for mercury, selenium and DDT, and the highest copper concentrations region-wide. (This copper value was still less than 50 percent higher than the control, so the significance of this result is uncertain.) No detectable pesticides except DDT were found.

7.5.6 Water and Sediment Toxicity

Water toxicity was tested with three species in samples from Suisun Creek at Cordelia (SUI010), at Rockville (SUI020), at Putah South Canal – Upstream (SUI060), and Wooden Valley Creek at Wooden Valley (SUI110). Sediment toxicity was tested in a sample from the Rockville site (SUI020). No significant toxicity was detected in any of these samples.

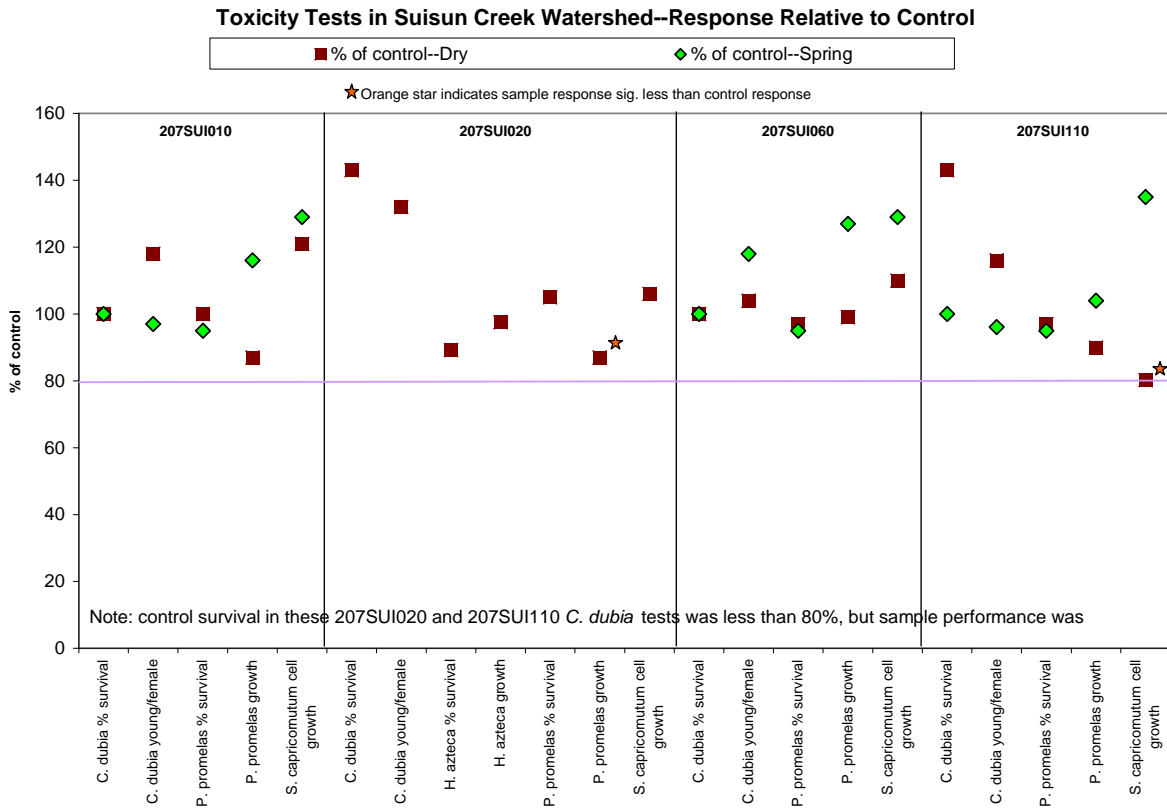


Figure 7-44. Toxicity tests in Suisun Creek

The graph shows results of toxicity tests with three species in water samples and one species in sediment samples from Suisun Creek. See text for description of species and endpoints. Stars indicate statistically significant differences from test controls. Samples with stars that were also below 80 percent of the control were considered toxic in this assessment.

7.5.7 Coliform Bacteria

Coliforms were not measured in the Suisun Creek watershed.

7.6 Arroyo Las Positas Watershed

7.6.1 Sites of Concern

Exceedances noted in the summary table (Table 7-17) are discussed in the relevant sub-section for this watershed. A weight-of-evidence approach was used to evaluate sites of concern. Individual exceedances of parameters, particularly those without regulatory objectives, are considered in context. In general, nickel and chromium levels are geologically elevated in the Bay Area, particularly where there are serpentine deposits. Nutrient levels (nitrate and phosphorus) are above U.S. EPA guidelines at most sites, but may still be relatively low.

Temperature, DO, pH, and specific conductance levels are of concern throughout this watershed. Measured conditions are unsuitable for salmonids, and benthic macroinvertebrate communities were degraded at all sampling sites. Nutrient concentrations were elevated throughout the watershed.

Elevated concentrations of oxadiazon were measured in clam tissues at Arroyo Las Positas at El Charro (ALP010). There were also numerous indicators of eutrophication at this site. At ALP150, DO saturation reached an unusual high of 395 percent, indicating eutrophic conditions.

Table 7-17. Summary of sites with exceedances in Arroyo Las Positas

| <i>Analyte</i> | <i>Standard</i> | <i># of sites</i> |
|--|-------------------|-------------------|
| FIELD MEASURES of 7 sites sampled | | |
| Temperature | maximum, salmonid | 5 |
| | MWAT, steelhead | 5 |
| Oxygen, dissolved | minimum, COLD | 4 |
| | minimum, WARM | 3 |
| | 3-month median | 1 |
| pH | range | 5 |
| CONVENTIONAL WATER QUALITY of 4 sites sampled | | |
| Nitrate as N | maximum | 4 |
| Phosphorus, total as P | maximum | 4 |
| WATER METALS of 3 sites sampled | | |
| Selenium | chronic | 2 |
| SEDIMENT CHEMISTRY of 1 site sampled | | |
| <i>Metals</i> | | |
| Chromium | TEC | 1 |
| Copper | TEC | 1 |
| Mercury | TEC | 1 |
| Nickel | TEC | 1 |

7.6.2 Water Quality in Relation to Land Use

Land use in the watershed is primarily urban, with some agriculture and mixed use. Measured nutrient concentrations were somewhat elevated, perhaps from residential fertilizer application. Riparian habitat restoration would appear to be the highest priority land management strategy to improve water quality. Riparian corridors throughout the watershed had poor physical habitat scores, and thresholds for temperature and dissolved oxygen were exceeded at all stream sites.

7.6.3 Macroinvertebrate Assemblages and Physical Habitat

Benthic assemblages at sites in the Arroyo Las Positas watershed exhibited uniformly low faunal diversity and high pollution tolerance. None of the 29 taxa identified as common intolerant taxa were found at any sites in the Arroyo Las Positas watershed. Taken together, the lack of these taxa and the abundance of midges and worms indicate that aquatic habitat conditions in Arroyo Las Positas were very poor.

All seven sites were classified in Assemblage Group 1 (the assemblage group dominated by highly degraded urban sites) and were closely clustered in the ordination graph and cluster analysis (Figure 7- and Figure 7-). Total taxa richness ranged between 11 and 16, well below the values observed at minimally disturbed sites, which ranged from 28 to 59 (Table 6-10). Most of these taxa were non-insects; the number of insect taxa ranged from 3 (ALP080) to 8 (ALP010).

The taxa present at sites throughout the mainstem of Arroyo Las Positas (ALP010, ALP040, ALP070, ALP080, ALP110) suggest that benthic invertebrate assemblages are significantly affected by poor water quality and habitat conditions.

- Chironomid midges (mostly Orthoclaadiinae) were the numerically dominant taxa at four of the five sites, and composed between 44 percent and 81 percent of organisms collected (Figure 7-).
- Oligochaete worms (dominated by the tubificid Naididae) occurred in high numbers at sites ALP070 and ALP110, where they made up 47 percent and 35 percent of total individuals.

Together, chironomids and oligochaetes composed the vast majority of organisms collected (82-94 percent). These two taxa are highly tolerant of pollution and are found in a wide range of lotic habitats, although oligochaetes are especially abundant in patches of fine sediment. Oligochaetes and midges were also the dominant taxa in low gradient effluent-dominated waterways in the lower Sacramento River watershed (de Vlaming et al 2004) and degraded urban streams in Santa Rosa (Sonoma County) (Sustainable Lands Stewardship Institute 2002).

Also common in the lower watershed sites was *Corbicula*, a non-native Asian clam. *Corbicula* was not collected in any of the other watersheds sampled by SWAMP. *Corbicula* is common in San Francisco Bay and the Sacramento/San Joaquin Delta, and has also been found in streams in the Santa Clara Valley (Carter and Fend 1997). Passive dispersal of *Corbicula* occurs only in the downstream direction by water currents, suggesting that the clam was introduced into the Arroyo Las Positas drainage by humans.

Baetis is a ubiquitous mayfly found in most flowing water habitat except the most severely degraded conditions; it has been found at all other sites monitored by SWAMP in the Bay Area except two channelized, urbanized sites in lower Permanente Creek (Santa Clara County) and three sites in the Arroyo Las Positas watershed: ALP040, ALP100, and ALP110.

Many beetles (Coleoptera) re-colonize very slowly following disturbance, because the larval and adult stages of many species are entirely aquatic. Only one individual beetle (Coleoptera) was collected from the entire watershed, a single riffle beetle (*Optioservus*) from ALP010. Stoneflies (Plecoptera), which require cool, clean water, were completely absent from all seven sites in the watershed.

Although all sites in the watershed were taxonomically similar, the site on upper Altamont Creek (ALP140) was the most dissimilar on the cluster analysis and ordination (Figure 7- and Figure 7-). While biological metrics at this site were generally similar to values at other sites in the watershed, some differences suggest that this site may possess slightly greater biological integrity. The mayfly *Baetis* sp. was the dominant taxa (72 percent of total organisms) at ALP140, while chironomids (14 percent) and oligochaetes (<1 percent) were less abundant (Figure 6-68). In addition to *Baetis*, three other EPT taxa were collected at ALP140: *Hydropsyche* sp. (Trichoptera: Hydropsychidae), *Hydroptila* sp. (Trichoptera: Hydroptilidae), and the uncommon, pollution intolerant caddisfly *Tinodes* sp. (Trichoptera: Psychomiidae). The tolerance value at ALP-140 was 5.1, well below values at other sites in the watershed (range: 5.7 to 6.1), but still greater than values at minimally disturbed sites (range 3.07 to 4.47).

Flow conditions in the Arroyo Las Positas watershed are generally good, but other habitat factors may limit benthic assemblages. Qualitative P-HAB scores for velocity/depth regimes and channel flow status were excellent across the seven sites, ranging from 13 to 18 and 16 to 19 (on a 0-20 scale), respectively. P-HAB scores relating to substrate quality, such as epifaunal substrate, embeddedness, and sediment deposition, indicated that substrate conditions are generally poor and high levels of fine sediment are present. Based on substrate size estimates, more than 50 percent of the area of the channel bed surface was composed of fine sediment at four of the seven sites. Deposition of fine sediment fills in crevices and interstitial spaces around gravel and cobble sized clasts, eliminating habitat required by many species of benthic insects.

Riparian conditions were also found to be poor throughout the watershed. Total riparian zone P-HAB scores ranged from 1 to 10 (0-20 scale) across the seven sites, while canopy cover was 5 percent or less at four sites. Much of the drainage network has been channelized, resulting in simplified channel forms and sparse or non-existent riparian vegetation. These changes in channel and riparian habitat may be partly responsible for high stream temperatures (see below, section 7.6.4)

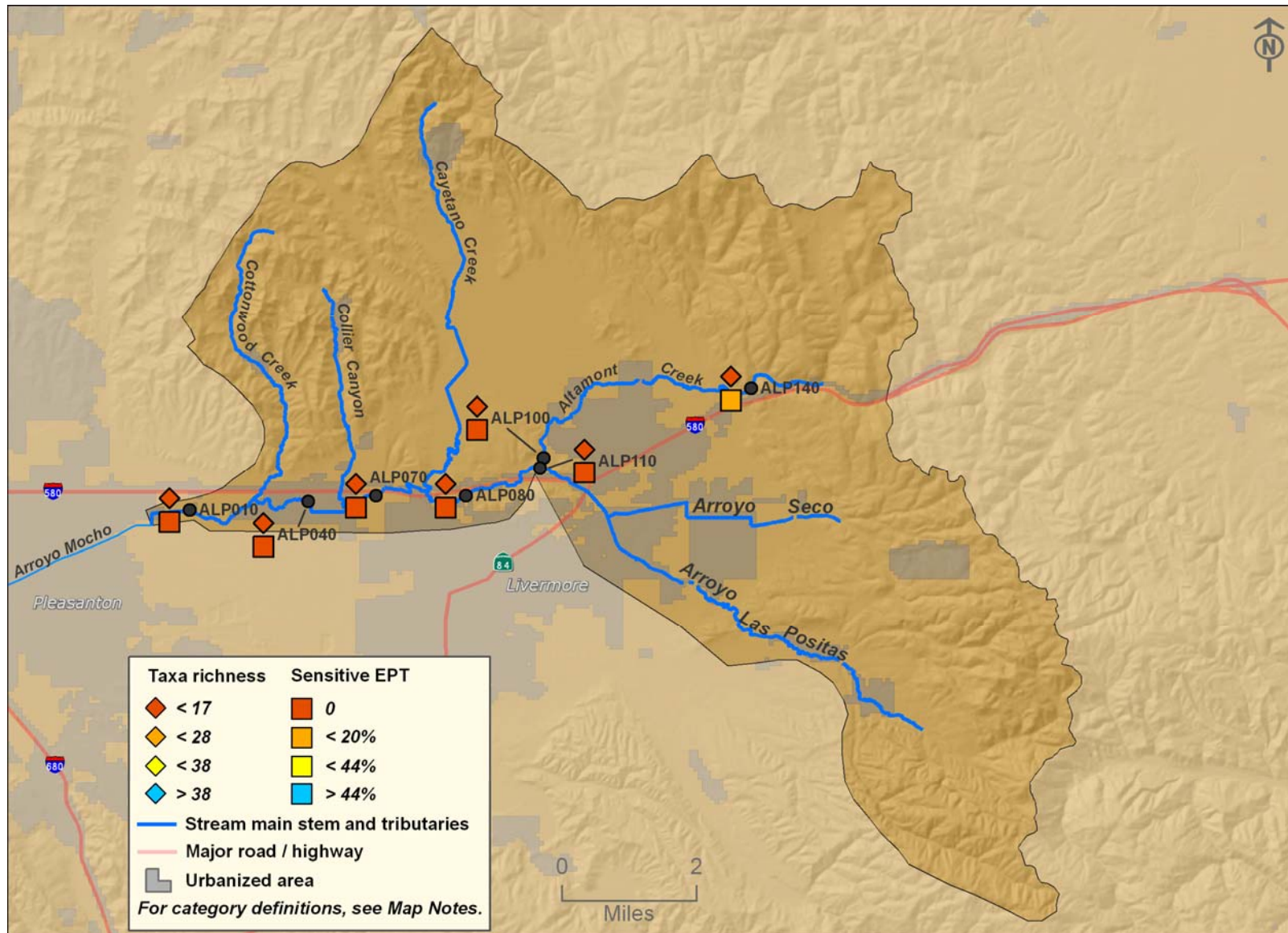


Figure 7-45. Map of Arroyo Las Positas benthic macroinvertebrate index scores

Table 7-18. Bioassessment sites in the Arroyo Las Positas watershed

| Site Code | Site Name | Stream |
|-----------|--------------------|--------------------|
| ALP010 | El Charro | Arroyo Las Positas |
| ALP040 | Airway Blvd. Exit | Arroyo Las Positas |
| ALP070 | Airway 2 | Arroyo Las Positas |
| ALP080 | N. Livermore Ave. | Arroyo Las Positas |
| ALP100 | Altamont Creek | Altamont Creek |
| ALP110 | Arroyo Las Positas | Arroyo Las Positas |
| ALP140 | Altamont Pass | Altamont Creek |

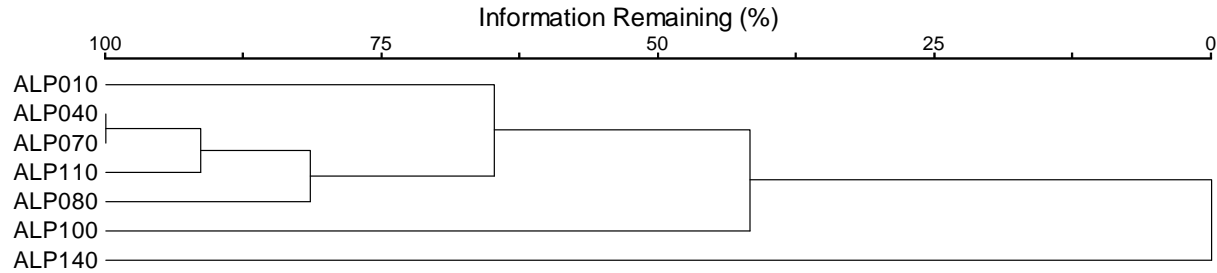


Figure 7-46. Cluster dendrogram of sites in the Arroyo Las Positas watershed

This cluster dendrogram groups sites based on similarity of macroinvertebrate taxa present in the Arroyo Las Positas watershed. Sites that are closely joined together have many similar taxa present, while groups with longer chain lengths are more dissimilar. The Information Remaining axis at top signifies how similar the sites are with respect to taxa presence.

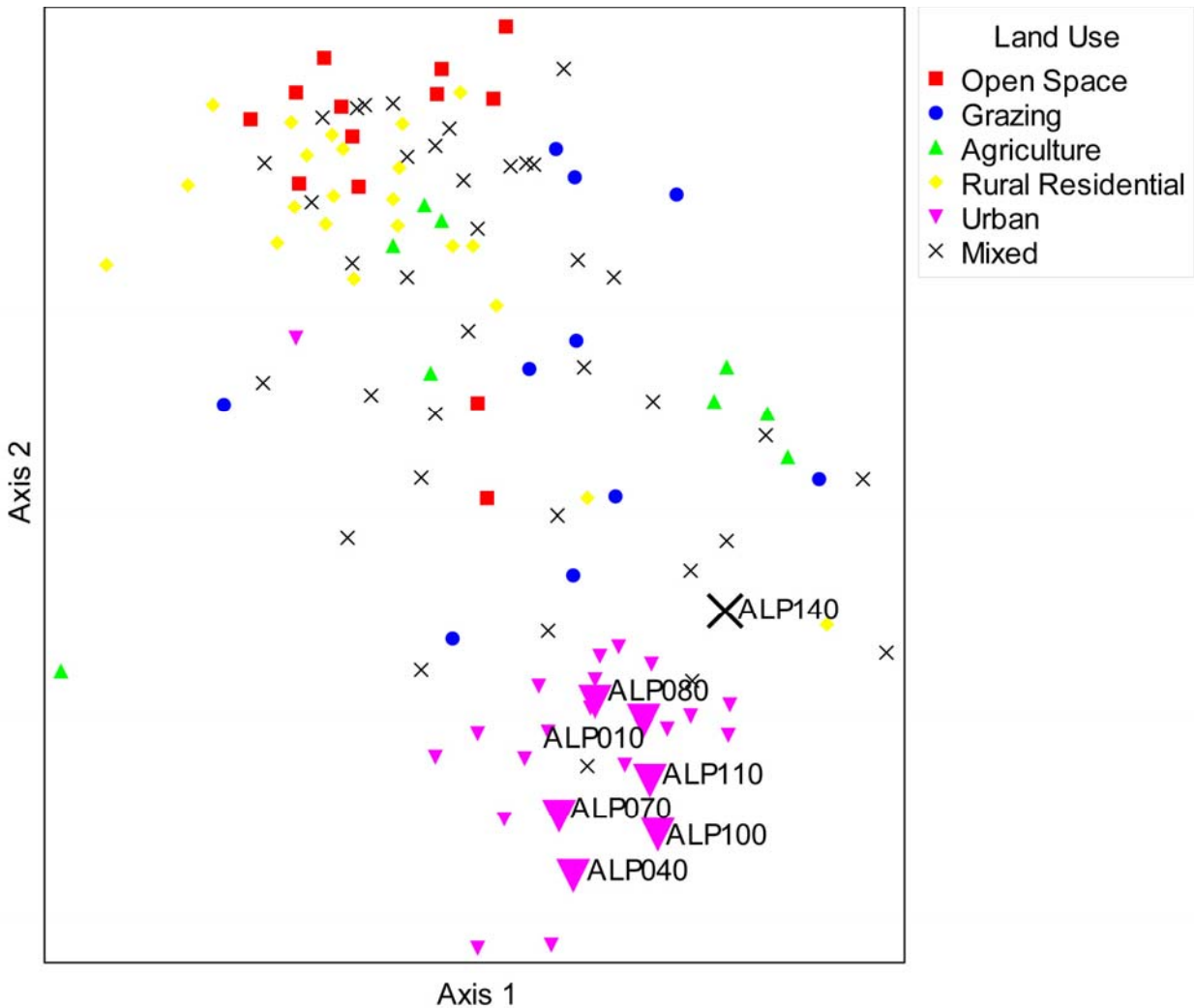


Figure 7-47. NMS ordination of sites in the Arroyo Las Positas watershed

The NMS ordination of sites by macroinvertebrate taxa abundance. Sites in the Arroyo Las Positas watershed are labeled and symbols are enlarged. For an explanation of ordination graphs, refer to Section 6.1.

Table 7-19. Biological metrics from the Arroyo Las Positas watershed

| | ALP010 | ALP040 | ALP070 | ALP080 | ALP100 | ALP110 | ALP140 |
|--------------------|--------|--------|--------|--------|--------|--------|--------|
| Coleoptera Taxa | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Diptera Taxa | 4 | 3 | 4 | 2 | 3 | 3 | 3 |
| Ephemeroptera Taxa | 1 | 0 | 1 | 1 | 0 | 0 | 1 |
| Plecoptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Trichoptera Taxa | 1 | 0 | 0 | 0 | 1 | 2 | 3 |
| Non-Insect Taxa | 8 | 7 | 9 | 10 | 7 | 7 | 5 |
| EPT Taxa | 2 | 0 | 1 | 1 | 1 | 2 | 4 |
| Taxa Richness | 16 | 11 | 14 | 13 | 11 | 12 | 12 |
| % EPT | 1 | 0 | 1 | 4 | 0 | 0 | 82 |
| % Sensitive EPT | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| % Chironomidae | 81 | 79 | 44 | 73 | 81 | 57 | 14 |

| | | | | | | | |
|----------------------|-----|-----|-----|-----|-----|-----|-----|
| % Coleoptera | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| % Oligochaeta | 3 | 15 | 47 | 9 | 2 | 35 | 0 |
| % Non-insect | 15 | 20 | 54 | 13 | 11 | 40 | 1 |
| % Intolerant | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| % Tolerant | 8 | 3 | 4 | 2 | 2 | 4 | 1 |
| Tolerance Value | 6.1 | 5.9 | 5.6 | 5.9 | 5.9 | 5.7 | 5.1 |
| % Predator | 1 | 1 | 1 | 2 | 0 | 2 | 2 |
| % Collector-filterer | 3 | 2 | 4 | 10 | 8 | 3 | 11 |
| %Collector-gatherer | 89 | 96 | 94 | 87 | 90 | 94 | 86 |
| % Scraper | 6 | 1 | 1 | 1 | 0 | 1 | 1 |
| % Shredder | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| % Other | 1 | 0 | 0 | 0 | 1 | 0 | 0 |

Table 7-20. Percentage abundance of taxonomic groups in Arroyo Las Positas

| | 1 | 2 | 3 | 4 | 5 |
|---------------|----------------------------|--------------------------------|---------------------------------|-----------------------------|--------------------------------|
| ALP010 | Chironomidae (81%) | <i>Physa/ Physella</i> (6%) | Turbellaria (4%) | Oligochaeta (3%) | <i>Simulium</i> sp. (2%) |
| ALP040 | Chironomidae (79%) | Oligochaeta (15%) | <i>Corbicula</i> (2%) | <i>Megadrili</i> (1%) | <i>Physa/ Physella</i> (1%) |
| ALP070 | Oligochaeta (47%) | Chironomidae (44%) | <i>Corbicula</i> (2%) | <i>Simulium</i> sp. (2%) | Turbellaria (2%) |
| ALP080 | Chironomidae (74%) | <i>Simulium</i> sp. (7%) | Turbellaria (7%) | Oligochaeta (2%) | Glossiphoniidae (1%) |
| ALP100 | Chironomidae (81%) | <i>Simulium</i> sp. (7%) | Turbellaria (7%) | Oligochaeta (2%) | Glossiphoniidae (1%) |
| ALP110 | Chironomidae (57%) | Oligochaeta (35%) | <i>Simulium</i> sp. (2%) | Sperchontidae (2%) | Ostracoda (1%) |
| ALP140 | <i>Baetis</i> sp. (72%) | Chironomidae (14%) | <i>Hydropsyche</i> sp. (10%) | <i>Argia</i> sp. (2%) | <i>Simulium</i> sp. (1%) |

Taxonomic Composition of Assemblages, Arroyo Las Positas Watershed

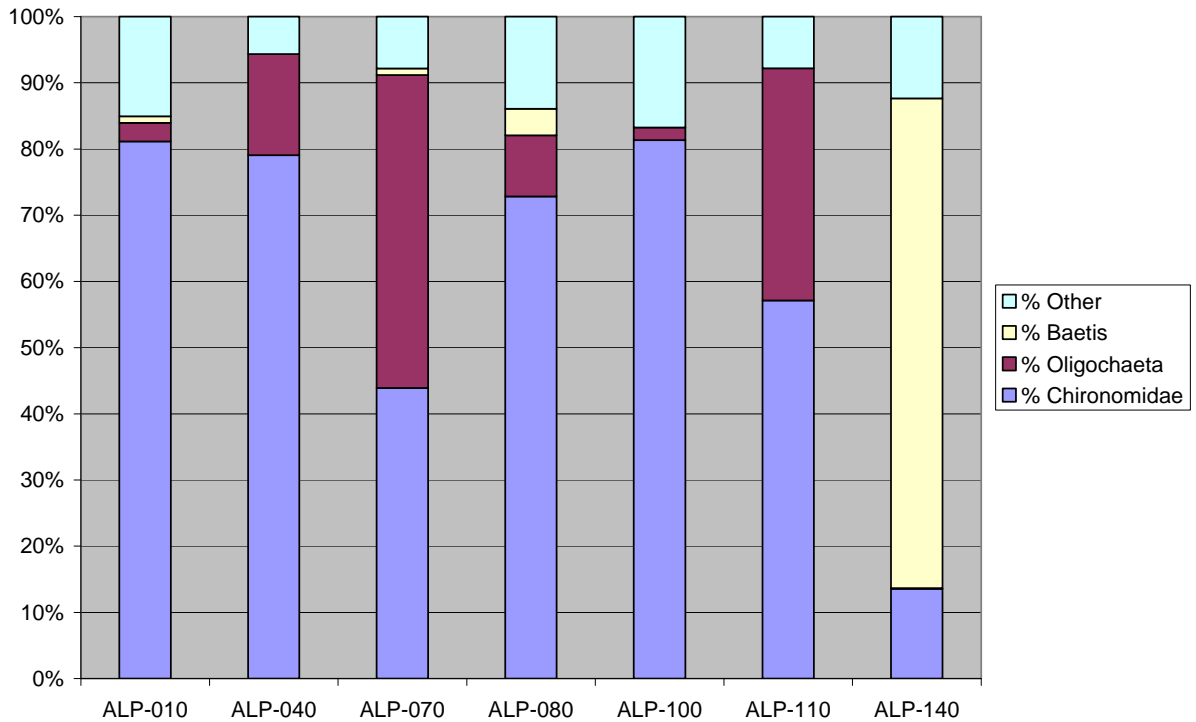


Figure 7-48. Taxonomic composition in the Arroyo Las Positas watershed

7.6.4 Basic Water Quality Field Measurements

Leidy *et al.* (2003) reports that Arroyo Las Positas was an historical migration route for steelhead but Gunther *et al.* (2000) determined that the creek had no current potential to support spawning and rearing.

Stream temperatures were high and dissolved oxygen low throughout the watershed in spring and dry season measurements (Figure 7-, Figure 7-, Figure 7-, and Appendix D).

- In spring, HWAT values exceeded the MWAT threshold for coho salmon (14.8°C) at Arroyo Las Positas at El Charro (ALP010), Airway Blvd. Exit (ALP040), and N. Livermore Ave. (ALP080).
- The higher MWAT threshold for steelhead (17°C) was exceeded in spring measurements from Altamont Creek at Springtown (ALP105), and Arroyo Seco (ALP150).
- The short-term maximum temperature for salmonid survival was exceeded in the spring at ALP040, ALP105, and ALP 150.
- In the summer, the MWAT threshold for steelhead (17°C) was exceeded at all of the above sites, as was the short-term maximum temperature for salmonid survival.

Dissolved oxygen concentrations failed to meet Basin Plan objectives (7 mg/L) in both spring and summer, with the exception of ALP010, which fell below 9 mg/L. Even in winter, DO at ALP150 fell below 7 mg/L (5.6 mg/L). Most notable were the maximum DO percent saturation values, which ranged from 112.9 (ALP010, summer) to 395 percent (ALP150, spring, 36.42 mg/L). Supersaturation above 120 percent is consistent with eutrophication and nutrient enrichment suggested by the high nitrate levels. The extreme value of 395 percent is the highest encountered in all SWAMP monitoring to date.

The Basin Plan objective for maximum pH (8.5) was exceeded at all five sites in at least one season, with a high measurement of 9.24 in the wet season at ALP150 (Figure 7-).

The Arroyo Las Positas watershed was the only one of the nine watersheds in this study that had a substantial number of exceedances for dissolved salt content, measured as specific conductance. The Basin Plan maximum for municipal drinking water taste was exceeded at all five sites in all three seasons. The higher objective for agricultural use was exceeded at Altamont Creek just downstream of the Spring (ALP105) in the spring and summer.

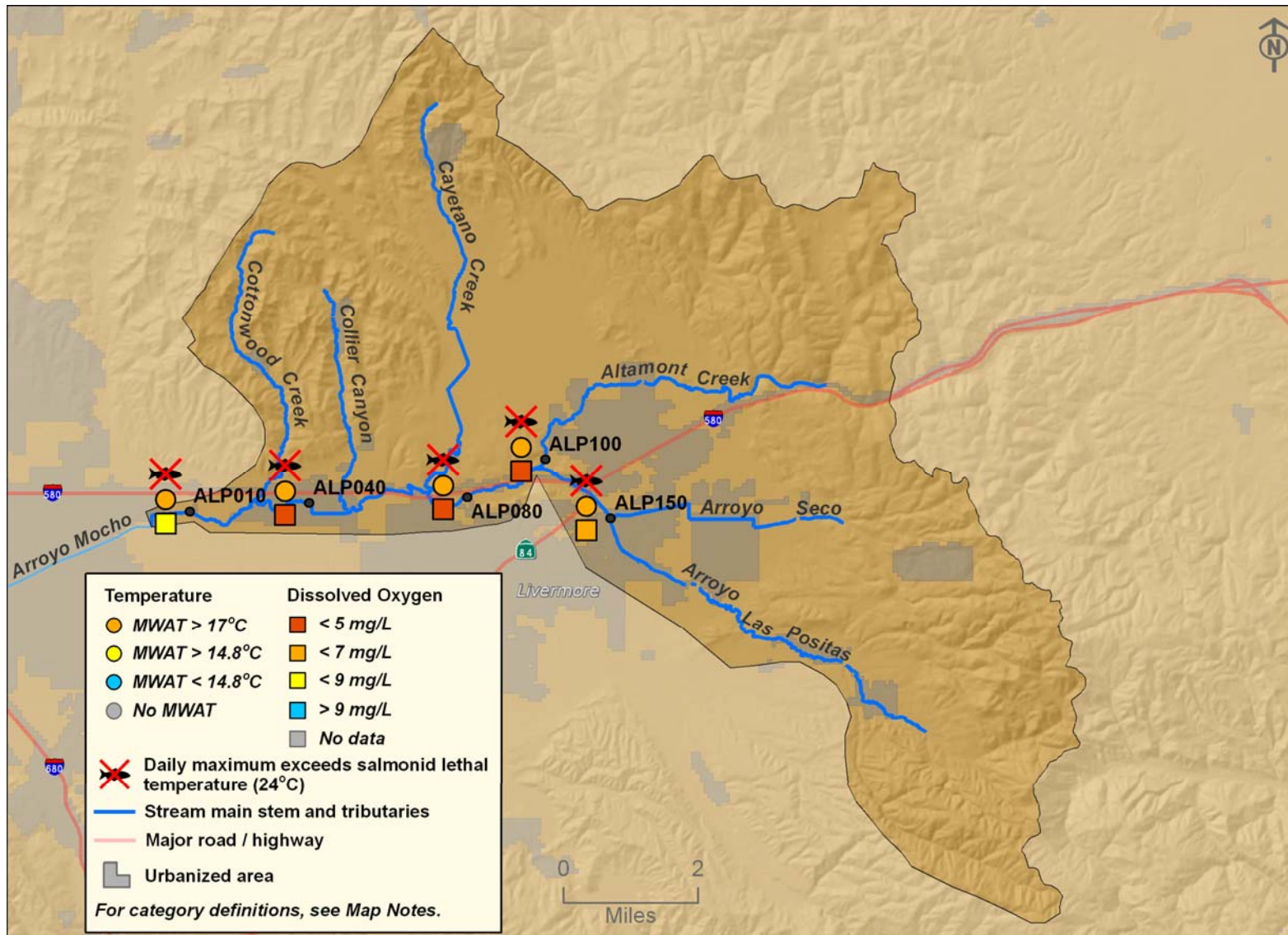


Figure 7-49. Map of temperature and DO monitoring in Arroyo Las Positas

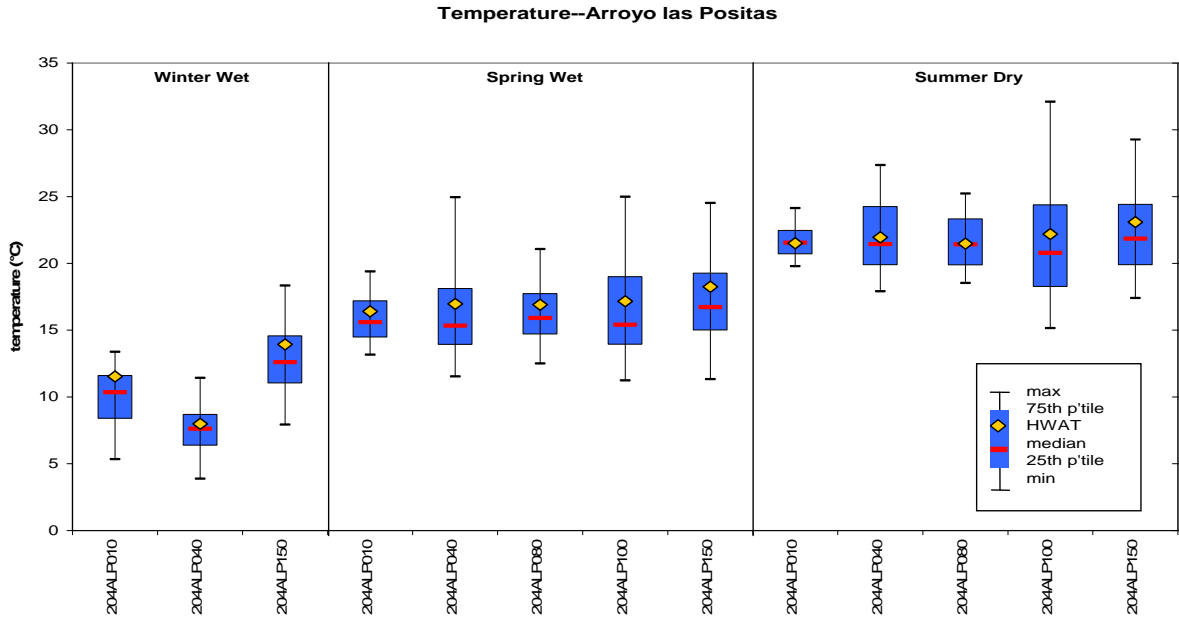


Figure 7-50. Temperature monitoring in Arroyo Las Positas

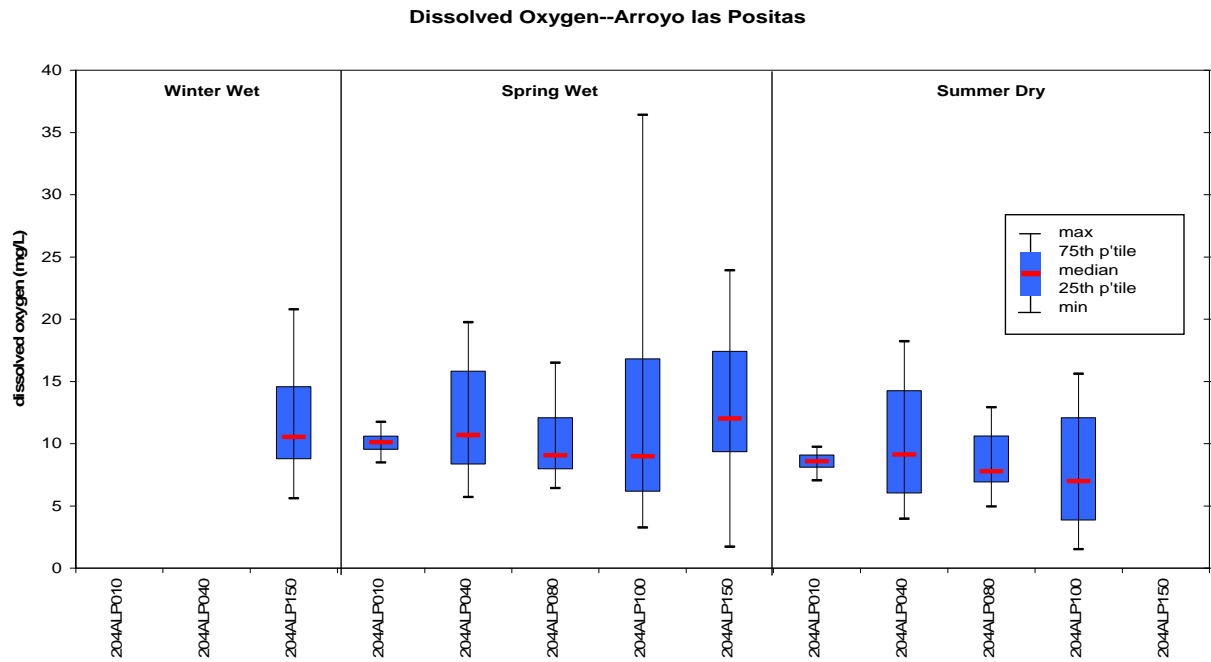


Figure 7-51. DO monitoring in Arroyo Las Positas

Data were rejected from ALP101 and ALP040 for the winter season and from ALP150 for the summer season.

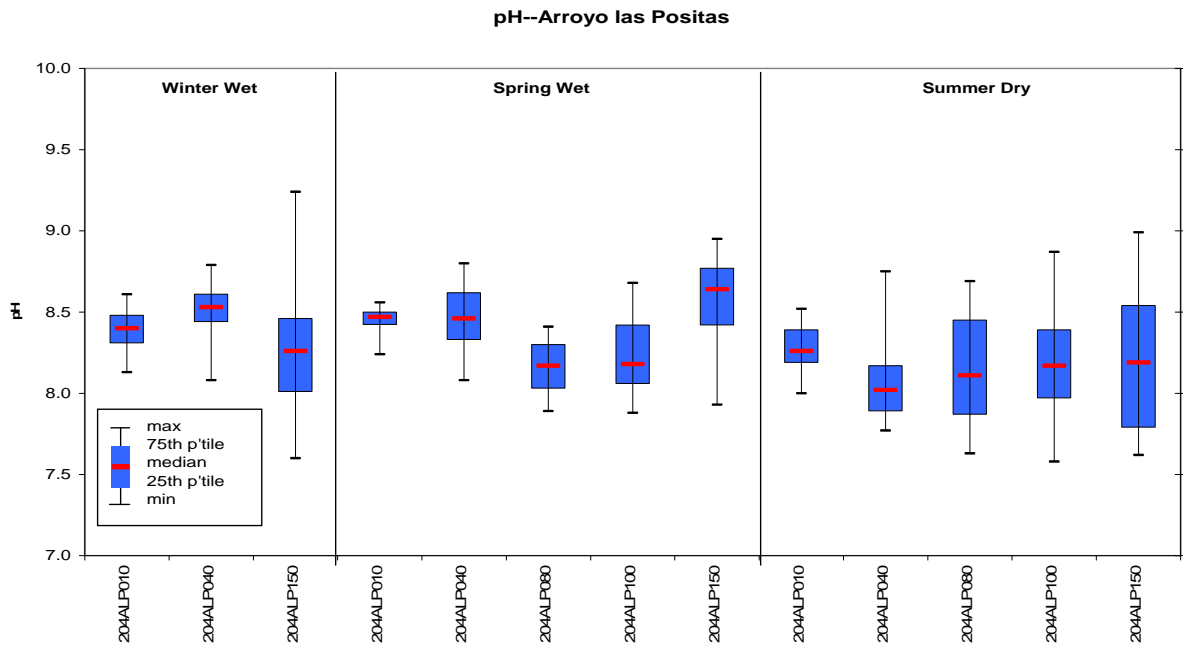


Figure 7-52. pH monitoring in Arroyo Las Positas

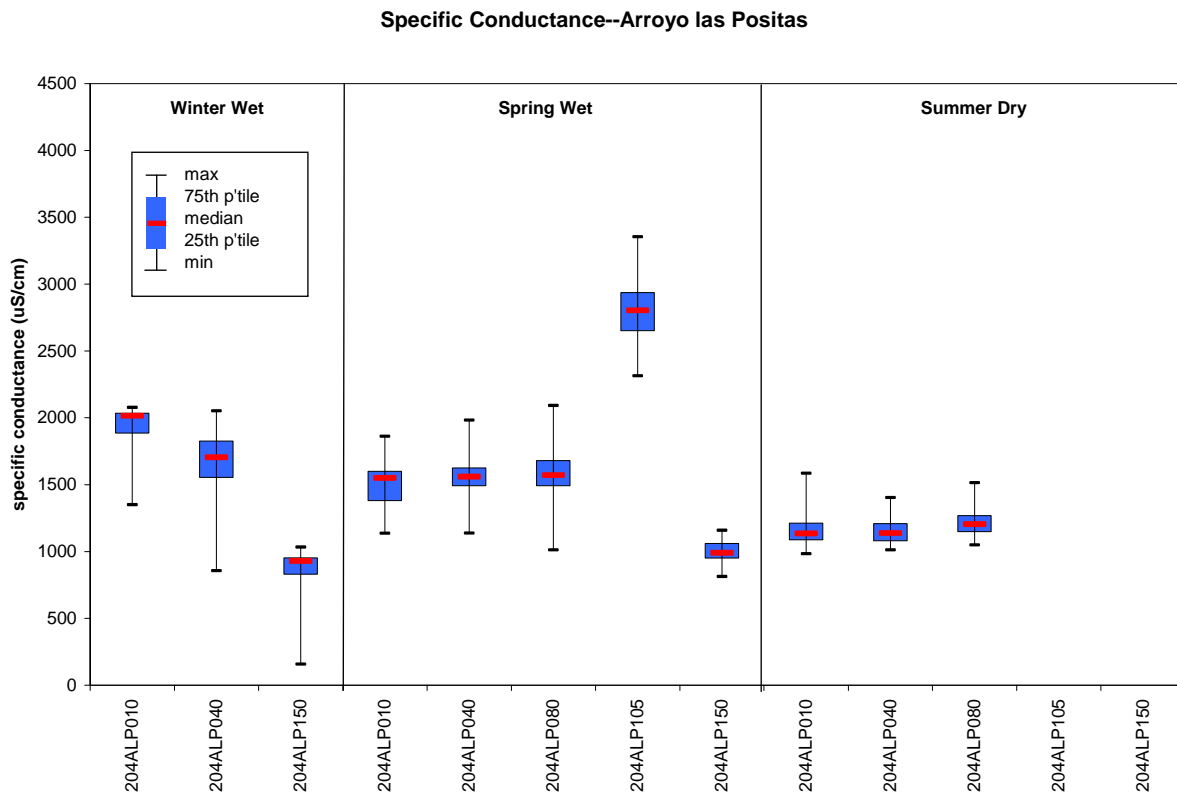


Figure 7-53. Specific conductance monitoring in Arroyo Las Positas

7.6.5 Water, Sediment, and Clam Tissue Chemistry

Water samples were analyzed for nutrients at four sites in the Arroyo Las Positas watershed: Arroyo Las Positas at El Charro (ALP010), Altamont Creek (ALP100), Arroyo Las Positas (ALP110) and Altamont Creek at Altamont Pass (ALP140). Nitrate concentrations were high at all sites, especially in the wet season. A high value of 8.04 mg/L was measured at ALP110 in spring. All measurements were significantly above the U.S. EPA reference guideline (Figures 6-15 to 6-17; Appendix G).

Total phosphorus measurements were not as conspicuously elevated as were the nitrate values, but most were significantly above 0.030, the U.S. EPA reference guideline. All but one orthophosphorus value exceeded the total phosphorus guideline.

Trace metals were measured at three sites: Arroyo Las Positas at El Charro (ALP010), Altamont Creek (ALP100), and Arroyo Las Positas (ALP110). Selenium exceeded the Basin Plan chronic aquatic life water quality objective (WQO) in the spring samples from ALP010 and ALP100. There were no other WQO exceedances in samples from this watershed.

Sediments were collected for analysis from Arroyo Las Positas at El Charro (ALP010). Probable effects concentrations were exceeded for the geologically abundant elements chromium and nickel, and the lower threshold effects concentrations were exceeded for copper and mercury. The sediment quality guideline quotient value was fairly low at 0.21 (Appendix H).

Tissues of clams deployed at Arroyo Las Positas at El Charro (ALP010) had values near the regional average for mercury, selenium, DDT, and PCBs. Copper was slightly elevated, and chlordane was not detected. Tissues from clams deployed at this site had a markedly high concentration of the herbicide oxadiazon (Figure 6-27).

7.6.6 Water and Sediment Toxicity

Three-species toxicity tests were conducted on water samples from Arroyo Las Positas at El Charro (ALP010), Altamont Creek (ALP100), and Arroyo Las Positas (ALP110). There were significant decreases in algal population growth in spring and dry season samples from ALP110, and in the spring sample from ALP010. The ALP010 site had the highly elevated clam tissue concentration of oxadiazon. No significant toxicity was observed to any other species (Figure 7-). An amphipod toxicity test was conducted on a sediment sample at ALP010; there were no significant observed effects on amphipods in this test.

Toxicity Tests in Arroyo las Positas Watershed--Response Relative to Control

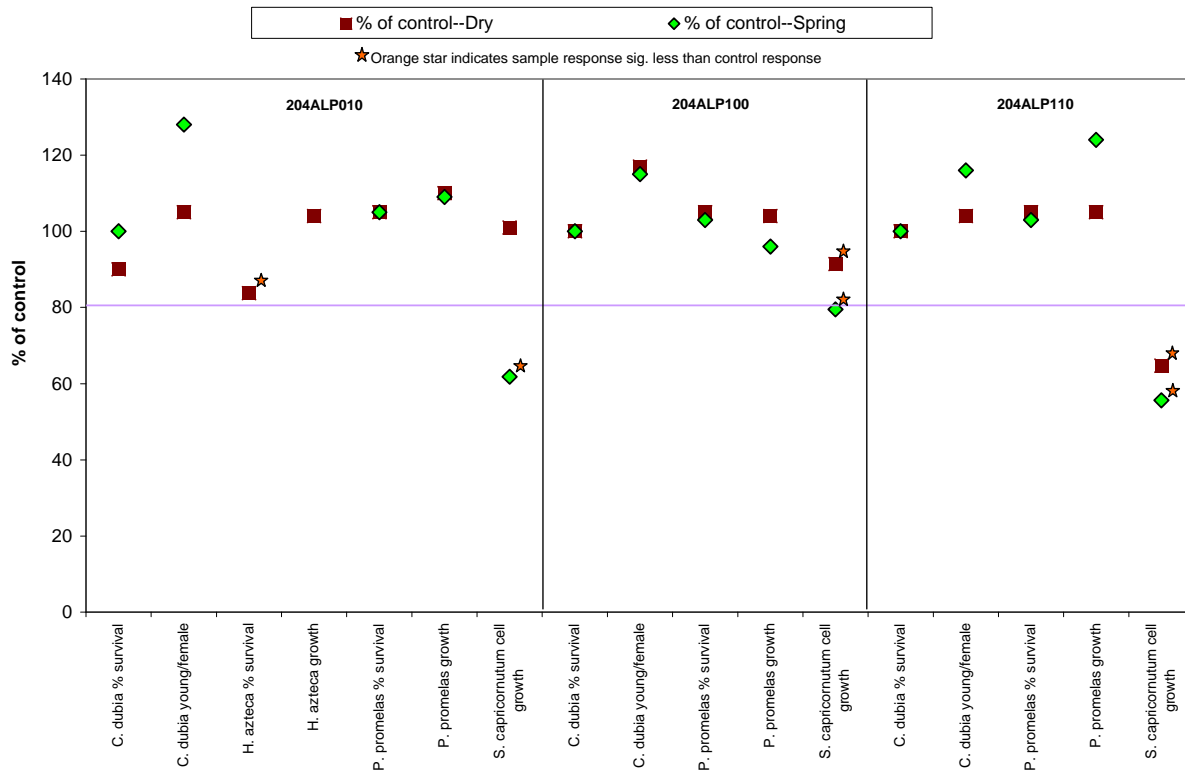


Figure 7-54. Toxicity tests in Arroyo Las Positas

Results of toxicity tests are shown, with three species in water samples and one species in sediment samples from Arroyo Las Positas. See text for description of species and endpoints. Stars indicate statistically significant differences from test controls. Samples with stars that were also below 80 percent of the control were considered toxic for this assessment.

7.6.7 Coliform Bacteria

Coliform bacteria were not measured in the Arroyo Las Positas watershed.

7.7 Pescadero Creek/Butano Creek Watershed

7.7.1 Sites of Concern

Exceedances noted in the summary tables (Table 7-21 and Table 7-22) are discussed in the relevant sub-section for this watershed. A weight-of-evidence approach was used to evaluate sites of concern. Individual exceedances of parameters, particularly those without regulatory objectives, are considered in context. In general, nickel and chromium levels are geologically elevated in the Bay Area, particularly where there are serpentine deposits. Nutrient levels (nitrate and phosphorus) are above U.S. EPA guidelines at most sites, but may still be relatively low.

While data collected in this study do not indicate any specific sites of serious concern for nutrients or contaminants, high coliform exceedances at PES060 (Community Church) suggest the need for further investigation and management.

Temperature and dissolved oxygen levels were supportive of salmonids at most sites. The significant relationship between riparian vegetation and stream temperature indicates that riparian zone management is a priority for continued protection of salmonid populations.

Table 7-21. Summary of sites with exceedances in Pescadero Creek watershed

| <i>Analyte</i> | <i>Standard</i> | <i># of sites</i> |
|--|-----------------|-------------------|
| FIELD MEASURES of 9 sites sampled | | |
| Temperature | MWAT, steelhead | 2 |
| | MWAT, coho | 4 |
| Oxygen, dissolved | minimum, COLD | 1 |
| | 3-month median | 1 |
| CONVENTIONAL WATER QUALITY of 4 sites sampled | | |
| Nitrate as N | maximum | 2 |
| Phosphorus, total as P | maximum | 4 |
| SEDIMENT CHEMISTRY of 1 site sampled | | |
| <i>Metals</i> | | |
| Chromium | TEC | 1 |
| SEDIMENT TOXICITY of 1 site sampled | | |
| <i>Hyalella</i> | growth | 1 |
| COLIFORMS of 6 sites sampled | | |
| <i>E. coli</i> | steady state | 1 |
| Fecal coliform | log mean | 1 |
| | 90th percentile | 1 |
| Total coliform | median | 6 |
| | maximum | 1 |

Table 7-22. Summary of sites with exceedances in Butano Creek watershed

| <i>Analyte</i> | <i>Standard</i> | <i># of sites</i> |
|--|-----------------|-------------------|
| FIELD MEASURES of 3 sites sampled | | |
| Oxygen, dissolved | minimum, COLD | 1 |
| CONVENTIONAL WATER QUALITY of 2 sites sampled | | |
| Nitrate as N | maximum | 2 |
| Phosphorus, total as P | maximum | 2 |
| SEDIMENT TOXICITY of 1 site sampled | | |
| <i>Hyalella</i> | growth | 1 |
| COLIFORMS of 2 sites sampled | | |
| Total coliform | median | 2 |

7.7.2 Water Quality in Relation to Land Use

Land use in the watershed is primarily open space, agriculture, rural residential, and mixed use. Agriculture is predominant in the broad valleys of the lower mainstem of Pescadero and Butano creeks. Benthic invertebrate assemblages at sites in the lower watershed are slightly impacted, although much better than assemblages in urban watersheds.

7.7.3 Macroinvertebrate Assemblages and Physical Habitat

Three groups of sites were identified from the cluster analysis dendrogram (Figure 7-) of benthic assemblages in the Pescadero Creek and Butano Creek watersheds:

- The Headwaters group (HW) is composed of sites on tributaries and headwater streams in the upper Pescadero Creek (PES170, PES180, PES200, PES210, PES240) and Butano Creek (BUT040, BUT050) watersheds.
- The Middle Mainstem Pescadero Creek group (MMP) includes five sites on the middle section of Pescadero Creek (PES100, PES120, PES140, PES160, PES190).
- The Lower Pescadero Creek and Butano Creek group (LPB) includes four sites on the lower mainstem of Pescadero Creek (PES050, PES060, PES070, PES095) and two sites on the mainstem of Butano Creek (BUT020, BUT030).
- Four dissimilar sites (BUT010, PES050, PES080, PES150) that were not classified into any of the three groups are apparent on the cluster dendrogram (Figure 7-) and the ordination plot (Figure 7-).

The Headwaters group

These sites are located on small, steep, perennial streams draining forested catchments with minimal human land use, other than limited timber harvest. Benthic assemblages at the seven sites were very diverse, with taxa richness values ranging from 40-55, all within the range of values (38-59) observed at minimally disturbed perennial streams (Table 6-10). In addition to high taxonomic diversity, sites in the Headwaters group exhibit high functional diversity, with scrapers making up 27 to 52 percent of organisms and collector-gatherers making up 19 to 42

percent of all organisms. Despite the small stream size and dense canopy cover common to these forested streams, the scraper functional feeding group was the dominant feeding group at five of seven sites. The high taxonomic and functional diversity, low abundance of common tolerant taxa, and presence of sensitive taxa indicates that these sites possessed excellent biological integrity relative to other streams in the Bay Area.

Abundant taxa included:

| | |
|--|-----|
| <i>Cinygmula</i> sp. (Heptageniidae) | 13% |
| <i>Optiosevus</i> sp. (Elmidae) | 10% |
| <i>Baetis</i> sp. (Batidae) | 7% |
| Chironomidae | 7% |
| <i>Calineuria californica</i> (Perlidae) | 5% |

The percentage of common tolerant taxa was greater than 20 percent at only two sites, BUT040 (37 percent) and PES180 (29 percent). Many rare taxa with specific habitat requirements were found at these sites, including *Pteronarcella* sp. and *Pteronarcys* sp. (Plecoptera: Pteronarcyidae), *Apatania* sp. (Trichoptera: Apataniidae), *Despaxia augusta* (Plecoptera: Leuctridae), the grappletail dragonfly *Octogomphus specularis* (Odonata: Gomphidae), *Hesperocanopa* sp. (Diptera: Tipulidae), *Oreogeton* sp. (Diptera: Empididae), and *Philorus* sp. (Diptera: Blephariceridae).

Also notable at four sites was the aquatic pillbug *Gnorimosphaeroma* sp. (Isopoda: Sphaeromatidae), common in estuarine and intertidal habitats of the Pacific Coast. *Gnorimosphaeroma* has been found in coastal freshwater streams, including low elevation tributaries to Drake's Estero in Point Reyes (Lee, 2001), but its occurrence in headwater tributaries of the Santa Cruz Mountains is remarkable. It was also found (in very high abundances) at the most downstream site on Pescadero Creek (PES050) and in Jones Gulch (PES150).

The Middle Mainstem Pescadero (MMP) group

Most of this portion of the creek flows in a narrow canyon through heavily forested state and county parkland. Sites in this group were located on the ordination plot (Figure 6-77) between the Headwaters group (HW) and the Lower Pescadero Creek and Butano Creek group (LPB). Like sites in the Headwaters group, sites in the Middle Mainstem Pescadero group were characterized by high taxonomic richness (42-45), high abundance of EPT, low abundance of common tolerant taxa, and presence of rare and intolerant taxa.

Common tolerant taxa made up less than 25 percent of individuals at all five sites (Table 6-26). Several rare taxa were found at these sites, including the riffle beetles *Cleptelmis addenda* and *Lara* sp. (Coleoptera: Elmidae), the beetle family Staphylinidae (Coleoptera), and *Pteronarcella* sp. (Plecoptera: Pteronarcyidae).

Benthic assemblages were dominated by mayflies (Ephemeroptera), including:

| | |
|---|-----|
| <i>Ephemerella</i> sp. (Ephemerellidae) | 16% |
| <i>Serratella</i> sp. (Ephemerellidae) | 9% |
| <i>Cinygmula</i> sp. (Heptageniidae) | 6% |
| <i>Baetis</i> sp. (Baetidae) | 6% |
| <i>Epeorus</i> sp. (Heptageniidae) | 5% |
| <i>Dipheter hageni</i> (Baetidae) | 4% |
| <i>Drunella</i> sp. (Ephemerellidae) | 4% |

Other abundant taxa included:

| | |
|---|-----|
| <i>Optioservus</i> sp. (Elmidae) | 7% |
| <i>Simulium</i> sp. (Diptera: Simuliidae) | 6% |
| <i>Malenka</i> sp. (Plecoptera: Nemouridae) | 4% |
| Chironomidae (Diptera) | 4%. |

Lower Pescadero-Butano group

Land use in the lower portion of the Pescadero Creek watershed is primarily agricultural, with minor residential and commercial land use near the town of Pescadero. Sites PES050 and PES060 are located downstream of the town, while PES070 and PES095 are located upstream, adjacent to row crop agricultural fields. BUT030 is located on Butano Creek downstream of the small residential community of Butano Park, while site BUT020 is farther downstream of agricultural land use, including greenhouse operations. BUT010, the farthest downstream site on Butano Creek, was taxonomically distinct from these sites and is discussed separately.

Within the Lower Pescadero-Butano group of sites, biological data suggest that upstream sites are affected by minor water quality impacts, while lower sites exhibit moderate reductions in biological integrity. Although tolerant taxa were numerically dominant, the presence of common intolerant taxa such as *Malenka* sp. (Plecoptera: Nemouridae) and *Drunella* sp. (Ephemeroptera: Ephemerellidae) at all six sites indicates at least a moderate level of biological integrity. The number of common intolerant taxa at upstream sites (PES095, BUT030) is higher than at downstream sites (PES050, PES060, PES070, BUT020), suggesting a downstream trend in water quality impacts. Similarly, the values of percent sensitive EPT at downstream sites (11-22 percent) are less than at upstream sites (35-47 percent), which in turn are less than sites in the Middle Mainstem Pescadero group (49-70 percent).

In general, benthic assemblages at sites in the Lower Pescadero-Butano group were numerically dominated by common tolerant taxa:

| | |
|----------------------------------|-----|
| Chironomidae | 26% |
| <i>Simulium</i> sp. (Simuliidae) | 10% |
| <i>Baetis</i> sp. (Baetidae) | 9% |
| <i>Optioservus</i> sp. (Elmidae) | 7% |

| | |
|--|-----|
| <i>Ephemerella</i> sp. (Ephemerellidae) | 6% |
| <i>Tricorythodes</i> sp. (Leptohyphidae) | 5% |
| <i>Dipheter hageni</i> (Baetidae) | 4% |
| <i>Cinygmula</i> sp. (Heptageniidae) | 3%. |

The aquatic pillbug *Gnorimosphaeroma* sp. (Isopoda: Sphaeromatidae), a common invertebrate in coastal brackish waters, was the dominant taxa at PES050, making up over 50 percent of organisms collected. Rare taxa found at these sites included beetles such as *Postelichus* sp. (Dryopidae), found only at PES060 in the Bay Area; *Brychius* sp. (Halipilidae); and *Laccobius* sp. (Hydrophilidae). Other rare, relatively intolerant taxa collected from these sites included *Molophilus* sp. (Diptera: Tipulidae), *Cultus* sp. (Plecoptera: Perlodidae), and *Brachycentrus* sp. (Trichoptera: Brachycentridae).

Anomalous sites

Four sites were noticeably dissimilar from all other sites in the watershed.

- Both the ordination plot (Figure 7-) and cluster dendrogram (Figure 7-) showed the site on Waterman Creek (PES230) to be considerably different from other nearby sites in the Headwaters group (HW). Several common intolerant taxa found at neighboring sites were absent or present in very low abundances at PES230, such as *Narpus* sp. (Coleoptera: Elmidae), and *Rhyacophila* sp. (Trichoptera: Rhyacophilidae). Most notably, three genera of Heptageniidae found at every site in the Headwaters group were absent from PES230: *Epeorus* sp., *Ironodes* sp., and *Rithrogena* sp. Heptageniid mayflies are the dominant algal grazing herbivores in minimally disturbed Bay Area streams. These taxonomic differences mirror differences in biological metrics between PES230 and other Headwater sites, especially EPT taxa richness (17 at PES230 versus 20-31 at other sites) and the percentage of collector-filterers (55 percent versus 19-42 percent). These differences are likely the result of local habitat conditions. The Waterman Creek sampling reach received very low P-HAB scores for embeddedness (4 out of 20), sediment deposition (5 out of 20), and channel alteration (2 out of 20) parameters.
- Jones Gulch (PES150) is a small tributary that flows through forested parkland. The only human land use in the watershed is a youth camp near the source of the creek. Many ubiquitous sensitive EPT taxa were absent from PES150, including the ephemereid mayflies *Drunella* sp. and *Serratella* sp.; the heptageniid mayflies *Cinygmula* sp., *Epeorus* sp., and *Ironodes* sp.; the chloroperlid stonefly *Suwallia* sp., and the caddisfly *Neophylax* sp. Physical habitat conditions at PES150 may be responsible for the taxonomic differences, as evidenced by the poor embeddedness (4 out of 20) and sediment deposition (5 out of 20) P-HAB scores. Additionally, water was observed to be brown-colored and foamy during the wet season, possibly suggesting an upstream source of organic pollution.
- Honsinger Creek (PES080) is another small tributary that flows across the broad alluvial valley of Pescadero Creek. Row crop agriculture is the dominant land use in the lower portion of the Honsinger Creek watershed. Relative to other sites in the Pescadero Creek watershed, many ubiquitous EPT taxa were absent from PES080, including three genera of ephemereid mayflies, the perlodid family of stoneflies, and the caddisflies *Hydropsyche* sp. (Hydropsychidae) and *Neophylax* sp. Whereas local physical habitat

conditions were generally excellent, water quality may be negatively impacting the biological integrity of Honsinger Creek.

- The lowest site on Butano Creek, BUT010, was characterized by extremely low benthic abundances (~50/sq. meter) and low diversity (16 taxa). The benthic assemblage was dominated by chironomids (42 percent), oligochaetes (12 percent), the elmid beetle *Optioservus* sp. (12 percent), *Baetis* sp. (8 percent), and the heptageniid *Cinygmula* sp. (7 percent). The channel bed is predominantly composed of sand, with minimal gravel. A combination of natural tectonic processes, human-induced erosion in the upper Butano Creek watershed, and local channel alteration have resulted in deposition of large amounts of fine sediment in this reach of Butano Creek (Curry 1985). The balance between sediment supply and transport capacity is such that the sand-grained bed is mobile for much of the year, resulting in unstable substrate for invertebrate colonization. This is reflected in the P-HAB scores for epifaunal substrate (3 out of 20), embeddedness (1 out of 20) and sediment deposition (1 out of 20). The lack of pollution sensitive and intolerant taxa suggests that water quality issues may be important as well, but the available biological information is inconclusive.

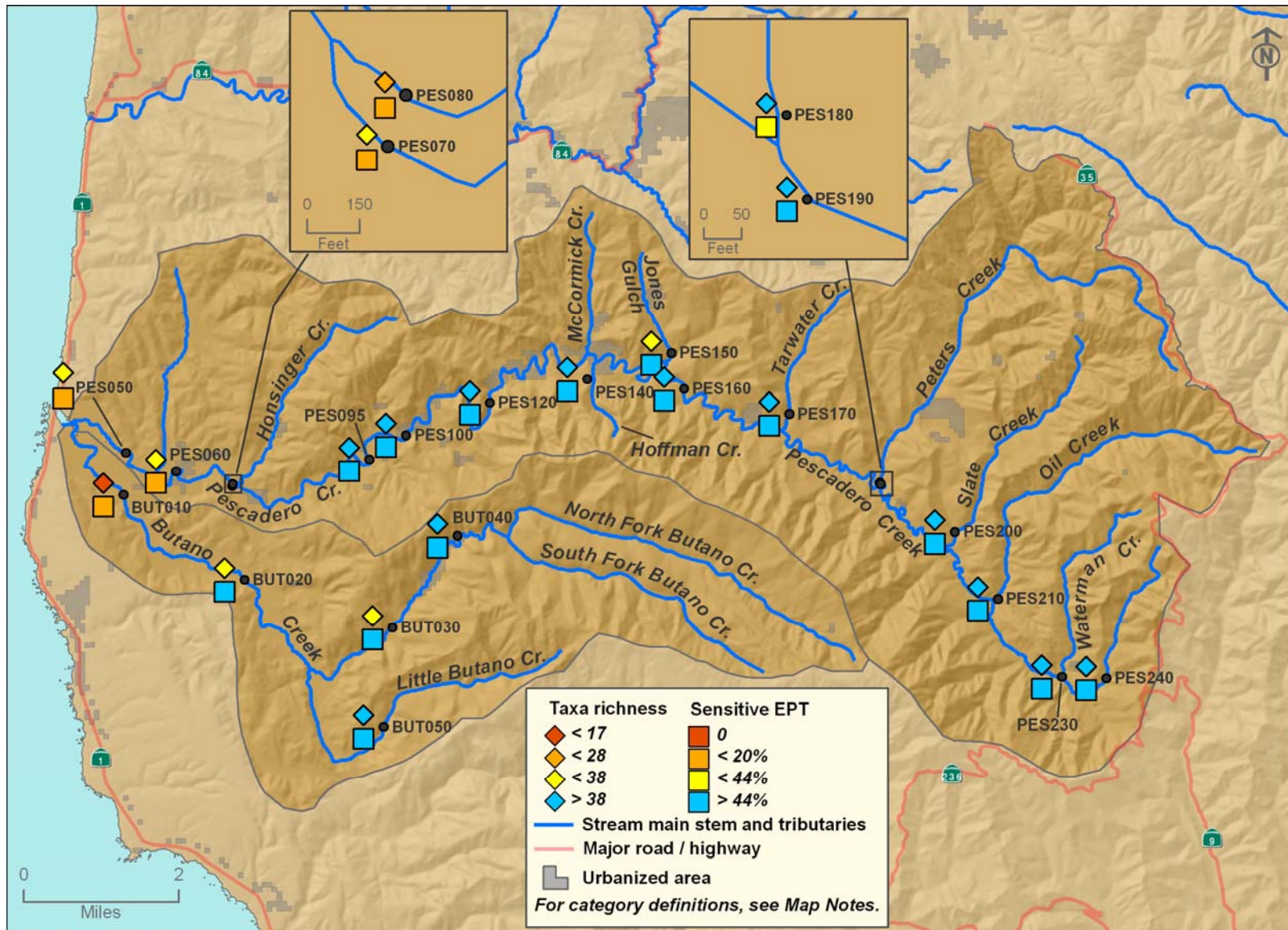


Figure 7-55. Map of Pescadero and Butano creeks benthic macroinvertebrate index scores

Table 7-23. Bioassessment sites in the Pescadero Creek/Butano Creek watershed

| Site ID | Site Name | Stream |
|---------|------------------------|---------------------|
| BUT010 | Lower Butano | Butano Creek |
| BUT020 | Cloverdale Coast Ranch | Butano Creek |
| BUT030 | Girl Scout Camp | Butano Creek |
| BUT040 | Butano Falls | Butano Creek |
| BUT050 | Butano State Park | Little Butano Creek |
| PES050 | Water Lane | Pescadero Creek |
| PES060 | Community Church | Pescadero Creek |
| PES070 | Cloverdale Road | Pescadero Creek |
| PES080 | Honsinger | Honsinger |
| PES095 | Pesky Ranch | Pescadero Creek |
| PES100 | USGS Gage | Pescadero Creek |
| PES120 | Loma Mar | Pescadero Creek |
| PES140 | Memorial Park | Pescadero Creek |
| PES150 | Jones Gulch | Jones Gulch |
| PES160 | Towne Fire Rd. | Pescadero Creek |
| PES170 | Tarwater | Tarwater Creek |
| PES180 | Peters Creek | Peters Creek |
| PES190 | Portola SP | Pescadero Creek |
| PES200 | Slate | Slate Creek |
| PES210 | Oil | Oil Creek |
| PES230 | Waterman | Waterman Creek |
| PES240 | Headwaters | Pescadero Creek |

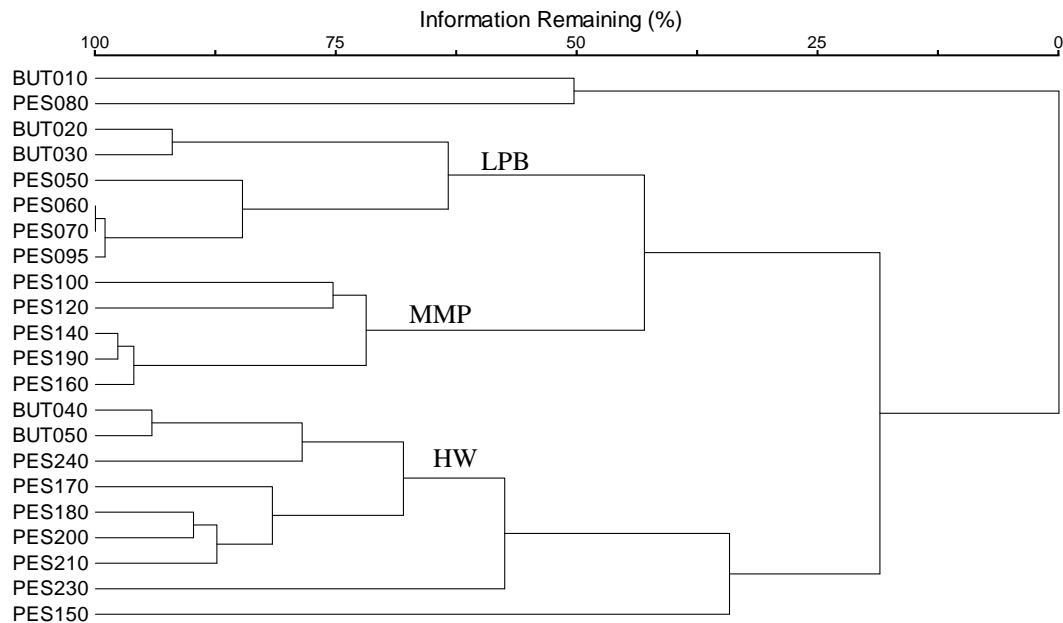


Figure 7-56. Cluster dendrogram of sites in the Pescadero Creek/Butano Creek watershed

The cluster dendrogram groups sites based on similarity of macroinvertebrate taxa present. The Information Remaining axis at the top signifies how similar the sites are with respect to taxa presence. Abbreviations: LPB, Lower Pescadero and Butano; MMP, Middle Mainstem Pescadero; HW, Headwaters.

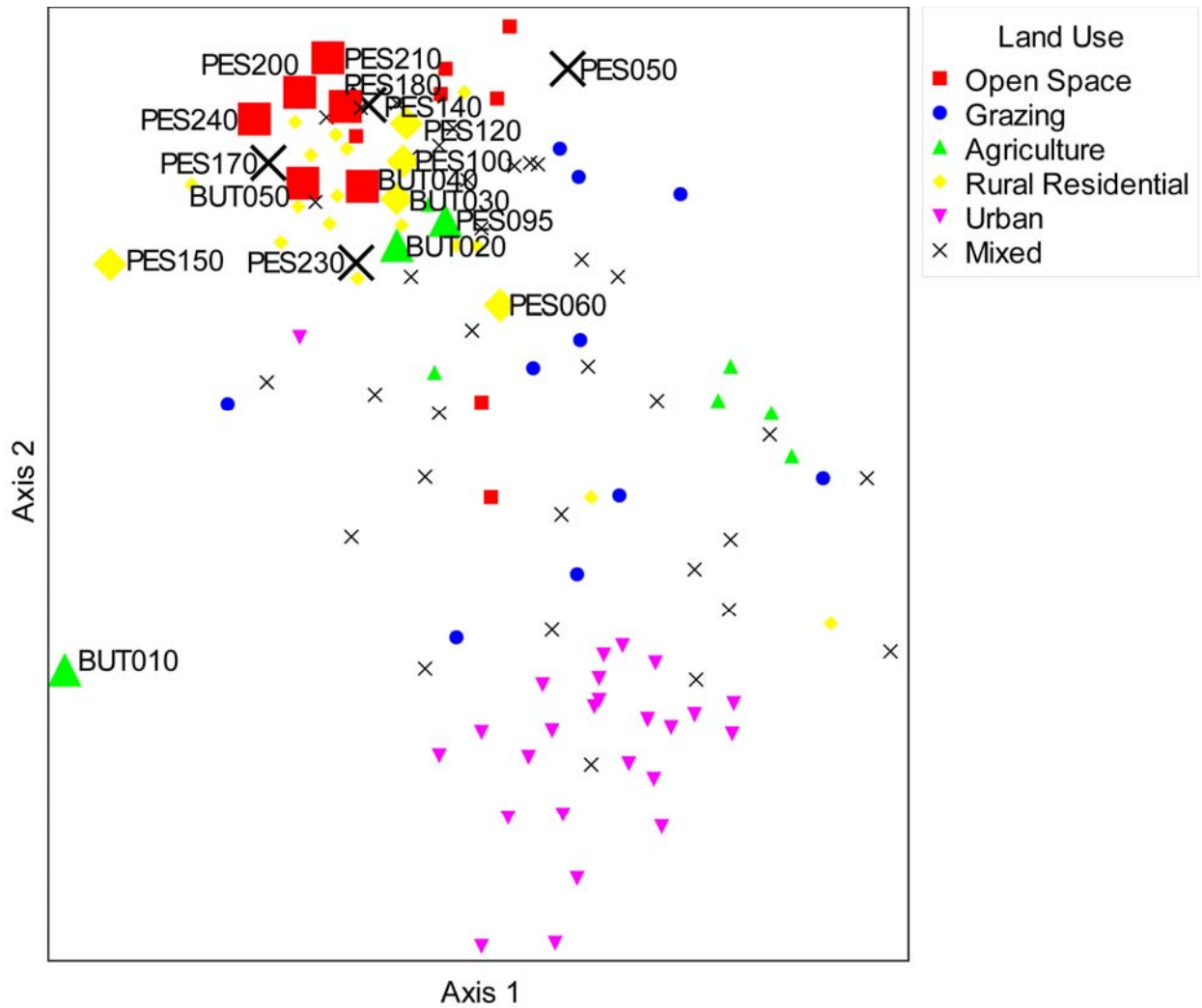


Figure 7-57. NMS ordination of sites in the Pescadero Creek/Butano Creek watershed

The NMS ordination of sites by macroinvertebrate taxa abundance. Sites in the Pescadero/Butano Creek watershed are labeled and symbols are enlarged. For an explanation of ordination graphs, refer to Section 6.1.

Table 7-24. Biological metrics in the Pescadero Creek/Butano Creek watershed

| | BUT 010 | BUT 020 | BUT 030 | BUT 040 | BUT 050 | PES 050 | PES 060 | PES 070 | PES 080 | PES 095 | PES 100 | PES 120 | PES 140 | PES 150 | PES 160 | PES 170 | PES 180 | PES 190 | PES 200 | PES 210 | PES 230 | PES 240 |
|--------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| Coleoptera Taxa | 2 | 2 | 5 | 8 | 7 | 3 | 5 | 3 | 3 | 5 | 5 | 7 | 6 | 6 | 7 | 6 | 10 | 7 | 8 | 9 | 8 | 8 |
| Diptera Taxa | 3 | 6 | 10 | 8 | 7 | 5 | 6 | 6 | 5 | 9 | 6 | 8 | 9 | 8 | 8 | 6 | 11 | 9 | 11 | 9 | 13 | 14 |
| Ephemeroptera Taxa | 4 | 10 | 10 | 10 | 12 | 9 | 10 | 9 | 7 | 10 | 11 | 10 | 11 | 7 | 10 | 10 | 12 | 11 | 10 | 11 | 7 | 11 |
| Plecoptera Taxa | 0 | 3 | 2 | 3 | 5 | 3 | 3 | 3 | 1 | 5 | 6 | 5 | 6 | 4 | 4 | 4 | 6 | 7 | 5 | 8 | 5 | 8 |
| Trichoptera Taxa | 2 | 7 | 7 | 8 | 7 | 5 | 8 | 10 | 4 | 10 | 9 | 8 | 9 | 5 | 9 | 6 | 8 | 7 | 9 | 9 | 5 | 12 |
| Non-Insect Taxa | 5 | 4 | 2 | 4 | 7 | 6 | 5 | 5 | 6 | 7 | 5 | 5 | 1 | 4 | 4 | 8 | 5 | 4 | 6 | 6 | 4 | 2 |
| EPT Taxa | 6 | 20 | 19 | 21 | 24 | 17 | 21 | 22 | 12 | 25 | 26 | 23 | 26 | 16 | 23 | 20 | 26 | 25 | 24 | 28 | 17 | 31 |
| Taxa Richness | 16 | 32 | 36 | 41 | 45 | 32 | 37 | 36 | 26 | 46 | 42 | 43 | 42 | 35 | 42 | 40 | 52 | 45 | 49 | 52 | 42 | 55 |
| % EPT | 25 | 41 | 67 | 62 | 70 | 22 | 42 | 38 | 33 | 55 | 76 | 78 | 84 | 39 | 69 | 69 | 69 | 73 | 48 | 41 | 47 | 62 |
| % Sensitive EPT | 10 | 22 | 47 | 37 | 41 | 11 | 16 | 19 | 18 | 35 | 58 | 54 | 70 | 38 | 54 | 37 | 41 | 49 | 31 | 23 | 40 | 47 |
| % Chironomidae | 42 | 37 | 12 | 9 | 6 | 14 | 36 | 29 | 46 | 29 | 4 | 2 | 6 | 10 | 5 | 3 | 12 | 5 | 7 | 5 | 26 | 8 |
| % Coleoptera | 12 | 8 | 7 | 15 | 10 | 5 | 9 | 7 | 4 | 6 | 9 | 7 | 7 | 5 | 17 | 15 | 8 | 13 | 34 | 35 | 18 | 17 |
| % Oligochaeta | 12 | 0 | 0 | 1 | 1 | 0 | 1 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 1 | 1 | 0 | 1 | 0 |
| % Non-insect | 20 | 3 | 0 | 2 | 2 | 56 | 5 | 1 | 5 | 1 | 1 | 2 | 0 | 39 | 1 | 9 | 1 | 2 | 2 | 12 | 2 | 2 |
| % Intolerant | 0 | 15 | 44 | 36 | 35 | 7 | 10 | 11 | 15 | 30 | 46 | 47 | 66 | 24 | 51 | 35 | 40 | 46 | 27 | 20 | 39 | 41 |
| % Tolerant | 8 | 3 | 0 | 0 | 1 | 56 | 3 | 1 | 1 | 1 | 1 | 1 | 0 | 39 | 0 | 3 | 1 | 0 | 1 | 12 | 1 | 2 |
| Tolerance Value | 5.3 | 4.8 | 3.4 | 3.5 | 3.5 | 6.4 | 4.8 | 4.8 | 5.0 | 4.0 | 3.2 | 3.1 | 2.1 | 5.3 | 2.6 | 3.3 | 3.4 | 3.1 | 3.3 | 4.1 | 3.7 | 3.2 |
| % Predator | 7 | 5 | 7 | 5 | 19 | 11 | 7 | 5 | 3 | 6 | 8 | 8 | 8 | 25 | 8 | 22 | 12 | 11 | 11 | 9 | 7 | 27 |
| % Coll.-filterer | 1 | 12 | 13 | 9 | 10 | 4 | 10 | 26 | 11 | 11 | 13 | 12 | 1 | 7 | 5 | 1 | 10 | 7 | 2 | 4 | 2 | 6 |
| % Coll.-gatherer | 67 | 59 | 52 | 41 | 26 | 68 | 65 | 50 | 59 | 65 | 57 | 57 | 64 | 46 | 41 | 19 | 42 | 36 | 26 | 26 | 55 | 26 |
| % Scraper | 19 | 21 | 16 | 29 | 31 | 16 | 14 | 12 | 13 | 11 | 15 | 11 | 21 | 10 | 37 | 48 | 27 | 38 | 44 | 52 | 20 | 33 |
| % Shredder | 6 | 3 | 11 | 8 | 14 | 1 | 4 | 6 | 13 | 6 | 7 | 4 | 5 | 12 | 8 | 11 | 5 | 6 | 17 | 8 | 15 | 9 |
| % Other | 1 | 0 | 1 | 7 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 9 | 1 | 0 | 1 | 0 | 5 | 3 | 0 | 1 | 1 | 0 |
| Com. Tol. (%) | 63 | 55 | 39 | 37 | 19 | 19 | 49 | 61 | 67 | 48 | 25 | 23 | 10 | 15 | 12 | 10 | 29 | 15 | 14 | 15 | 32 | 14 |
| Com. Int. Taxa | 3 | 9 | 14 | 21 | 19 | 4 | 9 | 10 | 6 | 15 | 16 | 16 | 18 | 12 | 15 | 17 | 21 | 21 | 21 | 20 | 17 | 21 |

7.7.4 Basic Water Quality Field Measurements

Stream temperatures exceeded some thresholds at about half of the continuously monitored sites, all during the dry season (Figure 7-, Figure 7-, and Appendix D). HWAT values exceeded the MWAT threshold for coho salmon (14.8°C) in Pescadero Creek at Memorial Park (PES140) and the Canyon Mouth (PES105). The higher MWAT threshold for steelhead (17°C) was exceeded in Pescadero Creek at the Community Church (PES060) and at the USGS Gage (PES100). In total, a temperature threshold was exceeded at four of eleven sites, but only in the dry season.

Dissolved oxygen concentrations fell below Basin Plan objectives in the dry season at eight sites: BUT010, PES060, PES100, PES140, PES150, PES170, PES190, and PES240 (see Figure 7-, Figure 7-, and Appendix D).

There were no measured exceedances of pH objectives, and specific conductance was low throughout the watershed.

Physical habitat parameters distinguished two sites in the lower watershed (PES060 and PES100) from three sites further upstream (PES140, PES150 and PES170). The lower sites had relatively low scores for channel alteration, riparian zone width, and canopy cover, while the upstream sites had very high scores for these parameters. There was a significant correlation between stream temperature and physical habitat parameters among these sites, as illustrated for canopy cover (Figure 7-). This trend may reflect a natural downstream pattern of decreased canopy cover with increasing stream width, however, and should not necessarily be considered a sign of poor habitat.

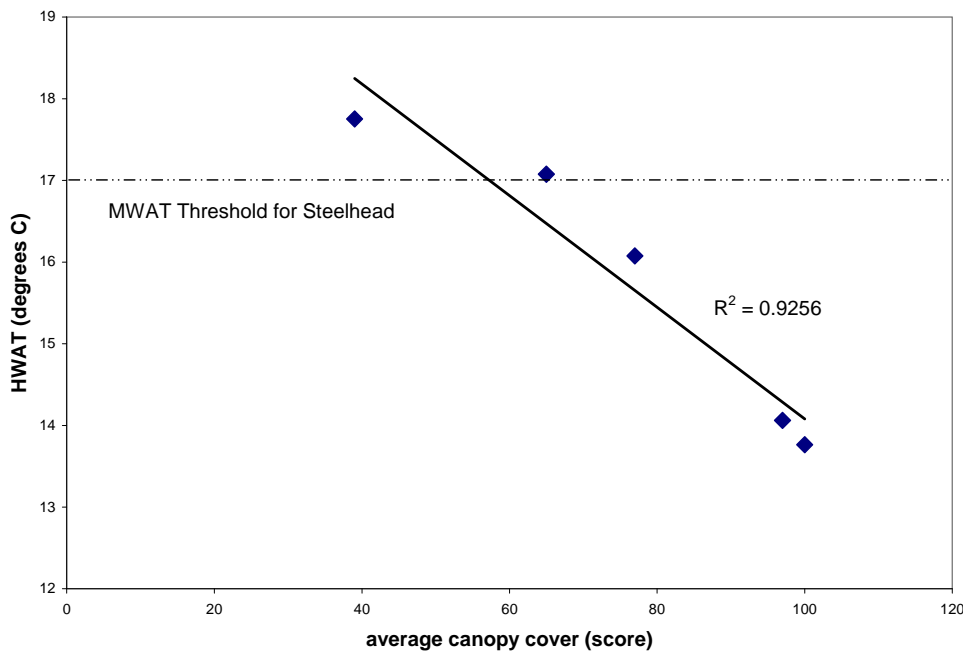


Figure 7-58. Riparian canopy cover and stream temperature in Pescadero Creek

MWAT is maximum weekly average temperature. HWAT is highest weekly average temperature for the period sampled.

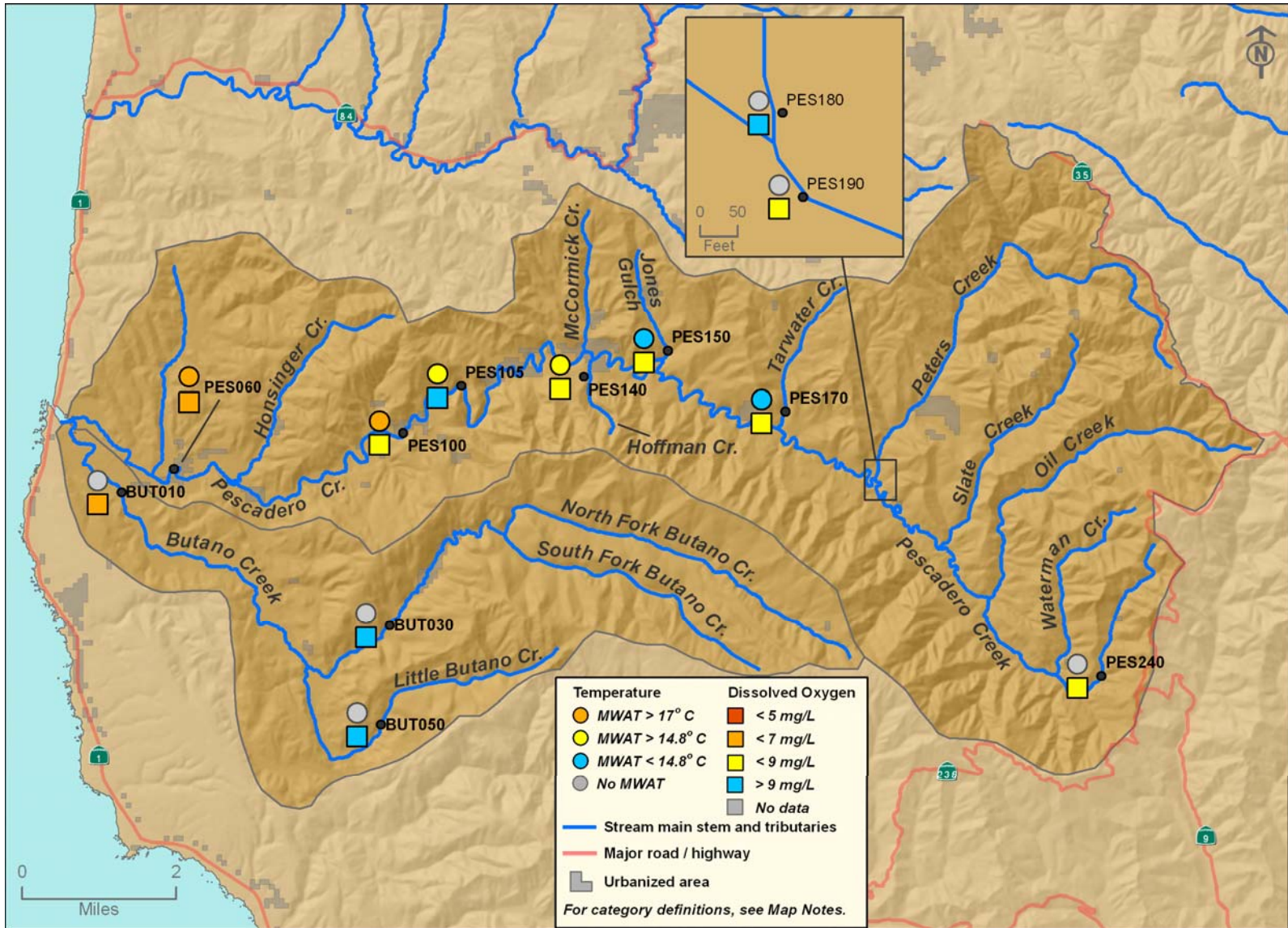


Figure 7-59. Map of temperature and DO levels in Pescadero and Butano creeks

Temperature--Pescadero/Butano Creeks

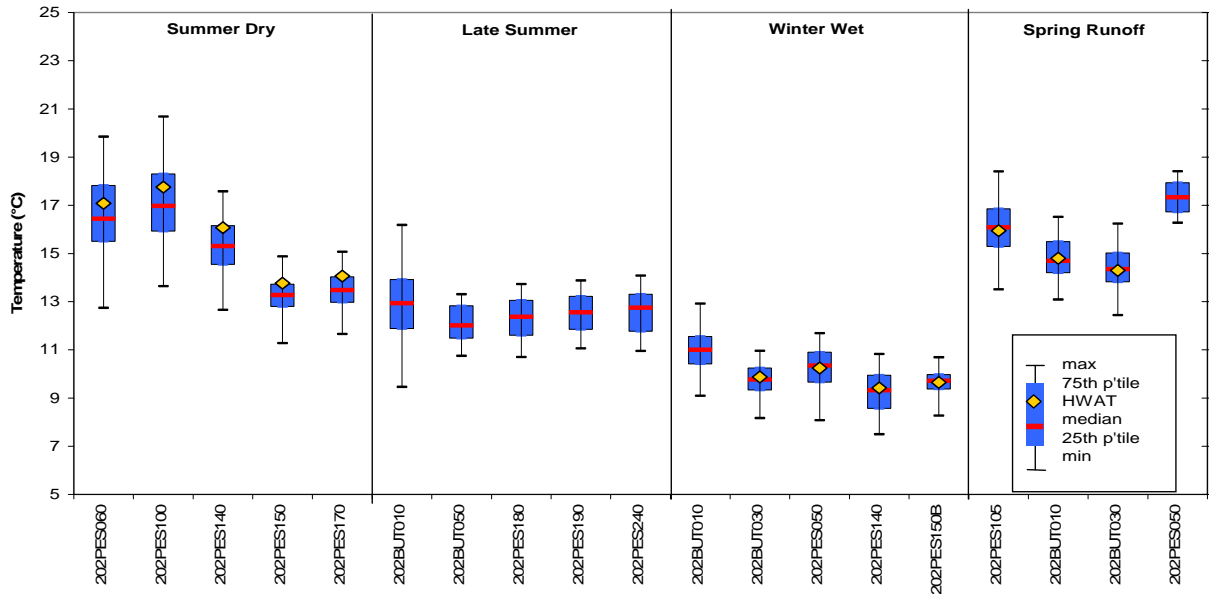


Figure 7-60. Temperature monitoring in Pescadero and Butano creeks

Dissolved Oxygen--Pescadero/Butano Creeks

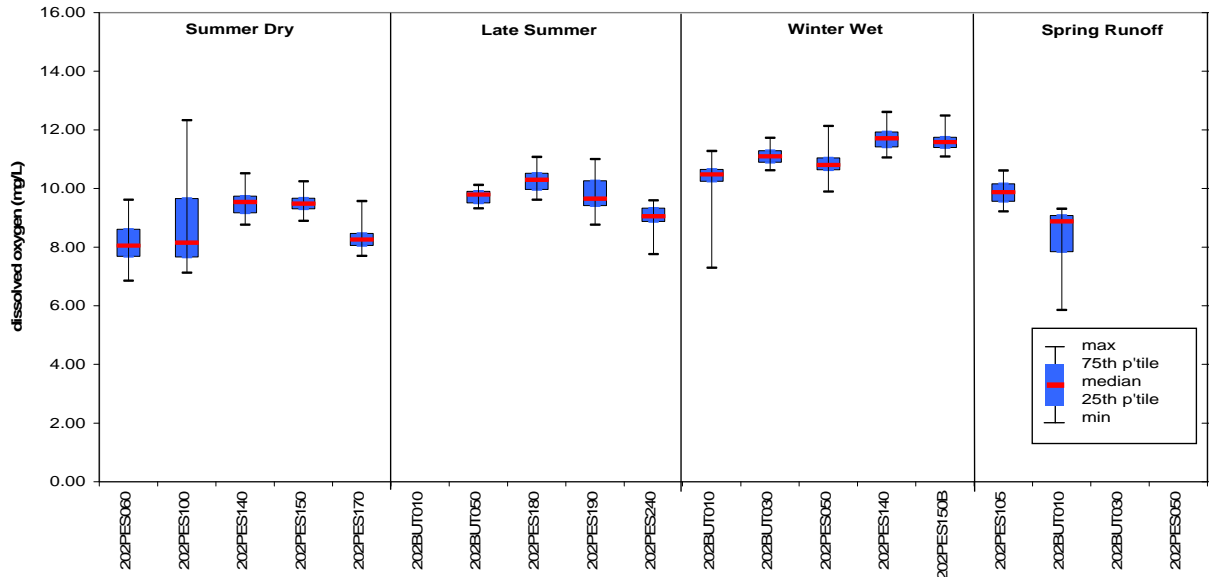


Figure 7-61. DO monitoring in Pescadero and Butano creeks

Data were rejected from BUT010 in the late summer and from BUT030 and PES050 in the spring.

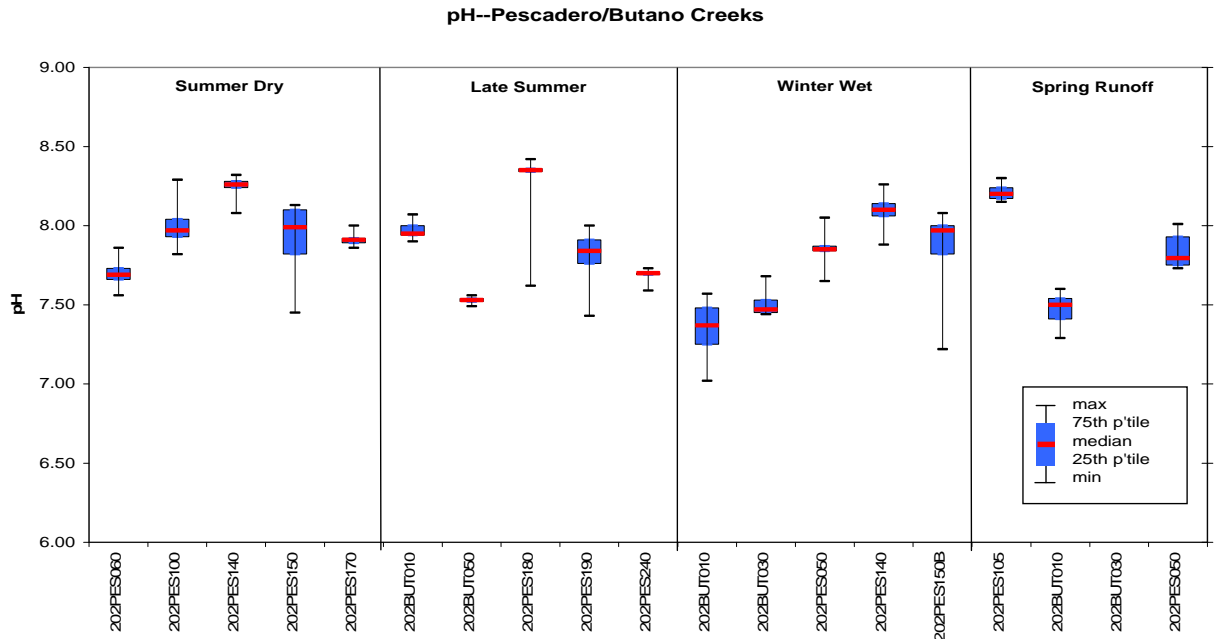


Figure 7-62. pH monitoring in Pescadero and Butano creeks

Data were rejected from BUT030 in the spring.

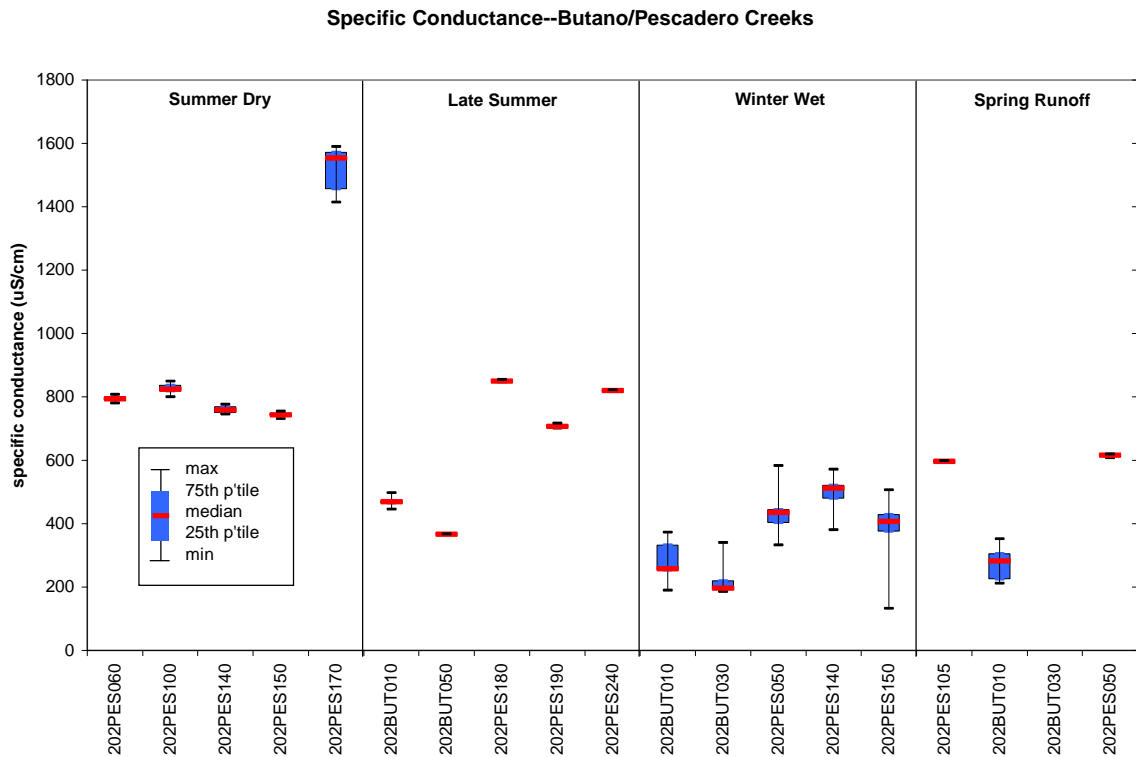


Figure 7-63. Specific conductance monitoring in Pescadero and Butano creeks

Data were rejected from BUT030 in the spring.

7.7.5 Water, Sediment, and Clam Tissue Chemistry

Nutrient and chlorophyll-a concentrations in water samples were relatively low throughout the watershed. Chlorophyll-a exceeded the 1.78 µg/L guideline value at one site (BUT010) in dry and spring seasons. About half of the nitrate values exceeded the U.S. EPA reference guideline of 0.155 mg/L.

Total phosphorus measurements were also moderate, 0.17 mg/L or less, but, most were significantly above the U.S. EPA reference guideline.

Orthophosphate concentrations were generally low to moderate, with all measurements less than 0.15. All samples in Pescadero Creek were above the 0.030 mg/L U.S. EPA total phosphorus guideline, but only one exceedance was observed in Butano Creek.

Metal and organic contaminants were measured in water samples from three sites: BUT010, PES050, and PES070. Land use at these sites was characterized as agricultural and mixed use. There were no exceedances of Basin Plan aquatic life water quality objectives in any samples from this watershed.

Sediments were collected for analysis from two sites, BUT010 and PES050. The only exceedance of a guideline value was chromium measured above the threshold effects concentration (TEC) at PES050. Nickel did not exceed any guideline values. The sediment quality guideline quotients values were the lowest measured in this study: 0.04 at BUT010 and 0.07 at PES050.

Tissues of clams deployed at these two sites had low concentrations of all measured compounds.

7.7.6 Water and Sediment Toxicity

Three-species toxicity tests were conducted on water samples from the same three sites as metal and organic chemistry: BUT010, PES050, and PES070. No significant toxicity was observed for any of the water samples (Figure 7-).

There were statistically significant effects on amphipod growth in sediments collected from BUT010 and PES050. While growth integrates a number of physiological processes, the relationship of the amphipod growth endpoint to specific stressors is not well established. These samples had the study's lowest concentrations of contaminants for which guideline values have been established, so it is not clear whether the depressed amphipod growth was related to chemical contamination.

Toxicity Tests in Pescadero/Butano Creeks Watershed--Response Relative to Control

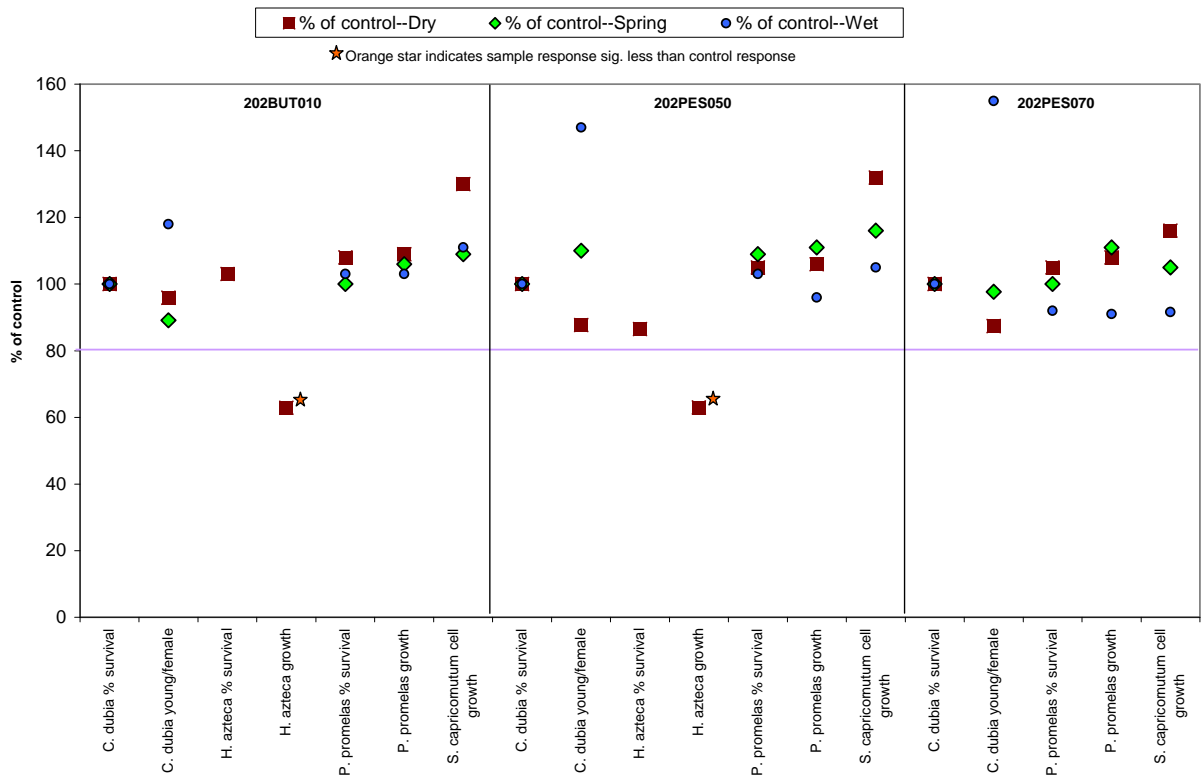


Figure 7-64. Toxicity tests from Pescadero and Butano creeks

Results of toxicity tests are shown, with three species in water samples and one species in sediment samples. See text for description of species and endpoints. Stars indicate statistically significant differences from test controls. Samples with stars that were also below 80 percent of the control were considered toxic for this assessment.

7.7.7 Coliform Bacteria

Samples were analyzed for coliform bacteria at eight sites in the watershed (Figure 3-7). As noted above, Basin Plan objectives for all three indicators, *E. coli*, fecal coliforms, and total coliforms were exceeded at only one site, Pescadero Creek at Community Church (PES060). Basin Plan objectives for total coliforms were exceeded at each of the eight sites: Butano Creek at Butano Falls (BUT040), Little Butano Creek at Butano State Park (BUT050), Pescadero Creek at Community Church (PES060), Memorial Park Swim 1 (PES134), Memorial Park Swim 2 (PES135), Jones Gulch (PES150), Iverson Trail (PES193), and Sequoia Nature Trail (PES194).

7.8 San Gregorio Creek Watershed

7.8.1 Sites of Concern

Exceedances noted in the summary table (Table 7-1725) are discussed in the relevant sub-section for this watershed. A weight-of-evidence approach was used to evaluate sites of concern. Individual exceedances of parameters, particularly those without regulatory objectives, are considered in context. In general, nickel and chromium levels are geologically elevated in the Bay Area, particularly where there are serpentine deposits. Nutrient levels (nitrate and phosphorus) are above U.S. EPA guidelines at most sites, but may still be relatively low.

Results of this survey did not indicate any specific sites of concern for nutrients or contaminants; however, coliform bacteria levels at Playbowl on La Honda Creek (SGR100) and San Gregorio Confluence (SGR079) did exceed coliform objectives and bear further investigation. Temperature and dissolved oxygen levels were supportive of salmonids at all sites except the downstream site (SGR010) during the dry season and Sky Londa (SGR120) in the fall. The unusual pattern of temperature, DO, and specific conductance noted at Sky Londa in the fall (see section 6.2.2.7) warrants further investigation.

Table 7-25. Summary of sites with exceedances in San Gregorio Creek watershed

| <i>Analyte</i> | <i>Standard</i> | <i># of sites</i> |
|--|-----------------|-------------------|
| FIELD MEASURES of 6 sites sampled | | |
| Temperature | MWAT, coho | 1 |
| | minimum, COLD | 1 |
| Oxygen, dissolved | minimum, WARM | 1 |
| | 3-month median | 1 |
| CONVENTIONAL WATER QUALITY of 4 sites sampled | | |
| Nitrate as N | maximum | 4 |
| Phosphorus, total as P | maximum | 4 |
| SEDIMENT CHEMISTRY of 1 site sampled | | |
| <i>Metals</i> | | |
| Chromium | TEC | 1 |
| Nickel | TEC | 1 |
| SEDIMENT TOXICITY of 1 site sampled | | |
| <i>Hyalella</i> | growth | 1 |
| COLIFORMS of 2 sites sampled | | |
| <i>E. coli</i> | steady state | 2 |
| Fecal coliform | log mean | 1 |
| Total coliform | median | 2 |
| | maximum | 1 |

7.8.2 Water Quality in Relation to Land Use

Land use in the watershed is primarily open space and rural residential. Only two sites (SGR010 and SGR040) are categorized as mixed use; the rest are categorized as rural residential. No trends between measured parameters and land use were apparent or likely be discerned, given the low variability among measured parameters in the watershed data set.

7.8.3 Macroinvertebrate Assemblages and Physical Habitat

Two groups of benthic assemblages were identified in the San Gregorio Creek watershed from the cluster dendrogram (Figure 7-) and ordination plot (Figure 7-):

- The Mainstem (MS) group is composed of two sites on San Gregorio Creek (SGR040 and SGR075) and one site on lower La Honda Creek (SGR080).
- The Tributary (TR) group includes sites on small tributary streams including El Corte de Madera Creek (SGR030), Harrington Creek (SGR060), Alpine Creek (SGR090, SGR150), Mindego Creek (SGR130), and upper La Honda Creek (SGR110, SGR120).
- The lowest site on San Gregorio Creek, SGR010, was sufficiently different from other sites that it is considered separately.

The Mainstem group

Land use in the San Gregorio Creek watershed is primarily protected open space, with small amounts of rural residential, commercial, grazing, and agriculture land use. Very high numbers of sensitive and intolerant organisms were present at these sites, suggesting excellent water quality and habitat conditions. The tolerance value of 2.1 at SGR040 is the lowest measured value in the Bay Area, while the percent intolerant values of 45-64 percent are well above the median value from minimally disturbed perennial sites (33 percent). Functional diversity was very high, with all five primary functional feeding groups well represented (Table 7-27). Although few rare taxa were found at these sites, the dominance of common intolerant taxa indicates high biological integrity.

Benthic assemblages at these sites were numerically dominated by EPT taxa (68-81 percent), including *Ephemerella* sp. (14 percent), *Cinygmula* sp. (10 percent), *Lepidostoma* sp. (9 percent), and *Epeorus* sp. (8 percent).

The Tributary group

These sites almost exclusively drain open space, with minor amounts of rural residential land use. Biological metric values, such as taxa richness (36-51) and tolerance value (2.8-4.3) generally fall within the range of minimally disturbed conditions (Table 6-10), suggesting that tributaries in the San Gregorio Creek watershed had excellent biological integrity and water quality conditions.

The benthic assemblages are dominated by mayflies, including:

| | |
|---|-----|
| <i>Baetis</i> sp. (Baetidae) | 20% |
| <i>Cinygmula</i> sp. (Heptageniidae) | 10% |
| <i>Epeorus</i> sp. (Heptageniidae) | 8% |
| <i>Ephemerella</i> sp. (Ephemerellidae) | 5% |
| <i>Serratella</i> sp. (Ephemerellidae) | 4% |
| <i>Drunella</i> sp. (Ephemerellidae) | 2% |

Other abundant taxa included:

| | |
|--|----|
| Chironomidae | 7% |
| <i>Optioservus</i> sp. (Elmidae) | 5% |
| <i>Calineuria californica</i> (Perlidae) | 4% |
| <i>Malenka</i> sp. (Nemouridae) | 4% |
| <i>Simulium</i> sp. (Simuliidae) | 4% |

Several rare taxa not found elsewhere in the Bay Area were collected at these sites, including the elmid beetle *Ampumixis dispar*, *Brachycera* sp. (Diptera: Brachycera), and the water mite *Testudacarus* sp. (Acari: Torrenticolidae).

Lower San Gregorio Creek Site

The richness and sensitivity of taxa at the lowest site on San Gregorio Creek, SGR010, suggests that this site provides a rare depiction of a minimally disturbed benthic assemblage from a large perennial stream in the Bay Area. The benthic assemblage here included many rare taxa not found at other sites in the watershed, such as the beetles *Oreodytes* sp. (Dytiscidae) and *Brychius* sp. (Halipidae), the caddisfly *Oecetis disjuncta* (Leptoceridae), *Atrichopogon* sp. (Diptera: Ceratopogonidae), and the water mite *Corticacarus* sp. (Hygrobatidae). These taxa are most likely present as a result of complex habitat, such as undercut banks and overhanging riparian vegetation, associated with minimally disturbed low gradient channels. Dominant taxa included Chironomidae (18 percent), *Optioservus* sp. (18 percent), and the mayflies *Ephemerella* sp. (8 percent) and *Dipheter hageni* (7 percent).

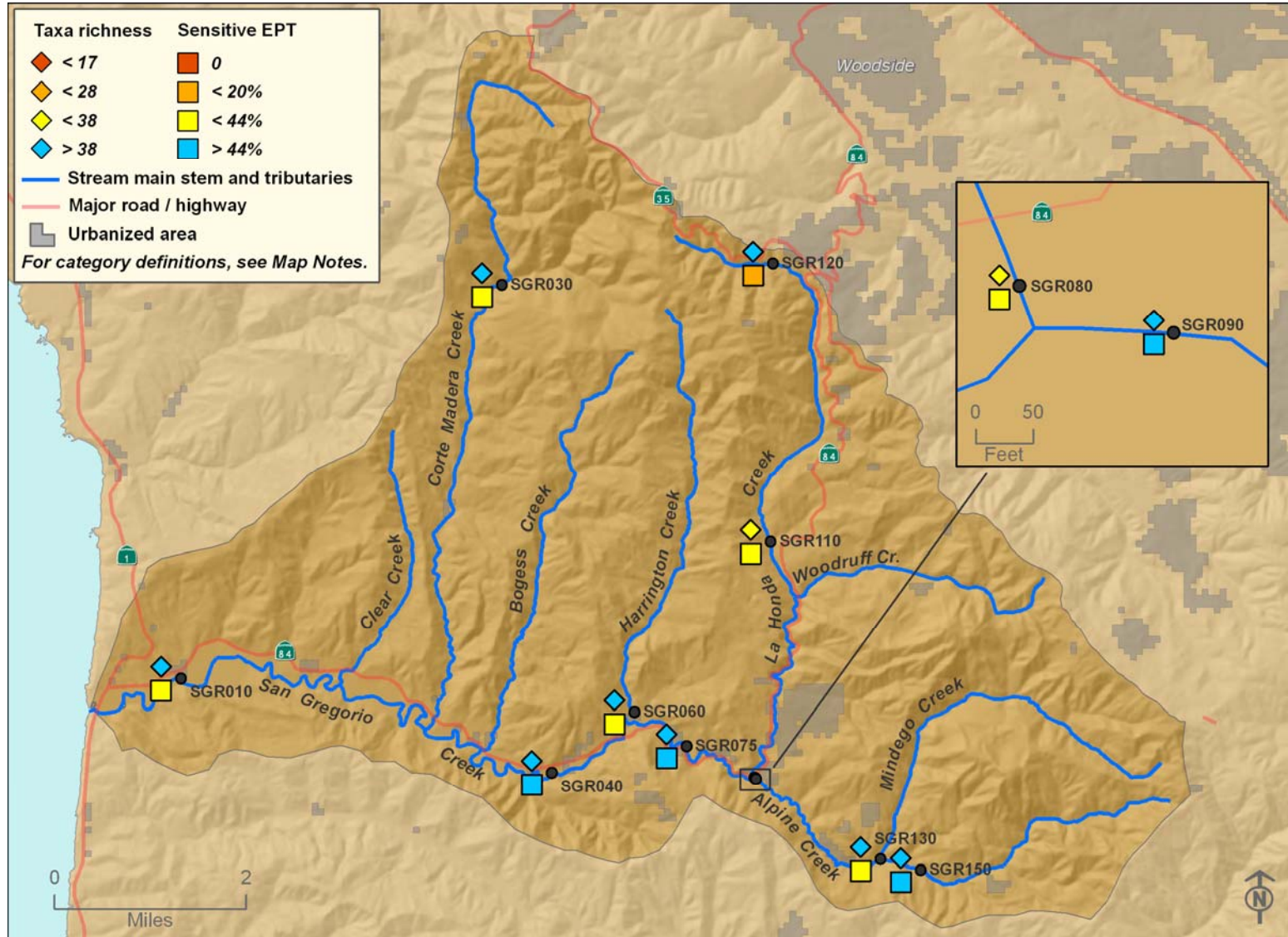


Figure 7-65. Map of San Gregorio Creek benthic macroinvertebrate index scores

Table 7-26. Bioassessment sites in the San Gregorio Creek watershed

| Site ID | Site Name | Stream |
|---------|--------------------|--------------------------|
| SGR010 | SG USGS Gage | San Gregorio Creek |
| SGR030 | Star Hill Rd. | El Corte de Madera Creek |
| SGR040 | Boysville | San Gregorio Creek |
| SGR060 | Harrington | Harrington Creek |
| SGR075 | Upper San Gregorio | San Gregorio Creek |
| SGR080 | La Honda Low | La Honda Creek |
| SGR090 | Alpine Low | Alpine Creek |
| SGR110 | Spanish Ranch | La Honda Creek |
| SGR120 | Sky Londa | La Honda Creek |
| SGR130 | Mindego | Mindego Creek |
| SGR150 | Heritage Grove | Alpine Creek |

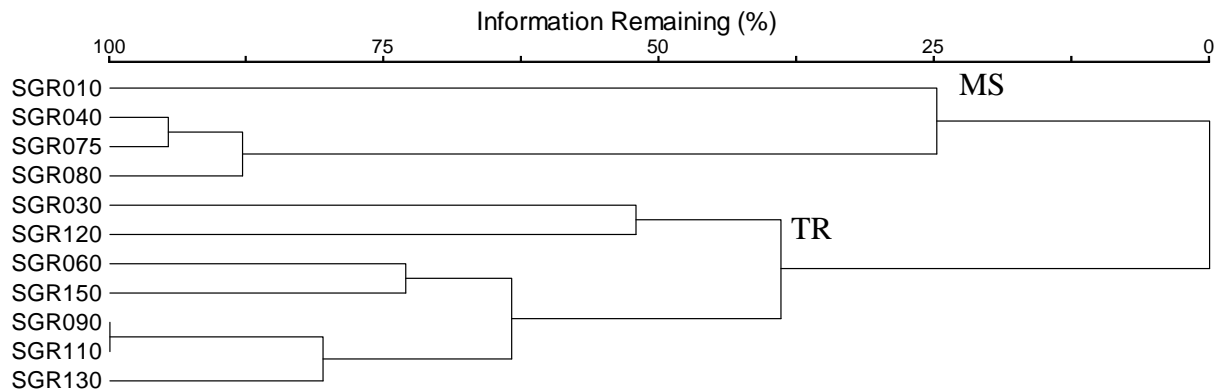


Figure 7-66. Cluster dendrogram of sites in the San Gregorio Creek watershed

This cluster dendrogram groups sites based on similarity of macroinvertebrate taxa present in the San Gregorio Creek watershed. The Information Remaining axis at the top signifies how similar the sites are with respect to taxa presence. Abbreviations: MS, Mainstem; TR, Tributary

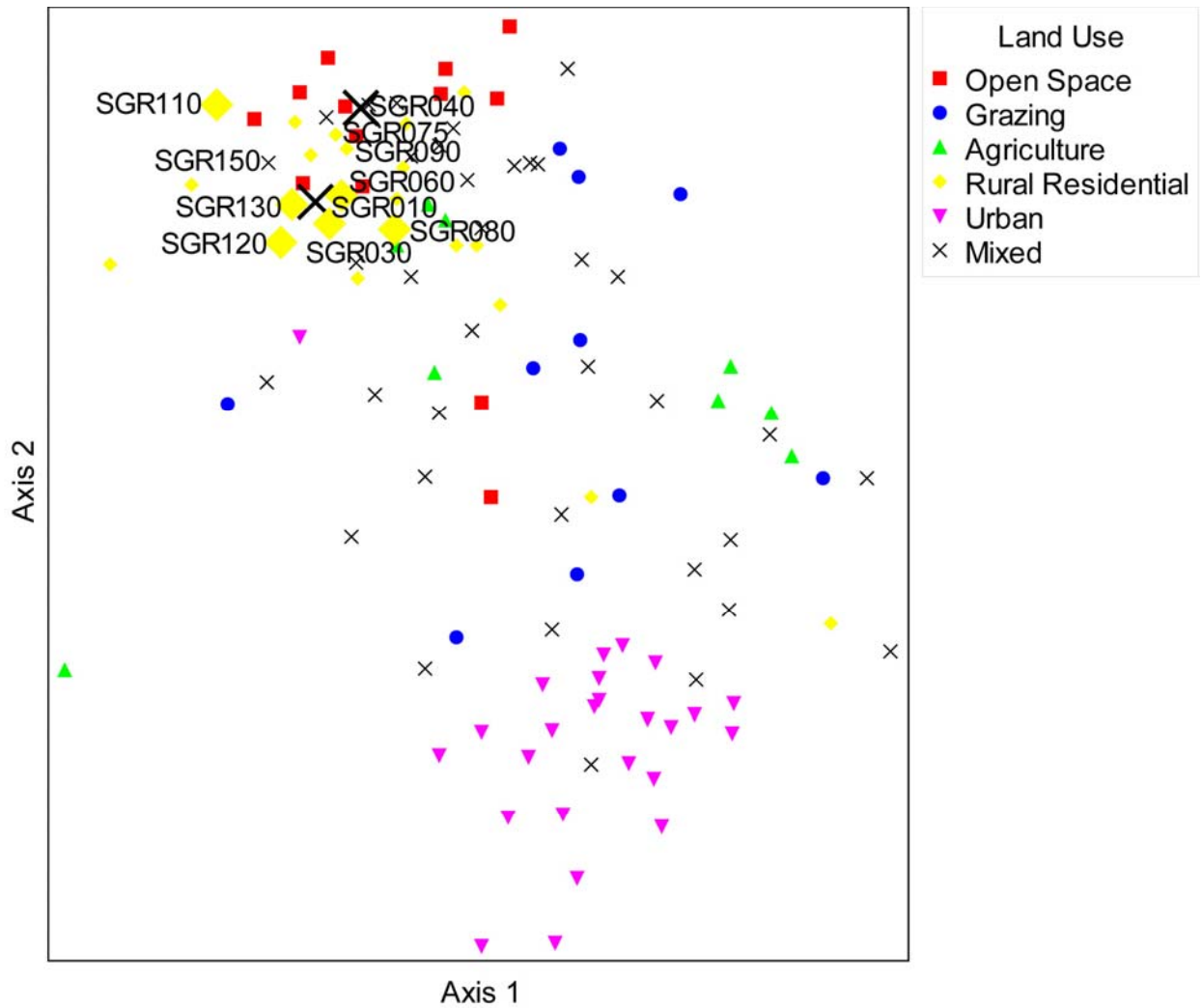


Figure 7-67. NMS ordination of sites in the San Gregorio Creek watershed

The NMS ordination of sites by macroinvertebrate taxa abundance. Sites in the San Gregorio Creek watershed are labeled and symbols are enlarged. For an explanation of ordination graphs, refer to Section 6.1.

Table 7-27. Biological metrics from San Gregorio Creek watershed

| | SGR010 | SGR030 | SGR040 | SGR060 | SGR075 | SGR080 | SGR090 | SGR110 | SGR120 | SGR130 | SGR150 |
|--------------------------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| Coleoptera Taxa | 7 | 8 | 5 | 8 | 5 | 5 | 6 | 6 | 5 | 7 | 9 |
| Diptera Taxa | 13 | 8 | 7 | 8 | 9 | 9 | 6 | 6 | 10 | 11 | 13 |
| Ephemeroptera Taxa | 11 | 10 | 11 | 11 | 12 | 9 | 11 | 10 | 10 | 11 | 9 |
| Plecoptera Taxa | 4 | 3 | 5 | 5 | 4 | 5 | 6 | 5 | 4 | 4 | 6 |
| Trichoptera Taxa | 10 | 8 | 10 | 5 | 8 | 6 | 7 | 7 | 9 | 6 | 6 |
| Non-Insect Taxa | 9 | 6 | 4 | 8 | 5 | 3 | 2 | 2 | 6 | 2 | 8 |
| EPT Taxa | 25 | 21 | 26 | 21 | 24 | 20 | 24 | 22 | 23 | 21 | 21 |
| Taxa Richness | 54 | 43 | 42 | 45 | 43 | 37 | 38 | 36 | 44 | 41 | 51 |
| % EPT | 44 | 76 | 81 | 69 | 79 | 68 | 84 | 73 | 69 | 71 | 75 |
| % Sensitive EPT | 32 | 35 | 65 | 41 | 58 | 46 | 50 | 43 | 17 | 38 | 53 |
| % Chironomidae | 18 | 5 | 1 | 7 | 4 | 6 | 5 | 4 | 13 | 7 | 7 |
| % Coleoptera | 19 | 11 | 6 | 5 | 9 | 4 | 3 | 19 | 7 | 8 | 8 |
| % Oligochaeta | 0 | 3 | 0 | 1 | 1 | 3 | 0 | 0 | 0 | 1 | 1 |
| % Non-insect | 10 | 4 | 1 | 3 | 2 | 3 | 0 | 0 | 1 | 2 | 3 |
| % Intolerant | 22 | 31 | 64 | 37 | 56 | 45 | 47 | 39 | 16 | 40 | 49 |
| % Tolerant | 9 | 1 | 1 | 2 | 1 | 0 | 0 | 0 | 1 | 1 | 1 |
| Tolerance Value | 4.3 | 3.7 | 2.1 | 3.6 | 2.5 | 3.2 | 2.8 | 2.9 | 4.3 | 3.3 | 2.8 |
| % Predator | 18 | 5 | 8 | 9 | 11 | 7 | 18 | 10 | 7 | 17 | 22 |
| % Collector-filterer | 4 | 2 | 7 | 14 | 5 | 21 | 10 | 6 | 5 | 16 | 3 |
| %Collector-gatherer | 44 | 63 | 26 | 45 | 39 | 34 | 28 | 37 | 72 | 28 | 38 |
| % Scraper | 24 | 17 | 30 | 27 | 26 | 21 | 32 | 41 | 10 | 35 | 31 |
| % Shredder | 9 | 4 | 26 | 3 | 14 | 12 | 11 | 5 | 7 | 3 | 5 |
| % Other | 1 | 9 | 3 | 1 | 4 | 6 | 1 | 1 | 0 | 0 | 1 |
| Common Tolerant (%) | 22 | 47 | 7 | 34 | 7 | 31 | 17 | 21 | 65 | 27 | 12 |
| Common Intolerant (Taxa) | 13 | 17 | 16 | 19 | 16 | 15 | 18 | 18 | 17 | 20 | 21 |

7.8.4 Basic Water Quality Field Measurements

Stream temperatures were measured at six sites (Figure 7-). HWAT values exceeded the MWAT threshold for coho salmon (14.8°C) at only one site, San Gregorio Creek at the USGS Gage (SGR010), during the dry season. Dissolved oxygen concentrations fell below 9 mg/L at sites SGR010, SGR040, SGR090, and SGR120, all during the dry season. All but SGR120 remained above cold water habitat limits of 7mg/L.

In the fall, SGR120 had an unusual pattern of midnight temperature spikes coincident with dramatic decreases of DO and increases of specific conductance, indicating periodic discharges (see Figures 6-13 and 6-14 and the discussion in section 6.2.2.7).

There were no measurements above pH guidelines.

Dissolved salt content, as measured by specific conductance, exceeded the objective for drinking water taste at three sites during the dry season, SGR010, SGR040, SGR100 (Figure 7- through Figure 7-; and Appendix D).

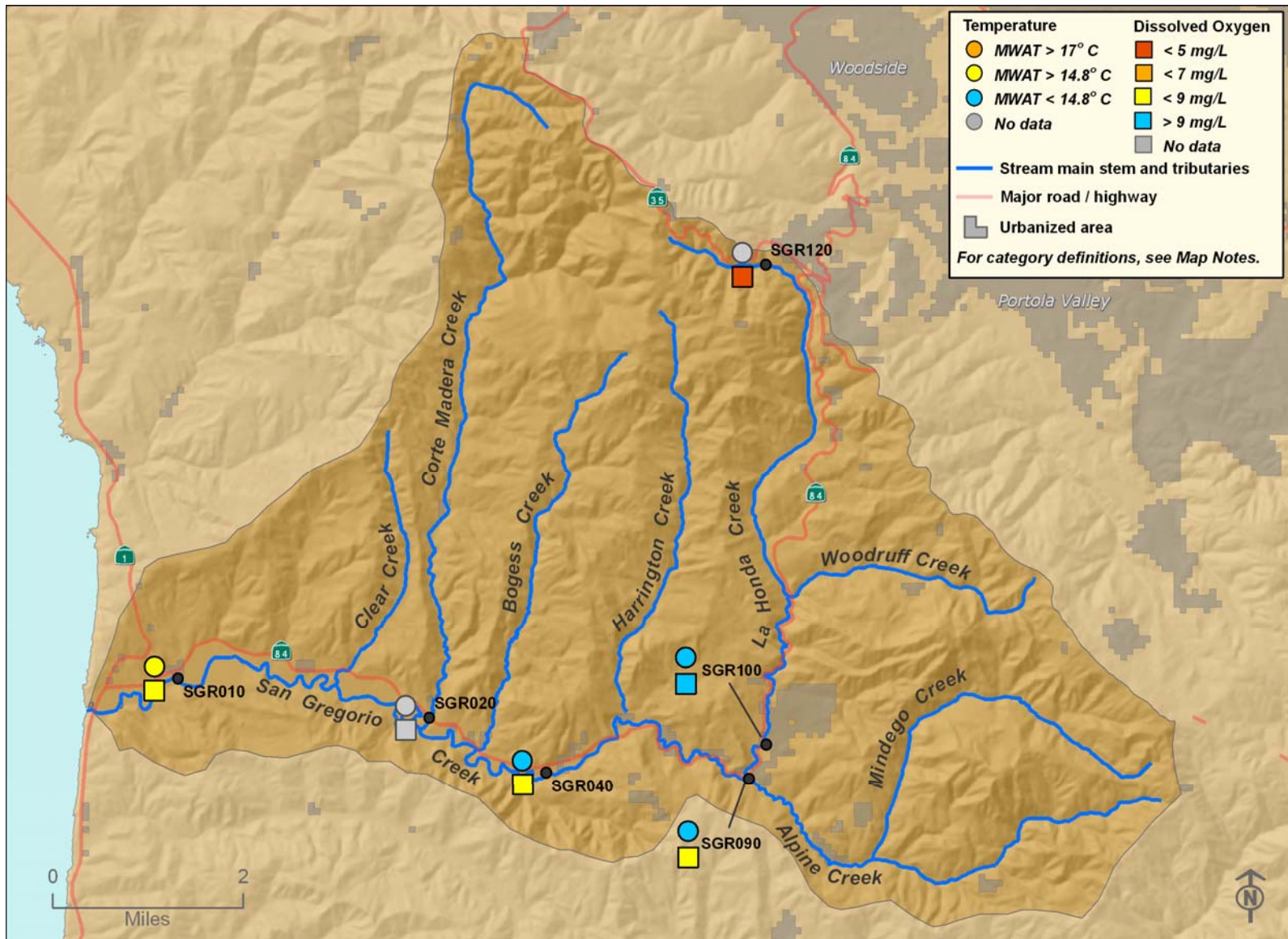


Figure 7-68. Map of temperature and DO levels in San Gregorio Creek watershed

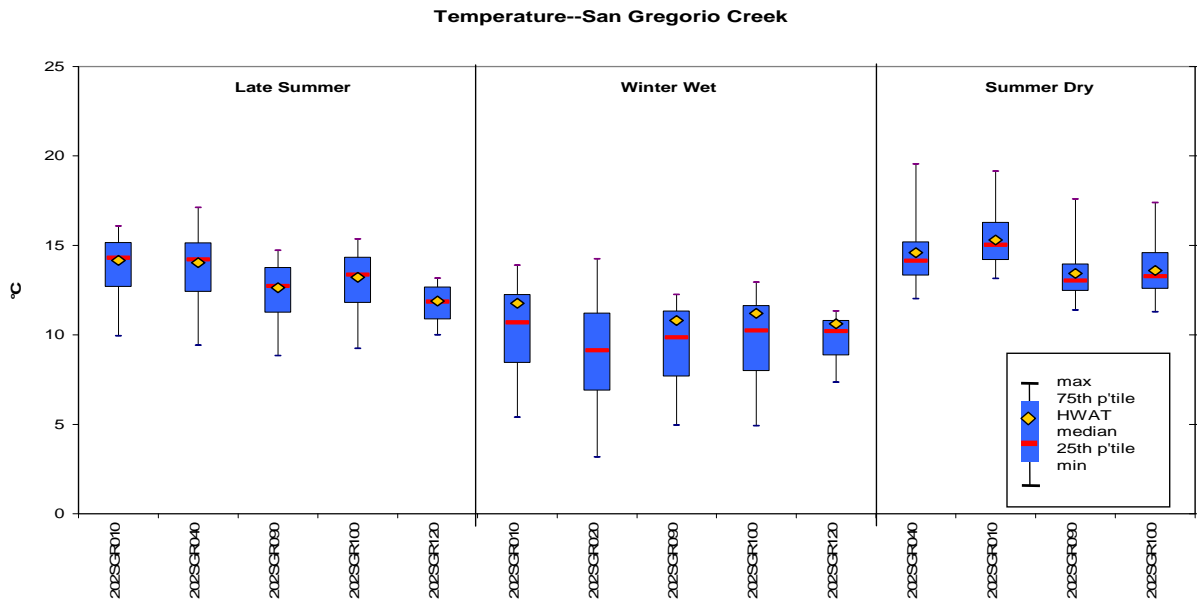


Figure 7-69. Temperature monitoring in San Gregorio Creek watershed

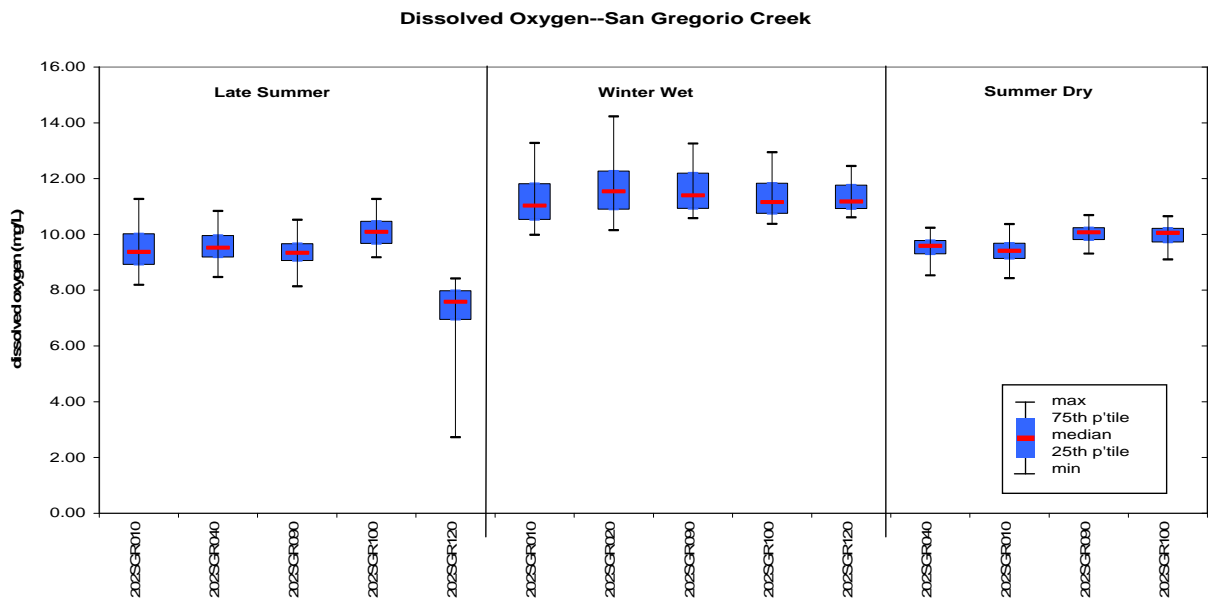


Figure 7-70. DO monitoring in San Gregorio Creek watershed

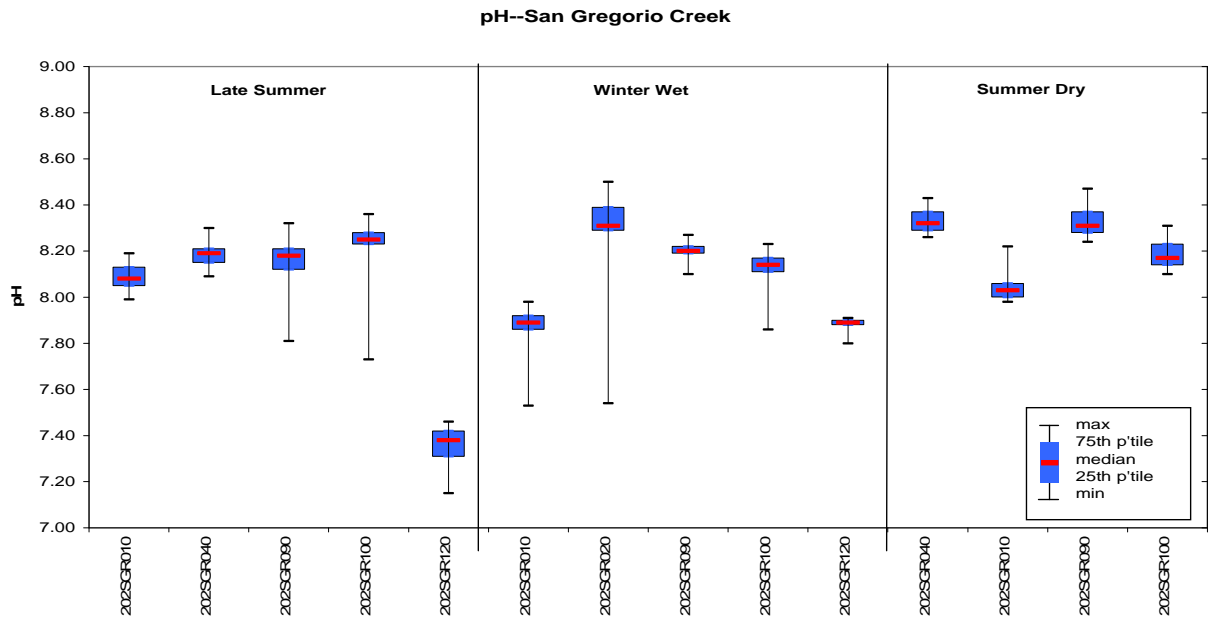


Figure 7-71. pH monitoring of sites in San Gregorio Creek watershed

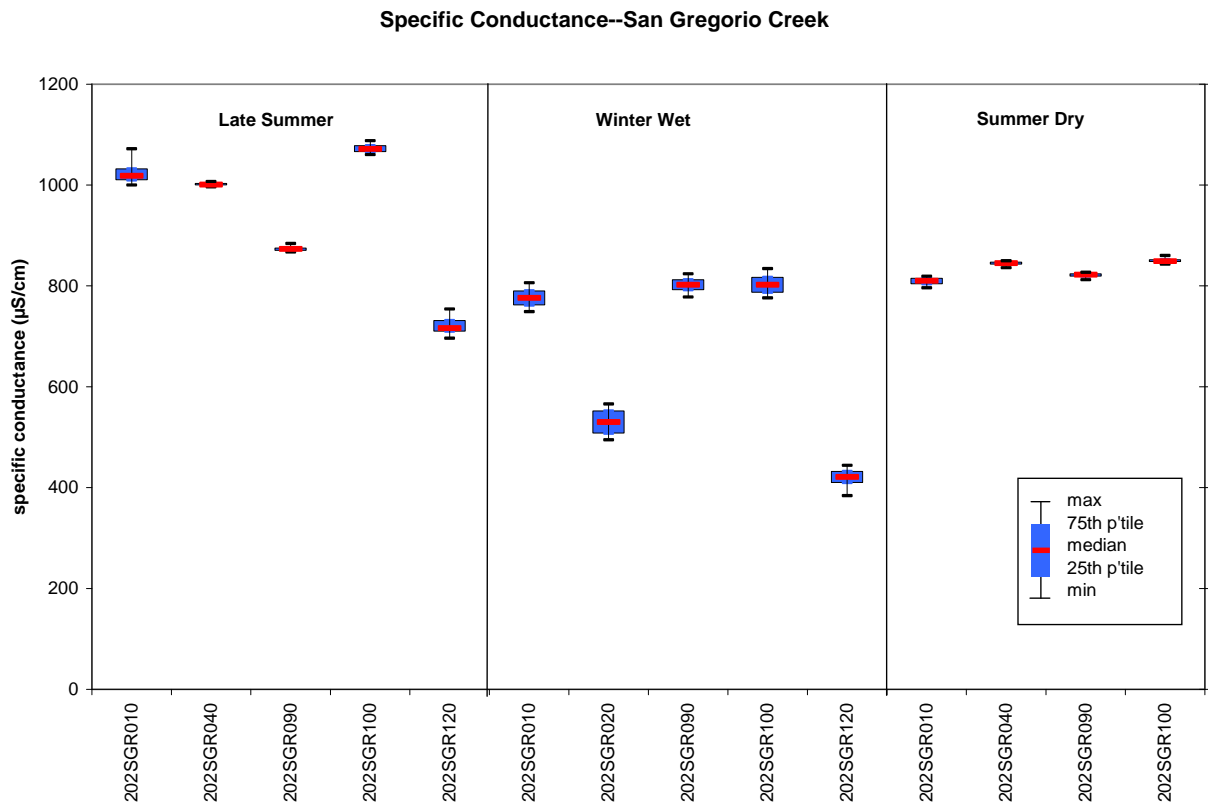


Figure 7-72. Specific conductance monitoring in San Gregorio Creek watershed

7.8.5 Water, Sediment, and Clam Tissue Chemistry

Nutrient concentrations in water samples were generally low, although about half of the samples exceeded the U.S. EPA reference guideline (Figures 6–15 to 6–17; Appendix G).

Chlorophyll-a values were also generally low, with only two measurements above the 1.78 µg/L guideline value (Appendix G).

Total phosphorus measurements were relatively high, all substantially above the 0.030 U.S. EPA reference guideline value. Orthophosphate values were also high, exceeding the total phosphorus guideline in all cases. The elevated phosphorus values are not of great concern, however, since San Gregorio Creek appears to be nitrogen-limited.

Measured contaminant concentrations in water and sediment were very low; there were no exceedances of aquatic life water quality objectives in any sample from this watershed. Metal and organic contaminants were analyzed in water from two sites: San Gregorio Creek at the USGS Gage (SGR010) and La Honda Creek at the Confluence (SGR080). The only sediment chemicals measured above guideline values were chromium and nickel, both above the threshold effects concentration (TEC). The SQGQ value for this sample was a very low 0.09.

Tissues of clams deployed at SGR010 had no apparent accumulation of any compounds over control values (Figure 6-21 to 6-27).

7.8.6 Water and Sediment Toxicity

Three-species toxicity tests were conducted on water samples from San Gregorio Creek at the USGS Gage (SGR010) and La Honda Creek at the Confluence (SGR080). None of the samples had significant toxicity.

There was a statistically significant effect on amphipod growth in sediment collected from SGR010. While growth integrates a number of physiological processes, the relationship of growth to specific stressors is not well established. These sediment samples had very low concentrations of contaminants for which guideline values have been established, so it is not clear whether the depressed amphipod growth was related to chemical contamination.

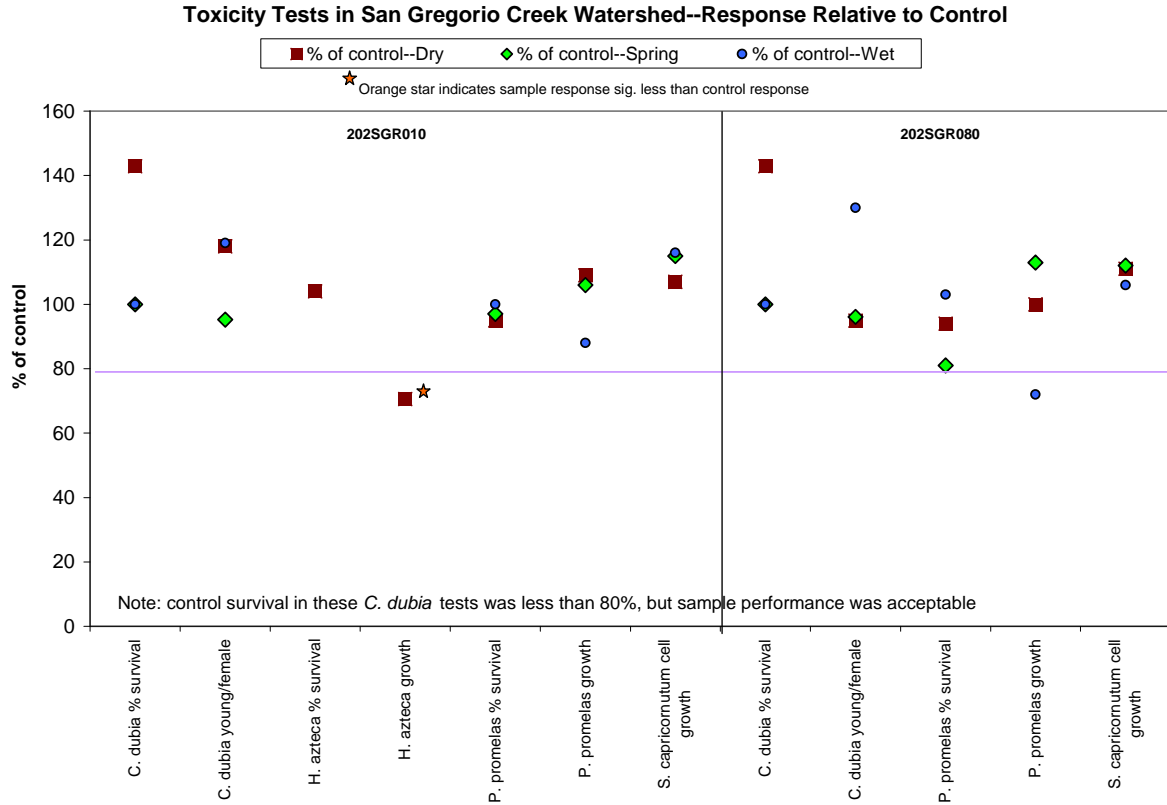


Figure 7-73. Toxicity tests in San Gregorio Creek

Results of toxicity tests are shown, with three species in water samples and one species in sediment samples from San Gregorio Creek. See text for description of species and endpoints. Stars indicate statistically significant differences from test controls. Samples with stars that were also below 80 percent of the control were considered toxic for this assessment.

7.8.7 Coliform Bacteria

Bacterial counts were made in water samples from two sites in the watershed (see Figures 6-28 to 6-30). Basin Plan objectives for all three indicators, *E. coli*, fecal coliforms, and total coliforms were slightly exceeded at San Gregorio Creek, San Gregorio confluence (SGR079). Basin Plan objectives for *E. coli* were also slightly exceeded, and total coliforms were exceeded at Playbowl on La Honda Creek (SGR100). While there is public access to these two sites, they are no longer used as community swimming holes.

In particular, there has been concern about leaking septic systems upstream of the confluence of La Honda Creek and Alpine Creek. The moderate to low counts for Jones Gulch should alleviate concern about potential contamination upstream of the site at the time of sampling.

7.9 Stevens Creek/Permanente Creek Watershed

7.9.1 Sites of Concern

Exceedances noted in the summary table (Table 7-28 and Table 7-29) are discussed in the relevant sub-section for this watershed. A weight-of-evidence approach was used to evaluate sites of concern. Individual exceedances of parameters, particularly those without regulatory objectives, are considered in context. In general, nickel and chromium levels are geologically elevated in the Bay Area, particularly where there are serpentine deposits. Nutrient levels (nitrate and phosphorus) are above U.S. EPA guidelines at most sites, but may still be relatively low.

Temperature and dissolved oxygen conditions throughout this watershed would make it difficult for these streams to support salmonid populations without further improvements. Nutrient and contaminant data indicate considerable inputs of metals, pesticides, and PAHs in the lower watershed. Toxicity results indicate the presence of constituents at toxic levels both at the upstream and downstream ends of the most urbanized areas. Contaminant concentrations near the creek mouths indicate that the urban areas are of concern for stream degradation and for transport of metals, PAHs, and legacy pesticides to the Bay. STE190 at the Cooley Picnic Area, with its elevated coliform counts, is a site of concern particularly because of its continued recreational use.

Table 7-28. Summary of sites with exceedances in Stevens Creek watershed

| <i>Analyte</i> | <i>Standard</i> | <i># of sites</i> |
|--|-------------------|-------------------|
| FIELD MEASURES of 3 sites sampled | | |
| Temperature | maximum, salmonid | 1 |
| | MWAT, steelhead | 3 |
| Oxygen, dissolved | minimum, COLD | 2 |
| | minimum, WARM | 1 |
| | 3-month median | 2 |
| pH | range | 1 |
| CONVENTIONAL WATER QUALITY of 3 sites sampled | | |
| Nitrate as N | maximum | 3 |
| Phosphorus, total as P | maximum | 3 |
| WATER TOXICITY of 2 sites sampled | | |
| <i>Ceriodaphnia</i> | survival | 1 |
| <i>Pimephales</i> | survival | 1 |
| <i>Selenastrum</i> | growth | 1 |
| SEDIMENT CHEMISTRY of 1 site sampled | | |
| <i>Metals</i> | | |
| Chromium | TEC | 1 |
| Nickel | TEC | 1 |
| <i>Organics</i> | | |
| DDE (sum op + pp) | TEC | 1 |
| PCB (total) | TEC | 1 |
| SEDIMENT TOXICITY of 1 sites sampled | | |
| <i>Hyalella</i> | growth | 1 |
| TISSUE CHEMISTRY of 1 sites sampled | | |
| Copper | EDL 85 | 1 |

| COLIFORMS of 4 sites sampled | | |
|-------------------------------------|-----------------|---|
| <i>E. coli</i> | steady state | 2 |
| Fecal coliform | log mean | 1 |
| | 90th percentile | 3 |
| Total coliform | median | 4 |
| | maximum | 4 |

Table 7-29. Summary of sites with exceedances in Permanente Creek watershed

| Analyte | Standard | # of sites |
|--|-------------------|-------------------|
| FIELD MEASURES of 2 sites sampled | | |
| Temperature | maximum, salmonid | 1 |
| | MWAT, steelhead | 1 |
| | MWAT, coho | 1 |
| Oxygen, dissolved | minimum, COLD | 1 |
| | minimum, WARM | 1 |
| | 3-month median | 1 |
| pH | range | 1 |
| CONVENTIONAL WATER QUALITY of 2 sites sampled | | |
| Nitrate as N | maximum | 2 |
| Phosphorus, total as P | maximum | 2 |
| WATER METALS of 2 sites sampled | | |
| Selenium, total | chronic | 1 |
| WATER TOXICITY of 2 sites sampled | | |
| <i>Ceriodaphnia</i> | survival | 1 |
| <i>Selenastrum</i> | growth | 2 |
| SEDIMENT CHEMISTRY of 1 site sampled | | |
| <u>Metals</u> | | |
| Zinc | TEC | 1 |
| <u>Organics</u> | | |
| Benz(a)anthracene | TEC | 1 |
| Benzo(a)pyrene | TEC | 1 |
| Chlordane | PEC | 1 |
| Chrysene | TEC | 1 |
| DDD (sum op + pp) | TEC | 1 |
| DDE (sum op + pp) | TEC | 1 |
| DDT (sum op + pp) | TEC | 1 |
| DDT (total) | TEC | 1 |
| Dibenz(a,h)anthracene | TEC | 1 |
| Dieldrin | TEC | 1 |
| Fluoranthene | TEC | 1 |
| PAH (total) | TEC | 1 |
| PCB (total) | TEC | 1 |
| Phenanthrene | TEC | 1 |
| Pyrene | TEC | 1 |
| SEDIMENT TOXICITY of 1 site sampled | | |
| <i>Hyalella</i> | growth | 1 |

| TISSUE CHEMISTRY of 1 site sampled | | |
|---|--------------|---|
| Copper | EDL 85 | 1 |
| COLIFORMS of 1 site sampled | | |
| <i>E. coli</i> | steady state | 1 |
| Total coliform | median | 1 |
| | maximum | 1 |

7.9.2 Water Quality in Relation to Land Use

Comparisons of contaminant profiles at the upstream and downstream edges of the urban areas indicate that urban land use is contributing lead, pesticides, and particularly PAHs into these two creeks. Chemical and toxicity data indicate that contaminants are present at elevated or toxic concentrations even at the upstream edge of the urbanized areas. Figure 3-9 and Figure 7- depict the urbanized areas and locates quarries in the watershed.

7.9.3 Macroinvertebrate Assemblages and Physical Habitat

Three distinct groups of sites are apparent on the cluster dendrogram (Figure 7-) and ordination plot (Figure 7-).

- The Lower Stevens Creek and Permanente Creek (LSP) group consists of seven sites: three on lower Permanente Creek (PER010, PER020, PER040), three on lower Stevens Creek (STE020, STE030, STE040), and one on Hale Creek (PER030), a tributary to Permanente Creek.
- The Middle Stevens Creek and Permanente Creek group (MSP) consists of sites on Stevens Creek (STE060, STE070) and Permanente Creek (PER050, PER070, PER080) at the base of the foothills on the southern edge of the urban boundary.
- The Upper Stevens Creek (US) group included three sites in the Stevens Creek canyon (STE100, STE110, STE120) draining protected parkland of the Santa Cruz Mountains.

Lower Stevens Creek and Permanente Creek group

All of the sites in this group drain the heavily urbanized Santa Clara Valley and were among the least diverse in the Bay Area, especially STE030 (7 taxa), PER020 (6 taxa), and PER040 (7 taxa). Taxa richness values, which ranged from 6 to 14, were all well below values from minimally disturbed sites (28-59). Other biological metrics were similarly very different from minimally disturbed conditions (Table 6–11). Benthic assemblages at sites in this group are severely altered, and show little resemblance to natural systems. Significant portions of these streams have been channelized, either with natural substrate banks or concrete. The structural and functional diversity of these streams has been significantly reduced, resulting in very low biological integrity.

Benthic assemblages at sites in the Lower Stevens Creek and Permanente Creek (LSP) group were completely dominated by two common tolerant taxa, Chironomidae (69 percent) and

Oligochaeta (23 percent). No other taxa were common to every site. The percentage of organisms belonging to common tolerant taxa ranged from 97 to 100 percent (Table 7-31).

Simulium sp. (Diptera: Simuliidae) was very abundant (21 percent) at Hale Creek (PER030), but was less abundant at sites in Stevens Creek and absent from sites in Permanente Creek, except for one individual from PER040. *Baetis* sp., the most pollution-tolerant EPT taxon, was found in low abundances at most sites and was absent from PER010 and PER020. No beetles were present at any of the sites, including the moderately intolerant family of riffle beetles (Elmidae). Stoneflies were completely absent, except for one individual *Calineuria californica* (Perlidae) from Hale Creek (PER030). Only two caddisflies were found, a single *Rhyacophila* sp. (Rhyacophilidae) at Hale Cr. (PER030) and one *Hydroptila* sp. (Hydroptilidae sp.) at STE020.

Middle Stevens Creek and Permanente Creek group

Other than PER080, which resembles minimally disturbed conditions, sites in the Middle Stevens Creek and Permanente Creek group (MSP) represent an intermediate level of biological integrity between sites in the Lower Stevens Creek and Permanente Creek (LSP) group and the Upper Stevens Creek (US) group.

These sites were numerically dominated by four common tolerant taxa:

| | |
|---------------------|-----|
| Chironomidae | 48% |
| <i>Baetis</i> sp. | 14% |
| Oligochaeta | 8% |
| <i>Simulium</i> sp. | 6% |

Benthic assemblages at all sites were dominated by common tolerant taxa (>87 percent), except at PER080, where only 32 percent of organisms were common tolerant taxa (see Table 7-31). Few common intolerant taxa were found at STE060 (none), STE070 (only 1), or PER050 (only 1), but PER070 and PER080 contained four and eight common intolerant taxa, respectively.

Sites on middle Stevens Creek (STE060, STE070) were vastly different from sites in the Upper Stevens Creek (US) group (STE100, STE110, STE120). Downstream differences suggest that Stevens Creek Reservoir has a dramatic effect on benthic assemblages. Site STE100 is located above the reservoir in Stevens Creek County Park, while STE070 is located directly downstream of the reservoir, but upstream of the urban boundary. At the upstream site, 44 taxa were found, of which 14 were common intolerant taxa; below the dam, taxa richness was only 17, and just one member of a common intolerant taxa was found. Other biological metrics follow this trend, with much lower values at downstream sites than upstream (Table 7-31). Within this group, at STE060 in the urban area, the benthic assemblage was very similar to STE070. High turbidity has been observed downstream of the reservoir throughout the year,

The West Fork of Permanente Creek (PER080), which primarily drains county open space, is the least disturbed site in the Permanente Creek watershed. Intolerant taxa found at PER080 include the mayfly *Paraleptophlebia* sp., the caddisflies *Lepidostoma* sp. and *Rhyacophila* sp., and the stonefly *Malenka* sp. Other intolerant taxa expected at intermittent minimally disturbed streams, such as *Drunella* sp. (Ephemerellidae) and *Sweltsa* sp. (Chloroperlidae), were absent. Metric

values at PER080, however, were within the range of values from intermittent minimally disturbed streams.

The East Fork of Permanente Creek (PER070) also primarily drains open space, except for a large quarry immediately adjacent to the stream. Taxa richness is higher at PER070 than at sites in urban areas, and several rare or intolerant taxa were collected here, including Aeshnidae (Odonata), *Lepidostoma* sp., and *Rhyacophila* sp. The benthic assemblage was primarily composed of common tolerant taxa (94 percent), however, and metric values were outside of the range of values from minimally disturbed streams.

Upper Stevens Creek group

Benthic assemblages at sites in the Upper Stevens Creek (US) group were taxonomically very diverse and contained many intolerant taxa. The taxonomic richness value of 58 at STE110 is the highest value from any site of the nine watersheds discussed in this report.

Less than half of the organisms at any of the three sites are common tolerant taxa. Abundant taxa included:

| | |
|--|-----|
| Chironomidae | 20% |
| <i>Ephemerella</i> sp. (Ephemeroptera: Ephemerellidae) | 12% |
| <i>Baetis</i> sp. (Ephemeroptera: Batidae) | 10% |
| <i>Calineuria californica</i> (Plecoptera: Perlidae) | 4% |
| <i>Malenka</i> sp. (Plecoptera: Nemouridae) | 3% |

Rare taxa found at one or more sites include sensitive taxa such as *Agathon* sp. and *Blepharicera* sp. (Diptera: Blephariceridae), *Ormosia* sp. (Diptera: Tipulidae), *Haploperla* sp. (Plecoptera: Chloroperlidae), *Pteronarcella* sp. and *Pteronarcys* sp. (Plecoptera: Pteronarcyidae), *Ecclisomyia* sp. (Trichoptera: Limnephilidae).

Although metric values from all three sites are within the range of conditions at minimally disturbed sites, there is some evidence that biological integrity at STE100 is slightly reduced relative to STE110 and STE120. Many common tolerant organisms, such as Chironomidae, *Simulium* sp., and Oligochaeta, were more abundant at STE100. Additionally, Plecoptera and Coleoptera taxa richness was reduced at STE100 relative to upstream sites. While STE110 and STE120 are located in protected parkland, STE100 is located downstream of some residences and adjacent to a paved road. STE100 received a lower sediment deposition P-HAB score (8 out of 20) and has a sandier bed than upstream sites, suggesting that increased levels of fine sediment at STE100 may be responsible for slight differences in benthic assemblages.

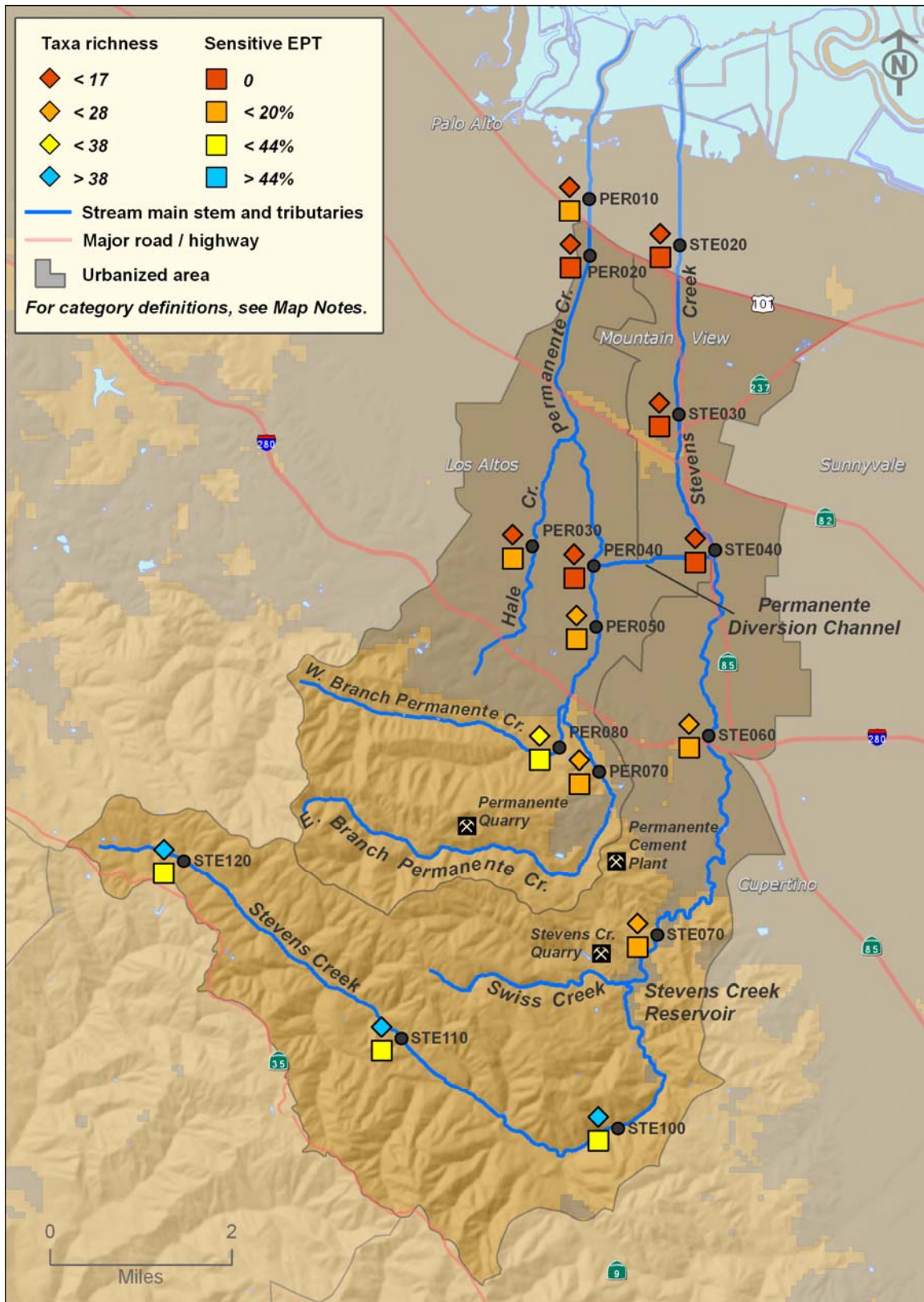


Figure 7-74. Map of Stevens/Permanente creeks benthic macroinvertebrate index scores

Table 7-30. Bioassessment sites in the Stevens Creek/Permanente Creek watershed

| Site ID | Site Name | Stream |
|---------|---------------------------------|----------------------------|
| PER010 | Lower Permanente | Permanente Creek |
| PER020 | Crittenden Middle School | Permanente Creek |
| PER030 | Hale Creek at Covington | Hale Creek |
| PER040 | Permanente at Diversion Channel | Permanente Creek |
| PER050 | Loyola Corners | Permanente Creek |
| PER070 | Rancho San Antonio | East Fork Permanente Creek |
| PER080 | West Fork Permanente | West Fork Permanente Creek |
| STE020 | La Avenida | Stevens Creek |
| STE030 | Landels School | Stevens Creek |
| STE040 | Below Diversion Channel | Stevens Creek |
| STE060 | Belleville/ Barranca | Stevens Creek |
| STE070 | Chestnut Picnic Area/ USGS Gage | Stevens Creek |
| STE100 | Moss Rock | Stevens Creek |
| STE110 | Upper Park 1 | Stevens Creek |
| STE120 | Upper Park 2 | Stevens Creek |

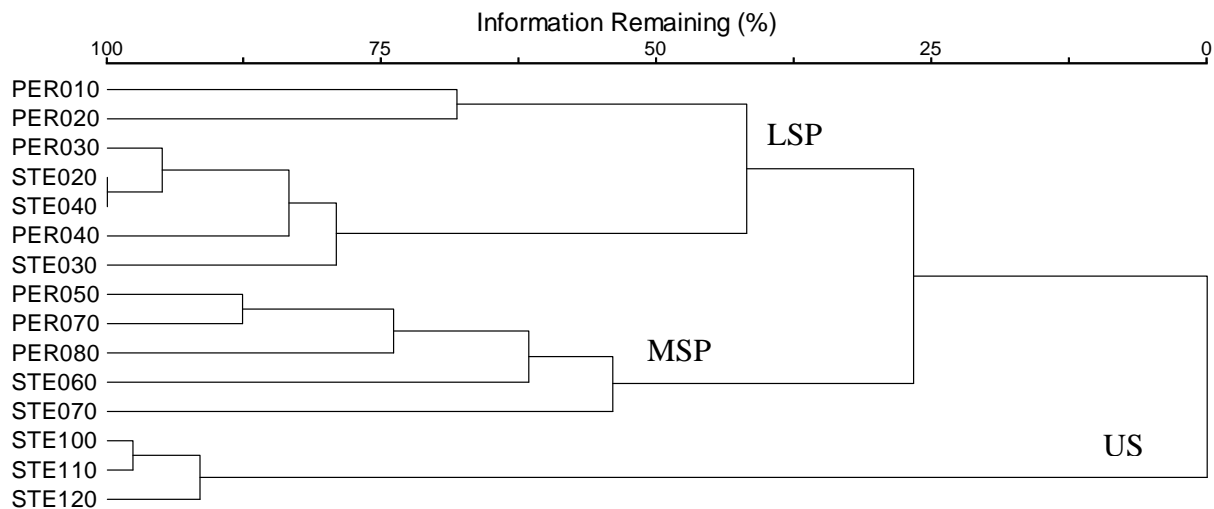


Figure 7-75. Cluster dendrogram for the Stevens Creek/Permanente Creek watershed

The cluster dendrogram is based on similarity of macroinvertebrate taxa present in the Stevens Creek/Permanente Creek watershed. The Information Remaining axis at the top signifies how similar the sites are with respect to taxa presence. Abbreviations: LSP, Lower Stevens/Permanente; MSP, Middle Stevens/Permanente; US, Upper Stevens.

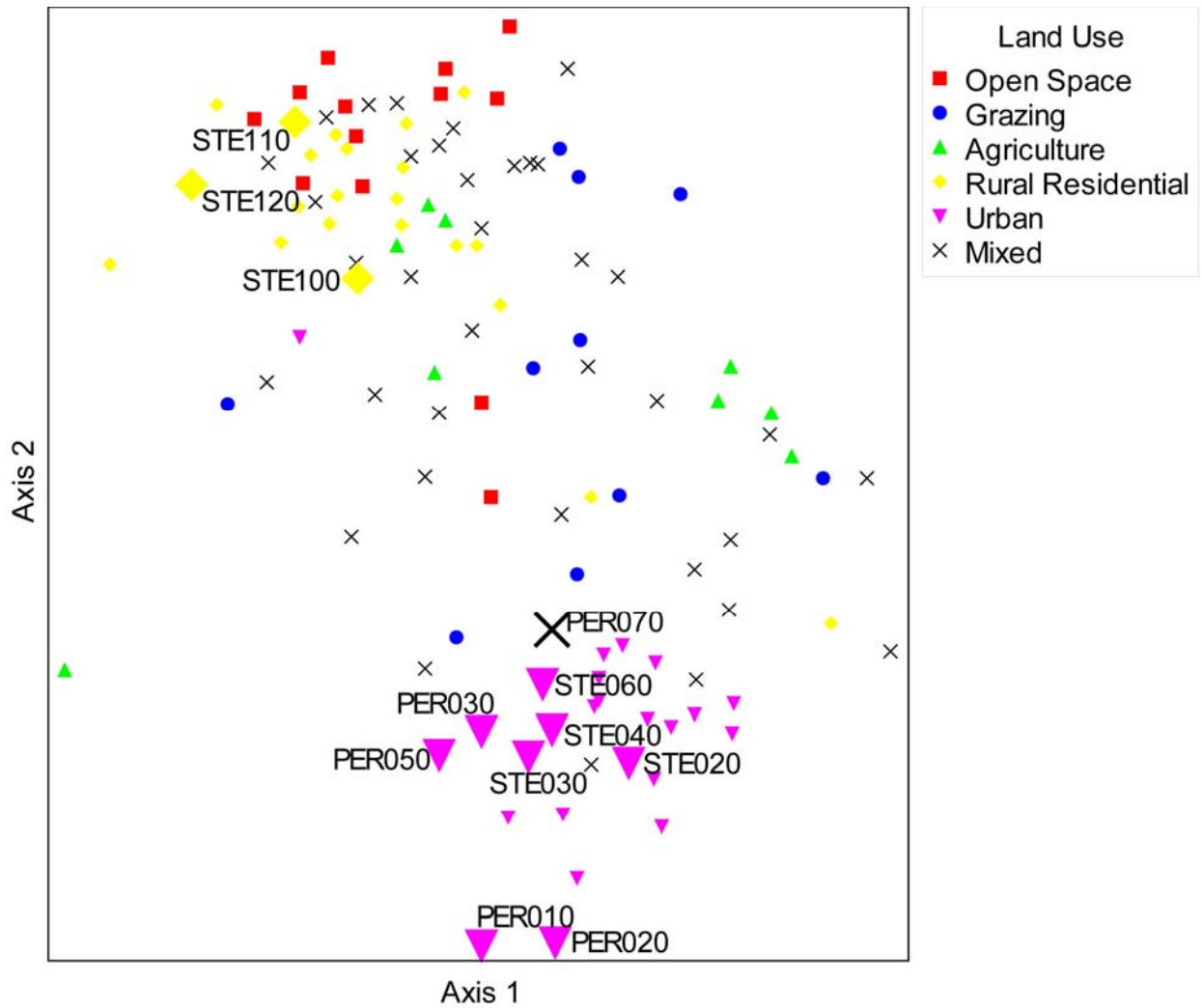


Figure 7-76. NMS ordination of sites in the Stevens Creek/Permanente Creek watershed

The NMS ordination of sites by macroinvertebrate taxa abundance. Sites in the Stevens Creek/Permanente Creek watershed are labeled and symbols are enlarged. For an explanation of ordination graphs, refer to Section 6.1.

Table 7-31. Biological metrics from the Stevens Creek/Permanente Creek watershed

| | STE 020 | STE 030 | STE 040 | STE 060 | STE 070 | STE 100 | STE 110 | STE 120 | PER 010 | PER 020 | PER 030 | PER 040 | PER 050 | PER 070 | PER 080 |
|--------------------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|
| Coleoptera Taxa | 0 | 0 | 0 | 1 | 0 | 5 | 9 | 9 | 0 | 0 | 0 | 0 | 2 | 1 | 1 |
| Diptera Taxa | 3 | 4 | 4 | 5 | 5 | 14 | 15 | 11 | 4 | 2 | 5 | 2 | 6 | 6 | 6 |
| Ephemeroptera Taxa | 1 | 1 | 1 | 1 | 1 | 10 | 11 | 10 | 2 | 0 | 1 | 1 | 1 | 3 | 4 |
| Plecoptera Taxa | 0 | 0 | 0 | 0 | 1 | 4 | 8 | 6 | 0 | 0 | 1 | 0 | 0 | 0 | 2 |
| Trichoptera Taxa | 1 | 0 | 0 | 2 | 2 | 6 | 8 | 5 | 0 | 0 | 1 | 0 | 1 | 6 | 6 |
| Non-Insect Taxa | 7 | 2 | 5 | 9 | 8 | 5 | 7 | 4 | 4 | 4 | 6 | 4 | 7 | 5 | 10 |
| EPT Taxa | 2 | 1 | 1 | 3 | 4 | 20 | 27 | 21 | 2 | 0 | 3 | 1 | 2 | 9 | 12 |
| Taxa Richness | 12 | 7 | 10 | 18 | 17 | 44 | 58 | 46 | 11 | 6 | 14 | 7 | 17 | 21 | 30 |
| % EPT | 0 | 3 | 3 | 1 | 33 | 53 | 54 | 65 | 0 | 0 | 1 | 1 | 16 | 21 | 70 |
| % Sensitive EPT | 0 | 0 | 0 | 0 | 4 | 44 | 43 | 32 | 0 | 0 | 0 | 0 | 0 | 1 | 21 |
| % Chironomidae | 80 | 71 | 61 | 67 | 53 | 30 | 14 | 13 | 16 | 89 | 72 | 91 | 50 | 61 | 10 |
| % Coleoptera | 0 | 0 | 0 | 2 | 0 | 2 | 17 | 10 | 0 | 0 | 0 | 0 | 1 | 1 | 0 |
| % Oligochaeta | 11 | 25 | 32 | 14 | 2 | 5 | 1 | 1 | 75 | 11 | 3 | 7 | 21 | 5 | 3 |
| % Non-insect | 13 | 25 | 34 | 18 | 10 | 7 | 4 | 2 | 83 | 11 | 5 | 8 | 25 | 11 | 14 |
| % Intolerant | 0 | 0 | 0 | 0 | 4 | 40 | 42 | 32 | 0 | 0 | 0 | 0 | 0 | 1 | 18 |
| % Tolerant | 2 | 0 | 2 | 3 | 4 | 1 | 1 | 0 | 8 | 0 | 1 | 1 | 3 | 6 | 4 |
| Tolerance Value | 5.9 | 5.7 | 5.7 | 5.9 | 5.5 | 3.7 | 3.1 | 3.5 | 5.4 | 5.9 | 6.0 | 5.9 | 5.7 | 5.8 | 4.1 |
| % Predator | 1 | 0 | 0 | 3 | 6 | 9 | 23 | 17 | 0 | 0 | 1 | 0 | 3 | 3 | 11 |
| % Collector-filterer | 6 | 1 | 2 | 11 | 3 | 3 | 4 | 8 | 0 | 0 | 21 | 0 | 7 | 5 | 4 |
| %Collector-gatherer | 93 | 99 | 97 | 82 | 84 | 73 | 38 | 56 | 99 | 100 | 77 | 99 | 87 | 86 | 70 |
| % Scraper | 0 | 0 | 1 | 3 | 0 | 7 | 26 | 14 | 1 | 0 | 0 | 1 | 2 | 4 | 1 |
| % Shredder | 0 | 0 | 0 | 0 | 4 | 8 | 6 | 4 | 0 | 0 | 0 | 0 | 0 | 1 | 13 |
| % Other | 0 | 0 | 0 | 0 | 2 | 1 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| Common Tolerant (%) | 100 | 100 | 99 | 95 | 87 | 44 | 21 | 37 | 99 | 100 | 97 | 99 | 97 | 94 | 32 |
| Common Intolerant (Taxa) | 0 | 0 | 0 | 0 | 1 | 14 | 22 | 16 | 1 | 0 | 2 | 0 | 1 | 4 | 8 |

7.9.4 Basic Water Quality Field Measurements

Continuous field measurements were made at six sites on Stevens and Permanente Creeks: Permanente Creek at Charleston Rd. (PER010), and at Loyola Corners (PER050); and Stevens Creek at La Avenida (STE020), Belleville/ Barranca (STE060), Chestnut Picnic Area/ USGS Gage, STE070, and at Moss Rock (STE100).

The two downstream sites, PER010 and STE020, are within the baylands wetland habitat. Continuous monitoring during winter 2002 overlapped with a high tide, confirming that PER010 is tidally influenced at least part of the year and thus not subject to the freshwater objectives.

Although coho salmon are not historically resident in these streams, the MWAT threshold for coho (14.8°C) was exceeded at STE020, STE060, STE070, and PER010 in the dry season, and at three of these sites (not STE060) in the wet season. Steelhead are currently resident in both Stevens and Permanente Creeks (Leidy *et al.* 2005). The higher MWAT threshold for steelhead (17°C) was exceeded at all four of these sites in the dry season, and at STE020 in the wet season. The maximum temperature for salmonid survival (24°C) was exceeded in the dry season at PER010 and STE020.

Dissolved oxygen concentrations failed to meet Basin Plan objectives (7 mg/L) at STE070 and PER010 in the both the dry season and the wet season. DO levels at STE070 are strongly dependent on the water released from Stevens Creek Reservoir directly upstream. Although PER010 is tidally influenced, it had DO values too low to support healthy aquatic life. STE060 and STE100 had DO values below 9mg/L during dry season measurements (Figure 7-, Figure 7-, and Appendix D).

The pH values at STE060 (8.65) and PER010 (8.70) were above the Basin Plan maximum.

Dissolved salt content, as measured by specific conductance, was above the objective for drinking water taste in both the wet and dry seasons at STE020 and PER010, consistent with their location in the baylands wetland habitat. A high tide sequence at PER010 accounts for the specific conductance above the higher value for the maximum for agricultural use in the wet season, with a very high reading of 14,541 µS/cm.

Dry season physical habitat scores were available for four sites on Stevens and Permanente Creeks: Permanente Creek at Charleston Rd. (PER010), and Stevens Creek at La Avenida (STE020), Belleville/Barranca (STE060), and the Chestnut Picnic Area/ USGS Gage, STE070. Site STE070 had excellent riparian habitat scores, but the scores for the other three sites ranged from fair to poor.

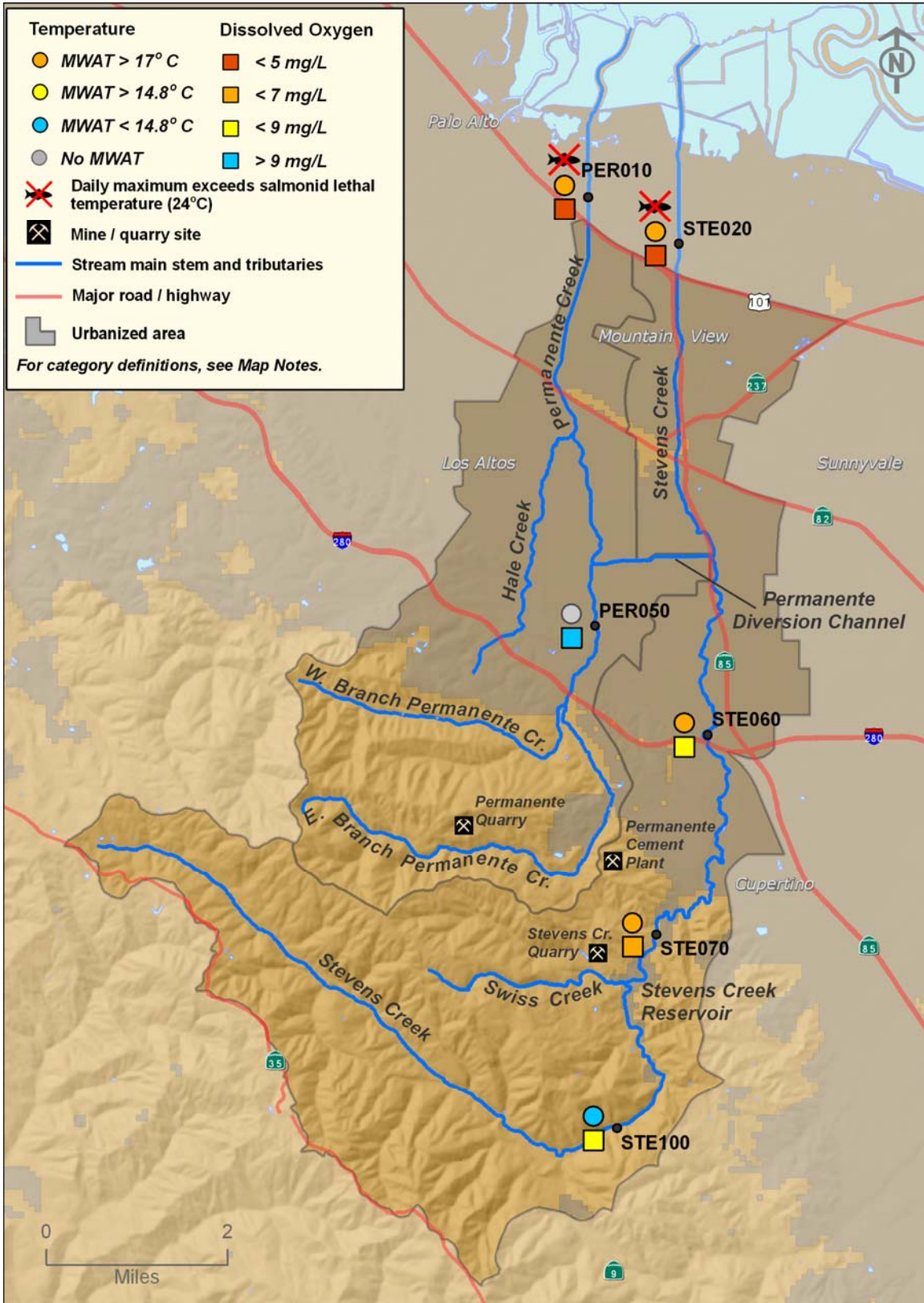


Figure 7-77. Map of temperature and DO levels in Stevens and Permanente creeks

Temperature--Stevens/Permanente Creeks

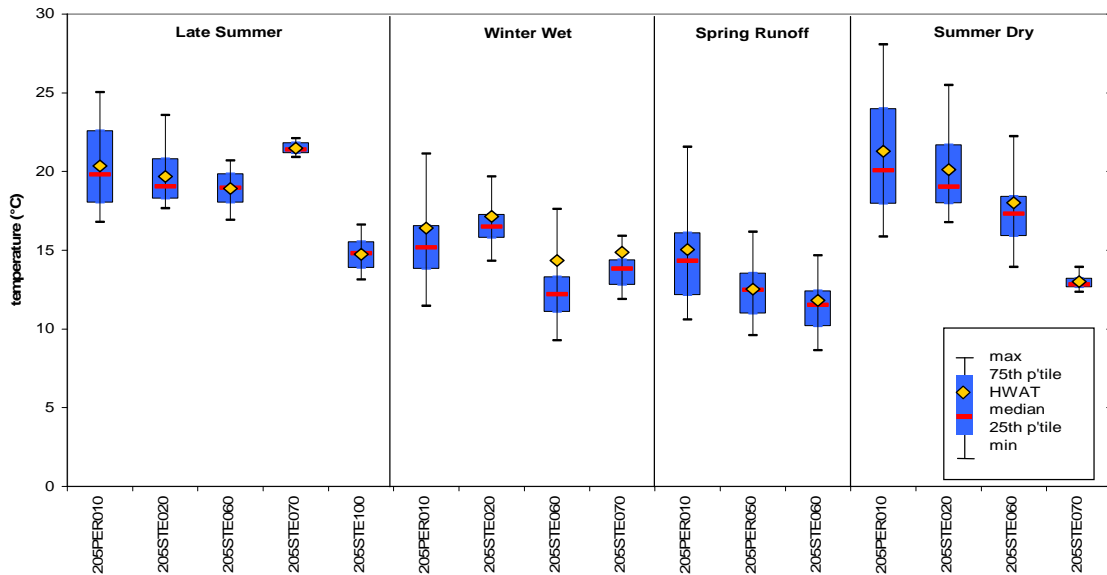


Figure 7-78. Temperature monitoring in Stevens and Permanente creeks

Dissolved Oxygen--Stevens/Permanente Creeks

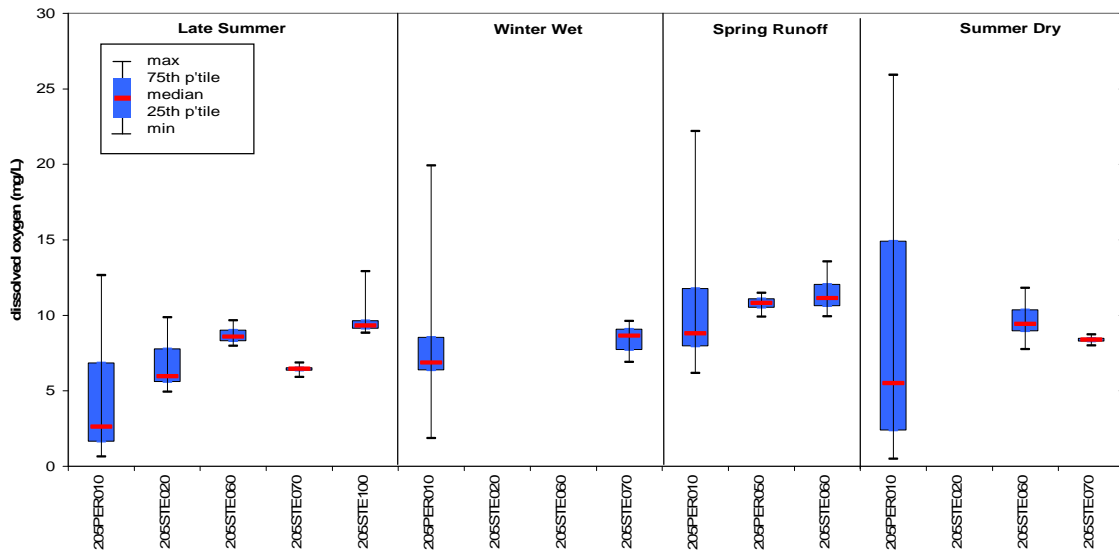


Figure 7-79. DO monitoring in Stevens and Permanente creeks

Data were rejected from STE020 and STE060 for the winter season and from STE020 for the early summer dry season.

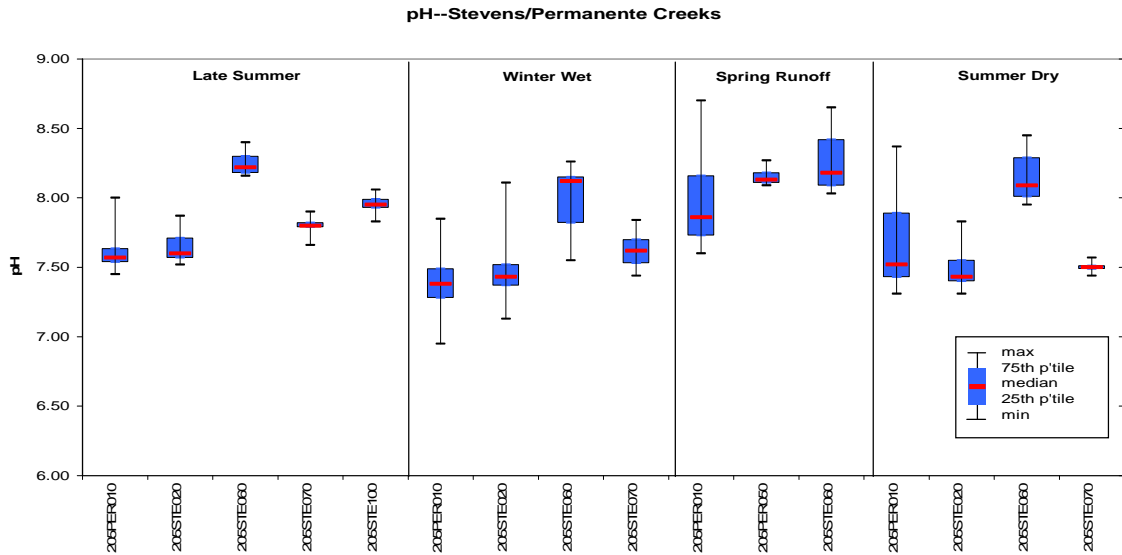


Figure 7-80. pH monitoring in Stevens and Permanente creeks

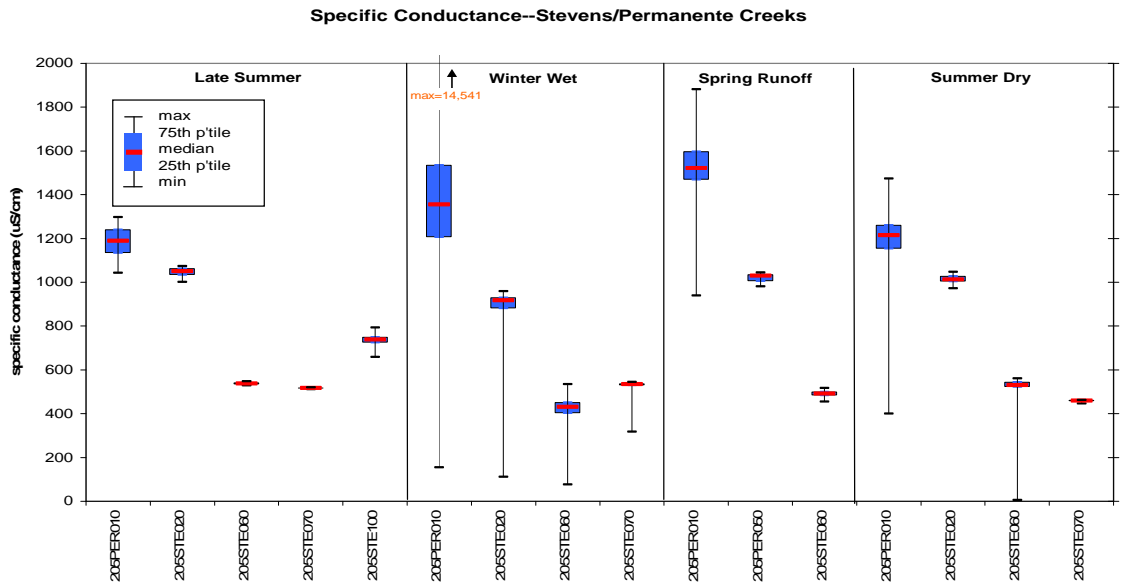


Figure 7-81. Specific conductance monitoring in Stevens and Permanente creeks

7.9.5 Water, Sediment, and Clam Tissue Chemistry

Nutrients were measured in water samples from PER010, PER070, STE020, STE060, and STE100. Nitrate exceeded 1 mg/L at PER010 in the wet season, PER070 in the spring and wet seasons, and STE020 in all three seasons. All but two nitrate samples exceeded the U.S. EPA reference guideline of 0.155 mg-N/L.

All total phosphorus and most orthophosphate values exceeded the U.S. EPA total phosphorus reference guideline of 0.030 mg/L, usually by a significant amount, with total phosphorus concentrations approaching, and in one case exceeding 0.5 mg/L.

Chlorophyll-a values were somewhat elevated relative to most of the other watersheds sampled, exceeding the 1.78 µg/L guideline value in the wet and spring seasons at PER010, and in all three seasons at STE020 and STE060. The high value was 22 µg/L at PER010 (Figures 6-15 to 6-17; Appendix G).

Trace metal and organic contaminants were measured in water samples at four sites. At the upstream Permanente Creek site (PER070), selenium was above the Basin Plan chronic water quality objective for aquatic life protection (WQO) during all three seasons (dry, wet, spring). There were no exceedances of WQOs for other chemicals in any water samples from this watershed.

Sediments were collected for analysis from the most downstream sites on Stevens and Permanente Creeks. The STE020 sediment had chromium, nickel, sum PCBs, and sum DDE above threshold effects concentrations (TECs). These concentrations were generally well below probable effects concentrations (PECs), and the PEC-based sediment quality guideline quotient was 0.12. The Permanente Creek sediment site (PER010) did not exceed the thresholds for chromium or nickel, but exceeded TEC for zinc and a list of organic compounds, including: sum PCBs, sum DDD, sum DDE, sum DDT, dieldrin, benz(a)anthracene, benzo(a)pyrene, chrysene, dibenz(a,h)anthracene, fluoranthene, phenanthrene, pyrene, and sum PAHs (Appendix H). This Permanente site had a sediment quality guideline quotient of 0.14 without DDE, and 0.35 with DDE (MacDonald *et al.*, 2000).

Tissues from clams deployed at STE020 and PER 010 had similar profiles, with PCB concentrations about twice the control value and elevated chlordanes, but average concentrations of the other analytes (Figures 6-21 to 6-27)

7.9.6 Water and Sediment Toxicity

Water samples from the Stevens and Permanente Creek sites (PER010, PER070, STE020, and STE060) had by far the most incidences of observed toxicity of any watershed in this study. The magnitude of toxicity was greatest for the alga *Selenastrum*

capricornatum, primarily in the downstream sites (PER010 and STE020), but also at PER070. Significant toxicity to fish at one upstream site (STE060) and invertebrates at two upstream sites (STE060 and PER070) was not reproduced in the downstream samples (Figure 7-). There were minor reductions in amphipod growth in sediments from the two downstream sites. The lack of mortality indicates that the long list of chemicals exceeding threshold effects concentrations (especially at PER010) did not produce bioavailable mixtures.

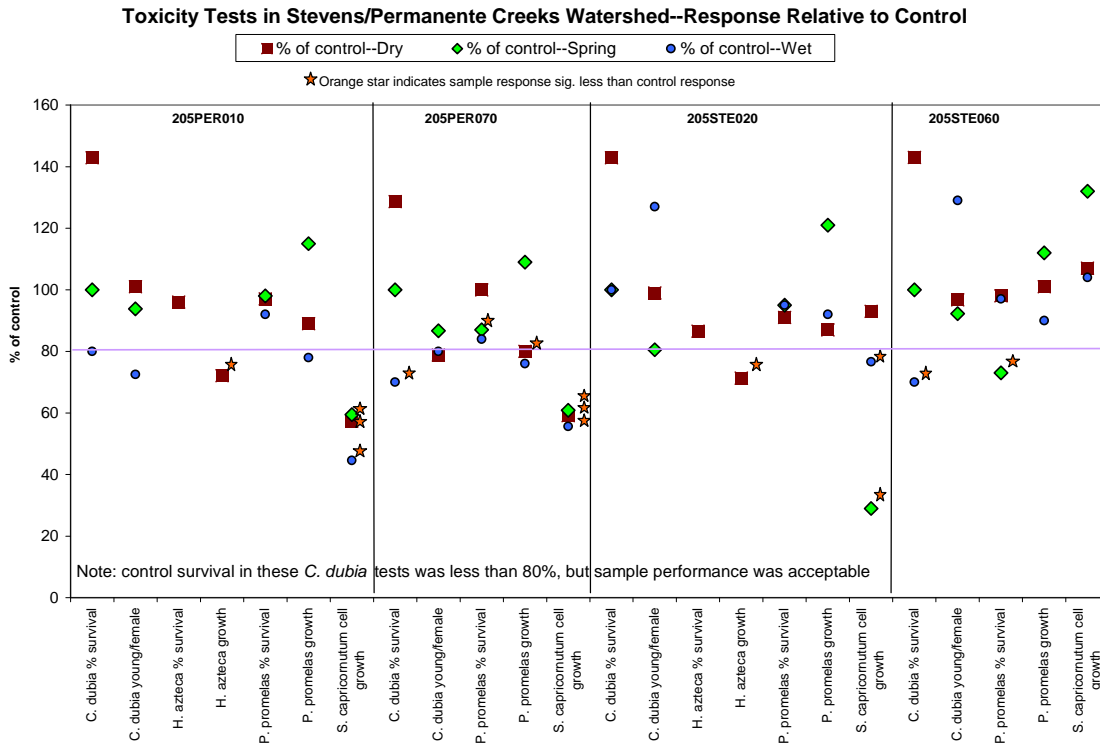


Figure 7-82. Toxicity tests in Stevens and Permanente creeks

Results from toxicity tests are shown, with three species in water samples and one species in sediment samples from Stevens and Permanente creeks. See text for description of species and endpoints. Stars indicate statistically significant differences from test controls. Samples with stars that were also below 80 percent of the control were considered toxic for this assessment.

7.9.7 Coliform Bacteria

Coliform bacteria were measured at five sites: Stevens Creek at La Avenida (STE020), at Chestnut Picnic Area/USGS Gage (STE070), above the reservoir (STE080), and at Camp Cooley (STE090); and Permanente Creek at Lower Meadow/West Branch (PER080). Basin Plan objectives for all three indicators, *E. coli*, fecal coliforms, and total coliforms were exceeded at two sites: STE020, and STE090. Camp Cooley (STE090) is an active swimming area despite posted swimming restrictions. Basin Plan objectives for *E. coli* and total coliforms were exceeded at PER080; and Basin Plan objectives for total coliforms only were exceeded at STE070 and STE080 (Figures 6-28 to 6-30).

Table of Contents

| | | |
|-----|---|-----|
| 8 | Conclusions and Recommendations | 8-1 |
| 8.1 | Conclusions..... | 8-1 |
| 8.2 | Reference Conditions..... | 8-2 |
| 8.3 | Evaluation of Monitoring Tools..... | 8-3 |
| 8.4 | Management Recommendations..... | 8-6 |
| 8.5 | Recommendations for Future Monitoring and Assessment..... | 8-7 |

8 Conclusions and Recommendations

Conclusions and recommendations in this section relate to the objectives of this study: new data are described and put in context; seasonal trends are evaluated; the relationships between water quality indicators, stressors, and land use are evaluated; reference conditions are discussed; monitoring tools are evaluated; and recommendations are made in regard to water quality management and future monitoring and assessment.

8.1 Conclusions

Land use

Land use is one of the most important factors affecting the composition and structure of benthic assemblages in streams in the San Francisco Bay Area. In general, benthic assemblages in urban areas are in poor condition. These communities exhibit low diversity and are numerically dominated by four pollutant-tolerant taxa of invertebrates: Oligochaeta, Chironomidae, *Simulium*, and *Baetis* sp. Although stressor identification is difficult, these degraded benthic assemblages are most likely due to a combination of factors including poor physical habitat, hydromodification, low dissolved oxygen, and elevated nutrients and chemical contaminants. Streams in open space areas or those with rural residential land uses generally contain benthic assemblages with more taxonomic and functional diversity indicating more healthy benthic communities. Streams with intact riparian habitat and less channel alteration had healthier benthic communities.

Flow Intermittency

One of the most important natural factors affecting benthic communities is flow intermittency. Benthic assemblages in streams that go dry in the summer are less diverse than in streams with perennial flow. The decreased diversity is a result of the strict aquatic habitat requirements of organisms such as aquatic beetles. While some streams are historically intermittent, water diversions can also effectively eliminate dry weather flows affecting aquatic communities.

Physical habitat

In this study, most of the watersheds had an overlapping pattern of habitat disturbances and water quality stressors with potential to affect aquatic life. Temperature and DO values often exceeded thresholds for fish and other aquatic life, even in areas with less intensive human land use activities. As with benthic communities, physical habitat parameters correlated significantly with temperature and DO. Sites where channels were less modified, riparian and stream bank vegetation was more extensive, and canopy cover was greatest had significantly lower stream temperatures and higher dissolved oxygen concentrations.

Nutrients and chemical contaminants

In addition to the effects of physical stressors on water quality, nutrients and chemical contaminants from areas of more intensive human activity, primarily urban areas, have significant impacts on stream ecosystems. Nutrients, herbicides, metals, PCBs, and PAHs tended to be higher as streams flowed through residential and urban areas. Most samples exceeded U.S. EPA criteria for nitrate, total phosphorus, and orthophosphate. Nitrate concentrations over all

seasons were highest and most variable in streams draining urban areas, followed by mixed land use, agricultural, rural residential, and open space. The mean nitrate level for urban streams was more than twice that for streams draining agricultural areas.

In general, land use and physical habitat alterations seemed to play a greater role in their impact on aquatic communities than chemical contaminants. Although nutrients were above U.S. EPA criteria in many cases, these criteria were based on the 25th percentile values of monitoring data from Southern and Central California reference streams, thus they may not be indicative of concentrations that cause eutrophication. A subsequent study conducted by Tetra Tech to identify more meaningful reference conditions in the same area seemed to indicate that these concentrations may be low (U.S. EPA 2003). U.S. EPA has encouraged states and tribes to establish more informed reference conditions and criteria.

Seasonal differences

In general, the number of exceedances of water quality thresholds for contaminants was low. In this study, the greatest number of exceedances was observed during the dry season. It is likely that the highest chemical concentrations are associated with discrete rain-related runoff events which were not sampled in this ambient monitoring program. Base flows during the winter and early spring are fed predominantly by seepage from soggy soils, while the base flows during summer are composed of more concentrated seepage heavily supplemented by urban runoff. Therefore, it is not unexpected that spring base flows had lower concentrations of contaminants than summer flows.

Toxicity

Stevens Creek had the highest number of toxic samples, but in general the level of toxicity in water samples was moderate. Grab samples taken three times a year during ambient flow conditions may not be reflective of chronic conditions. Sediment samples act as a better integrator of temporal variability. In this study, except for one sample at San Leandro Creek where mortality was 100 percent, sediment toxicity was also moderate. Sediment chemistry showed few exceedances of the sediment guidelines, and there was no clear relationship between sediment chemistry and sediment toxicity.

8.2 Reference Conditions

It is important to identify reference sites and reference conditions in order to develop meaningful tools for management. Understanding the physical and biological characteristics of minimally impacted creeks in the region is essential in interpreting monitoring data. For instance, in the San Francisco Bay Area, chromium and nickel are geologically high. This knowledge is important in differentiating between natural background concentrations of metals and those that may be elevated due to anthropogenic causes.

Minimally disturbed conditions in the Bay Area generally exist in steep, mountainous streams. Biological conditions at minimally disturbed streams can serve as standards that can be applied to most stream types in the Bay Area, except for low-gradient sandy streams. A subset of biological metrics for benthic invertebrates, such as taxa richness, Diptera taxa richness, percent

EPT, and tolerance value may serve as useful non-redundant indicators of human disturbance for intermittent and perennial streams. In order to establish appropriate biological criteria, further work is needed to compile all existing bioassessment data in the Bay Area and analyze it with the techniques used in this analysis. Given the great extent of urban land use in the Bay Area, investigations should aim to identify the least disturbed urban “reference” sites to develop feasible restoration targets for streams in urban areas.

Identification of reference conditions for nutrient monitoring is important to view SWAMP data in a broader context. In this study, only two sites designated as open space were monitored for nutrients. Although these sites exceeded U.S. EPA criteria, they were well below the maximum values for minimally impacted streams in the Tetra Tech study (U.S. EPA 2003). Currently, the State is developing nutrient criteria. Collection of nutrient data from reference sites in this region is important to develop meaningful criteria that apply to the Bay Area.

8.3 Evaluation of Monitoring Tools

Monitoring tools were evaluated based on the following criteria:

- ability to produce interpretable data
- cost compared to the value of data
- availability of water quality thresholds (guidelines or objectives) to evaluate results
- ability to integrate the effect of different stressors
- ability to reflect a relevant indication of beneficial use impact

Stream bioassessments

Of all of the tools used in this study, macroinvertebrate bioassessments provided the most integrative and relevant indicator of ecosystem health. Although an Index of Biological Integrity (IBI) has not been developed for the Bay Area, a comparison of metrics made it possible to distinguish between degraded and non-degraded sites. This indicator was shown to be a valuable tool for evaluating ecosystem effects based on various land uses and alterations in physical habitat. Currently, Bay Area macroinvertebrate data from local agencies and volunteer groups that used the same protocol are being entered into the California Environmental Data Analysis System (CalEDAS) database. This will allow for evaluation of whether Northern or Southern California IBIs, a combination of the two, or a separate IBI is appropriate for the Bay Area.

At this time, tools for macroinvertebrate bioassessments are the most developed type of stream bioassessment. The development of periphyton bioassessment tools for this area is recommended to evaluate the potential effects of nutrient input or other factors that may cause an increase in the biomass or change in the taxonomic makeup of periphyton. In addition, this tool could be used to indicate toxic effects of herbicides, copper, or other chemicals. Periphyton bioassessments would be a better indicator than chlorophyll analysis or the *Selenastrum* toxicity test, which were used in this study, because they can integrate impacts over time and are more ecologically relevant. SWAMP is starting to conduct studies to develop this tool; however, it will take some time before this is a useful assessment tool in California. Although fish bioassessments have been conducted by local agencies and U.S.EPA in this area, their usefulness in evaluating aquatic life impacts is questionable. Low diversity in fish species found in this region makes it difficult to differentiate between impacted and un-impacted fish communities.

However, aside from an indicator of impact, the identification of fish species that inhabit Bay Area streams has value in itself.

Physical habitat assessment

Evaluation of physical habitat is a part of the bioassessment procedure. This assessment procedure was valuable in determining the potential cause of high temperatures and low dissolved oxygen and contributed to interpretation of macroinvertebrate bioassessment metrics. An updated more quantitative protocol is currently being drafted.

Continuous monitoring

Continuous monitoring of basic water quality parameters using data-logging sondes was shown in this study to be a very valuable monitoring tool. Continuous monitoring revealed diurnal patterns that cannot be detected with grab samples. These patterns not only indicated the potential level of impact on salmonid populations but also the mediating factors (photosynthesis or temperature) for low dissolved oxygen concentrations. Sondes, however, are expensive and difficult to deploy for long periods, due to the need for periodic maintenance. Seven days is the minimum deployment period needed to calculate Maximum Weekly Average Temperatures (MWATs). In this study, sensors were not deployed year-round, so the maximum yearly temperature could not be determined; thus, an annual MWAT could not be calculated. When temperature is a cause for concern, small temperature data loggers are the most valuable and economical tools to use. Small digital temperature loggers are inexpensive and can be deployed over the period of a year, annual maximum temperatures and the duration of those temperatures can be measured, and the potential for effects on salmonid populations can be better evaluated.

Nutrients

Measurement of nutrients was valuable in order to determine concentrations in areas with various land uses. Several TMDL nutrient targets for the region will be available in coming years. Currently, the state is developing nutrient criteria. Determining background concentrations for the San Francisco Bay Region is important when considering the development of water quality objectives. In this study, the fact that nitrates were significantly higher in urban areas than in agricultural areas was a surprising and useful finding for management. The concentrations of chlorophyll-a in water samples did not have a relationship to nutrients or other measurements of eutrophication. Since primary productivity in shallow streams is dominated by benthic algae, periphyton bioassessments would be a better indicator of the effect of nutrients in these systems.

Water contaminants

Water column chemistry (metals and organics) was measured to compare concentrations to Basin Plan objectives and other water quality guidelines. In general, concentrations of contaminants were low. It is likely that the highest chemical concentrations are associated with discrete rain-related runoff events; however, these were not sampled in this ambient monitoring program. Although it is useful to evaluate whether water quality objectives for contaminants are exceeded, collecting grab samples during ambient conditions are not representative of worst case or chronic conditions, which have the most impact on aquatic life, and thus may not produce information worth the cost. This is especially true for organics which are expensive to analyze and may not

have associated water quality thresholds. Evaluating storm events for concentrations and loads of contaminants of concern may be a better use of resources.

Toxicity tests

U.S. EPA three-species toxicity tests (*Ceriodaphnia* survival and reproduction, fathead minnow survival and growth, and *Selenastrum* growth) were conducted on water collected at the same time as water collected for chemical analysis. Results showed that diazinon-associated *Ceriodaphnia* toxicity had declined in urban creeks since management measures were taken to control diazinon use. This information was valuable in the development of the pesticide TMDL for the region. The other two toxicity tests were used because different species are sensitive to different chemicals. Using these three tests was also part of the weight-of-evidence approach. Toxicity was only found once for fathead minnow. *Selenastrum* showed the most toxicity of all three species, yet *Selenastrum* data can be difficult to interpret due to variability in the test. This is predominantly due to the difficulty in keeping laboratory cultures and the variability in cell growth rates. Nevertheless, until additional bioassessment tools are developed, these tools should be considered for inclusion in ambient monitoring programs.

Sediment

Measurement of sediment chemistry along with sediment toxicity tests and macroinvertebrate bioassessments could provide a triad weight-of-evidence approach to determining environmental impact and potential stressors. From a temporal perspective, sediment evaluations provide a more integrative measurement than aquatic grab samples. Sediment assessments are becoming more valuable with the replacement of pyrethroids for orthophosphate pesticides. Whereas water column toxicity tests were useful in detecting the effects of orthophosphate pesticides, which are water soluble, pyrethroids are predominantly found adhered to the sediment and are hard to detect in water column tests. Pyrethroids can have severe benthic impacts, which can be predicted by the results of bulk sediment toxicity tests. In this study pyrethroids were not measured, but have been included in subsequent years' studies.

Although regulatory objectives for sediment chemistry have not been developed, published guidelines are useful in sediment evaluations. In collecting sediment for analysis, care must be taken to find depositional areas with fine grain sediment. Although a sediment-triad approach using bioassessment, sediment chemistry, and toxicity is theoretically ideal, depositional areas with fine grain sediment are not the appropriate habitat for current benthic macroinvertebrate assessments.

Clam Tissue

In this study SWAMP deployed clams at sites located at the bottom of watersheds for approximately three months to provide an integrative measurement of chemicals that would be entering the receiving water. Clam tissue concentrations were useful for identifying oxadiazon as a potential stressor. This tool identified oxadiazon as ten times higher than the control at Arroyo Las Positas. Oxadiazon was also found at lower levels in sediment and water. The clam deployment results did not resolve major differences between sites or between sites and controls. In addition, collecting, deploying, retrieving, and analyzing the clams before and after deployment was very expensive. This tool is not recommended for ambient studies due to the high cost and limited benefit.

Trash

SWAMP in the San Francisco Region has developed a quantitative trash assessment procedure and has used that procedure to evaluate trash in various watersheds in the region (see <http://www.waterboards.ca.gov/sanfranciscobay/basinplan.htm#monitoring>). Trash is increasingly being recognized as a water quality problem and regulated as such. This assessment procedure is useful to set priorities for management actions, identify sources, and show trends.

Study design

The study design used in this monitoring program was two-tiered, using bioassessments at all sites as the primary tool for evaluating aquatic life impacts and physical habitat assessments as the primary evaluation of stress. Biological measurements (toxicity tests) and measurements of other stressors including temperature, DO, nutrients and contaminants were performed at a subset of stations. Integrator sites were placed at the bottom of watersheds to evaluate the cumulative effect of the watershed on the receiving waterbody. A deterministic design was used to locate sites near confluences and in areas with different land use impacts. In general, this approach was useful in meeting our objectives within the resources that were available. As stated above, bioassessments and physical habitat assessment were some of the most valuable tools used in assessing impacts on aquatic life and identifying stressors. Although a probabilistic approach may be used, the deterministic approach used in this study was helpful in identifying water quality characteristics associated with specific land uses. Although integrator sites were located at the bottom of watersheds to integrate upstream water quality characteristics, the spatial extent they represented was unclear. Integrator sites have been useful in the Central Coast Region in showing temporal trends of nutrients being discharged to the Pacific.

In general, a weight-of-evidence approach is most useful in evaluating potential water quality impacts. Evaluation of the costs and benefits of the available tools is necessary. It is important to measure indicators such as bioassessment to determine beneficial use impact, as well as stressors such as physical habitat, temperature, and dissolved oxygen to identify the potential cause of impacts. The combination of these tools is necessary for effective water quality management.

8.4 Management Recommendations

This study found that benthic invertebrate assemblages were strongly affected by land use. Urban areas, in particular, had uniformly degraded conditions. Low biological integrity in urban areas is most likely due to a variety of factors, including poor physical habitat (particularly riparian habitat and channel alteration), low dissolved oxygen, and chemical contaminants. There were significant correlations between riparian habitat, temperature, and DO. Temperature and DO values often exceeded thresholds for fish and other aquatic life, even in areas with less intensive human land use activities. Improvement and maintenance of riparian habitat as a primary management focus is likely to lower temperatures, elevate DO, and improve the health of benthic communities. The role of water management also needs to be assessed to determine how alternative strategies might improve conditions for aquatic life.

Nutrients, herbicides, metals, PCBs and PAHs tended to be higher as streams flowed through residential and urban areas. Many of the contaminants found in streams in this study are released

from a multitude of small sources, such as cars, lawn-care activities, and transport of urban soils contaminated with legacy pesticides from decades ago. Management activities here should also include riparian habitat improvement to allow for the absorption and degradation of pollutants through retention in vegetated areas. Designing riparian corridors to limit transport of these materials is likely to be among the easiest and more successful strategies to reduce contaminants, improve temperature and dissolved oxygen conditions, and maintain the essential ecological values of riparian areas. Management measures should also concentrate on preventing pollution from entering creeks by: 1) educating urban citizens on the use of pesticides and fertilizers, pet care and car washing, 2) regulating pollution sources and source materials, and 3) using pervious surface and vegetation in the design of green buildings, parking lots, green swales and decentralized landscape for the treatment of stormwater.

8.5 Recommendations for Future Monitoring and Assessment

This study has provided a survey of nine watersheds at a screening level. Most water quality parameters were measured in only four or five locations in watersheds covering 50 to over 100 square miles. This study has been useful in providing indications of impacts, potential stressors, and associated land uses. Managing these stressors will require monitoring that focuses on specific parameters and their spatial patterns relative to land use, so that water quality can be more effectively managed.

Stream temperature is one of the key stressors highlighted by the results of these surveys, and there appears to be a significant relationship between temperature and certain features of riparian habitats. We recommend that inexpensive temperature data loggers be deployed for extended periods at sites with salmonid populations and at sites with elevated temperatures due to alterations in physical habitat. This will assist in identifying areas where riparian improvements could make streams more supportive of fish and other aquatic organisms. Flow measurements are also needed to evaluate natural flow conditions and to determine the role water management may have on water quality.

Toxicity in San Leandro and Stevens Creek should be further investigated to determine if the toxicity is persistent and to identify causes. Follow-up monitoring should also be conducted at bathing areas that exceeded coliform objectives. Additional monitoring should be conducted at Arroyo Las Positas to determine whether oxadiazon is still elevated and to identify the source.

All macroinvertebrate bioassessment data collected in the region using the California Stream Bioassessment Procedure (CSBP) type sampling is being entered into a common database. Data analysis should be conducted to determine whether the Northern or Southern California IBI may be used for interpretation, or whether a Bay Area specific IBI is needed. In addition, specific emphasis should be put on identifying urban areas that are least impacted by land use and characterizing benthic communities at these sites. This assessment is necessary to develop management actions for urban areas and to identify the highest achievable metrics for tiered aquatic life use assessments.

Long-term monitoring sites should be located at impacted sites to determine trends and at reference sites to determine variability in natural conditions and the role of climate change. Sites

should also be located in areas representative of different land uses to determine the level of impact associated with that land use. It is especially important to locate sites where there is a plan to change land use to determine baseline conditions and to track water quality impacts that may be caused by these changes.

A watershed monitoring coalition should be formed in this region in order to develop the most meaningful information in the most efficient manner. SWAMP may best be used to provide a regional context for local monitoring programs.

9 References

- Allan, J.D. 1995. *Stream Ecology*. Chapman and Hall, London.
- APHA. 1998. *Standard Methods for the Examination of Water and Wastewater*, 20th ed. L.S. Clesceri, A.E. Greenberg, A.D. Eaton (eds). American Public Health Association, American Water Works Association, and Water Environment Federation, Washington, DC.
- Bailey, H.C., J.L. Miller, M.J. Miller, L.C. Wiborg, L. Deanovic, and T. Shed. 1997. “Joint acute toxicity of diazinon and chlorpyrifos to *Ceriodaphnia dubia*.” *Environmental Toxicology and Chemistry* **16**, 2304–2308.
- Barbour, M.T., J. Gerritsen, B.D. Snyder and J.B. Stribling. 1999a. *Rapid bioassessment protocols for use in streams and wadeable rivers: periphyton, benthic macroinvertebrates and fish*, Second edition. EPA 841-B-99-002. U.S. Environmental Protection Agency, Office of Water, Washington D.C.
- Barbour, M.T., J. Gerritsen, B.D. Snyder and J.B. Stribling. 1999b. *Revision to Rapid bioassessment protocols for use in stream and rivers: periphyton, BMIs and fish*. EPA 841-D-97-002. U.S. Environmental Protection Agency, Office of Water, Washington DC.
- Breaux, A., M. Born, L. Suer, S. Cochran, R. Looker. 2005. *Macroinvertebrates in Wildcat and San Leandro Creeks*, Draft Final. San Francisco Bay Regional Water Quality Control Board, Oakland, CA.
- Brodberg, R.K., G.A. Pollock. 1999. *Prevalence of selected target chemical contaminants in sport fish from two California lakes: public health designed screening study*. Final project report. EPA Assistance Agreement No. CX 825856-01-0. Pesticide and Environmental Toxicology Section, Office of Environmental Health Hazard Assessment, California Environmental Protection Agency, Sacramento, CA.
- Bush, L. 1995. *Final report of the Marin Coastal Watershed Enhancement Project*. University of California Cooperative Extension, Novato, CA.
- Carter and Fend. 1997. The distribution and abundance of lotic macroinvertebrates during spring 1997 in seven streams of the Santa Clara Valley area, California. U.S. Geological Survey Open File Report 00-68.
- CDFG. 1999. *California stream bioassessment procedure*. California Aquatic Bioassessment Laboratory, California Department of Fish and Game, Rancho Cordova, CA.
- Choy, M. L. 2004. A comparison of the effects of regulated and non-regulated hydrologic regimes on fine sediment deposition and benthic macroinvertebrate distributions. Senior Thesis, University of California, Berkeley.

- County of Marin. 2000. *Interim health advisory for sport fish from Tomales Bay*. Marin County Department of Health and Human Services, and Marin County Community Development Agency/Environmental Health Services Division, in cooperation with the State Office of Environmental Health Hazard Assessment.
- Curry, R., R. Houghton, T. Kidwell, and P. Tang, 1985. *Pescadero Marsh management: a plan for persistence and productivity*. University of California, Santa Cruz, CA.
- CVRWQCB. 2003. *A Compilation of Water Quality Goals*. Prepared by J.B. Marshack, Central Valley Regional Water Quality Control Board, Sacramento, CA.
- de Vlaming, V., D. Markiewicz, K. Goding, A. Morrill, and J. Rowan. 2004. *Macroinvertebrate assemblages of the San Joaquin River Watershed*. Aquatic Toxicology Laboratory, University of California, Davis and Central Valley Regional Water Quality Control Board, Sacramento, CA.
- Fairey, R., E.R. Long, C.A. Roberts, B.S. Anderson, B.M. Phillips, J.W. Hunt, H.M. Puckett, C.J. Wilson, G. Kapahi, and M. Stephenson. 2001. "A recommended method for calculation of sediment quality guideline quotients." *Environmental Toxicology and Chemistry* 20: 2276-2286.
- Fend, S. V., J. L. Carter, and F. R. Kearns. 2005. Relationships of field habitat measurements, visual habitat indices and land cover to benthic macroinvertebrates in urbanized streams of the Santa Clara Valley, California. In: Brown, L. R., R. Gray, R. M. Hughes, and M. R. Meador, eds., *The effects of urbanization on stream ecosystems (Symposium 47)*. Bethesda, Maryland: American Fisheries Society.
- Gassel, M., S. Klasing, and R. K. Brodberg. 2004. *Health advisory: Guidelines for consumption of fish and shellfish from Tomales Bay (Marin County)*. Pesticide and Environmental Toxicology Section, Office of Environmental Health Hazard Assessment, Oakland, CA.
- Gunther, A.J., J.A. Davis, D. Hardin, J. Gold, D. Bell, J. Crick, G. Scelfo, J. Sericano, M. Stephenson. 1999. "Long term bioaccumulation monitoring with transplanted bivalves in San Francisco Bay." *Marine Pollution Bulletin* 38(3): 170-181.
- Gunther, A. J., J. Hagar, and P. Salop. 2000. *An assessment of the potential for restoring a viable steelhead trout population in the Alameda Creek Watershed*. Alameda Creek Fisheries Restoration Workgroup.
- Harrington, J. and M. Born. 1999. *Measuring the Health of California Streams and Rivers*. Second edition. Sustainable Land Stewardship International Institute: Sacramento, CA.
- Harrington, J.M. 1999. *California stream bioassessment procedures*. California Department of Fish and Game, Water Pollution Control Laboratory. Rancho Cordova, CA.

- Hem, J.D. 1985. *Study and interpretation of the chemical characteristics of natural water*, Third edition. U.S. Geological Survey Water Supply Paper 2254, Alexandria, VA.
- Hornberger, M.I., S.N. Luoma, A. van Geen, C. Fuller, R. Anima. 1998. “Historical trends of metals in the sediments of San Francisco Bay, California.” *Marine Chemistry*, June 25, 1998.
- Kearns, F.R., V.H. Resh, M.T. Barbour, J.L. Carter, and S.V. Fend. 2001. “The relationship between observer-based physical habitat assessments and biological condition in urban streams.” Poster presented at the North American Benthological Society Annual Meeting, La Crosse, WI.
- Ketcham, B.J. 2001. *Point Reyes National Seashore water quality monitoring report: May 1999-May 2001*. National Park Service, Point Reyes National Seashore, Point Reyes Station, CA.
- Lee, J. and B.J. Ketcham. 2001. *Benthic macroinvertebrate analysis of twelve sites in the Point Reyes National Seashore using the California Stream Bioassessment Procedure*. National Park Service, Point Reyes National Seashore, Point Reyes Station, CA.
- Leidy, R.A., G.S. Becker, B.N. Harvey. 2003. *Historical distribution and current status of Steelhead (Oncorhynchus mykiss), Coho salmon (O. kisutch), and Chinook salmon (O. tshawytscha) in streams of the San Francisco Estuary, California*. Center for Ecosystem Management and Restoration, Oakland, CA.
- Leidy, R.A., G.S. Becker, B.N. Harvey. 2005. *Historical distribution and current status of steelhead/rainbow trout (Oncorhynchus mykiss) in streams of the San Francisco Estuary, California*. Center for Ecosystem Management and Restoration, Oakland, CA.
- MacDonald, D.D., G.G. Ingersoll and T.A. Berger. 2000a. “Development and evaluation of consensus-based sediment quality guidelines for freshwater ecosystems.” *Archives of Environmental Contamination and Toxicology* 39(1): 20-31.
- MacDonald, D.D., L.M. Dipinto, J. Field, C.G. Ingersoll, E.R. Long, R.C. Swartz, 2000b. “Development and evaluation of consensus-based sediment effect concentrations for polychlorinated biphenyls.” *Environmental Toxicology and Chemistry* 19:1403-1413.
- McCune, B. and J.B. Grace. 2002. *Analysis of Ecological Communities*. MjM Software Design, Gleneden Beach, OR.
- NCRWQCB. 2001. *Water Quality Control Plan for the North Coast Region*. North Coast Regional Water Quality Control Board, Santa Rosa, CA.
- Newcombe, C. and J. Jensen. 1996. “Channel Suspended Sediment and Fisheries: A Synthesis for Quantitative Assessment of Risk and Impact.” *North American Journal of Fisheries Management*. November 1996.

- Nolte, George S. Consulting Engineers. 1965. *Master Drainage and Sediment Control Plan: Lagunitas and Walker Creek Watersheds*. Marin County Soil Conservation Service, Point Reyes Station, CA.
- Ode, P.R. 2005. Projects Director, Aquatic Bioassessment Laboratory, California Department of Fish and Game, Rancho Cordova, CA. Personal communication, November 1, 2005.
- Ode, P.R., A.C. Rehn, and J.T. May. 2005. "A quantitative tool for assessing the integrity of southern coastal California streams." *Environmental Management* 35:493-504.
- OEHHA. 2004. *Draft Health Advisory: Guidelines for Consumption of Fish and Shellfish from Tomales Bay (Marin County)*. Prepared by M. Gassel, Pesticide and Environmental Toxicology Section, Office of Environmental Health Hazard Assessment, Oakland, CA.
- PAN. 2004. Pesticide Action Network data base available online at http://www.pesticideinfo.org/Detail_ChemUse.jsp?Rec_Id=PC15#TopSites . Accessed 2/28/2007.
- Phillips, B.M, B.S. Anderson, J.W. Hunt, B. Thompson, S. Lowe, R. Hoenicke, R. Tjeerdema. 2003. "Causes of sediment toxicity to *Mytilus galloprovincialis* in San Francisco Bay." *Archives of Environmental Contamination and Toxicology* 45:492-497.
- Poole, G., J. Dunham, M. Hicks, D. Keenan, J. Lockwood, E. Materna, D. McCullough, C. Mebane, J. Risley, S. Sauter, S. Spalding, and D. Sturdevant. 2001. *Technical Synthesis: scientific issues relating to temperature criteria for salmon, trout, and char native to the Pacific Northwest*. A summary report submitted to the Policy Workgroup of the EPA Region 10 Water Temperature Criteria Guidance Project. EPA 910-R-01-007. U. S. Environmental Protection Agency, Region 10 Office of Water, Seattle, WA.
- Puckett, M. 2002. *Quality Assurance Program Plan for the State of California's Surface Water Ambient Monitoring Program ("S.W.A.M.P")*. Also referred to as the Quality Assurance Management Plan, or the QAMP. Prepared by M. Puckett, California Department of Fish and Game, Monterey, CA. Available online at <http://www.waterboards.ca.gov/swamp/qapp.html#swampqapp>
- Rantz, S.E. 1971. "Isohyetal map of San Francisco Bay Region, California." U.S. Geological Survey, Menlo Park, CA.
- Rasmussen, D. 1997. *Toxic Substances Monitoring Program 1994-95 Data Report*. Division of Water Quality State Water Resources Control Board, Sacramento, CA.
- Rasmussen, D. 2000. *State Mussel Watch 1995-1997 Data Report*. Division of Water Quality, State Water Resources Control Board, Sacramento, CA.

- Redfield, A.C. 1958. "The biological control of chemical factors in the environment." *American Scientist* 46: 205-222.
- Rosenberg, D., and V. Resh. 1993. *Freshwater biomonitoring and benthic macroinvertebrates*. Chapman and Hall, New York City, NY.
- SCBWMI (Santa Clara Basin Watershed Management Initiative). 2003. *Watershed Characteristics Report (unabridged)*. Palo Alto, CA.
- SFBRWQCB. 1995. *San Francisco Bay Basin (Region 2) Water Quality Control Plan*. California Regional Water Quality Control Board, San Francisco Bay Region, Oakland, CA. Available at: <http://www.waterboards.ca.gov/sanfranciscobay/basinplan.htm>
- SFBRWQCB. 2002. *Surface Water Ambient Monitoring Program (SWAMP) Final Workplan 2001-2002*. (August 2001, revised August 2002.) San Francisco Bay Regional Water Quality Control Board, Oakland, CA. Available at: <http://www.waterboards.ca.gov/sanfranciscobay/monitoring.html>
- SFBRWQCB. 2004. *Surface Water Ambient Monitoring Program (SWAMP) Draft Workplan; Long-term 5-year Workplan and Annual 03-04 Workplan*. San Francisco Bay Regional Water Quality Control Board, Oakland, CA. Available at: <http://www.waterboards.ca.gov/sanfranciscobay/monitoring.html>
- SFBRWQCB. 2005. *Chemical concentrations in fish tissues from selected reservoirs and coastal areas in the San Francisco Bay Region; Final report*. Prepared by University of California, Davis, California Dept. of Fish and Game, and San Francisco Bay Regional Water Quality Control Board. San Francisco Bay Regional Water Quality Control Board, Oakland, CA. Available at: <http://www.waterboards.ca.gov/sanfranciscobay/monitoring.html>
- SFBRWQCB. 2005a. *Pathogens in Tomales Bay watershed—Total Maximum Daily Load (TMDL) Staff Report*, September 14. San Francisco Bay Regional Water Quality Control Board, Oakland, CA. Available at: <http://www.waterboards.ca.gov/sanfranciscobay/tmdlmain.htm>
- SFBRWQCB. 2005b. *Diazinon and pesticide-related toxicity in San Francisco Bay Area urban creeks Total Maximum Daily Load (TMDL) Staff Report*, November 9. San Francisco Bay Regional Water Quality Control Board, Oakland, CA. <http://www.waterboards.ca.gov/sanfranciscobay/tmdlmain.htm>
- Sigler, J., T. Bjorunn, F. Everest. 1984. "Effects of Chronic Turbidity on Density and Growth of Steelheads and Coho Salmon." *Transactions of the American Fisheries Society*, 113:142-150. American Fisheries Society.

- Stevens, H.H., Jr., J.F. Ficke and G.F. Smoot. 1975. "Water temperature, influential factors, field measurement and data presentation" in *Techniques of water-resources investigations of the US Geological Survey*. Book 1, collection of water data by direct measurement, chapter D1. U.S. Geological Survey, Arlington, VA,
- Sullivan, J. J., K.S. Goh. 2000. "Evaluation and validation of a commercial ELISA for diazinon in surface waters." *Journal of Agricultural and Food Chemistry*. 2000 48:4071-4078.
- Sullivan, K., D.J. Martin, R.D. Cardwell, J.E. Toll, S. Duke. 2000. *An analysis of the effects of temperature on salmonids of the Pacific Northwest with implications for selecting temperature criteria*. Sustainable Ecosystem Institute, Portland, OR.
- Sustainable Lands Stewardship Institute. 2002. *Biological and physical habitat assessment of selected sites in the Santa Rosa watershed*. Sustainable Lands Stewardship Institute, Sacramento, CA.
- Swartz, R.C., F.A. Cole, J.O. Lamberson, S.P. Ferraro, D.W. Schults, W.A. DeBen, H. Lee II, R.J. Ozretich. 1994. "Sediment toxicity, contamination and amphipod abundance at a DDT- and dieldrin-contaminated site in San Francisco Bay." *Environmental Toxicology and Chemistry* 13:949-962.
- Thompson, B., B. Anderson, J.W. Hunt, K. Taberski, and B. Phillips. 1999. "Relationships between sediment contamination and toxicity in San Francisco Bay." *Marine Environmental Research* 48: 1-25.
- Thompson, B., R. Hoenicke, J.A. Davis, and A. Gunther. 2000. "An overview of contaminant-related issues identified by monitoring in San Francisco Bay." *Environmental Monitoring and Assessment* 64(1):409-419.
- Thursby, G.B., J. Heltshe, K.J. Scott. 1997. "Revised approach to toxicity test acceptability criteria using a statistical performance assessment." *Environmental Toxicology and Chemistry* 16:1322-1329.
- URS Greiner Woodward Clyde. 1999. *Volunteer monitoring in San Leandro Creek: Technical summary report 1995-1998*. Alameda Countywide Clean Water Program, Hayward, CA.
- U.S. Fish and Wildlife Service. 1997. *Draft California freshwater shrimp (Syncaris pacifica Holmes 1895) recovery plan*. Portland, OR: U.S. Fish and Wildlife Service, Region 1.
- U.S. EPA. 1977. *Temperature criteria for freshwater fish: protocol and procedures*. Ecological Research Series. EPA-600/3-77-061 (NTIS PB270032). Prepared by W.A. Brungs and B.R. Jones. U.S. Environmental Protection Agency, Washington, D.C.
- U.S. EPA. 1986. *Ambient water quality criteria for bacteria – 1986*. EPA 440/5-84-002. Office of Water, Regulations and Standards, Criteria and Standards Division, Washington, D.C.

- U.S. EPA. 1991. *Monitoring guidelines to evaluate effects of forestry activities on streams in the Pacific Northwest and Alaska*. EPA/910/9-91-001. Prepared by L. H. MacDonald, A. Smart, and R.C. Wissmar. U.S. Environmental Protection Agency, Seattle, WA.
- U.S. EPA. 1994. *Short-term methods for estimating the chronic toxicity of effluents and receiving waters to freshwater organism* (Third Edition). EPA/600/4-91/002. Klemm, D.J. and G.E. Morrison (eds.). Environmental Monitoring Systems Laboratory, U. S. Environmental Protection Agency, Cincinnati, OH.
- U.S. EPA. 1997. *Volunteer stream monitoring: A methods manual*. EPA 841-B-97-003. Office of Water 4503F, Washington, D.C.
- U.S. EPA. 2000a. *Ambient water quality criteria recommendations—information supporting the development of state and tribal nutrient criteria for rivers and streams in nutrient Ecoregion III*. EPA 822-B-00-016. December 2000. U.S. Environmental Protection Agency, Office of Water, Office of Science and Technology, Health and Ecological Criteria Division, Washington, D.C.
- U.S. EPA. 2000b. *Methods for measuring the toxicity and bioaccumulation of sediment-associated contaminants with freshwater invertebrates*. Second edition. EPA/600/R-99/064. U.S. Environmental Protection Agency, Office of Water and Office of Research and Development, Washington, D.C.
- U.S. EPA. 2001. *National Primary Drinking Water Regulations; Arsenic and Clarifications to Compliance and New Source Contaminants Monitoring; Final Rule*. 40 CFR 141.62(b)(16).
- U.S. EPA. 2003. "2003 Progress Report Ecoregion 6 (Southern and Central California Oak and Chaparral Woodlands) Pilot Study for the Development of Nutrient Criteria." Prepared for U.S. EPA Region IX Technical Advisory Group and California State Water Resources Control Board Regional Board Advisory Group by Tetra Tech, Inc., Lafayette, CA.
- Weissich, P. R. 2005. Effects of Santa Clara Valley area dams on benthic macroinvertebrate assemblages. Masters Thesis, California State University at Hayward.
- Whitfield, M. 1974. "The hydrolysis of ammonium ions in sea water; a theoretical study." *Journal of the Marine Biological Association of the United Kingdom* 54:565-580.
- Whyte, D.C. 2002. *Staff Summary Report. California Regional Water Quality Control Board, San Francisco Bay Region*. October 16, 2002. San Francisco Bay Regional Water Quality Control Board, Oakland, CA. Available at <http://www.swrcb.ca.gov/rwqcb2/Agenda/10-16-02/10-16-02-5hssr.doc>
- Wilde, F.D., D.B. Radtke, J. Gibbs, and R.T. Iwatsubo. 1998. "Preparations for water sampling" in U.S. Geological Survey *Techniques of Water-Resources Investigations*, book 9, chap. A1. Available at <http://pubs.water.usgs.gov/twri9A1/> . Accessed 5/13/2004.

Wright, J.F., D. Moss, P.D. Armitage, and M.T. Furse. 1984. “A preliminary classification of running-water sites in Great Britain based on macro-invertebrate species and the prediction of community type using environmental data.” *Freshwater Biology* 14:221-256.

YSI. 1999. *6-Series Environmental Monitoring Systems Manual*. YSI, Inc., Yellow Springs, OH.

Zumwalt, C.W. 1972. “Consideration of vegetation and soils on the Tomales Bay Watershed” in *Tomales Bay Environmental Study*, R. Corwin (ed). The Conservation Foundation, Washington D.C.

Appendix A: Land uses by watershed and site

| Station Code | Station Name | Land Use Category |
|----------------------------------|--------------------------|--------------------------|
| <i>Walker Creek Watershed</i> | | |
| WLK030 | Keys at Tomales | Rural Residential |
| WLK050 | Keys @ Irvin Road | Mixed |
| WLK090 | Walker Creek | Mixed |
| WLK100 | Chileno Canyon | Grazing |
| WLK120 | Chileno Valley | Mixed |
| WLK130 | Laguna Lake | Mixed |
| WLK140 | Walker Canyon | Grazing |
| WLK160 | Walker Creek Ranch | Mixed |
| WLK162 | Turtle Pond | Mixed |
| WLK170 | Verde Canyon | Grazing |
| WLK180 | Gambonini Mine | Mixed |
| WLK190 | Salmon Creek | Grazing |
| WLK200 | Soulajule | Mixed |
| WLK230 | Arroyo Sausal | Grazing |
| WLK240 | Cheese Factory | Grazing |
| <i>Lagunitas Creek Watershed</i> | | |
| LAG040 | Olema Low | Mixed |
| LAG050 | Bear Valley Road Bridge | Mixed |
| LAG060 | Vedanta | Agriculture |
| LAG075 | Truttman | Agriculture |
| LAG085 | Five Brooks | Agriculture |
| LAG090 | Giacomini Gulch | Agriculture |
| LAG100 | Blue Line | Agriculture |
| LAG115 | Hagmaier Pond | Agriculture |
| LAG120 | Green Bridge (Hwy 1) | Mixed |
| LAG130 | Gallagher's Ranch | Mixed |
| LAG150 | Halleck | Mixed |
| LAG160 | Nicasio | Rural Residential |
| LAG165 | Below Tocaloma | Mixed |
| LAG170 | Tocaloma | Open Space |
| LAG180 | Cheda | Mixed |
| LAG185 | Swimming Hole @ SPTaylor | Mixed |
| LAG190 | Devil's Gulch | Mixed |
| LAG210 | Taylor Park | Mixed |
| LAG220 | Irving Bridge | Mixed |
| LAG230 | Inkwells | Mixed |
| LAG240 | White Horse Bridge | Rural Residential |
| LAG250 | Papermill Cr. Saloon | Rural Residential |
| LAG270 | Creamery Gulch | Rural Residential |
| LAG290 | Water Treatment Plant | Rural Residential |
| LAG300 | Woodacre Creek | Rural Residential |
| LAG320 | Shafter Bridge | Mixed |
| LAG330 | Big Carson 1 | Open Space |
| LAG335 | Big Carson 2 | Open Space |
| LAG380 | Little Carson | Open Space |
| LAG390 | Cataract | Open Space |

Appendix A: Land uses by watershed and site

| Station Code | Station Name | Land Use Category |
|------------------------------------|------------------------------------|--------------------------|
| <i>San Leandro Creek Watershed</i> | | |
| SLE030 | Empire Road | Urban |
| SLE050 | San Leandro BART | Urban |
| SLE070 | Root Park | Urban |
| SLE090 | Chabot City Park | Mixed |
| SLE170 | Redwood Park (West Fork) | Mixed |
| SLE180 | East Fork Redwood Creek | Open Space |
| SLE190 | Canyon School | Rural Residential |
| SLE200 | Huckleberry Preserve | Rural Residential |
| SLE210 | Moraga | Urban |
| SLE220 | Indian | Grazing |
| SLE230 | Kaiser Creek at Callahan | Open Space |
| <i>Wildcat Creek Watershed</i> | | |
| WIL020 | Richmond Parkway | Urban |
| WIL030 | 3rd St. | Urban |
| WIL050 | Davis Park | Urban |
| WIL060 | Vale Road | Urban |
| WIL070 | Alvarado Park | Mixed |
| WIL100 | Jewel Lake Outlet | Mixed |
| WIL130 | Lone Oak | Mixed |
| WIL170 | Botanic Garden | Mixed |
| WIL180 | Big Springs Picnic Area | Open Space |
| WIL190 | Poosum Picnic Area | Mixed |
| <i>San Pablo Creek Watershed</i> | | |
| SPA020 | 3rd St. Bridge | Urban |
| SPA050 | 20th St. at Road 20 | Urban |
| SPA060 | San Pablo City Park | Urban |
| SPA070 | Cemetery Bridge | Urban |
| SPA100 | Appian Way | Urban |
| SPA110 | De Anza School | Urban |
| SPA130 | Castro Ranch | Mixed |
| SPA140 | Xmas Tree Farm | Mixed |
| SPA150 | Briones 1 | Grazing |
| SPA160 | Briones 2 | Open Space |
| SPA170 | Bear Creek Road | Mixed |
| SPA200 | Lauterwasser | Urban |
| SPA220 | Orinda Village | Urban |
| SPA235 | Camino Encinas Low | Urban |
| SPA240 | Camino Encinas | Urban |
| <i>Suisun Creek Watershed</i> | | |
| 207SUI010 | Cordelia | Mixed |
| 207SUI020 | Rockville | Agriculture |
| 207SUI050 | Putah South Canal - Downstream | Agriculture |
| 207SUI060 | Putah South Canal - Upstream | Agriculture |
| 207SUI090 | Capp Confluence | Agriculture |
| 207SUI110 | Wooden Valley | Mixed |
| 207SUI125 | Wolf | Mixed |
| 207SUI130 | Lake Curry Road | Agriculture |
| 207SUI180 | White Creek | Mixed |
| 207SUI185 | East Wooden Valley above White Cr. | Agriculture |
| 207SUI210 | E Wooden Valley | Agriculture |
| 207SUI250 | Gordon Valley | Grazing |
| 207SUI260 | Upper Suisun | Grazing |

Appendix A: Land uses by watershed and site

| Station Code | Station Name | Land Use Category |
|---|------------------------------|--------------------------|
| <i>Arroyo Las Positas Creek Watershed</i> | | |
| ALP010 | El Charro | Urban |
| ALP040 | Airway Blvd. Exit | Urban |
| ALP070 | Airway 2 | Urban |
| ALP080 | N. Livermore Ave. | Urban |
| ALP100 | Altamont Creek | Mixed |
| ALP105 | Altamont-Springtown | Urban |
| ALP110 | Arroyo Seco | Mixed |
| ALP140 | Altamont Pass | Agriculture |
| ALP150 | Above Arroyo Seco | Mixed |
| <i>Pescadero Creek Watershed</i> | | |
| PES050 | Water Lane | Mixed |
| PES060 | Community Church | Mixed |
| PES070 | Cloverdale Rd | Agriculture |
| PES080 | Honsinger | Agriculture |
| PES095 | Pesky Ranch | Agriculture |
| PES100 | PES USGS Gage | Rural Residential |
| PES105 | Canyon Mouth | Rural Residential |
| PES120 | Loma Mar | Mixed |
| PES134 | Memorial Park Swim 1 | Mixed |
| PES135 | Memorial Park Swim 2 | Mixed |
| PES140 | Water Treatment Plant | Mixed |
| PES150 | Jones Gulch | Rural Residential |
| PES160 | Towne Fire Rd. (Steve Young) | Mixed |
| PES170 | Tarwater Creek | Mixed |
| PES180 | Peters Creek | Open Space |
| PES190 | Portola State Park | Open Space |
| PES193 | Iverson Trail | Open Space |
| PES194 | Sequoia Nature Trail | Open Space |
| PES200 | Slate Creek | Open Space |
| PES210 | Oil Creek | Open Space |
| PES230 | Waterman Creek | Mixed |
| PES240 | Headwaters | Open Space |
| <i>Butano Creek Watershed</i> | | |
| BUT010 | Bean Hollow | Agriculture |
| BUT020 | Cloverdale Coast Ranch | Agriculture |
| BUT030 | Girl Scout Camp | Rural Residential |
| BUT040 | Butano Falls | Open Space |
| BUT050 | Butano State Park | Open Space |
| <i>San Gregorio Creek Watershed</i> | | |
| SGR010 | San Gregorio USGS Gage | Mixed |
| SGR020 | El Corte de Madera Low | Mixed |
| SGR030 | Star Hill Rd. | Rural Residential |
| SGR040 | Boysville | Mixed |
| SGR060 | Harrington Creek | Rural Residential |
| SGR075 | Upper San Gregorio | Rural Residential |
| SGR079 | San Gregorio confluence | Rural Residential |
| SGR080 | La Honda at Confluence | Rural Residential |
| SGR090 | Alpine at Confluence | Rural Residential |
| SGR100 | Playbowl | Rural Residential |
| SGR110 | Spanish Ranch | Rural Residential |
| SGR120 | Sky Londa | Rural Residential |
| SGR130 | Mindego Creek | Rural Residential |
| SGR150 | Heritage Grove | Rural Residential |

Appendix A: Land uses by watershed and site

| Station Code | Station Name | Land Use Category |
|-----------------------------------|--------------------------------|--------------------------|
| <i>Stevens Creek Watershed</i> | | |
| STE020 | La Avenida | Urban |
| STE030 | Landels School | Urban |
| STE040 | Below Diversion Channel | Urban |
| STE060 | "Belleville"/Barranca | Urban |
| STE070 | Chestnut Picnic Area | Mixed |
| STE080 | Above the Reservoir | Mixed |
| STE090 | Camp Cooley | Mixed |
| STE100 | Moss Rock | Rural Residential |
| STE110 | Upper Stevens 1 | Rural Residential |
| STE120 | Upper Stevens 2 | Rural Residential |
| <i>Permanente Creek Watershed</i> | | |
| PER010 | Charleston Rd | Urban |
| PER020 | Crittendon Middle School | Urban |
| PER030 | Hale Creek @ Covington | Urban |
| PER040 | Permanente @ Diversion Channel | Urban |
| PER050 | Loyola Corners | Urban |
| PER070 | Rancho San Antonio | Mixed |
| PER080 | Lower Meadow/West Branch | Mixed |

Appendix B: Station locations by watershed and station ID

| Station Code | Station Name | Waterbody | Latitude | Longitude |
|----------------------------------|---------------------------|--------------------|----------|------------|
| <i>Walker Creek Watershed</i> | | | | |
| 201WLK030 | Keys at Tomales | Keys Creek | 38.24111 | -122.90431 |
| 201WLK050 | Keys @ Irvin Road | Keys Creek | 38.24353 | -122.89806 |
| 201WLK090 | Walker Creek | Walker Creek | 38.22778 | -122.90806 |
| 201WLK100 | Chileno Canyon | Chileno Creek | 38.21450 | -122.85794 |
| 201WLK120 | Chileno Valley | Chileno Creek | 38.20764 | -122.79456 |
| 201WLK130 | Laguna Lake | Lake | 38.20639 | -122.76975 |
| 201WLK140 | Walker Canyon | Walker Creek | 38.21192 | -122.86006 |
| 201WLK160 | Walker Creek Ranch | Walker Creek | 38.17544 | -122.82044 |
| 201WLK162 | Turtle Pond | Pond | 38.16782 | -122.81734 |
| 201WLK170 | Verde Canyon | Verde Canyon | 38.16444 | -122.81136 |
| 201WLK180 | Gambonini Mine | Salmon Creek | 38.16175 | -122.78033 |
| 201WLK190 | Salmon Creek | Salmon Creek | 38.16458 | -122.77194 |
| 201WLK200 | Soulajoule | Sausal, Arroyo | 38.15758 | -122.78408 |
| 201WLK230 | Arroyo Sausal | Sausal, Arroyo | 38.14342 | -122.72472 |
| 201WLK240 | Cheese Factory | Sausal, Arroyo | 38.13589 | -122.70914 |
| <i>Lagunitas Creek Watershed</i> | | | | |
| 201LAG040 | Olema Low | Olema Creek | 38.05731 | -122.80311 |
| 201LAG050 | Bear Valley Road Bridge | Olema Creek | 38.04181 | -122.78969 |
| 201LAG060 | Vedanta | Olema Creek | 38.03808 | -122.78531 |
| 201LAG075 | Truttman | Olema Creek | 38.02306 | -122.77917 |
| 201LAG085 | Five Brooks | Olema Creek | 37.99905 | -122.75533 |
| 201LAG100 | Blue Line | Olema Creek | 37.99614 | -122.75578 |
| 201LAG115 | Hagmeier Pond | Hagmeier Pond | 37.97382 | -122.72711 |
| 201LAG120 | Green Bridge (Hwy 1) | Olema Creek | 38.06470 | -122.80860 |
| 201LAG130 | Gallagher's Ranch | Lagunitas Creek | 38.08064 | -122.78450 |
| 201LAG150 | Halleck | Halleck Creek | 38.06669 | -122.70264 |
| 201LAG160 | Nicasio | Nicasio Creek | 38.06156 | -122.69958 |
| 201LAG165 | Below Tocaloma | Lagunitas Creek | 38.05839 | -122.76522 |
| 201LAG170 | Tocaloma | Lagunitas Creek | 38.04970 | -122.75945 |
| 201LAG180 | Cheda | Lagunitas Creek | 38.03722 | -122.74611 |
| 201LAG185 | Swimming Hole @ SPTaylor | Lagunitas Creek | 38.02872 | -122.74326 |
| 201LAG190 | Devils Gulch | Devils Gulch | 38.02964 | -122.73636 |
| 201LAG210 | Taylor Park | Lagunitas Creek | 38.01861 | -122.73306 |
| 201LAG220 | Irving Bridge | Lagunitas Creek | 38.01611 | -122.72297 |
| 201LAG230 | Inkwells | San Geronimo Creek | 38.00500 | -122.70833 |
| 201LAG240 | White Horse Bridge | San Geronimo Creek | 38.00703 | -122.70569 |
| 201LAG250 | Papermill Cr. Saloon | San Geronimo Creek | 38.01472 | -122.68917 |
| 201LAG270 | Creamery Gulch | San Geronimo Creek | 38.01356 | -122.66664 |
| 201LAG290 | Lag Water Treatment Plant | San Geronimo Creek | 38.01305 | -122.65055 |
| 201LAG300 | Woodacre Creek | San Geronimo Creek | 38.01275 | -122.64689 |
| 201LAG320 | Shafter Bridge | Lagunitas Creek | 38.00453 | -122.70878 |
| 201LAG330 | Big Carson 1 | Big Carson Creek | 37.99194 | -122.66972 |
| 201LAG335 | Big Carson 2 | Big Carson Creek | 37.99222 | -122.66000 |
| 201LAG380 | Little Carson | Big Carson Creek | 37.96722 | -122.64944 |
| 201LAG390 | Cataract | Cataract Creek | 37.93250 | -122.63556 |

Appendix B: Station locations by watershed and station ID

| Station Code | Station Name | Waterbody | Latitude | Longitude |
|------------------------------------|----------------------------------|---------------------|----------|------------|
| <i>San Leandro Creek Watershed</i> | | | | |
| 204SLE030 | Empire Road | San Leandro Creek | 37.72556 | -122.18361 |
| 204SLE050 | San Leandro BART | San Leandro Creek | 37.72611 | -122.16389 |
| 204SLE070 | Root Park | San Leandro Creek | 37.72706 | -122.15733 |
| 204SLE090 | Chabot City Park | San Leandro Creek | 37.73139 | -122.13206 |
| 204SLE170 | Redwood Park (West Fork) | San Leandro Creek | 37.80028 | -122.14472 |
| 204SLE180 | East Fork Redwood Creek | Redwood Creek | 37.80111 | -122.14472 |
| 204SLE190 | Canyon School | San Leandro Creek | 37.82778 | -122.16397 |
| 204SLE200 | Huckleberry Preserve | San Leandro Creek | 37.84150 | -122.18725 |
| 204SLE210 | Moraga Creek | Moraga Creek | 37.84333 | -122.14972 |
| 204SLE220 | Indian Creek | Indian Creek | 37.81833 | -122.14583 |
| 204SLE230 | Kaiser Creek at Callahan | Kaiser Creek | 37.79778 | -122.07528 |
| <i>Wilcat Creek Watershed</i> | | | | |
| 206WIL020 | Richmond Parkway | Wildcat Creek | 37.95778 | -122.37389 |
| 206WIL030 | 3rd St. | Wildcat Creek | 37.96028 | -122.36750 |
| 206WIL050 | Davis Park | Wildcat Creek | 37.96136 | -122.35250 |
| 206WIL060 | Vale Road | Wildcat Creek | 37.95319 | -122.33836 |
| 206WIL070 | Alvarado Park | Wildcat Creek | 37.95233 | -122.32039 |
| 206WIL100 | Jewel Lake Outlet | Wildcat Creek | 37.91306 | -122.27000 |
| 206WIL130 | Lone Oak (130) | Wildcat Creek | 37.90361 | -122.25889 |
| 206WIL170 | Botanic Garden | Wildcat Creek | 37.89333 | -122.24333 |
| 206WIL180 | Big Springs Picnic Area | Wildcat Creek | 37.88944 | -122.23361 |
| 206WIL190 | Possum Picnic Area | Wildcat Creek | 37.88444 | -122.22944 |
| <i>San Pablo Creek Watershed</i> | | | | |
| 206SPA020 | 3rd St. Bridge | San Pablo Creek | 37.96750 | -122.36583 |
| 206SPA050 | 20th St. at Road 20 | San Pablo Creek | 37.96333 | -122.34833 |
| 206SPA060 | San Pablo City Park | San Pablo Creek | 37.96250 | -122.34639 |
| 206SPA070 | Cemetery Bridge | San Pablo Creek | 37.96278 | -122.33278 |
| 206SPA100 | Appian Way | San Pablo Creek | 37.97031 | -122.30642 |
| 206SPA110 | De Anza School | San Pablo Creek | 37.96889 | -122.29042 |
| 206SPA130 | Castro Ranch | San Pablo Creek | 37.95592 | -122.26992 |
| 206SPA140 | Xmas Tree Farm | Bear Creek | 37.92194 | -122.16333 |
| 206SPA150 | Briones 1 | Bear Creek | 37.92611 | -122.15889 |
| 206SPA160 | Briones 2 | Bear Creek | 37.92722 | -122.15250 |
| 206SPA170 | Bear Creek Road | San Pablo Creek | 37.90111 | -122.20722 |
| 206SPA200 | Lauterwasser Creek | Lauterwasser Creek | 37.89139 | -122.19389 |
| 206SPA220 | Orinda Village | San Pablo Creek | 37.88611 | -122.19278 |
| 206SPA235 | Camino Encinas Low | San Pablo Creek | 37.87623 | -122.18028 |
| 206SPA240 | Camino Encinas | San Pablo Creek | 37.87250 | -122.17861 |
| <i>Suisun Creek Watershed</i> | | | | |
| 207SUI010 | Cordelia | Suisun Creek | 38.21833 | -122.10444 |
| 207SUI020 | Rockville | Suisun Creek | 38.24472 | -122.11194 |
| 207SUI050 | Putah South Canal - Downstream | Suisun Creek | 38.27483 | -122.12186 |
| 207SUI060 | Putah South Canal (upstream) | Suisun Creek | 38.27458 | -122.12275 |
| 207SUI090 | Capp Confluence | Suisun Creek | 38.32831 | -122.13708 |
| 207SUI110 | Wooden Valley | Wooden Valley Creek | 38.33086 | -122.13858 |
| 207SUI125 | Middle Suisun | Suisun Creek | 38.33361 | -122.12800 |
| 207SUI130 | Lake Curry Road | Suisun Creek | 38.34956 | -122.13061 |
| 207SUI180 | White Creek | White Creek | 38.35689 | -122.17086 |
| 207SUI185 | East Wooden Valley above White C | Wooden Valley Creek | 38.35742 | -122.17120 |
| 207SUI210 | E Wooden Valley | Suisun Creek | 38.36333 | -122.17444 |
| 207SUI250 | Gordon Valley | Suisun Creek | 38.37361 | -122.14500 |
| 207SUI260 | Upper Suisun | Suisun Creek | 38.37528 | -122.14222 |

Appendix B: Station locations by watershed and station ID

| Station Code | Station Name | Waterbody | Latitude | Longitude |
|---|------------------------------|--------------------|----------|------------|
| <i>Arroyo Las Positas Creek Watershed</i> | | | | |
| 204ALP010 | El Charro | Arroyo Las Positas | 37.69708 | -121.84964 |
| 204ALP040 | Airway Blvd. Exit | Arroyo Las Positas | 37.69856 | -121.81728 |
| 204ALP070 | Airway 2 | Arroyo Las Positas | 37.69961 | -121.79886 |
| 204ALP080 | N. Livermore Ave. | Arroyo Las Positas | 37.69942 | -121.77442 |
| 204ALP100 | Altamont Creek | Altamont Creek | 37.70739 | -121.75325 |
| 204ALP105 | Altamont-Springtown | Altamont Creek | 37.71533 | -121.74780 |
| 204ALP110 | Arroyo las Positas | Arroyo Seco | 37.70528 | -121.75417 |
| 204ALP140 | Altamont Pass | Arroyo Las Positas | 37.72175 | -121.69658 |
| 204ALP150 | Above Arroyo Seco | Arroyo Seco | 37.69441 | -121.73578 |
| <i>Pescadero Creek Watershed</i> | | | | |
| 202PES050 | Water Lane | Pescadero Creek | 37.25800 | -122.39492 |
| 202PES060 | Community Church | Pescadero Creek | 37.25449 | -122.38314 |
| 202PES070 | Cloverdale Rd | Pescadero Creek | 37.25164 | -122.36999 |
| 202PES080 | Honsinger | Honsinger Creek | 37.25200 | -122.36980 |
| 202PES095 | Pesky Ranch | Pescadero Creek | 37.25649 | -122.33777 |
| 202PES100 | PES USGS Gage | Pescadero Creek | 37.26098 | -122.32909 |
| 202PES105 | Canyon Mouth | Pescadero Creek | 37.26966 | -122.31538 |
| 202PES120 | Loma Mar | Pescadero Creek | 37.26695 | -122.30935 |
| 202PES134 | Memorial Park Swim 1 | Pescadero Creek | 37.27345 | -122.29180 |
| 202PES135 | Memorial Park Swim 2 | Pescadero Creek | 37.27334 | -122.29003 |
| 202PES140 | Water Treatment Plant | Hoffman Creek | 37.27119 | -122.28647 |
| 202PES150 | Jones Gulch | Jones Gulch | 37.27605 | -122.26669 |
| 202PES160 | Towne Fire Rd. (Steve Young) | Pescadero Creek | 37.26927 | -122.26368 |
| 202PES170 | Tarwater Creek | Tarwater Creek | 37.26443 | -122.23911 |
| 202PES180 | Peters Creek | Peters Creek | 37.25137 | -122.21792 |
| 202PES190 | Portola State Park | Pescadero Creek | 37.25106 | -122.21783 |
| 202PES193 | Iverson Trail | Pescadero Creek | 37.25731 | -122.23508 |
| 202PES194 | Sequoia Nature Trail | Pescadero Creek | 37.25030 | -122.21972 |
| 202PES200 | Slate Creek | Slate Creek | 37.24190 | -122.20041 |
| 202PES210 | Oil Creek | Oil Creek | 37.22935 | -122.19052 |
| 202PES230 | Waterman Creek | Waterman Creek | 37.21446 | -122.17565 |
| 202PES240 | Headwaters | Pescadero Creek | 37.21410 | -122.16521 |
| <i>Butano Creek Watershed</i> | | | | |
| 202BUT010 | Bean Hollow | Butano Creek | 37.25003 | -122.39553 |
| 202BUT020 | Cloverdale Coast Ranch | Butano Creek | 37.23395 | -122.36719 |
| 202BUT030 | Girl Scout Camp | Little Butano Cree | 37.22474 | -122.33254 |
| 202BUT040 | Butano Falls | Little Butano Cree | 37.24195 | -122.31719 |
| 202BUT050 | Butano State Park | Butano Creek | 37.20607 | -122.33483 |
| <i>San Gregorio Creek Watershed</i> | | | | |
| 202SGR010 | San Gregorio USGS Gage | San Gregorio Creek | 37.32583 | -122.38583 |
| 202SGR020 | El Corte de Madera Low | Corte Madera Creek | 37.31962 | -122.33811 |
| 202SGR030 | Star Hill Rd. | Corte Madera Creek | 37.38509 | -122.32473 |
| 202SGR040 | Boysville | San Gregorio Creek | 37.31120 | -122.31595 |
| 202SGR060 | Harrington Creek | Harrington Creek | 37.32027 | -122.30026 |
| 202SGR075 | Upper San Gregorio | San Gregorio Creek | 37.31512 | -122.29048 |
| 202SGR079 | San Gregorio confluence | San Gregorio Creek | 37.31003 | -122.27776 |
| 202SGR080 | La Honda at Confluence | Honda Creek, La | 37.31024 | -122.27774 |
| 202SGR090 | Alpine at Confluence | Alpine Creek | 37.31005 | -122.27740 |
| 202SGR100 | Playbowl | Honda Creek, La | 37.31524 | -122.27408 |
| 202SGR110 | Spanish Ranch | Honda Creek, La | 37.34591 | -122.27425 |
| 202SGR120 | Sky Londa | Honda Creek, La | 37.38799 | -122.27345 |
| 202SGR130 | Mindego Creek | Mindego Creek | 37.29775 | -122.25393 |
| 202SGR150 | Heritage Grove | Alpine Creek | 37.29606 | -122.24639 |

Appendix B: Station locations by watershed and station ID

| Station Code | Station Name | Waterbody | Latitude | Longitude |
|-----------------------------------|--------------------------------|------------------|----------|------------|
| <i>Steven's Creek Watershed</i> | | | | |
| 205STE020 | La Avenida | Stevens Creek | 37.41357 | -122.06865 |
| 205STE030 | Landels School | Stevens Creek | 37.38658 | -122.06917 |
| 205STE040 | Below Diversion Channel | Stevens Creek | 37.36475 | -122.06224 |
| 205STE060 | "Belleville"/Barranca | Stevens Creek | 37.33503 | -122.06384 |
| 205STE070 | Chestnut Picnic Area | Stevens Creek | 37.30320 | -122.07456 |
| 205STE080 | Above the Reservoir | Stevens Creek | 37.28367 | -122.07694 |
| 205STE090 | Camp Cooley | Stevens Creek | 37.28038 | -122.07334 |
| 205STE100 | Moss Rock | Stevens Creek | 37.27226 | -122.08278 |
| 205STE110 | Upper Stevens 1 | Stevens Creek | 37.28705 | -122.12601 |
| 205STE120 | Upper Stevens 2 | Stevens Creek | 37.31578 | -122.16917 |
| <i>Permanente Creek Watershed</i> | | | | |
| 205PER010 | Charleston Rd | Permanente Creek | 37.42118 | -122.08673 |
| 205PER020 | Crittendon Middle School | Permanente Creek | 37.41206 | -122.08663 |
| 205PER030 | Hale Creek @ Covington | Hale Creek | 37.36565 | -122.09877 |
| 205PER040 | Permanente @ Diversion Channel | Permanente Creek | 37.36245 | -122.08656 |
| 205PER050 | Loyola Corners | Permanente Creek | 37.35264 | -122.08617 |
| 205PER070 | Rancho San Antonio | Permanente Creek | 37.32941 | -122.08586 |
| 205PER080 | Lower Meadow/West Branch | Permanente Creek | 37.33335 | -122.09381 |

Appendix C-1: Data batches missing blank(s)

| Batch ID | BatchQualCode | BatchValCode | Notes | Lab |
|----------------|---------------|--------------|--|----------|
| 012703-TDS | A,VMD | VAC,VI | QAO: no blank or CRM | |
| 012903-TDS | A,VMD | VAC,VI | QAO: no blank or CRM | |
| AMS10401-1 | A,VMD | VAC,VI | QAO: no blk, DUP, CRM | AMS |
| AMSDOC21102-1 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSDOC21302-1 | A,VMD | VAC,VI | LAB:all samples were lab filtered within 3 days of collectionQA: no blk, DUP | AMS |
| AMSDOC2902-1 | A,VMD | VAC,VI | QAO: no blk | AMS |
| AMSDOC2902-2 | A,VMD | VAC,VI | QAO: no blk | AMS |
| AMSDOC51102-1 | A,VMD | VAC,VI | QAO: no blk, DUP | |
| AMSDOC5302-1 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSDOC5302-2 | A,VMD | VAC,VI | QAO: no blk | AMS |
| AMSDOC5402-1 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSDOC5402-2 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSDOC5402-3 | A,VMD | VAC,VI | QAO: no blk, DUP | |
| AMSDOC62302-1 | A,VMD | VAC,VI | QAO: no blk, DUP | |
| AMSDOC62402-1 | A,VMD | VAC,VI | QAO: no blk, DUP | |
| AMSDOC62402-2 | A,VMD | VAC,VI | QAO: no blk, DUP | |
| AMSTOC51102-1 | A,VMD | VAC,VI | QAO: no blk, DUP | |
| L-012503-OCH | A,VMD | VAC,VI | QAO: no QC (blank, MS/MSD, LCS) for extraction date 1/31/03; MS/MSD RPD for p'p'-DDD was above QC limit | |
| L-012503-OCH | A,VMD | VAC,VI | QAO: no QC (blank, MS/MSD, LCS) for extraction date 1/31/03; MS/MSD RPD for p'p'-DDD was above QC limit | |
| L-012503-OP | A,VMD | VAC,VI | QAO: no QC (blank, MS/MSD, LCS) for extraction date 1/31/03; Both MS/MSD RPDs above 25% for coumaphos, demeton-s, tetrachlorvinphos | |
| L-012503-OP | A,VMD | VAC,VI | QAO: no QC (blank, MS/MSD, LCS) for extraction date 1/31/03; Both MS/MSD RPDs above 25% for coumaphos, demeton-s, tetrachlorvinphos | |
| L-012503-PAH | A,MD | VAC,VI | LAB:holding times exceeded; low surrogate %Rs<50% for many samples, MS/D, LCS %R & RPD outside criteria for several cmpds; See QA CodesQAO:no LCS or MS/MSD for 1/25/03, no QC for 1/31/03, no blank for 1/29/03 | |
| L-012503-PAH | A,MD | VAC,VI | LAB:holding times exceeded; low surrogate %Rs<50% for many samples, MS/D, LCS %R & RPD outside criteria for several cmpds; See QA CodesQAO:no LCS or MS/MSD for 1/25/03, no QC for 1/31/03, no blank for 1/29/03 | |
| L-012503-PCB | A,VMD | VAC,VI | QAO: no QC(blank, MS/MSD, LCS) for extraction date 1/31/03 | |
| L-012503-PCB | A,VMD | VAC,VI | QAO: no QC(blank, MS/MSD, LCS) for extraction date 1/31/03 | |
| L-012503-TRIAZ | A,VMD | VAC,VI | QAO: no QC(blank, MS/MSD, LCS) for extraction date 1/31/03, per laboratory surrogates are reported with the OP pest batch L-012503-OP, MS/MSD RPD out for Ametryn | |
| L-012503-TRIAZ | A,VMD | VAC,VI | QAO: no QC(blank, MS/MSD, LCS) for extraction date 1/31/03, per laboratory surrogates are reported with the OP pest batch L-012503-OP, MS/MSD RPD out for Ametryn | |
| L-041802-OPP | A,VMD | VAC,VI | QAO: no blk for 4/15/2002 and no LCS for 4/12/2002 | DFG-WPCL |
| L-041802-OPP | A,VMD | VAC,VI | QAO: no blk for 4/15/2002 and no LCS for 4/12/2002 | |
| L-041802-PAH | A,VMD | VAC,VI | QAO: no blank,MS/MSD,LCS for extraction date 4/15/2002;MS/MSD RPD for napthalene above QC limit | DFG-WPCL |
| L-041802-PAH | A,VMD | VAC,VI | QAO: no blank,MS/MSD,LCS for extraction date 4/15/2002;MS/MSD RPD for napthalene above QC limit | |
| L-042702-OCH | A,VMD | VAC,VI | QAO: no CRM, no Blk for 4/15/02, no LCS for 4/12/02 | DFG-WPCL |
| L-042702-OCH | A,VMD | VAC,VI | QAO: no CRM, no Blk for 4/15/02, no LCS for 4/12/02 | |
| L-042702-PCB | A,VMD | VAC,VI | QAO: no CRM, no Blk for 4/15/02, no LCS for 4/12/02 | DFG-WPCL |
| L-042702-PCB | A,VMD | VAC,VI | QAO: no CRM, no Blk for 4/15/02, no LCS for 4/12/02 | |
| L-050302-PAH | A,VMD | VAC,VI | QAO: no blank, MS/MSD, LCS for 4/22/02, no blank for 4/23/02 | |
| L-050602-TR | A,VMD | VAC,VI | QAO: no blank was performed, no QC since screening data only | |

Appendix C-1: Data batches missing blank(s)

| Batch ID | BatchQualCode | BatchValCode | Notes | Lab |
|----------------------|---------------|--------------|--|----------|
| L-050602-TR | A,VMD | VAC,VI | QAO: no blank was performed, no QC since screening data only | |
| L-070602-OPP | A,VMD | VAC,VI | QAO: no LCS for 6/22/02, no blk, LCS, MS/MSD for 6/26/02, no blk, MS/MSD for 6/30/02, MS/MSD RPDs above QC limit of 25% | |
| L-070602-OPP | A,VMD | VAC,VI | QAO: no LCS for 6/22/02, no blk, LCS, MS/MSD for 6/26/02, no blk, MS/MSD for 6/30/02, MS/MSD RPDs above QC limit of 25% | |
| L-071902-OCH | A,VMD | VAC,VI | QAO: associated blk, MS/MSD and LCS reported in batch L-072002-OCH | |
| L-071902-OCH | A,VMD | VAC,VI | QAO: associated blk, MS/MSD and LCS reported in batch L-072002-OCH | |
| L-071902-PCB | A,VMD | VAC,VI | QAO: no blank or LCS for 6/26 or 6/30, no MS/MSD for 6/22 or 6/30 | |
| L-071902-PCB | A,VMD | VAC,VI | QAO: no blank or LCS for 6/26 or 6/30, no MS/MSD for 6/22 or 6/30 | |
| L-101001-OCH | A,MD | VAC,VI | QAO: no blank for 10/7/2001, MS/MSD RPD for endosulfan I above QC limit of 25% | DFG-WPCL |
| L-101001-OPP | A,VMD | VAC,VI | QAO: no MSD; no blank for 10/7/2001 | DFG-WPCL |
| L-101001-PAH | A,VMD | VAC,VI | QAO: no blank for 10/7/2001 | DFG-WPCL |
| L-101001-PCB | A,VMD | VAC,VI | QAO: no MSD; no blank for 10/7/2001 | DFG-WPCL |
| L39801_BS179_KR_CONG | A,VMD | VAC,VI | QAO: no QC for 11/20/01, samples added to batch after extraction process began; LCS %R outside QC limit for PCB-29, PCB-56, PCB-60; MS/MSD %R PCB-029 and PCB-056 | |
| L39801BS179_KRPEST | A,VMD | VAC,VI | QAO: no QC for 11/20/01, samples added to batch after extraction process began; LCS %R outside QC limit for several compounds; CRM %R outside QC limit for trans chlordane, alpha HCH, cis Nonachlor, MS/MSD %R and RPD outside QC limit for several compounds | |
| L-418-01-ALK | A,VMD | VAC,VI | QAO: no blank | DFG-WPCL |
| L-418-01-NO2 | A,VMD | VAC,VI | QAO: no blank | DFG-WPCL |
| L-424-01-ALK | A,VMD | VAC,VI | QAO: no blank | DFG-WPCL |
| L-424-01-NO2 | A,VMD | VAC,VI | QAO: no blank or MS | DFG-WPCL |
| L-430-01-ALK | A,VMD | VAC,VI | QAO: no blank | DFG-WPCL |
| L-430-01-NH3 | A,VMD | VAC,VI | QAO: no blank | DFG-WPCL |
| L-430-01-NO2 | A,VMD | VAC,VI | QAO: no blank or MS | DFG-WPCL |
| L-442-01-ALK | A,VMD | VAC,VI | QAO: no blank | |
| L-442-01-NO2 | A,VMD | VAC,VI | QAO: no blank or MS | |
| L-442-01-TDS | A,VMD | VAC,VI | QAO: no blank, MS, DUP | |
| L-445-01-ALK | A,VMD | VAC,VI | QAO: no blank | |
| L-445-01-NO2 | A,VMD | VAC,VI | QAO: no blank or MS | |
| L-445-01-TDS | A,VMD | VAC,VI | QAO: no blank | |
| R2-013002-Chl | A,VMD | VAC,VI | LAB:all samples were field filtered on the date of collectionQAO: no blank or DUP | |
| R2-013002-Ph | A,VMD | VAC,VI | LAB:all samples were field filtered on the date of collectionQAO: no blk or DUP | |
| R2-071002-ICP | A,VMD | VAC,VI | LAB:all samples were acidified within 2 days of collectionQAO: no blank or CRM; MS/MSD %Rs outside QC limit of 75-125 | |
| R2-100101-Chl | A,VMD | VAC,VI | LAB:all samples were field filtered on the date of collection QAO: no blank or DUP | |
| R2-100101-Ph | A,VMD | VAC,VI | LAB:all samples were field filtered on the date of collectionQAO: no blk or DUP | |
| R2-101001-Chl | A,VMD | VAC,VI | LAB:all samples were field filtered on the date of collection QAO: no blk or DUP | |
| R2-101001-ICP | A,VMD | VAC,VI | LAB:sample was acidified within 2 days of collection;QAO: no blank, MS %R for Cu and Cr outside QC limit of 75-125 | |
| R2-101001-Ph | A,VMD | VAC,VI | LAB:all samples were field filtered on the date of collectionQAO: no blk or DUP | |
| R29-101102-ICP | A,VMD | VAC,VI | LAB:All samples were acidified within 2 days of collectionQAO: no blank | |
| R29-101102-ICP | A,VMD | VAC,VI | LAB:All samples were acidified within 2 days of collectionQAO: no blank | |

Appendix C-2: Laboratory blanks detected

| Analyte | Result | Unit | MDL | RL | Detected | AnalysisDate | MethodName | Laboratory | Batch ID |
|-----------------------|--------|-------|-------|-------|----------|--------------|---------------|------------|----------------|
| Alkalinity as CaCO3 | -88 | mg/L | 3 | 10 | DNQ | 6/24/2002 | QC 10303311A | DFG-WPCL | L-295-02-ALK |
| Alkalinity as CaCO3 | -88 | mg/L | 3 | 10 | DNQ | 4/29/2002 | QC 10303311A | DFG-WPCL | L-181-02-ALK |
| Alkalinity as CaCO3 | -88 | mg/L | 3 | 10 | DNQ | 6/24/2002 | QC 10303311A | DFG-WPCL | L-288-02-ALK |
| Alkalinity as CaCO3 | -88 | mg/L | 3 | 10 | DNQ | 4/16/2002 | QC 10303311A | DFG-WPCL | L17902-ALK |
| Alkalinity as CaCO3 | -88 | mg/L | 3 | 10 | DNQ | 4/16/2002 | QC 10303311A | DFG-WPCL | L-177-01-ALK |
| Alkalinity as CaCO3 | -88 | mg/L | 3 | 10 | DNQ | 4/16/2002 | QC 10303311A | DFG-WPCL | L-173-01-ALK |
| Alkalinity as CaCO3 | 1 | mg/L | 0.5 | 1 | | 2/1/2002 | DFG-WPCL | DFG-WPCL | L-058-01-ALK |
| Alkalinity as CaCO3 | 2 | mg/L | 0.5 | 1 | | 1/28/2002 | DFG-WPCL | DFG-WPCL | L-052-01-ALK |
| Alkalinity as CaCO3 | 3 | mg/L | 0.5 | 1 | | 1/28/2002 | DFG-WPCL | DFG-WPCL | L-048-01-ALK |
| Ammonia as N | -88 | mg/L | 0.05 | 0.1 | DNQ | 6/21/2002 | EPA 350.3 | DFG-WPCL | L-288-02-NH3 |
| Ammonia as N | -88 | mg/L | 0.05 | 0.1 | DNQ | 4/13/2002 | EPA 350.3 | DFG-WPCL | L17902-NH3 |
| Ammonia as N | -88 | mg/L | 0.05 | 0.1 | DNQ | 4/12/2002 | EPA 350.3 | DFG-WPCL | L-173-01-NH3 |
| Ammonia as N | -88 | mg/L | 0.05 | 0.1 | DNQ | 6/25/2002 | EPA 350.3 | DFG-WPCL | L-295-02-NH3 |
| Ammonia as N | -88 | mg/L | 0.05 | 0.1 | DNQ | 4/19/2002 | EPA 350.3 | DFG-WPCL | L-181-02-NH3 |
| Ammonia as N | 0.08 | mg/L | 0.05 | 0.1 | DNQ | 1/28/2002 | EPA 350.3 | DFG-WPCL | L-048-01-NH3 |
| Ammonia as N | 0.08 | mg/L | 0.05 | 0.1 | DNQ | 2/4/2002 | EPA 350.3 | DFG-WPCL | L-058-01-NH3 |
| Ammonia as N | -88 | mg/L | 0.05 | 0.1 | DNQ | 4/13/2002 | EPA 350.3 | DFG-WPCL | L-177-01-NH3 |
| Ammonia as N | 0.08 | mg/L | 0.05 | 0.1 | DNQ | 1/30/2002 | EPA 350.3 | DFG-WPCL | L-052-01-NH3 |
| Chloride | -88 | mg/L | 0.15 | 0.25 | DNQ | 6/20/2002 | EPA 300.0 | DFG-WPCL | L-288-02-CL |
| Chloride | -88 | mg/L | 0.15 | 0.25 | DNQ | 4/10/2002 | EPA 300.0 | DFG-WPCL | L-173-01-CL |
| Chloride | -88 | mg/L | 0.15 | 0.25 | DNQ | 4/15/2002 | EPA 300.0 | DFG-WPCL | L-177-01-CL |
| Chloride | -88 | mg/L | 0.15 | 0.25 | DNQ | 6/21/2002 | EPA 300.0 | DFG-WPCL | L-295-02-CL |
| Chloride | -88 | mg/L | 0.15 | 0.25 | DNQ | 4/18/2002 | EPA 300.0 | DFG-WPCL | L-181-02-CL |
| Chloride | -88 | mg/L | 0.15 | 0.25 | DNQ | 4/16/2002 | EPA 300.0 | DFG-WPCL | L17902-CL |
| Chlorophyll a | 0.75 | µg/L | 0.5 | 1 | DNQ | 4/16/2002 | SM 10200 H-2b | SFL | R12-041602-Chl |
| Chlorophyll a | 1.25 | µg/L | 0.5 | 2 | DNQ | 4/23/2002 | SM 10200 H-2b | SFL | R12-042302-Chl |
| Chlorophyll a | 1.25 | µg/L | 0.5 | 2 | DNQ | 7/17/2002 | SM 10200 H-2b | SFL | R12-071702-Chl |
| Chlorophyll a | 0.75 | µg/L | 0.5 | 1 | DNQ | 7/10/2002 | SM 10200 H-2b | SFL | R12-071002-Chl |
| Chlorophyll a | 1.25 | µg/L | 0.5 | 2 | DNQ | 4/23/2002 | SM 10200 H-2b | SFL | R12-042302-Chl |
| Chlorophyll a | 1.25 | µg/L | 0.5 | 2 | DNQ | 7/17/2002 | SM 10200 H-2b | SFL | R12-071702-Chl |
| Chlorophyll a | 0.9 | µg/L | 0.5 | 1.3 | DNQ | 4/17/2002 | SM 10200 H-2b | SFL | R2-041702-Chl |
| Chlorophyll a | 1.25 | µg/L | 0.5 | 2 | DNQ | 2/7/2002 | SM 10200 H-2b | SFL | R12-020702-Chl |
| Chlorophyll a | 0.75 | µg/L | 0.5 | 1 | DNQ | 4/16/2002 | SM 10200 H-2b | SFL | R12-041602-Chl |
| Chlorophyll a | 1.25 | µg/L | 0.5 | 2 | DNQ | 4/19/2002 | SM 10200 H-2b | SFL | R12-041902-Chl |
| Chlorophyll a | 1.25 | µg/L | 0.5 | 2 | DNQ | 7/17/2002 | SM 10200 H-2b | SFL | R12-071702-Chl |
| Chlorophyll a | 0.75 | µg/L | 0.5 | 1 | DNQ | 7/10/2002 | SM 10200 H-2b | SFL | R12-071002-Chl |
| Chlorophyll a | 1.25 | µg/L | 0.5 | 2 | DNQ | 7/17/2002 | SM 10200 H-2b | SFL | R12-071702-Chl |
| Copper | 0.004 | mg/Kg | 0.003 | 0.01 | DNQ | 3/1/2002 | EPA 200.8 | MPSL-DFG | 2002Dig7 |
| Copper | 0.005 | mg/Kg | 0.003 | 0.01 | DNQ | 2/20/2003 | EPA 200.8 | MPSL-DFG | 2003Dig06 |
| Copper | 0.006 | mg/Kg | 0.003 | 0.01 | DNQ | 3/1/2002 | EPA 200.8 | MPSL-DFG | 2002Dig7 |
| Copper | 0.006 | mg/Kg | 0.003 | 0.01 | DNQ | 2/20/2003 | EPA 200.8 | MPSL-DFG | 2003Dig06 |
| C1 -Dibenzothiophenes | 5.42 | ng/g | 1.95 | 1.95 | | 4/15/2002 | DFG-WPCL | DFG-WPCL | L-021902-PAH |
| C2 -Dibenzothiophenes | 4.02 | ng/g | 1.95 | 1.95 | | 4/15/2002 | DFG-WPCL | DFG-WPCL | L-021902-PAH |
| Hardness as CaCO3 | -88 | mg/L | 0.5 | 1 | DNQ | 4/11/2002 | SM 2340 C | DFG-WPCL | L-173-01-H |
| Hardness as CaCO3 | -88 | mg/L | 0.5 | 1 | DNQ | 6/20/2002 | SM 2340 C | DFG-WPCL | L-288-02-H |
| Hardness as CaCO3 | -88 | mg/L | 0.5 | 1 | DNQ | 4/15/2002 | SM 2340 C | DFG-WPCL | L17902-H |
| Hardness as CaCO3 | -88 | mg/L | 0.5 | 1 | DNQ | 4/17/2002 | SM 2340 C | DFG-WPCL | L-181-02-H |
| Hardness as CaCO3 | -88 | mg/L | 0.5 | 1 | DNQ | 7/1/2002 | SM 2340 C | DFG-WPCL | L-295-02-H |
| Hardness as CaCO3 | -88 | mg/L | 0.5 | 1 | DNQ | 4/15/2002 | SM 2340 C | DFG-WPCL | L-177-01-H |
| HCH, gamma | 0.0015 | µg/L | 0.001 | 0.002 | DNQ | 5/3/2002 | EPA 8081AM | DFG-WPCL | L-050302-OCH |
| Manganese | 0.005 | mg/Kg | 0.003 | 0.01 | DNQ | 2/20/2003 | EPA 200.8 | MPSL-DFG | 2003Dig06 |
| Manganese | 0.009 | mg/Kg | 0.003 | 0.01 | DNQ | 3/1/2002 | EPA 200.8 | MPSL-DFG | 2002Dig7 |
| Manganese | 0.009 | mg/Kg | 0.003 | 0.01 | DNQ | 3/1/2002 | EPA 200.8 | MPSL-DFG | 2002Dig7 |
| Manganese | 0.006 | mg/Kg | 0.003 | 0.01 | DNQ | 2/20/2003 | EPA 200.8 | MPSL-DFG | 2003Dig06 |
| Naphthalene | 2.47 | ng/g | 1.52 | 1.52 | | 4/15/2002 | DFG-WPCL | DFG-WPCL | L-010902-PAH |
| C1 - Naphthalenes | 2.9 | ng/g | 1.52 | 1.52 | | 4/15/2002 | DFG-WPCL | DFG-WPCL | L-010902-PAH |
| C1 - Naphthalenes | 7.3 | ng/g | 1.95 | 1.95 | | 4/15/2002 | DFG-WPCL | DFG-WPCL | L-021902-PAH |
| C2 - Naphthalenes | 1.81 | ng/g | 1.52 | 1.52 | | 4/15/2002 | DFG-WPCL | DFG-WPCL | L-010902-PAH |
| C3 - Naphthalenes | 3.18 | ng/g | 1.95 | 1.95 | | 4/15/2002 | DFG-WPCL | DFG-WPCL | L-021902-PAH |
| Nickel | 0.007 | mg/Kg | 0.006 | 0.01 | DNQ | 3/1/2002 | EPA 200.8 | MPSL-DFG | 2002Dig7 |
| Nitrate as N | -88 | mg/L | 0.005 | 0.01 | DNQ | 6/21/2002 | QC 10107041B | DFG-WPCL | L-295-02-NO3 |

Appendix C-2: Laboratory blanks detected

| Analyte | Result | Unit | MDL | RL | Detected | AnalysisDate | MethodName | Laboratory | Batch ID |
|--------------------------|--------|-------|-------|------|----------|--------------|---------------|------------|------------------------------|
| Nitrate as N | -88 | mg/L | 0.005 | 0.01 | DNQ | 6/20/2002 | QC 10107041B | DFG-WPCL | L-288-02-NO3 |
| Nitrate as N | -88 | mg/L | 0.005 | 0.01 | DNQ | 4/10/2002 | QC 10107041B | DFG-WPCL | L-173-01-NO3 |
| Nitrate as N | -88 | mg/L | 0.005 | 0.01 | DNQ | 4/18/2002 | QC 10107041B | DFG-WPCL | L-181-02-NO3 |
| Nitrate as N | -88 | mg/L | 0.005 | 0.01 | DNQ | 4/12/2002 | QC 10107041B | DFG-WPCL | L17902-NO3 |
| Nitrate as N | -88 | mg/L | 0.005 | 0.01 | DNQ | 4/12/2002 | QC 10107041B | DFG-WPCL | L-177-01-NO3 |
| Nitrite as N | -88 | mg/L | 0.005 | 0.01 | DNQ | 4/12/2002 | QC 10107041B | DFG-WPCL | L17902-NO2 |
| Nitrite as N | -88 | mg/L | 0.005 | 0.01 | DNQ | 4/17/2002 | QC 10107041B | DFG-WPCL | L-181-02-NO2 |
| Nitrite as N | 0.02 | mg/L | 0.01 | 0.03 | DNQ | 2/1/2002 | FR 8507 | DFG-WPCL | L-058-01-NO2 |
| Nitrite as N | -88 | mg/L | 0.005 | 0.01 | DNQ | 4/12/2002 | QC 10107041B | DFG-WPCL | L-177-01-NO2 |
| Nitrite as N | -88 | mg/L | 0.005 | 0.01 | DNQ | 6/21/2002 | QC 10107041B | DFG-WPCL | L-295-02-NO2 |
| Nitrite as N | -88 | mg/L | 0.005 | 0.01 | DNQ | 4/10/2002 | QC 10107041B | DFG-WPCL | L-173-01-NO2 |
| Nitrite as N | -88 | mg/L | 0.005 | 0.01 | DNQ | 6/19/2002 | QC 10107041B | DFG-WPCL | L-288-02-NO2 |
| Nitrogen, Total Kjeldahl | -88 | mg/L | 0.25 | 0.5 | DNQ | 5/2/2002 | EPA 351.3 | DFG-WPCL | L17902-TKN |
| Nitrogen, Total Kjeldahl | -88 | mg/L | 0.25 | 0.5 | DNQ | 6/27/2002 | EPA 351.3 | DFG-WPCL | L-295-02-TKN |
| Nitrogen, Total Kjeldahl | -88 | mg/L | 0.25 | 0.5 | DNQ | 5/2/2002 | EPA 351.3 | DFG-WPCL | L-181-02-TKN |
| Nitrogen, Total Kjeldahl | -88 | mg/L | 0.25 | 0.5 | DNQ | 6/26/2002 | EPA 351.3 | DFG-WPCL | L-288-02-TKN |
| Nitrogen, Total Kjeldahl | -88 | mg/L | 0.25 | 0.5 | DNQ | 4/19/2002 | EPA 351.3 | DFG-WPCL | L-173-01-TKN |
| Nitrogen, Total Kjeldahl | -88 | mg/L | 0.25 | 0.5 | DNQ | 4/24/2002 | EPA 351.3 | DFG-WPCL | L-177-01-TKN |
| OrthoPhosphate as P | -88 | mg/L | 0.005 | 0.01 | DNQ | 6/21/2002 | QC 10115011M | DFG-WPCL | L-295-02-OPO4 |
| OrthoPhosphate as P | -88 | mg/L | 0.005 | 0.01 | DNQ | 4/12/2002 | QC 10115011M | DFG-WPCL | L17902-OPO4 |
| OrthoPhosphate as P | -88 | mg/L | 0.005 | 0.01 | DNQ | 4/17/2002 | QC 10115011M | DFG-WPCL | L-181-02-OPO4 |
| OrthoPhosphate as P | -88 | mg/L | 0.005 | 0.01 | DNQ | 4/12/2002 | QC 10115011M | DFG-WPCL | L-177-01-OPO4 |
| OrthoPhosphate as P | -88 | mg/L | 0.005 | 0.01 | DNQ | 4/10/2002 | QC 10115011M | DFG-WPCL | L-173-01-OPO4 |
| OrthoPhosphate as P | -88 | mg/L | 0.005 | 0.01 | DNQ | 6/19/2002 | QC 10115011M | DFG-WPCL | L-288-02-OPO4 |
| PCB 028 | 0.165 | ng/g | 0.15 | 0.6 | DNQ | 8/22/2003 | EPA 8082M | DFG-WPCL | L_121_03_BS 241_KR_CONGENERS |
| PCB 031 | 0.192 | ng/g | 0.15 | 0.6 | DNQ | 8/22/2003 | EPA 8082M | DFG-WPCL | L_121_03_BS 241_KR_CONGENERS |
| PCB 044 | 0.182 | ng/g | 0.15 | 0.6 | DNQ | 8/22/2003 | EPA 8082M | DFG-WPCL | L_121_03_BS 241_KR_CONGENERS |
| PCB 052 | 0.292 | ng/g | 0.15 | 0.6 | DNQ | 8/22/2003 | EPA 8082M | DFG-WPCL | L_121_03_BS 241_KR_CONGENERS |
| PCB 066 | 0.158 | ng/g | 0.15 | 0.6 | DNQ | 8/22/2003 | EPA 8082M | DFG-WPCL | L_121_03_BS 241_KR_CONGENERS |
| PCB 095 | 0.236 | ng/g | 0.15 | 0.6 | DNQ | 8/22/2003 | EPA 8082M | DFG-WPCL | L_121_03_BS 241_KR_CONGENERS |
| PCB 101 | 0.265 | ng/g | 0.15 | 0.6 | DNQ | 8/22/2003 | EPA 8082M | DFG-WPCL | L_121_03_BS 241_KR_CONGENERS |
| PCB 110 | 0.348 | ng/g | 0.15 | 0.6 | DNQ | 8/22/2003 | EPA 8082M | DFG-WPCL | L_121_03_BS 241_KR_CONGENERS |
| PCB 110 | 0.261 | ng/g | 0.25 | 0.25 | | 12/11/2001 | DFG-WPCL | DFG-WPCL | L39801_BS179_KR_CONG |
| PCB 118 | 0.269 | ng/g | 0.15 | 0.6 | DNQ | 8/22/2003 | EPA 8082M | DFG-WPCL | L_121_03_BS 241_KR_CONGENERS |
| Pheophytin a | 0.9 | µg/L | 0.5 | 1.3 | DNQ | 4/17/2002 | SM 10200 H-2a | SFL | R2-041702-Ph |
| Pheophytin a | 0.75 | µg/L | 0.5 | 1 | DNQ | 7/10/2002 | SM 10200 H-2a | SFL | R12-071002-Ph |
| Pheophytin a | 1.25 | µg/L | 0.5 | 2 | DNQ | 4/23/2002 | SM 10200 H-2a | SFL | R12-042302-Ph |
| Pheophytin a | 1.25 | µg/L | 0.5 | 2 | DNQ | 4/23/2002 | SM 10200 H-2a | SFL | R12-042302-Ph |
| Pheophytin a | 0.75 | µg/L | 0.5 | 1 | DNQ | 7/10/2002 | SM 10200 H-2a | SFL | R12-071002-Ph |
| Pheophytin a | 1.25 | µg/L | 0.5 | 2 | DNQ | 4/19/2002 | SM 10200 H-2a | SFL | R12-041902-Ph |
| Pheophytin a | 1.25 | µg/L | 0.5 | 2 | DNQ | 7/17/2002 | SM 10200 H-2a | SFL | R12-071702-Ph |
| Pheophytin a | 1.25 | µg/L | 0.5 | 2 | DNQ | 7/17/2002 | SM 10200 H-2a | SFL | R12-071702-Ph |
| Pheophytin a | 1.25 | µg/L | 0.5 | 2 | DNQ | 7/17/2002 | SM 10200 H-2a | SFL | R12-071702-Ph |
| Pheophytin a | 0.75 | µg/L | 0.5 | 1 | DNQ | 4/16/2002 | SM 10200 H-2a | SFL | R12-041602-Ph |
| Pheophytin a | 1.65 | µg/L | 0.5 | 2.8 | DNQ | 2/7/2002 | SM 10200 H-2a | SFL | R12-020702-Ph |
| Pheophytin a | 0.75 | µg/L | 0.5 | 1 | DNQ | 4/16/2002 | SM 10200 H-2a | SFL | R12-041602-Ph |
| Sulfate | -88 | mg/L | 0.37 | 1 | DNQ | 4/18/2002 | EPA 300.0 | DFG-WPCL | L-181-02-SO4 |
| Sulfate | -88 | mg/L | 0.37 | 1 | DNQ | 4/15/2002 | EPA 300.0 | DFG-WPCL | L-177-01-SO4 |
| Sulfate | -88 | mg/L | 0.37 | 1 | DNQ | 6/20/2002 | EPA 300.0 | DFG-WPCL | L-288-02-SO4 |
| Sulfate | -88 | mg/L | 0.37 | 1 | DNQ | 4/10/2002 | EPA 300.0 | DFG-WPCL | L-173-01-SO4 |
| Sulfate | -88 | mg/L | 0.37 | 1 | DNQ | 6/21/2002 | EPA 300.0 | DFG-WPCL | L-295-02-SO4 |
| Sulfate | -88 | mg/L | 0.37 | 1 | DNQ | 4/16/2002 | EPA 300.0 | DFG-WPCL | L17902-SO4 |
| Total Aluminum | 0.98 | mg/Kg | 0.05 | 0.1 | DNQ | 2/20/2003 | EPA 200.8 | MPSL-DFG | 2003Dig06 |
| Total Aluminum | 0.105 | µg/L | 0.1 | 0.3 | DNQ | 4/8/2003 | EPA 1638M | MPSL-DFG | ICP040803 |
| Total Aluminum | 1.28 | mg/Kg | 0.05 | 0.1 | | 3/1/2002 | MPSL-DFG | MPSL-DFG | 2002Dig7 |
| Total Aluminum | 1.0093 | mg/Kg | 0.05 | 0.1 | | 2/20/2003 | MPSL-DFG | MPSL-DFG | 2003Dig06 |
| Total Aluminum | 0.929 | mg/Kg | 0.05 | 0.1 | | 3/1/2002 | MPSL-DFG | MPSL-DFG | 2002Dig7 |
| Total Boron | -88 | mg/L | 0.02 | 0.1 | NR | 7/1/2002 | EPA 200.7 | Babcock | R2-070102-B |
| Total Cadmium | 0.002 | µg/L | 0.002 | 0.05 | DNQ | 4/8/2003 | EPA 1638M | MPSL-DFG | ICP040803 |
| Total Copper | 0.023 | µg/L | 0.01 | 0.01 | | 5/30/2002 | MPSL-DFG | MPSL-DFG | R2-053002-ICP |
| Total Copper | 0.063 | µg/L | 0.003 | 0.01 | | 4/8/2003 | MPSL-DFG | MPSL-DFG | ICP040803 |

Appendix C-2: Laboratory blanks detected

| | | | | | | | | | |
|------------------------|---------------|-------------|------------|-----------|-----------------|---------------------|-------------------|-------------------|-----------------|
| Total Dissolved Solids | -88 | mg/L | 10 | 10 | DNQ | 6/21/2002 | SM 2540 C | DFG-WPCL | L-295-02-TDS |
| Total Dissolved Solids | -88 | mg/L | 10 | 10 | DNQ | 4/15/2002 | SM 2540 C | DFG-WPCL | L-177-01-TDS |
| Analyte | Result | Unit | MDL | RL | Detected | AnalysisDate | MethodName | Laboratory | Batch ID |
| Total Dissolved Solids | -88 | mg/L | 10 | 10 | DNQ | 4/11/2002 | SM 2540 C | DFG-WPCL | L-173-01-TDS |
| Total Dissolved Solids | -88 | mg/L | 10 | 10 | DNQ | 6/21/2002 | SM 2540 C | DFG-WPCL | L-288-02-TDS |
| Total Dissolved Solids | -88 | mg/L | 10 | 10 | DNQ | 4/15/2002 | SM 2540 C | DFG-WPCL | L17902-TDS |
| Total Dissolved Solids | -88 | mg/L | 10 | 10 | DNQ | 4/18/2002 | SM 2540 C | DFG-WPCL | L-181-02-TDS |
| Total Lead | 0.006 | µg/L | 0.002 | 0.05 | DNQ | 4/8/2003 | EPA 1638M | MPSL-DFG | ICP040803 |
| Total Organic Carbon | 47.8 | mg/L | 1 | 1 | | 6/24/2002 | AMS | AMS | AMSTOC62402-2 |
| Total Phosphorus as P | -88 | mg/L | 0.03 | 0.05 | DNQ | 6/21/2002 | EPA 365.3 | DFG-WPCL | L-295-02-TPHOS |
| Total Phosphorus as P | -88 | mg/L | 0.03 | 0.05 | DNQ | 6/19/2002 | EPA 365.3 | DFG-WPCL | L-288-02-TPHOS |
| Total Phosphorus as P | -88 | mg/L | 0.03 | 0.05 | DNQ | 4/12/2002 | EPA 365.3 | DFG-WPCL | L-177-01-TPHOS |
| Total Phosphorus as P | -88 | mg/L | 0.03 | 0.05 | DNQ | 4/10/2002 | EPA 365.3 | DFG-WPCL | L-173-01-TPHOS |
| Total Phosphorus as P | -88 | mg/L | 0.03 | 0.05 | DNQ | 4/12/2002 | EPA 365.3 | DFG-WPCL | L17902-TPHOS |
| Total Phosphorus as P | -88 | mg/L | 0.03 | 0.05 | DNQ | 4/17/2002 | EPA 365.3 | DFG-WPCL | L-181-02-TPHOS |
| Total Silver | 0.018 | µg/L | 0.008 | 0.1 | DNQ | 4/8/2003 | EPA 1638M | MPSL-DFG | ICP040803 |
| Total Zinc | 0.034 | µg/L | 0.02 | 0.06 | DNQ | 4/8/2003 | EPA 1638M | MPSL-DFG | ICP040803 |

Appendix C-3: Batches containing surrogate recoveries outside control limits

| Surrogate | Site | Batch ID | %R | Laboratory |
|-------------------------------------|-----------|--------------|------|------------|
| Benz(a)anthracene-d12(Surrogate) | 202BUT010 | L-012503-PAH | 45.1 | DFG-WPCL |
| Benz(a)anthracene-d12(Surrogate) | 202SGR010 | L-012503-PAH | 43.5 | DFG-WPCL |
| Benz(e)pyrene-d12(Surrogate) | LCS | L-012503-PAH | 47 | DFG-WPCL |
| Benz(e)pyrene-d12(Surrogate) | 202SGR080 | L-012503-PAH | 37.9 | DFG-WPCL |
| Benz(e)pyrene-d12(Surrogate) | 202PES070 | L-012503-PAH | 22 | DFG-WPCL |
| Benz(e)pyrene-d12(Surrogate) | 205STE060 | L-012503-PAH | 37.7 | DFG-WPCL |
| Benz(e)pyrene-d12(Surrogate) | LabBlank | L-060203-PAH | 40.4 | DFG-WPCL |
| Benz(e)pyrene-d12(Surrogate) | 205PER070 | L-012503-PAH | 41.2 | DFG-WPCL |
| Benz(e)pyrene-d12(Surrogate) | 202BUT010 | L-012503-PAH | 22.1 | DFG-WPCL |
| Benz(e)pyrene-d12(Surrogate) | 202SGR010 | L-012503-PAH | 19.8 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | LabBlank | L-060203-PAH | 10.8 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | 202SGR080 | L-012503-PAH | 10.5 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | 206PET310 | L-012503-PAH | 46.1 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | 202BUT010 | L-012503-PAH | 4.89 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | LCS | L-060203-PAH | 18 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | 205STE020 | L-012503-PAH | 39.2 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | 202PES070 | L-012503-PAH | 2.88 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | 205STE060 | L-012503-PAH | 9.56 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | 202SGR010 | L-012503-PAH | 6.37 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | 205PER070 | L-012503-PAH | 13.8 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | 206PET010 | L-012503-PAH | 41 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | LabBlank | L-012503-PAH | 27.4 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | LCS | L-012503-PAH | 8.12 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | CRM | L-010902-PAH | 22.1 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | 201LAG120 | L-060203-PAH | 35.9 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | LabBlank | L-021902-PAH | 45.2 | DFG-WPCL |
| Benzo(g,h,i)perylene-d12(Surrogate) | LCS | L-010902-PAH | 22.2 | DFG-WPCL |
| Naphthalene-d8(Surrogate) | LabBlank | L-021902-PAH | 46 | DFG-WPCL |
| Perylene-d12(Surrogate) | 205PER070 | L-012503-PAH | 31.3 | DFG-WPCL |
| Perylene-d12(Surrogate) | 202BUT010 | L-012503-PAH | 13.9 | DFG-WPCL |
| Perylene-d12(Surrogate) | LCS | L-012503-PAH | 33.2 | DFG-WPCL |
| Perylene-d12(Surrogate) | LabBlank | L-060203-PAH | 33.8 | DFG-WPCL |
| Perylene-d12(Surrogate) | 202PES070 | L-012503-PAH | 17 | DFG-WPCL |
| Perylene-d12(Surrogate) | LabBlank | L-012503-PAH | 46.2 | DFG-WPCL |
| Perylene-d12(Surrogate) | 202PES070 | L-070602-PAH | 49.6 | DFG-WPCL |
| Perylene-d12(Surrogate) | 202SGR080 | L-012503-PAH | 25.9 | DFG-WPCL |
| Perylene-d12(Surrogate) | 202SGR010 | L-012503-PAH | 15.3 | DFG-WPCL |
| Perylene-d12(Surrogate) | 205STE060 | L-012503-PAH | 29.8 | DFG-WPCL |

Appendix C-4: Data batches missing MS/MSDs

| Batch ID | BatchQualCode | BatchValCode | Notes | Lab |
|----------------------|---------------|--------------|---|----------|
| 23ELGC | A,VMD | VAC,VI | QAO: No MS/MSD; LCM %R for chlorpyrifos above QAPP QC limit of 80-120 | UCD-GC |
| 3ELGC | A,VMD | VAC,VI | QAO: no MS/MSD; LCM %R for chlorpyrifos above QAPP QC limit of 80-120 | UCD-GC |
| L-041802-PAH | A,VMD | VAC,VI | QAO: no blank,MS/MSD,LCS for extraction date 4/15/2002;MS/MSD RPD for naphthalene above QC limit | DFG-WPCL |
| L-058-01-aNO3 | A,VMD | VAC,VI | QAO: no CRM or MS/MSD | DFG-WPCL |
| L-093001-OCH | A,VMD | VAC,VI | QAO: no LCS or MS/MSD | DFG-WPCL |
| L-093001-OPP | A,VMD | VAC,VI | QAO: no LCS or MS/MSD | DFG-WPCL |
| L-093001-PAH | A,VMD | VAC,VI | QAO: no LCS or MS/MSD, a DUP was performed | DFG-WPCL |
| L-093001-PCB | A,VMD | VAC,VI | QAO: no LCS or MS/MSD | DFG-WPCL |
| L-101001-OCH | A,MD | VAC,VI | QAO: no blank for 10/7/2001, MS/MSD RPD for endosulfan I above QC limit of 25% | DFG-WPCL |
| L39801_BS179_KR_CONG | A,VMD | VAC,VI | QAO: no QC for 11/20/01; LCS %R outside QC limit for PCB-29, PCB-56, PCB-60; MS/MSD %R PCB-029 and PCB-056 | DFG-WPCL |
| L39801BS179_KRPEST | A,VMD | VAC,VI | QAO: no QC for 11/20/01; LCS %R outside QC limit for several compounds; MS/MSD %R and RPD outside QC limit for several compounds | DFG-WPCL |
| R2-071002-ICP | A,VMD | VAC,VI | QAO: no blank or CRM; MS/MSD %Rs outside QC limit of 75-125 | MPSL-DFG |
| 39ELGC | A,VMD | VAC,VI | QAO: no MS/MSD or LCM | UCD-GC |
| L-070602-OPP | A,VMD | VAC,VI | QAO: no LCS for 6/22/02, no blk, LCS, MS/MSD for 6/26/02, no blk, MS/MSD for 6/30/02, MS/MSD RPDs above QC limit of 25% | DFG-WPCL |
| L-071902-OCH | A,VMD | VAC,VI | QAO: associated blk, MS/MSD and LCS reported in batch L-072002-OCH | DFG-WPCL |
| L-071902-PCB | A,VMD | VAC,VI | QAO: no blank or LCS for 6/26 or 6/30, no MS/MSD for 6/22 or 6/30 | DFG-WPCL |
| 012403-TKN-1 | A,VMD | VAC,VI | QAO: no MS/MSD | DFG-WPCL |
| 021103-TPHOS-1 | A,VMD | VAC,VI | QAO: no MS/MSD | DFG-WPCL |
| 23ELGC | A,VMD | VAC,VI | QAO: No MS/MSD; LCM %R for chlorpyrifos above QAPP QC limit of 80-120 | UCD-GC |
| 39ELGC | A,VMD | VAC,VI | QAO: no MS/MSD or LCM | UCD-GC |
| L-012503-OCH | A,VMD | VAC,VI | QAO: no QC (blank, MS/MSD, LCS) for extraction date 1/31/03; MS/MSD RPD for p'p'-DDD was above QC limit | DFG-WPCL |
| L-012503-OP | A,VMD | VAC,VI | QAO: no QC (blank, MS/MSD, LCS) for extraction date 1/31/03; Both MS/MSD RPDs above 25% for coumaphos, demeton-s, tetrachlorvinphos | DFG-WPCL |
| L-012503-PAH | A,MD | VAC,VI | QAO: no LCS or MS/MSD for 1/25/03, no QC for 1/31/03, no blank for 1/29/03 | DFG-WPCL |
| L-012503-PCB | A,VMD | VAC,VI | QAO: no QC(blank, MS/MSD, LCS) for extraction date 1/31/03 | DFG-WPCL |
| L-012503-TRIAZ | A,VMD | VAC,VI | QAO: no QC(blank, MS/MSD, LCS) for extraction date 1/31/03, MS/MSD RPD out for Ametryn | DFG-WPCL |
| L-041802-PAH | A,VMD | VAC,VI | QAO: no blank,MS/MSD,LCS for extraction date 4/15/2002;MS/MSD RPD for naphthalene above QC limit | DFG-WPCL |
| L-050302-PAH | A,VMD | VAC,VI | QAO: no blank, MS/MSD, LCS for 4/22/02, no blank for 4/23/02 | DFG-WPCL |
| L-070602-OPP | A,VMD | VAC,VI | QAO: no LCS for 6/22/02, no blk, LCS, MS/MSD for 6/26/02, no blk, MS/MSD for 6/30/02, MS/MSD RPDs above QC limit of 25% | DFG-WPCL |
| L-071902-OCH | A,VMD | VAC,VI | QAO: associated blk, MS/MSD and LCS reported in batch L-072002-OCH | DFG-WPCL |
| L-071902-PCB | A,VMD | VAC,VI | QAO: no blank or LCS for 6/26 or 6/30, no MS/MSD for 6/22 or 6/30 | DFG-WPCL |
| R2-070102-B | A,VMD | VAC,VI | QAO: no CRM, MS/MSD on non-SWAMP sample, lab would not provide confirmation of prep and digest methods and dates | Babcock |

Appendix C-5: Data batches containing MS/MSD % Recovery and RDP values outside of acceptance criteria

| Analyte | StationCode | Sample Date | MatrixName | LabBatch | MS %R | MSD %R | RPD | Laboratory |
|-----------------------------|-------------|-------------|-------------|-------------------------------|---------|----------|--------|------------|
| (o,p') DDT | 204ALP010 | 9/18/2001 | sediment | L39801BS178_KRPEST | 50.7256 | 65.63707 | 25.629 | DFG-WPCL |
| (p,p') DCBP | 206WIL020 | 6/17/2002 | sediment | L_121_03_BS_241_KR_PESTICIDES | 0 | 0 | 0 | DFG-WPCL |
| (p,p') DDD | 202SGR080 | 1/22/2003 | samplewater | L-012503-OCH | 75 | 108.5 | 36.512 | DFG-WPCL |
| 2-Methylnaphthalene | 201LAG270 | 10/1/2001 | sediment | L-010902-PAH | 65.1163 | 45.73643 | 34.965 | DFG-WPCL |
| alpha Chlordene | 204ALP010 | 9/18/2001 | sediment | L39801BS178_KRPEST | 58.7071 | 86.61519 | 38.409 | DFG-WPCL |
| Ametryn | 204SMA020 | 1/23/2003 | samplewater | L-012503-Triaz | 73.00 | 102.00 | 33.14 | DFG-WPCL |
| Anthracene | 206WIL020 | 6/17/2002 | sediment | L-060203-PAH | 164.82 | 123.0986 | 28.981 | DFG-WPCL |
| Benz(a)anthracene | 206WIL020 | 6/17/2002 | sediment | L-060203-PAH | 0 | 0 | 0 | DFG-WPCL |
| Benzo(a)pyrene | 206WIL020 | 6/17/2002 | sediment | L-060203-PAH | 164.74 | 93.78531 | 54.892 | DFG-WPCL |
| Benzo(b)fluoranthene | 206WIL020 | 6/17/2002 | sediment | L-060203-PAH | 0 | 0 | 0 | DFG-WPCL |
| Benzo(k)fluoranthene | 204ALP010 | 9/18/2001 | sediment | L-021902-PAH | 50 | 35.19553 | 34.754 | DFG-WPCL |
| Benzo(k)fluoranthene | 206WIL020 | 6/17/2002 | sediment | L-060203-PAH | 122.599 | 88.20225 | 32.634 | DFG-WPCL |
| Biphenyl | 201LAG270 | 10/1/2001 | sediment | L-010902-PAH | 68.9922 | 44.18605 | 43.836 | DFG-WPCL |
| C1- Naphthalenes | 204ALP010 | 9/18/2001 | sediment | L-021902-PAH | 206.881 | 197.1429 | 4.8204 | DFG-WPCL |
| C1- Naphthalenes | 206WIL020 | 6/17/2002 | sediment | L-060203-PAH | 193.296 | 205.6497 | 6.1931 | DFG-WPCL |
| C1- Phenanthrene/Anthracene | 201LAG270 | 10/1/2001 | sediment | L-010902-PAH | 51.5385 | 30.23256 | 52.111 | DFG-WPCL |
| C2-Naphthalenes | 201LAG270 | 10/1/2001 | sediment | L-010902-PAH | 83.7209 | 50.3876 | 49.711 | DFG-WPCL |
| Chlorpyrifos | 204ALP010 | 9/18/2001 | samplewater | 1ELGC | 148 | NA | NA | UCD-GC |
| Chlorpyrifos | 204SLE230 | 9/19/2001 | samplewater | 2ELGC | 140 | NA | NA | UCD-GC |
| Chlorpyrifos | 201LAG270 | 10/1/2001 | sediment | L39801BS179_KRPEST | 49.4832 | 21.28744 | 79.682 | DFG-WPCL |
| Chlorpyrifos | 201LAG270 | 10/1/2001 | sediment | L39801BS179_KRPEST | 49.4832 | 21.28744 | 79.682 | DFG-WPCL |
| Chrysene | 206WIL020 | 6/17/2002 | sediment | L-060203-PAH | 116.571 | 48.31461 | 82.793 | DFG-WPCL |
| Coumaphos | 202PES070 | 1/22/2003 | samplewater | L-012503-OP | 75.5 | 103 | 30.812 | DFG-WPCL |
| delta HCH | 204ALP010 | 9/18/2001 | sediment | L39801BS178_KRPEST | 0 | 0 | 0 | DFG-WPCL |
| delta HCH | 201LAG270 | 10/1/2001 | sediment | L39801BS179_KRPEST | 0 | 0 | 0 | DFG-WPCL |
| Demeton-s | 202PES070 | 1/22/2003 | samplewater | L-012503-OP | 69 | 98 | 34.731 | DFG-WPCL |
| Diazinon | 205STE060 | 1/23/2003 | samplewater | 63ELGC | 153 | NA | NA | UCD-GC |
| Dichlorvos | 207SUI010 | 4/18/2002 | samplewater | L-041802-OPP | 101 | 72 | 34 | DFG-WPCL |
| Dichlorvos | 202BUT010 | 4/16/2002 | samplewater | L-050302-OPP | 74.5 | 114 | 41.91 | DFG-WPCL |
| Dioxathion | 202SGR080 | 6/18/2002 | samplewater | L-070602-OPP | 110 | 67 | 48 | DFG-WPCL |
| Dissolved Cadmium | 207KIR115 | 1/21/2003 | samplewater | ICP040803 | 88.00 | 72.00 | 6.27 | MPSL-DFG |
| Endosulfan I | 207SUI010 | 10/2/2001 | samplewater | L-101001-OCH | 79.3 | 111 | 33.316 | DFG-WPCL |
| Endosulfan sulfate | 204ALP010 | 9/18/2001 | sediment | L39801BS178_KRPEST | 43.9974 | 45.17375 | 2.6385 | DFG-WPCL |
| Endosulfan sulfate | 201LAG270 | 10/1/2001 | sediment | L39801BS179_KRPEST | 35.3609 | 33.33401 | 5.9011 | DFG-WPCL |
| Fenclorophos | 202SGR080 | 6/18/2002 | samplewater | L-070602-OPP | 119 | 70.5 | 51.187 | DFG-WPCL |
| Fensulfothion | 202PES070 | 1/22/2003 | samplewater | L-012503-OP | 75 | 99.5 | 28.08 | DFG-WPCL |
| Fensulfothion | 202SGR080 | 6/18/2002 | samplewater | L-070602-OPP | 107 | 82.7 | 25.619 | DFG-WPCL |
| Fluoranthene | 206WIL020 | 6/17/2002 | sediment | L-060203-PAH | 115.06 | 70.62147 | 47.866 | DFG-WPCL |
| gamma Chlordene | 202PES050 | 4/16/2002 | samplewater | L-050302-OCH | 120 | 72.2 | 49.74 | DFG-WPCL |
| gamma HCH | 202PES050 | 4/16/2002 | samplewater | L-050302-OCH | 90 | 56 | 46 | DFG-WPCL |
| Heptachlor | 204ALP010 | 9/18/2001 | sediment | L39801BS178_KRPEST | 0 | 12.99871 | 200 | DFG-WPCL |
| Heptachlor | 201LAG270 | 10/1/2001 | sediment | L39801BS179_KRPEST | 17.3256 | 0 | 200 | DFG-WPCL |
| Heptachlor epoxide | 204ALP010 | 9/18/2001 | sediment | L39801BS178_KRPEST | 131.32 | 151.3889 | 14.198 | DFG-WPCL |
| Methyl Chlorpyrifos | 202SGR080 | 6/18/2002 | samplewater | L-070602-OPP | 112 | 69 | 48 | DFG-WPCL |
| Mevinphos | 202PES070 | 1/22/2003 | samplewater | L-012503-OP | 84.5 | 61.5 | 31.507 | DFG-WPCL |
| Naled(Dibrom) | 202SGR080 | 6/18/2002 | samplewater | L-070602-OPP | 90 | 120 | 28 | DFG-WPCL |
| Naphthalene | 201LAG040 | 4/18/2002 | samplewater | L-041802-PAH | 87.3 | 114 | 26.528 | DFG-WPCL |
| Nitrite as N | 204SMA020 | 1/23/2003 | samplewater | 012403-NO2 | 103 | 76 | 30 | DFG-WPCL |
| OrthoPhosphate as P | 204SMA020 | 1/23/2003 | samplewater | 012403-OPO4 | 124 | 152 | 21 | DFG-WPCL |
| Oxadiazon | 206WIL020 | 6/17/2002 | sediment | L_121_03_BS_241_KR_PESTICIDES | 271.63 | 169.6713 | 46.208 | DFG-WPCL |
| Oxadiazon | 202PES050 | 4/16/2002 | samplewater | L-050302-OCH | 119 | 82.7 | 35.994 | DFG-WPCL |
| PCB 029 | 204ALP010 | 9/18/2001 | sediment | L39801_BS178_KR_CONG | 36.1 | 48.3 | 28.91 | DFG-WPCL |
| PCB 029 | 201LAG270 | 10/1/2001 | sediment | L39801_BS179_KR_CONG | 61.2 | 38.1 | 46.526 | DFG-WPCL |
| PCB 056 | 204ALP010 | 9/18/2001 | sediment | L39801_BS178_KR_CONG | 42.7 | 55.5 | 26.069 | DFG-WPCL |
| PCB 056 | 201LAG270 | 10/1/2001 | sediment | L39801_BS179_KR_CONG | 57.8 | 43.1 | 29.138 | DFG-WPCL |
| PCB 095 | 202SGR010 | 4/16/2002 | samplewater | L-050302-PCB | 83.5 | 109 | 26.494 | DFG-WPCL |
| PCB 101 | 202SGR010 | 4/16/2002 | samplewater | L-050302-PCB | 83.3 | 109 | 26.729 | DFG-WPCL |
| PCB 110 | 206WIL020 | 6/17/2002 | sediment | L_121_03_BS_241_KR_CONGENERS | 75.3 | 104.5 | 32.481 | DFG-WPCL |
| PCB 118 | 205PER070 | 4/11/2002 | samplewater | L-042702-PCB | 88 | 118 | 29.126 | DFG-WPCL |
| PCB 158 | 202SGR010 | 4/16/2002 | samplewater | L-050302-PCB | 71.4 | 93.3 | 26.594 | DFG-WPCL |
| Perylene | 206WIL020 | 6/17/2002 | sediment | L-060203-PAH | 31.4607 | 44.63277 | 34.621 | DFG-WPCL |
| Phenanthrene | 201LAG270 | 10/1/2001 | sediment | L-010902-PAH | 48.4615 | 22.48062 | 73.245 | DFG-WPCL |
| Phenanthrene | 206WIL020 | 6/17/2002 | sediment | L-060203-PAH | 170.76 | 205.0847 | 18.265 | DFG-WPCL |
| Phorate | 202SGR080 | 6/18/2002 | samplewater | L-070602-OPP | 65.3 | 106 | 47.519 | DFG-WPCL |
| Phosmet | 202PES070 | 1/22/2003 | samplewater | L-012503-OP | 66 | 87 | 27.451 | DFG-WPCL |
| Pyrene | 206WIL020 | 6/17/2002 | sediment | L-060203-PAH | 97.619 | 57.38636 | 51.911 | DFG-WPCL |
| Terbufos | 202SGR080 | 6/18/2002 | samplewater | L-070602-OPP | 57.1 | 100 | 54.615 | DFG-WPCL |
| Tetrachlorvinphos | 202PES070 | 1/22/2003 | samplewater | L-012503-OP | 79.5 | 113 | 34.805 | DFG-WPCL |

Appendix C-5: Data batches containing MS/MSD % Recovery and RDP values outside of acceptance criteria

| | | | | | | | | |
|-----------------|-----------|-----------|-------------|---------------|--------|--------|--------|----------|
| Total Aluminum | 201LAG130 | 4/8/2002 | samplewater | R2-071002-ICP | 21.00 | 23.00 | 9.09 | MPSL-DFG |
| Total Arsenic | 201LAG130 | 4/8/2002 | samplewater | R2-071002-ICP | 137.00 | 144.00 | 4.98 | MPSL-DFG |
| Total Boron | 207SUI020 | 6/17/2002 | samplewater | R2-071002-B | 73.00 | NA | NA | SFL |
| Total Cadmium | 201LAG130 | 4/8/2002 | samplewater | R2-071002-ICP | 100.00 | 150.00 | 40.00 | MPSL-DFG |
| Total Chromium | 201LAG130 | 4/8/2002 | samplewater | R2-071002-ICP | 1.00 | 0.00 | 200.00 | MPSL-DFG |
| Total Chromium | 204ALP010 | 9/18/2001 | samplewater | R2-101001-ICP | 37.79 | NA | NA | MPSL-DFG |
| Total Copper | 204ALP010 | 9/18/2001 | samplewater | R2-101001-ICP | 51.11 | NA | NA | MPSL-DFG |
| Total Manganese | 201LAG130 | 4/8/2002 | samplewater | R2-071002-ICP | 1.60 | 0.40 | 120.00 | MPSL-DFG |
| Total Nickel | 201LAG130 | 4/8/2002 | samplewater | R2-071002-ICP | 10.40 | 12.00 | 14.29 | MPSL-DFG |
| Total Selenium | 201LAG130 | 4/8/2002 | samplewater | R2-071002-ICP | 132.00 | 132.00 | 0.00 | MPSL-DFG |
| Total Silver | 201LAG130 | 4/8/2002 | samplewater | R2-071002-ICP | 150.00 | 150.00 | 0.00 | MPSL-DFG |

Appendix C-6: Data batches missing CRM, SRM, LCS or LCM

| Batch ID | BatchQualCode | BatchValCode | Notes | Lab |
|----------------------|---------------|--------------|---|----------|
| AMS102901-1 | A,VMD | VAC,VI | QAO: no DUP or SRM | AMS |
| AMS10401-1 | A,VMD | VAC,VI | QAO: no blk, DUP, CRM | AMS |
| AMS92901-1 | A,VMD | VAC,VI | QAO: no SRM | AMS |
| AMS62702-1 | A,VMD | VAC,VI | QAO: no SRM | AMS |
| 012703-TDS | A,VMD | VAC,VI | QAO: no blank or CRM | DFG-WPCL |
| 012903-TDS | A,VMD | VAC,VI | QAO: no blank or CRM | DFG-WPCL |
| AMS62702-1 | A,VMD | VAC,VI | QAO: no SRM | AMS |
| 23ELGC | A,VMD | VAC,VI | QAO: No MS/MSD; LCM %R for chlorpyrifos above QAPP QC limit of 80-120 | UCD-GC |
| 3ELGC | A,VMD | VAC,VI | QAO: no MS/MSD; LCM %R for chlorpyrifos above QAPP QC limit of 80-120 | UCD-GC |
| L-041802-OPP | A,VMD | VAC,VI | QAO: no blk for 4/15/2002 and no LCS for 4/12/2002 | DFG-WPCL |
| L-041802-PAH | A,VMD | VAC,VI | QAO: no blank,MS/MSD,LCS for extraction date 4/15/2002;MS/MSD RPD for naphthalene above QC limit | DFG-WPCL |
| L-042702-OCH | A,VMD | VAC,VI | QAO: no CRM, no Blk for 4/15/02, no LCS for 4/12/02 | DFG-WPCL |
| L-042702-PCB | A,VMD | VAC,VI | QAO: no CRM, no Blk for 4/15/02, no LCS for 4/12/02 | DFG-WPCL |
| L-093001-OCH | A,VMD | VAC,VI | QAO: no LCS or MS/MSD | DFG-WPCL |
| L-093001-OPP | A,VMD | VAC,VI | QAO: no LCS or MS/MSD | DFG-WPCL |
| L-093001-PAH | A,VMD | VAC,VI | QAO: no LCS or MS/MSD, a DUP was performed | DFG-WPCL |
| L-093001-PCB | A,VMD | VAC,VI | QAO: no LCS or MS/MSD | DFG-WPCL |
| L39801_BS179_KR_CONG | A,VMD | VAC,VI | QAO: no QC for 11/20/01; LCS %R outside QC limit for PCB-29, PCB-56, PCB-60; MS/MSD %R PCB-029 and PCB-056 | DFG-WPCL |
| L39801BS179_KRPEST | A,VMD | VAC,VI | QAO: no QC for 11/20/01; LCS %R outside QC limit for several compounds; CRM %R outside QC limit for trans chlordane, alpha HCH, cis Nonachlor | DFG-WPCL |
| 39ELGC | A,VMD | VAC,VI | QAO: no MS/MSD or LCM | UCD-GC |
| L-070602-OPP | A,VMD | VAC,VI | QAO: no LCS for 6/22/02, no blk, LCS, MS/MSD for 6/26/02, no blk, MS/MSD for 6/30/02, MS/MSD RPDs above QC limit of 25% | DFG-WPCL |
| L-071902-OCH | A,VMD | VAC,VI | QAO: associated blk, MS/MSD and LCS reported in batch L-072002-OCH | DFG-WPCL |
| L-071902-PCB | A,VMD | VAC,VI | QAO: no blank or LCS for 6/26 or 6/30, no MS/MSD for 6/22 or 6/30 | DFG-WPCL |
| 23ELGC | A,VMD | VAC,VI | QAO: No MS/MSD; LCM %R for chlorpyrifos above QAPP QC limit of 80-120 | UCD-GC |
| 25ELGC | A,VMD | VAC,VI | QAO: no MS, LCM %R for diazinon above QC limit of 80-120 | UCD-GC |
| 39ELGC | A,VMD | VAC,VI | QAO: no MS/MSD or LCM | UCD-GC |
| 66ELGC | A,VMD | VAC,VI | QAO: no MS, LCM %R outside QC limit of 80-120, 0%R for diazinon, 123% for chlorpyrifos | UCD-GC |
| L-012503-OCH | A,VMD | VAC,VI | QAO: no QC (blank, MS/MSD, LCS) for extraction date 1/31/03; MS/MSD RPD for p,p'-DDD was above QC limit | DFG-WPCL |
| L-012503-OP | A,VMD | VAC,VI | QAO: no QC (blank, MS/MSD, LCS) for extraction date 1/31/03; Both MS/MSD RPDs above 25% for coumaphos, demeton-s, tetrachlorvinphos | DFG-WPCL |
| L-012503-PAH | A,MD | VAC,VI | QAO: no LCS or MS/MSD for 1/25/03, no QC for 1/31/03, no blank for 1/29/03 | DFG-WPCL |
| L-012503-PCB | A,VMD | VAC,VI | QAO: no QC (blank, MS/MSD, LCS) for extraction date 1/31/03 | DFG-WPCL |
| L-012503-TRIAZ | A,VMD | VAC,VI | QAO: no QC (blank, MS/MSD, LCS) for extraction date 1/31/03, MS/MSD RPD out for Ametryn | DFG-WPCL |
| L-041802-OPP | A,VMD | VAC,VI | QAO: no blk for 4/15/2002 and no LCS for 4/12/2002 | DFG-WPCL |
| L-041802-PAH | A,VMD | VAC,VI | QAO: no blank,MS/MSD,LCS for extraction date 4/15/2002;MS/MSD RPD for naphthalene above QC limit | DFG-WPCL |
| L-042702-OCH | A,VMD | VAC,VI | QAO: no CRM, no Blk for 4/15/02, no LCS for 4/12/02 | DFG-WPCL |
| L-042702-PCB | A,VMD | VAC,VI | QAO: no CRM, no Blk for 4/15/02, no LCS for 4/12/02 | DFG-WPCL |
| L-050302-PAH | A,VMD | VAC,VI | QAO: no blank, MS/MSD, LCS for 4/22/02, no blank for 4/23/02 | DFG-WPCL |
| L-070602-OPP | A,VMD | VAC,VI | QAO: no LCS for 6/22/02, no blk, LCS, MS/MSD for 6/26/02, no blk, MS/MSD for 6/30/02, MS/MSD RPDs above QC limit of 25% | DFG-WPCL |
| L-071902-OCH | A,VMD | VAC,VI | QAO: associated blk, MS/MSD and LCS reported in batch L-072002-OCH | DFG-WPCL |
| L-071902-PCB | A,VMD | VAC,VI | QAO: no blank or LCS for 6/26 or 6/30, no MS/MSD for 6/22 or 6/30 | DFG-WPCL |

Appendix C-7: Data batches containing LCS, LCM and CRM % Recovery outliers

| Analyte | Sample Type | Batch ID | %R | Laboratory |
|-------------------------|-------------|-------------------------------|-----|------------|
| Benzo(g,h,i)perylene | LCS | L-012503-PAH | 47 | DFG-WPCL |
| Benzo(g,h,i)perylene | CRM | L-021902-PAH | 176 | DFG-WPCL |
| Benzo(k)fluoranthene | LCS | L-021902-PAH | 28 | DFG-WPCL |
| Benzo(k)fluoranthene | LCS | L-012503-PAH | 48 | DFG-WPCL |
| trans-Chlordane | CRM | L_121_03_BS 241_KR_PESTICIDES | 168 | DFG-WPCL |
| trans-Chlordane | CRM | L39801BS179_KRPEST | 306 | DFG-WPCL |
| trans-Chlordane | CRM | L39801BS178_KRPEST | 401 | DFG-WPCL |
| Chlorpyrifos | LCM | 3ELGC | 136 | UCD-GC |
| Chlorpyrifos | LCS | L39801BS179_KRPEST | 26 | DFG-WPCL |
| Chlorpyrifos | LCM | 66ELGC | 123 | UCD-GC |
| Chlorpyrifos | LCM | 23ELGC | 125 | UCD-GC |
| Chlorpyrifos | LCM | 1ELGC | 134 | UCD-GC |
| Chlorpyrifos | LCM | 2ELGC | 137 | UCD-GC |
| DDT(p,p') | CRM | L39801BS178_KRPEST | 194 | DFG-WPCL |
| Diazinon | LCM | 1ELGC | 130 | UCD-GC |
| Diazinon | LCM | 2ELGC | 128 | UCD-GC |
| Diazinon | LCM | 25ELGC | 131 | UCD-GC |
| Diazinon | LCM | 66ELGC | -12 | UCD-GC |
| Endosulfan sulfate | LCS | L39801BS178_KRPEST | 38 | DFG-WPCL |
| Endosulfan sulfate | LCS | L39801BS179_KRPEST | 33 | DFG-WPCL |
| HCH, alpha | CRM | L39801BS179_KRPEST | -48 | DFG-WPCL |
| HCH, alpha | CRM | L_121_03_BS 241_KR_PESTICIDES | -47 | DFG-WPCL |
| HCH, alpha | CRM | L39801BS178_KRPEST | -47 | DFG-WPCL |
| HCH, delta | LCS | L39801BS178_KRPEST | -4 | DFG-WPCL |
| HCH, delta | LCS | L39801BS179_KRPEST | -4 | DFG-WPCL |
| Heptachlor | LCS | L39801BS179_KRPEST | -5 | DFG-WPCL |
| Heptachlor | LCS | L39801BS178_KRPEST | 18 | DFG-WPCL |
| Indeno(1,2,3-c,d)pyrene | LCS | L-012503-PAH | 158 | DFG-WPCL |
| Indeno(1,2,3-c,d)pyrene | CRM | L-010902-PAH | 151 | DFG-WPCL |
| Indeno(1,2,3-c,d)pyrene | CRM | L-021902-PAH | 294 | DFG-WPCL |
| Manganese | CRM | 2002Dig7 | 65 | MPSL-DFG |
| C1 - Naphthalenes | LCS | L-010902-PAH | 209 | DFG-WPCL |
| C1 - Naphthalenes | LCS | L-021902-PAH | 249 | DFG-WPCL |
| C1 - Naphthalenes | LCS | L-012503-PAH | 220 | DFG-WPCL |
| C1 - Naphthalenes | LCS | L-060203-PAH | 187 | DFG-WPCL |
| cis-Nonachlor | CRM | L39801BS179_KRPEST | -53 | DFG-WPCL |
| PCB 029 | LCS | L39801_BS179_KR_CONG | 33 | DFG-WPCL |
| PCB 056 | LCS | L39801_BS179_KR_CONG | 36 | DFG-WPCL |
| PCB 060 | LCS | L39801_BS179_KR_CONG | 48 | DFG-WPCL |
| Perylene | CRM | L-021902-PAH | 45 | DFG-WPCL |
| Perylene | CRM | L-060203-PAH | 47 | DFG-WPCL |
| Silver | CRM | 2002Dig7 | 128 | MPSL-DFG |

Appendix C-8: Batches containing field duplicated RPD outliers

| Analyte | Site | Field Sample | Field Duplicate | Unit | RPD | Laboratory |
|----------------------------------|-----------|--------------|-----------------|------|-----|------------|
| Aluminum | 206PET310 | 18.7 | 6.03 | µg/L | 102 | MPSL-DFG |
| Benzo(b)fluoranthene | 204ALP010 | 9.24 | 6.91 | ng/g | 29 | DFG-WPCL |
| Benzo(b)fluoranthene | 205PER010 | 0.035 | -0.02 | µg/L | 200 | DFG-WPCL |
| Benzo(b)fluoranthene | 206WIL020 | -0.02 | 0.035 | µg/L | 200 | DFG-WPCL |
| Benzo(g,h,i)perylene | 205PER010 | 0.057 | -0.02 | µg/L | 200 | DFG-WPCL |
| Benzo(g,h,i)perylene | 206WIL020 | -0.02 | 0.035 | µg/L | 200 | DFG-WPCL |
| Benzo(k)fluoranthene | 204ALP010 | -0.02 | 0.035 | µg/L | 200 | DFG-WPCL |
| Benzo(k)fluoranthene | 204ALP010 | -1.83 | 5.45 | ng/g | 200 | DFG-WPCL |
| Benzo(k)fluoranthene | 205PER010 | 0.035 | -0.02 | µg/L | 200 | DFG-WPCL |
| Benzo(k)fluoranthene | 205PER010 | 131 | 99.8 | ng/g | 27 | DFG-WPCL |
| Benzo(k)fluoranthene | 206WIL020 | -0.02 | 0.035 | µg/L | 200 | DFG-WPCL |
| Chrysene | 204ALP010 | -0.02 | 0.035 | µg/L | 200 | DFG-WPCL |
| C2-Chrysenes | 204ALP010 | 5.83 | 4.12 | ng/g | 34 | DFG-WPCL |
| C3-Chrysenes | 204ALP010 | 6.7 | -1.61 | ng/g | 200 | DFG-WPCL |
| Dibenz(a,h)anthracene | 204ALP010 | 4.07 | 2.96 | ng/g | 32 | DFG-WPCL |
| Dibenzothiophene | 204ALP010 | 4.49 | -1.61 | ng/g | 200 | DFG-WPCL |
| 2,6, - Dimethylnaphthalene | 205PER010 | 9.17 | 15.3 | ng/g | 50 | DFG-WPCL |
| Fluoranthene | 205PER010 | 468 | -2 | ng/g | 200 | DFG-WPCL |
| C3-Naphthalenes | 205PER010 | 17.1 | 25.8 | ng/g | 40 | DFG-WPCL |
| C4-Naphthalenes | 205PER010 | 5.77 | 8.37 | ng/g | 37 | DFG-WPCL |
| PCB 027 | 204ALP010 | 0.212 | 0.564 | ng/g | 91 | DFG-WPCL |
| PCB 028 | 204ALP010 | -0.191 | 0.167 | ng/g | 200 | DFG-WPCL |
| PCB 031 | 204ALP010 | 0.233 | -0.166 | ng/g | 200 | DFG-WPCL |
| PCB 033 | 204ALP010 | 1.05 | 0.322 | ng/g | 106 | DFG-WPCL |
| PCB 066 | 204ALP010 | -0.191 | 0.181 | ng/g | 200 | DFG-WPCL |
| PCB 070 | 204ALP010 | -0.191 | 0.19 | ng/g | 200 | DFG-WPCL |
| PCB 087 | 204ALP010 | 0.198 | -0.166 | ng/g | 200 | DFG-WPCL |
| Perylene | 206WIL020 | 0.08 | -0.02 | µg/L | 200 | DFG-WPCL |
| Phenanthrene | 206WIL020 | 0.035 | -0.02 | µg/L | 200 | DFG-WPCL |
| Pyrene | 205PER010 | 0.035 | -0.02 | µg/L | 200 | DFG-WPCL |
| Suspended Sediment Concentration | 205PER010 | 2.9 | -0.1 | mg/L | 200 | AMS |
| 2,3,5-Trimethylnaphthalene | 204ALP010 | 0.035 | -0.02 | µg/L | 200 | DFG-WPCL |
| 2,3,5-Trimethylnaphthalene | 205PER010 | 3.76 | 5.41 | ng/g | 36 | DFG-WPCL |

Appendix C-9: Batches missing laboratory duplicates

| Batch ID | BatchQualCode | BatchValCode | Notes | Lab |
|---------------|---------------|--------------|--|----------|
| AMS102901-1 | A,VMD | VAC,VI | QAO: no DUP or SRM | AMS |
| AMS10401-1 | A,VMD | VAC,VI | QAO: no blk, DUP, CRM | AMS |
| AMS10901-1 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMS12902-1 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMS12902-2 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMS12902-3 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMS2102-1 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMS41102-4 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMS41102-5 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMS41201-1 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMS41202-2 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMS91901-1 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMSDOC21102-1 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSDOC21302-1 | A,VMD | VAC,VI | LAB:all samples were lab filtered within 3 days of collectionQAO: no blk, DUP | AMS |
| AMSDOC5302-1 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSDOC5402-1 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSDOC5402-2 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSTOC2902-1 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMSTOC2902-2 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMSTOC5302-2 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMSTOC5402-1 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMSTOC5402-2 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| L-093001-PAH | A,VMD | VAC,VI | QAO: no LCS or MS/MSD, a DUP was performed | DFG-WPCL |
| L-424-01-TDS | A,VMD | VAC,VI | QAO: no DUP | DFG-WPCL |
| L-430-01-TKN | A,VMD | VAC,VI | QAO: no MS or DUP | DFG-WPCL |
| L-442-01-TDS | A,VMD | VAC,VI | QAO: no blank, MS, DUP | DFG-WPCL |
| L-442-01-TSS | A,VMD | VAC,VI | QAO: no MS or DUP | DFG-WPCL |
| R2-013002-Chl | A,VMD | VAC,VI | LAB:all samples were field filtered on the date of collectionQAO: no blank or DUP | MPSL-DFG |
| R2-013002-Ph | A,VMD | VAC,VI | LAB:all samples were field filtered on the date of collectionQAO: no blk or DUP | MPSL-DFG |
| R2-100101-Chl | A,VMD | VAC,VI | LAB:all samples were field filtered on the date of collection QAO: no blank or DUP | MPSL-DFG |
| R2-100101-Ph | A,VMD | VAC,VI | LAB:all samples were field filtered on the date of collectionQAO: no blk or DUP | MPSL-DFG |
| R2-101001-Chl | A,VMD | VAC,VI | LAB:all samples were field filtered on the date of collection QAO: no blk or DUP | MPSL-DFG |
| R2-101001-Ph | A,VMD | VAC,VI | LAB:all samples were field filtered on the date of collectionQAO: no blk or DUP | MPSL-DFG |
| AMS62102-2 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMS62102-3 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMSDOC62402-1 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSDOC62402-2 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| 021403-4 | A,VMD | VAC,VI | LAB:all samples were acidified within 48 hours of collectionQAO: no DUP | AMS |
| 021403-6 | A,VMD | VAC,VI | LAB:all samples were lab filtered within 48 hours of collectionQAO: no DUP | AMS |
| 63ELGC | A,VMD | VAC,VI | QAO: no DUP, MS %R for diazinon outside QC limit of 80-120 | UCD-GC |
| AMS41201-1 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMS41202-2 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMS41802-1 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMS62102-1 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMS62102-2 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMSDOC51102-1 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSDOC5402-1 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSDOC5402-2 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSDOC5402-3 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSDOC62302-1 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSDOC62402-1 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSTOC51102-1 | A,VMD | VAC,VI | QAO: no blk, DUP | AMS |
| AMSTOC5402-1 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| AMSTOC5402-2 | A,VMD | VAC,VI | QAO: no DUP | AMS |
| SSC03-0002 | A,VMD | VAC,VI | LAB:all samples were lab filtered on the analysis dateQAO: no DUP | MPSL-DFG |

Appendix C-10: Data batches with laboratory duplicates containing RPD outliers

| Analyte | Site | Parent Value | Duplicate Value | Unit | RPD | Laboratory | Batch ID |
|----------------------------|-----------|--------------|-----------------|-------|-----|------------|----------------------|
| Acenaphthene | 201LAG120 | 3.76 | -1.45 | ng/g | 200 | DFG-WPCL | L-060203-PAH |
| Aluminum | 403STCNRB | 23234 | 40000 | mg/Kg | 57 | MPSL-DFG | 2002Dig7 |
| Benz(a)anthracene | 201LAG120 | 2.87 | -1.45 | ng/g | 200 | DFG-WPCL | L-060203-PAH |
| Benzo(a)pyrene | 201LAG120 | 1.64 | -1.45 | ng/g | 200 | DFG-WPCL | L-060203-PAH |
| Benzo(b)fluoranthene | 201LAG120 | 6.11 | 2.6 | ng/g | 80 | DFG-WPCL | L-060203-PAH |
| Benzo(e)pyrene | 206SPA020 | 27.9 | 21.5 | ng/g | 26 | DFG-WPCL | L-010902-PAH |
| Benzo(g,h,i)perylene | 206SPA020 | 25.6 | 3.41 | ng/g | 153 | DFG-WPCL | L-010902-PAH |
| Benzo(g,h,i)perylene | 206SPA020 | 33.3 | 6.56 | ng/g | 134 | DFG-WPCL | L-010902-PAH |
| Benzo(k)fluoranthene | 204ALP010 | -1.58 | 5.45 | ng/g | 200 | DFG-WPCL | L-021902-PAH |
| Biphenyl | 206SPA020 | 3.82 | 31.6 | ng/g | 157 | DFG-WPCL | L-010902-PAH |
| Biphenyl | 201LAG120 | 4.65 | -2.38 | ng/g | 200 | DFG-WPCL | L-060203-PAH |
| Chromium | 403STCNRB | 15.5 | 22.9 | mg/Kg | 42 | MPSL-DFG | 2002Dig7 |
| Chrysene | 201LAG120 | 3.83 | 11.5 | ng/g | 100 | DFG-WPCL | L-060203-PAH |
| C1-Chrysenes | 201LAG120 | 2.62 | 13 | ng/g | 133 | DFG-WPCL | L-060203-PAH |
| C2-Chrysenes | 201LAG120 | 1.65 | 9.35 | ng/g | 140 | DFG-WPCL | L-060203-PAH |
| C3-Chrysenes | 204ALP010 | 5.77 | -1.61 | ng/g | 200 | DFG-WPCL | L-021902-PAH |
| C3-Chrysenes | 201LAG120 | -1.45 | 5.92 | ng/g | 200 | DFG-WPCL | L-060203-PAH |
| Dibenz(a,h)anthracene | 201LAG120 | 14.1 | 2.6 | ng/g | 138 | DFG-WPCL | L-060203-PAH |
| Dibenzothiophene | 204ALP010 | 3.58 | -1.61 | ng/g | 200 | DFG-WPCL | L-021902-PAH |
| C1-Dibenzothiophenes | 201LAG120 | 2.05 | 70.6 | ng/g | 189 | DFG-WPCL | L-060203-PAH |
| C1-Dibenzothiophenes | 206SPA020 | 44.8 | -1.45 | ng/g | 200 | DFG-WPCL | L-010902-PAH |
| C2-Dibenzothiophenes | 206SPA020 | 52.6 | -1.45 | ng/g | 200 | DFG-WPCL | L-010902-PAH |
| C2-Dibenzothiophenes | 206SPA020 | 73.8 | 1.95 | ng/g | 190 | DFG-WPCL | L-010902-PAH |
| C3-Dibenzothiophenes | 206SPA020 | 65.8 | 50 | ng/g | 27 | DFG-WPCL | L-010902-PAH |
| 2,6-Dimethylnaphthalene | 204ALP010 | -1.61 | 3.45 | ng/g | 200 | DFG-WPCL | L-021902-PAH |
| 2,6-Dimethylnaphthalene | 201LAG120 | 35.7 | 4.64 | ng/g | 154 | DFG-WPCL | L-060203-PAH |
| Fluoranthene | 201LAG120 | 1.86 | 5.21 | ng/g | 95 | DFG-WPCL | L-060203-PAH |
| C1-Fluoranthene/Pyrenes | 201LAG120 | 3.02 | 12.9 | ng/g | 124 | DFG-WPCL | L-060203-PAH |
| C1-Fluoranthene/Pyrenes | 204ALP010 | 1.97 | 2.79 | ng/g | 34 | DFG-WPCL | L-021902-PAH |
| Fluorene | 206SPA020 | 2.52 | -1.45 | ng/g | 200 | DFG-WPCL | L-010902-PAH |
| Fluorene | 206SPA020 | -2.38 | 2.37 | ng/g | 200 | DFG-WPCL | L-010902-PAH |
| C2-Fluorenes | 204ALP010 | 3.62 | 6.06 | ng/g | 50 | DFG-WPCL | L-021902-PAH |
| C2-Fluorenes | 206SPA020 | 4.82 | 7.44 | ng/g | 43 | DFG-WPCL | L-010902-PAH |
| C3-Fluorenes | 206SPA020 | 7.28 | 22.1 | ng/g | 100 | DFG-WPCL | L-010902-PAH |
| Heptachlor epoxide | 204ALP010 | 2.21 | 1.58 | ng/g | 33 | DFG-WPCL | L39801BS178_KRPEST |
| Indeno(1,2,3-c,d)pyrene | 204ALP010 | 6.91 | 3.09 | ng/g | 76 | DFG-WPCL | L-021902-PAH |
| Indeno(1,2,3-c,d)pyrene | 201LAG120 | 9.56 | 4.19 | ng/g | 78 | DFG-WPCL | L-060203-PAH |
| Mercury | 403STCNRB | 0.169 | 0.104 | mg/Kg | 48 | MPSL-DFG | 2002Dig7Hg |
| 1-Methylnaphthalene | 201LAG120 | 46.4 | 4.03 | ng/g | 168 | DFG-WPCL | L-060203-PAH |
| 1-Methylnaphthalene | 206SPA020 | -2.34 | 4.61 | ng/g | 200 | DFG-WPCL | L-010902-PAH |
| 2-Methylnaphthalene | 201LAG120 | 6.11 | 10.6 | ng/g | 54 | DFG-WPCL | L-060203-PAH |
| 2-Methylnaphthalene | 206SPA020 | 6.11 | 47.9 | ng/g | 155 | DFG-WPCL | L-010902-PAH |
| 1-Methylphenanthrene | 201LAG120 | 2.01 | 15.5 | ng/g | 154 | DFG-WPCL | L-060203-PAH |
| Naphthalene | 201LAG120 | 4.98 | 61.2 | ng/g | 170 | DFG-WPCL | L-060203-PAH |
| Naphthalene | 206SPA020 | 15.1 | 8.95 | ng/g | 51 | DFG-WPCL | L-010902-PAH |
| C1-Naphthalenes | 206SPA020 | 19.4 | -1.61 | ng/g | 200 | DFG-WPCL | L-010902-PAH |
| C1-Naphthalenes | 204ALP010 | 2.28 | 77.6 | ng/g | 188 | DFG-WPCL | L-021902-PAH |
| C2-Naphthalenes | 206SPA020 | 17.6 | 12 | ng/g | 38 | DFG-WPCL | L-010902-PAH |
| C2-Naphthalenes | 204ALP010 | 6.4 | 2.22 | ng/g | 97 | DFG-WPCL | L-021902-PAH |
| C2-Naphthalenes | 201LAG120 | 14 | 66.7 | ng/g | 131 | DFG-WPCL | L-060203-PAH |
| C3-Naphthalenes | 206SPA020 | 11.6 | 41.8 | ng/g | 113 | DFG-WPCL | L-010902-PAH |
| C3-Naphthalenes | 204ALP010 | 2.22 | 8.05 | ng/g | 113 | DFG-WPCL | L-021902-PAH |
| C3-Naphthalenes | 201LAG120 | 10.6 | 2.08 | ng/g | 134 | DFG-WPCL | L-060203-PAH |
| C4-Naphthalenes | 206SPA020 | -2.38 | 2.23 | ng/g | 200 | DFG-WPCL | L-010902-PAH |
| C4-Naphthalenes | 206SPA020 | 2.83 | 12.3 | ng/g | 125 | DFG-WPCL | L-010902-PAH |
| PCB 027 | 204ALP010 | 0.564 | 0.316 | ng/g | 56 | DFG-WPCL | L39801_BS178_KR_CONG |
| PCB 028 | 204ALP010 | 0.167 | -0.169 | ng/g | 200 | DFG-WPCL | L39801_BS178_KR_CONG |
| PCB 031 | 204ALP010 | 0.249 | -0.166 | ng/g | 200 | DFG-WPCL | L39801_BS178_KR_CONG |
| PCB 033 | 204ALP010 | -0.169 | 0.322 | ng/g | 200 | DFG-WPCL | L39801_BS178_KR_CONG |
| PCB 066 | 204ALP010 | 0.181 | -0.169 | ng/g | 200 | DFG-WPCL | L39801_BS178_KR_CONG |
| Analyte | Site | Parent Value | Duplicate Value | Unit | RPD | Laboratory | Batch ID |
| Phenanthrene | 201LAG120 | 49 | 14.8 | ng/g | 107 | DFG-WPCL | L-060203-PAH |
| Phenanthrene | 206SPA020 | 21.2 | 27.7 | ng/g | 26 | DFG-WPCL | L-010902-PAH |
| C1-Phenanthrene/Anthracene | 201LAG120 | 70.1 | 22.5 | ng/g | 103 | DFG-WPCL | L-060203-PAH |
| C1-Phenanthrene/Anthracene | 204ALP010 | 7.65 | 4.81 | ng/g | 46 | DFG-WPCL | L-021902-PAH |
| C2-Phenanthrene/Anthracene | 206SPA020 | 17.9 | 10.4 | ng/g | 53 | DFG-WPCL | L-010902-PAH |

Appendix C-10: Data batches with laboratory duplicates containing RPD outliers

| | | | | | | | |
|----------------------------|-----------|------|-------|------|-----|----------|--------------|
| C2-Phenanthrene/Anthracene | 206SPA020 | 24.1 | 62.4 | ng/g | 88 | DFG-WPCL | L-010902-PAH |
| C3-Phenanthrene/Anthracene | 201LAG120 | 3.67 | 18.6 | ng/g | 134 | DFG-WPCL | L-060203-PAH |
| C4-Phenanthrene/Anthracene | 206SPA020 | 23.9 | 5.27 | ng/g | 128 | DFG-WPCL | L-010902-PAH |
| C4-Phenanthrene/Anthracene | 206SPA020 | 17.6 | -1.45 | ng/g | 200 | DFG-WPCL | L-010902-PAH |
| Pyrene | 201LAG120 | 8.52 | 50.3 | ng/g | 142 | DFG-WPCL | L-060203-PAH |
| Pyrene | 206SPA020 | 37.7 | 2.58 | ng/g | 174 | DFG-WPCL | L-010902-PAH |
| 2,3,5-Trimethylnaphthalene | 201LAG120 | 1.86 | 7.93 | ng/g | 124 | DFG-WPCL | L-060203-PAH |

Appendix D-1: Summary of continuous field measures in Walker Creek Watershed

| | Station | Winter | | Spring | | | Late Spring | | Late Summer | | | |
|------------------------------|---|--------------|--------------|--------------|--------------|--------------|-------------|-----------|--------------|--------------|--------------|--------------|
| | | WLK160 | WLK180 | WLK100 | WLK140 | WLK200 | WLK160 | WLK180 | WLK140 | WLK160 | WLK180 | WLK200 |
| | Start Date | 1/21/2002 | 2/15/2002 | 3/5/2002 | 3/5/2002 | 3/5/2002 | 5/23/2002 | 5/23/2002 | 8/16/2002 | 8/16/2002 | 8/16/2002 | 8/16/2002 |
| | End Date | 2/5/2002 | 2/23/2002 | 3/19/2002 | 3/19/2002 | 3/19/2002 | 5/30/2002 | 5/30/2002 | 8/21/2002 | 8/21/2002 | 8/21/2002 | 8/21/2002 |
| | Data points | 1445 | 790 | 1343 | 1346 | 1344 | 672 | 671 | 490 | 476 | 469 | 464 |
| Temperature (°C) | Min. | 5.66 | 7.82 | 7.54 | 6.99 | 9.90 | 12.05 | 11.50 | 13.99 | 13.57 | 13.91 | 13.40 |
| | 0.25 | 7.17 | 9.94 | 10.35 | 9.58 | 10.59 | 14.40 | 13.99 | 14.68 | 14.39 | 14.84 | 13.71 |
| | Median | 8.04 | 11.22 | 11.15 | 10.84 | 11.27 | 15.54 | 15.11 | 15.47 | 15.12 | 15.20 | 13.81 |
| | 0.75 | 8.88 | 11.91 | 12.48 | 11.50 | 11.88 | 17.13 | 16.52 | 16.07 | 16.42 | 16.01 | 14.15 |
| | Max. | 10.72 | 14.33 | 14.20 | 13.02 | 13.19 | 20.92 | 18.42 | 17.72 | 18.64 | 17.12 | 15.42 |
| | HWAT | 8.1 | 11.1 | 11.8 | 11.1 | 11.6 | 15.9 | 15.2 | | | | |
| | QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 7.32 | 7.36 | 7.63 | 6.63 | 7.08 | 7.42 | 7.30 | 7.55 | 7.64 | 7.08 | 7.05 |
| | 0.25 | 7.66 | 7.42 | 7.70 | 7.42 | 7.21 | 7.47 | 7.63 | 7.60 | 7.71 | 7.16 | 7.07 |
| | Median | 7.70 | 7.57 | 7.76 | 7.47 | 7.32 | 7.74 | 7.65 | 7.65 | 7.76 | 7.20 | 7.10 |
| | 0.75 | 7.76 | 7.67 | 7.86 | 7.55 | 7.54 | 8.71 | 7.75 | 7.74 | 7.91 | 7.27 | 7.12 |
| | Max. | 8.31 | 8.02 | 8.09 | 7.81 | 8.13 | 9.07 | 7.99 | 7.84 | 8.18 | 7.37 | 7.31 |
| | Data Quality Objective: +/- 0.50 pH units | 0.10 | 0.02 | 0.02 | 0.02 | 0.10 | 0.13 | 0.16 | 0.10 | 0.11 | 0.06 | 0.03 |
| | QA Qualifier | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 3 | Valid | Valid | Estimated, 3 | Estimated, 3 | Estimated, 3 | Estimated, 3 |
| Dissolved Oxygen (%) | Min. | 96.60 | 95.40 | 96.60 | 97.50 | 89.90 | 82.90 | 63.20 | 90.00 | 90.00 | 43.60 | 80.10 |
| | 0.25 | 105.30 | 97.70 | 101.30 | 99.70 | 99.08 | 86.90 | 86.80 | 91.20 | 91.80 | 64.60 | 80.80 |
| | Median | 108.00 | 100.20 | 103.10 | 100.90 | 100.60 | 95.10 | 88.20 | 92.80 | 93.60 | 68.30 | 82.20 |
| | 0.75 | 110.40 | 101.10 | 109.45 | 106.70 | 108.83 | 114.10 | 94.80 | 96.70 | 99.53 | 70.80 | 85.90 |
| | Max. | 121.30 | 107.90 | 123.90 | 126.90 | 134.90 | 127.30 | 104.90 | 101.00 | 107.10 | 79.40 | 94.00 |
| | Data Quality Objective: +/- 5.0% | 11.00 | 2.40 | 3.00 | 3.90 | 5.00 | 5.60 | 2.90 | 0.70 | 0.40 | 1.80 | 0.60 |
| | QA Qualifier | rejected | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | rejected | Valid | Estimated, 3 | Estimated, 3 | Estimated, 3 | Estimated, 3 |
| Dissolved Oxygen (mg/L) | Min. | 11.22 | 10.35 | 10.29 | 10.54 | 9.86 | 7.98 | 5.93 | 8.77 | 8.81 | 4.20 | 8.31 |
| | 0.25 | 12.40 | 10.82 | 11.03 | 11.03 | 10.94 | 8.79 | 8.77 | 9.14 | 9.18 | 6.43 | 8.38 |
| | Median | 12.83 | 11.00 | 11.46 | 11.39 | 11.13 | 9.48 | 9.05 | 9.40 | 9.52 | 6.84 | 8.51 |
| | 0.75 | 13.26 | 11.12 | 12.13 | 11.91 | 11.84 | 11.03 | 9.47 | 9.61 | 9.88 | 7.13 | 8.83 |
| | Max. | 14.22 | 12.22 | 13.58 | 13.94 | 14.43 | 12.57 | 10.58 | 9.94 | 10.37 | 8.10 | 9.42 |
| | Data Quality Objective: +/- 0.5 mg/l | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% |
| | QA Qualifier | rejected | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | rejected | Valid | Estimated, 3 | Estimated, 3 | Estimated, 3 | Estimated, 3 |
| Specific Conductance (µS/cm) | Min. | 155 | rejected | 211 | 146 | 152 | NA | 261 | 162 | 161 | 283 | 161 |
| | 0.25 | 159 | rejected | 261 | 155 | 155 | NA | 267 | 163 | 162 | 284 | 161 |
| | Median | 163 | rejected | 266 | 158 | 155 | NA | 268 | 164 | 163 | 285 | 162 |
| | 0.75 | 165 | rejected | 270 | 161 | 156 | NA | 269 | 164 | 163 | 286 | 162 |
| | Max. | 171 | rejected | 285 | 172 | 158 | NA | 273 | 167 | 165 | 289 | 162 |
| | Data Quality Objective: +/- 50 uS/cm | 3 | 0.11 | 4 | 29 | 2 | NA | 30 | 2 | 6 | 0 | 8 |
| | QA Qualifier | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | NA | Valid | Estimated, 3 | Estimated, 3 | Estimated, 3 | Estimated, 3 |

Appendix D-2: Summary of continuous field measures in Lagunitas Creek Watershed

| | Station | Winter Wet | | | Spring Wet | | | | |
|------------------------------|---|------------|-----------|-----------|------------|-----------|-----------|-------------|-----------|
| | | LAG050 | LAG160 | LAG270 | LAG050 | LAG085 | LAG160 | LAG165 | LAG270 |
| | Start Date | 2/15/2002 | 1/21/2002 | 1/6/2002 | 5/23/2002 | 5/9/2002 | 5/23/2002 | 5/23/2002 | 5/8/2002 |
| | End date | 3/1/2002 | 2/5/2002 | 1/21/2002 | 5/30/2002 | 5/21/2002 | 5/30/2002 | 5/30/2002 | 5/21/2002 |
| | Data points | 1336 | 1457 | 1415 | 669 | 1159 | 654 | 669 | 1153 |
| Temperature (°C) | Min. | 9.05 | 4.07 | 5.93 | 11.35 | 9.26 | 12.26 | 11.60 | 10.08 |
| | 0.25 | 9.81 | 5.82 | 7.28 | 13.62 | 10.85 | 13.40 | 13.65 | 11.59 |
| | Median | 10.51 | 6.40 | 9.69 | 14.76 | 11.50 | 14.01 | 14.36 | 12.02 |
| | 0.75 | 11.06 | 7.26 | 10.97 | 15.93 | 12.67 | 14.57 | 15.04 | 12.66 |
| | Max. | 12.35 | 8.81 | 12.35 | 19.40 | 14.22 | 15.98 | 17.07 | 13.84 |
| | HWAT | | 6.9 | 11.0 | 14.8 | 11.9 | | 14.3 | 12.4 |
| | QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 7.03 | 7.29 | 7.70 | 6.60 | 7.62 | 7.05 | NA | 7.74 |
| | 0.25 | 7.10 | 7.65 | 7.91 | 7.51 | 7.69 | 7.13 | NA | 7.82 |
| | Median | 7.20 | 7.67 | 7.95 | 7.53 | 7.72 | 7.18 | NA | 7.83 |
| | 0.75 | 7.27 | 7.75 | 7.97 | 7.63 | 7.79 | 7.32 | NA | 7.86 |
| | Max. | 7.41 | 8.15 | 8.00 | 7.71 | 7.88 | 7.52 | NA | 7.91 |
| | Data Quality Objective: +/- 0.50 pH units | 0.08 | 0.09 | 0.09 | 0.15 | 0.05 | 0.18 | NA | 0.03 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Valid | Rejected, 6 | Valid |
| Dissolved Oxygen (%) | Min. | 91.90 | 95.90 | 58.90 | 72.00 | 90.60 | 0.90 | 89.30 | 79.20 |
| | 0.25 | 98.30 | 99.40 | 66.10 | 86.30 | 91.60 | 23.85 | 90.90 | 83.50 |
| | Median | 99.30 | 102.60 | 76.00 | 88.40 | 92.80 | 51.40 | 92.90 | 85.40 |
| | 0.75 | 100.80 | 105.60 | 86.05 | 95.60 | 97.20 | 66.90 | 97.00 | 89.90 |
| | Max. | 103.30 | 123.00 | 104.10 | 103.40 | 103.10 | 81.80 | 100.90 | 95.10 |
| | Data Quality Objective: +/- 5.0% | 3.50 | 5.00 | 34.60 | 3.00 | 0.10 | 3.40 | 3.90 | 2.50 |
| | QA Qualifier | Valid | Valid | Rejected | Valid | Valid | Valid | Valid | Valid |
| Dissolved Oxygen (mg/L) | Min. | 10.14 | 11.29 | 7.03 | 7.01 | 9.57 | 0.10 | 8.87 | 8.63 |
| | 0.25 | 10.93 | 12.24 | 7.93 | 8.87 | 10.00 | 2.41 | 9.35 | 8.99 |
| | Median | 11.15 | 12.69 | 8.56 | 9.12 | 10.16 | 5.26 | 9.55 | 9.23 |
| | 0.75 | 11.26 | 13.15 | 9.54 | 9.60 | 10.44 | 6.88 | 9.86 | 9.57 |
| | Max. | 11.56 | 14.76 | 11.18 | 10.27 | 10.96 | 8.76 | 10.61 | 10.16 |
| | Data Quality Objective: +/- 0.5 mg/l | NA | NA | NA | NA | NA | NA | NA | NA |
| | QA Qualifier | Valid, 1 | Valid, 1 | Rejected | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 |
| Specific Conductance (µS/cm) | Min. | 99.00 | 206.00 | 213.00 | 277.00 | 238.00 | 299.00 | 216.00 | 398.00 |
| | 0.25 | 119.00 | 220.00 | 259.00 | 289.00 | 243.00 | 306.00 | 217.00 | 450.00 |
| | Median | 147.00 | 224.00 | 291.00 | 290.00 | 248.00 | 317.00 | 218.00 | 454.00 |
| | 0.75 | 161.00 | 229.00 | 316.00 | 296.00 | 253.00 | 338.00 | 220.00 | 459.00 |
| | Max. | 195.00 | 236.00 | 332.00 | 528.00 | 292.00 | 410.00 | 221.00 | 463.00 |
| | Data Quality Objective: +/- 50 uS/cm | 20.00 | 17.00 | 17.00 | 13.00 | 2.00 | 14.00 | 3.00 | 7.00 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid |

Appendix D-2: Summary of continuous field measures in Lagunitas Creek Watershed

| | | Summer Dry | | | | |
|------------------------------|---|------------|-----------|-----------|-----------|-----------|
| Station | | LAG050 | LAG085 | LAG165 | LAG190 | LAG270 |
| Start Date | | 8/1/2002 | 8/1/2002 | 8/1/2002 | 8/1/2002 | 8/1/2002 |
| End date | | 8/13/2002 | 8/13/2002 | 8/13/2002 | 8/13/2002 | 8/13/2002 |
| Data points | | 1147 | 1148 | 1137 | 1148 | 1131 |
| Temperature (°C) | Min. | 13.05 | 13.14 | 13.18 | 11.65 | 13.23 |
| | 0.25 | 15.29 | 15.16 | 14.80 | 13.97 | 14.97 |
| | Median | 16.09 | 15.87 | 15.49 | 14.91 | 15.74 |
| | 0.75 | 17.46 | 16.69 | 16.26 | 15.78 | 16.36 |
| | Max. | 19.98 | 18.40 | 17.17 | 17.59 | 18.09 |
| | HWAT | 16.5 | 16.1 | 15.4 | 14.9 | 15.4 |
| | QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 7.42 | 7.56 | 7.57 | 8.00 | 7.49 |
| | 0.25 | 7.47 | 7.67 | 7.61 | 8.04 | 7.70 |
| | Median | 7.52 | 7.70 | 7.63 | 8.05 | 7.78 |
| | 0.75 | 7.66 | 7.78 | 7.66 | 8.08 | 7.83 |
| | Max. | 7.92 | 7.93 | 7.72 | 8.13 | 7.97 |
| | Data Quality Objective: +/- 0.50 pH units | 0.10 | 0.12 | 0.03 | 0.06 | 0.11 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid |
| Dissolved Oxygen (%) | Min. | 72.10 | 70.10 | 86.40 | 89.10 | 56.00 |
| | 0.25 | 77.00 | 75.50 | 88.30 | 94.40 | 63.80 |
| | Median | 79.50 | 78.20 | 90.30 | 95.60 | 68.70 |
| | 0.75 | 92.95 | 86.40 | 93.40 | 96.90 | 74.30 |
| | Max. | 109.60 | 97.20 | 97.30 | 99.50 | 88.90 |
| | Data Quality Objective: +/- 5.0% | 3.60 | 0.30 | 0.80 | 0.80 | 2.00 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid |
| Dissolved Oxygen (mg/L) | Min. | 7.07 | 6.83 | 8.51 | 8.83 | 5.44 |
| | 0.25 | 7.61 | 7.48 | 8.81 | 9.41 | 6.36 |
| | Median | 8.00 | 7.89 | 9.06 | 9.67 | 6.86 |
| | 0.75 | 9.04 | 8.47 | 9.30 | 9.90 | 7.43 |
| | Max. | 10.29 | 9.37 | 9.83 | 10.55 | 8.76 |
| | Data Quality Objective: +/- 0.5 mg/l | NA | NA | NA | NA | NA |
| | QA Qualifier | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 |
| Specific Conductance (µS/cm) | Min. | 319.00 | 300.00 | 184.00 | 360.00 | 488.00 |
| | 0.25 | 321.00 | 303.00 | 185.00 | 361.00 | 491.00 |
| | Median | 322.00 | 304.00 | 186.00 | 362.00 | 497.00 |
| | 0.75 | 324.00 | 308.00 | 186.00 | 364.00 | 506.00 |
| | Max. | 340.00 | 310.00 | 187.00 | 366.00 | 514.00 |
| | Data Quality Objective: +/- 50 uS/cm | 2.00 | 6.00 | 8.00 | 0.00 | 6.00 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid |

Appendix D-3: Summary of continuous field measures in San Leandro Creek Watershed

| | Station | Winter Wet | | | Spring Wet | | | | |
|------------------------------|---|------------|----------|----------|------------|----------|----------|----------|----------|
| | | SLE050 | SLE190 | SLE210 | SLE050 | SLE170 | SLE180 | SLE190 | SLE210 |
| | Start Date | 02/15/02 | 01/18/02 | 01/04/02 | 04/11/02 | 04/11/02 | 04/11/02 | 04/11/02 | 04/11/02 |
| | End Date | 03/01/02 | 02/05/02 | 01/18/02 | 04/23/02 | 04/23/02 | 04/23/02 | 04/23/02 | 04/23/02 |
| | Data points | 1339 | 1707 | 1355 | 1153 | 1151 | 1151 | 1162 | 1160 |
| Temperature (°C) | Min. | 9.37 | 3.51 | 4.51 | 10.09 | 6.97 | 7.31 | 7.66 | 8.47 |
| | 0.25 | 10.74 | 5.03 | 7.41 | 12.12 | 9.36 | 9.56 | 9.36 | 10.28 |
| | Median | 11.79 | 5.69 | 9.48 | 13.14 | 10.63 | 10.92 | 10.35 | 11.32 |
| | 0.75 | 12.53 | 6.36 | 10.78 | 14.98 | 12.18 | 12.38 | 11.73 | 12.92 |
| | Max. | 14.54 | 8.75 | 12.15 | 17.96 | 15.40 | 15.80 | 14.66 | 16.61 |
| | HWAT | 12.1 | 6.3 | 10.7 | 14.0 | 11.2 | 11.5 | 11.0 | 11.8 |
| | QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 7.52 | 8.11 | 8.06 | 7.19 | 7.80 | 7.56 | 7.65 | 7.84 |
| | 0.25 | 7.88 | 8.25 | 8.16 | 7.34 | 7.95 | 7.64 | 8.00 | 8.02 |
| | Median | 8.01 | 8.27 | 8.20 | 7.46 | 7.97 | 7.66 | 8.01 | 8.04 |
| | 0.75 | 8.16 | 8.30 | 8.23 | 7.77 | 8.02 | 7.73 | 8.02 | 8.07 |
| | Max. | 8.41 | 8.31 | 8.28 | 8.22 | 8.11 | 7.84 | 8.06 | 8.66 |
| | Data Quality Objective: +/- 0.50 pH units | 0.07 | 0.09 | 0.09 | 0.02 | 0.04 | 0.06 | 0.03 | 0.05 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid |
| Dissolved Oxygen (%) | Min. | 63.80 | 72.40 | 100.90 | 12.30 | 95.30 | 90.00 | 92.00 | 86.10 |
| | 0.25 | 79.10 | 80.20 | 104.80 | 38.30 | 97.80 | 92.00 | 94.20 | 88.90 |
| | Median | 88.60 | 88.70 | 108.00 | 50.80 | 99.60 | 93.20 | 95.30 | 90.75 |
| | 0.75 | 99.20 | 96.80 | 109.50 | 60.20 | 101.20 | 97.00 | 96.20 | 94.60 |
| | Max. | 125.30 | 109.10 | 111.30 | 96.00 | 105.40 | 105.50 | 98.60 | 101.90 |
| | Data Quality Objective: +/- 5.0% | 1.70 | 17.50 | 17.50 | 0.40 | 4.20 | 0.10 | 2.70 | 0.30 |
| | QA Qualifier | Valid | Rejected | Rejected | Valid | Valid | Valid | Valid | Valid |
| Dissolved Oxygen (mg/L) | Min. | 7.17 | 9.22 | 11.58 | 1.24 | 9.66 | 9.26 | 9.41 | 8.85 |
| | 0.25 | 8.60 | 10.34 | 12.15 | 4.11 | 10.55 | 10.00 | 10.22 | 9.64 |
| | Median | 9.60 | 10.86 | 12.33 | 5.30 | 11.11 | 10.41 | 10.67 | 9.99 |
| | 0.75 | 10.78 | 12.08 | 12.57 | 6.28 | 11.55 | 10.83 | 10.99 | 10.26 |
| | Max. | 13.14 | 13.29 | 13.09 | 9.75 | 12.34 | 11.58 | 11.58 | 11.59 |
| | Data Quality Objective: +/- 0.5 mg/l | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A |
| | QA Qualifier | Valid, 1 | Rejected | Rejected | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 |
| Specific Conductance (µS/cm) | Min. | 65 | 396 | 388 | 137 | 345 | 538 | 670 | 18 |
| | 0.25 | 416 | 636 | 484 | 551 | 458 | 568 | 775 | 629 |
| | Median | 567 | 655 | 529 | 609 | 462 | 568 | 782 | 656 |
| | 0.75 | 604 | 688 | 558 | 624 | 464 | 569 | 790 | 667 |
| | Max. | 620 | 701 | 573 | 654 | 468 | 573 | 800 | 678 |
| | Data Quality Objective: +/- 50 uS/cm | 45 | 10 | 10 | 3 | 3 | 9 | 7 | 10 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid |

Appendix D-3: Summary of continuous field measures in San Leandro Creek Watershed

| | | Summer Dry | | | | Fall Dry | | | |
|------------------------------|---|------------|----------|----------|----------|--------------|--------------|--------------|--------------|
| | Station | SLE070 | SLE170 | SLE190 | SLE210 | SLE090 | SLE170 | SLE190 | SLE210 |
| | Start Date | 05/31/02 | 05/31/02 | 05/31/02 | 05/31/02 | 10/15/2002 | 10/15/2002 | 10/15/2002 | 10/15/2002 |
| | End Date | 06/05/02 | 06/18/02 | 06/18/02 | 06/18/02 | 10/18/02 | 10/18/02 | 10/18/02 | 10/18/02 |
| | Data points | 373 | 1723 | 1722 | 1724 | 263 | 262 | 260 | 265 |
| Temperature (°C) | Min. | 13.20 | 10.85 | 10.95 | 12.67 | 14.63 | 11.99 | 11.53 | 11.57 |
| | 0.25 | 14.33 | 12.42 | 12.70 | 14.19 | 14.90 | 12.14 | 11.68 | 11.88 |
| | Median | 15.31 | 13.32 | 13.56 | 15.19 | 15.02 | 12.28 | 11.77 | 11.96 |
| | 0.75 | 16.19 | 14.19 | 14.36 | 16.42 | 15.10 | 12.53 | 11.97 | 12.21 |
| | Max. | 18.36 | 17.14 | 16.76 | 19.62 | 15.25 | 13.13 | 12.32 | 12.76 |
| | HWAT | | 13.9 | 14.4 | 16.2 | | | | |
| | QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 7.39 | 7.47 | 7.85 | 7.83 | 7.14 | 7.00 | 7.79 | 7.27 |
| | 0.25 | 7.56 | 7.71 | 7.94 | 7.91 | 7.20 | 7.18 | 7.85 | 7.32 |
| | Median | 7.65 | 7.74 | 7.97 | 7.98 | 7.22 | 7.20 | 7.86 | 7.34 |
| | 0.75 | 7.78 | 7.77 | 8.00 | 8.07 | 7.26 | 7.21 | 7.86 | 7.36 |
| | Max. | 8.49 | 7.85 | 8.06 | 8.20 | 7.53 | 7.24 | 7.87 | 7.40 |
| | Data Quality Objective: +/- 0.50 pH units | 0.13 | 0.03 | 0.01 | 0.07 | 0.05 | 0.06 | 0.10 | 0.06 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 |
| Dissolved Oxygen (%) | Min. | 11.50 | 85.50 | 83.10 | 67.60 | 33.10 | 40.40 | 80.60 | 9.80 |
| | 0.25 | 42.20 | 90.10 | 87.20 | 80.08 | 51.45 | 51.90 | 81.60 | 13.80 |
| | Median | 51.10 | 91.70 | 89.00 | 85.20 | 62.40 | 54.10 | 82.00 | 15.50 |
| | 0.75 | 58.70 | 93.70 | 90.70 | 97.00 | 78.90 | 55.70 | 82.50 | 21.80 |
| | Max. | 118.40 | 98.60 | 95.50 | 118.50 | 83.50 | 58.80 | 84.40 | 29.80 |
| | Data Quality Objective: +/- 5.0% | 2.60 | 2.60 | 2.90 | 0.20 | 4.30 | 4.50 | 2.10 | 0.30 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid |
| Dissolved Oxygen (mg/L) | Min. | 1.14 | 8.43 | 8.14 | 6.88 | 3.33 | 4.30 | 8.71 | 1.05 |
| | 0.25 | 4.17 | 9.31 | 8.95 | 8.01 | 5.17 | 5.54 | 8.82 | 1.48 |
| | Median | 5.14 | 9.62 | 9.26 | 8.65 | 6.28 | 5.74 | 8.85 | 1.67 |
| | 0.75 | 5.99 | 9.90 | 9.54 | 9.57 | 7.97 | 5.97 | 8.89 | 2.35 |
| | Max. | 11.13 | 10.52 | 10.28 | 11.57 | 8.40 | 6.28 | 9.06 | 3.18 |
| | Data Quality Objective: +/- 0.5 mg/l | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A |
| | QA Qualifier | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 |
| Specific Conductance (µS/cm) | Min. | 521 | 454 | 869 | 650 | 559 | 460 | 974 | 669 |
| | 0.25 | 557 | 457 | 879 | 677 | 539 | 459 | 970 | 667 |
| | Median | 612 | 458 | 893 | 681 | 566 | 461 | 976 | 670 |
| | 0.75 | 624 | 460 | 906 | 684 | 575 | 461 | 978 | 673 |
| | Max. | 629 | 467 | 913 | 702 | 591 | 465 | 979 | 680 |
| | Data Quality Objective: +/- 50 uS/cm | 15 | 15 | 29 | 17 | 3 | 4 | 10 | 1 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid |

Appendix D-4: Summary of continuous field measures in San Pablo and Wildcat Creek Watersheds

| | | Winter Wet Season | | | | Spring Runoff Season | | | | | | |
|------------------------------|---|-------------------|--------------|--------------|-------------|----------------------|-----------|--------------|--------------|--------------|--------------|-----------|
| Station | | WIL020 | SPA050 | SPA220 | SPA240 | WIL020 | WIL100 | SPA050 | SPA070 | SPA200 | SPA220 | SPA235 |
| Start Date | | 2/19/2002 | 11/21/2001 | 1/18/2002 | 1/4/2002 | 5/9/2002 | 4/11/2002 | 5/9/2002 | 3/5/2002 | 3/5/2002 | 3/5/2002 | 5/9/2002 |
| End Date | | 03/01/02 | 12/4/2001 | 2/5/2002 | 1/18/2002 | 5/21/2002 | 4/23/2002 | 5/21/2002 | 3/19/2002 | 3/19/2002 | 3/19/2002 | 5/21/2002 |
| Data points | | 972 | 1273 | 1727 | 1338 | 1155 | 1161 | 1155 | 1322 | 1316 | 1312 | 1128 |
| Temperature (°C) | Min. | 10.42 | 9.86 | 5.04 | 4.29 | 13.38 | 10.98 | 12.67 | 8.62 | 6.64 | 7.06 | 10.51 |
| | 0.25 | 11.67 | 11.04 | 6.52 | 7.61 | 14.73 | 12.53 | 13.95 | 9.92 | 8.57 | 8.89 | 12.27 |
| | Median | 12.41 | 11.57 | 7.12 | 9.18 | 15.17 | 13.29 | 14.51 | 10.68 | 9.63 | 9.85 | 13.12 |
| | 0.75 | 12.97 | 13.16 | 7.69 | 10.64 | 15.63 | 14.48 | 15.34 | 11.65 | 10.76 | 11.01 | 14.25 |
| | Max. | 14.63 | 16.29 | 10.17 | 12.27 | 17.11 | 17.17 | 16.37 | 13.03 | 12.19 | 12.68 | 15.89 |
| | HWAT | 12.6 | 12.8 | 7.6 | 10.6 | 15.3 | 14.0 | 14.9 | 11.2 | 10.3 | 10.5 | 13.8 |
| QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | |
| pH | Min. | 7.73 | 7.57 | 7.97 | 8.05 | 7.22 | 7.64 | 7.57 | 7.54 | 7.50 | 7.79 | 7.70 |
| | 0.25 | 8.10 | 7.65 | 8.10 | 8.27 | 7.55 | 7.81 | 7.63 | 7.75 | 7.73 | 8.14 | 7.87 |
| | Median | 8.21 | 7.75 | 8.35 | 8.29 | 7.60 | 7.84 | 7.75 | 7.81 | 7.76 | 8.23 | 7.92 |
| | 0.75 | 8.29 | 7.86 | 8.39 | 8.31 | 7.64 | 7.88 | 7.79 | 7.87 | 7.81 | 8.27 | 8.00 |
| | Max. | 8.56 | 8.07 | 8.56 | 8.44 | 7.89 | 7.96 | 7.90 | 8.04 | 8.11 | 8.45 | 8.20 |
| | Data Quality Objective: +/- 0.50 pH units | 0.04 | 0.29 | 0.09 | 0.09 | 2.65 | 0.02 | 0.05 | 0.13 | 0.04 | 0.05 | 0.02 |
| QA Qualifier | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Rejected, 6 | Valid | Valid | Valid | Estimated, 2 | Estimated, 2 | Estimated, 2 | Valid |
| Dissolved Oxygen (%) | Min. | 70.50 | 78.50 | 101.10 | 99.30 | 3.70 | 68.70 | 60.50 | 84.10 | 91.90 | 95.90 | 68.50 |
| | 0.25 | 81.88 | 85.10 | 105.30 | 100.50 | 32.90 | 77.80 | 75.80 | 92.80 | 97.60 | 101.00 | 75.40 |
| | Median | 98.10 | 87.70 | 107.50 | 101.40 | 42.00 | 81.90 | 78.80 | 95.60 | 98.80 | 101.90 | 81.30 |
| | 0.75 | 120.05 | 89.90 | 109.30 | 103.10 | 51.30 | 85.60 | 86.50 | 98.50 | 100.40 | 103.40 | 89.43 |
| | Max. | 145.20 | 106.80 | 119.20 | 109.60 | 78.50 | 91.00 | 104.00 | 131.20 | 106.80 | 111.80 | 101.60 |
| | Data Quality Objective: +/- 5.0% | 0.50 | 49.50 | 7.20 | 7.20 | 5.00 | 5.00 | 2.00 | 0.50 | 4.20 | 2.50 | 0.70 |
| QA Qualifier | Estimated, 2 | Rejected | Rejected | Rejected | Valid | Valid | Valid | Estimated, 2 | Estimated, 2 | Estimated, 2 | Valid | |
| Dissolved Oxygen (mg/L) | Min. | 7.46 | 8.61 | 11.73 | 10.69 | 0.37 | 7.14 | 6.25 | 9.02 | 10.14 | 10.35 | 7.02 |
| | 0.25 | 8.84 | 9.02 | 12.60 | 11.16 | 3.32 | 8.03 | 7.73 | 10.35 | 10.94 | 11.17 | 7.93 |
| | Median | 10.52 | 9.41 | 13.00 | 11.68 | 4.24 | 8.48 | 8.05 | 10.52 | 11.25 | 11.58 | 8.44 |
| | 0.75 | 12.70 | 9.75 | 13.41 | 12.31 | 5.15 | 8.89 | 8.69 | 10.97 | 11.58 | 11.94 | 9.40 |
| | Max. | 15.04 | 11.27 | 14.36 | 13.85 | 7.67 | 9.76 | 10.22 | 14.11 | 12.40 | 12.78 | 10.41 |
| | Data Quality Objective: +/- 0.5 mg/l | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% |
| QA Qualifier | Estimated, 2 | Rejected | Rejected | Rejected | Valid, 1 | Valid, 1 | Valid, 1 | Estimated, 1 | Estimated, 1 | Estimated, 1 | Valid, 1 | |
| Specific Conductance (µS/cm) | Min. | 434 | 245 | 242 | 425 | 250 | 683 | 415 | 154 | 206 | 119 | 274 |
| | 0.25 | 541 | 517 | 555 | 514 | 720 | 719 | 1250 | 779 | 682 | 402 | 496 |
| | Median | 599 | 736 | 596 | 549 | 810 | 731 | 1297 | 1022 | 879 | 480 | 691 |
| | 0.75 | 627 | 902 | 615 | 572 | 820 | 745 | 1308 | 1233 | 990 | 509 | 701 |
| | Max. | 649 | 1189 | 650 | 588 | 830 | 759 | 1332 | 1521 | 1514 | 649 | 767 |
| | Data Quality Objective: +/- 50 µS/cm | 3 | 19 | 11 | 10 | 10 | 22 | 24 | 14 | 3 | 24 | 10 |
| QA Qualifier | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Valid | Valid | Valid | Estimated, 2 | Estimated, 2 | Estimated, 2 | Valid | |

Appendix D-4: Summary of continuous field measures in San Pablo and Wildcat Creek Watersheds

| | Station | Summer Dry Season | | | | | Late Summer Dry | | | | | |
|------------------------------|---|-------------------|-----------|-----------|-----------|-----------|-----------------|--------------|--------------|--------------|--------------|--------------|
| | | WIL100 | SPA050 | SPA070 | SPA200 | SPA220 | SPA070 | SPA220 | SPA235 | SPA240 | WIL070 | WIL100 |
| | Start Date | 5/31/2002 | 7/18/2002 | 7/18/2002 | 7/3/2002 | 7/3/2002 | 10/18/02 | 10/17/02 | 10/18/02 | 10/18/02 | 10/18/02 | 10/18/02 |
| | End Date | 6/18/2002 | 07/30/02 | 07/30/02 | 7/16/2002 | 7/16/2002 | 10/24/02 | 10/24/02 | 10/24/02 | 10/24/02 | 10/19/02 | 10/24/02 |
| | Data points | 1725 | 1145 | 1147 | 1242 | 1243 | 560 | 560 | 7/16/01 | 567 | 88 | 543 |
| Temperature (°C) | Min. | 12.70 | 15.27 | 14.92 | 14.95 | 14.55 | 12.96 | 12.04 | 11.21 | 11.50 | 12.84 | 10.23 |
| | 0.25 | 14.44 | 16.07 | 15.68 | 16.39 | 15.80 | 13.33 | 12.83 | 12.20 | 12.17 | 13.11 | 11.59 |
| | Median | 15.36 | 16.62 | 16.19 | 17.51 | 16.87 | 13.84 | 13.10 | 12.66 | 12.63 | 13.13 | 11.96 |
| | 0.75 | 16.29 | 17.56 | 16.73 | 18.60 | 17.78 | 14.21 | 13.84 | 13.64 | 13.49 | 13.44 | 12.89 |
| | Max. | 19.21 | 24.52 | 17.61 | 21.27 | 19.68 | 14.89 | 15.22 | 15.62 | 15.61 | 14.03 | 14.39 |
| | HWAT | 16.4 | 16.9 | 16.4 | 18.3 | 17.6 | | | | | | |
| | QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 4.66 | 7.10 | 7.86 | 7.86 | 7.71 | 7.89 | 8.12 | 7.52 | 7.76 | 7.91 | 7.44 |
| | 0.25 | 4.73 | 7.62 | 7.93 | 7.94 | 8.27 | 7.91 | 8.14 | 7.84 | 7.79 | 7.93 | 7.47 |
| | Median | 4.75 | 7.64 | 7.96 | 7.97 | 8.29 | 7.92 | 8.16 | 7.90 | 7.83 | 7.95 | 7.54 |
| | 0.75 | 4.77 | 7.67 | 8.00 | 8.02 | 8.31 | 7.93 | 8.17 | 7.96 | 7.86 | 7.95 | 7.63 |
| | Max. | 4.86 | 8.22 | 8.08 | 8.10 | 8.37 | 7.96 | 8.22 | 8.08 | 7.92 | 7.95 | 7.70 |
| | Data Quality Objective: +/- 0.50 pH units | 1.05 | 0.05 | 0.03 | 0.00 | 0.06 | 0.06 | 0.10 | 0.06 | 0.00 | 0.96 | 0.00 |
| | QA Qualifier | Rejected | Valid | Valid | Valid | Valid | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Rejected | Estimated, 2 |
| Dissolved Oxygen (%) | Min. | 44.20 | 1.40 | 67.70 | 81.60 | 74.80 | 76.10 | 88.00 | 53.50 | 77.50 | 68.70 | 14.70 |
| | 0.25 | 56.70 | 70.70 | 77.60 | 85.00 | 81.80 | 79.68 | 90.50 | 70.90 | 81.10 | 70.68 | 18.15 |
| | Median | 62.70 | 74.90 | 82.50 | 88.15 | 84.80 | 81.50 | 91.75 | 78.00 | 85.90 | 74.00 | 26.00 |
| | 0.75 | 69.40 | 88.20 | 93.20 | 93.90 | 87.80 | 84.10 | 94.03 | 84.55 | 90.40 | 75.53 | 36.50 |
| | Max. | 81.30 | 102.10 | 104.90 | 100.40 | 96.70 | 88.20 | 100.80 | 97.90 | 103.00 | 78.30 | 49.30 |
| | Data Quality Objective: +/- 5.0% | 2.20 | 1.50 | 0.80 | 1.10 | 2.10 | 2.30 | 2.80 | 5.80 | 5.70 | 2.90 | 0.00 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Valid | Valid | rejected | rejected | Valid | Valid |
| Dissolved Oxygen (mg/L) | Min. | 4.64 | 0.13 | 6.53 | 7.51 | 7.00 | 7.80 | 8.85 | 5.67 | 8.21 | 7.06 | 1.61 |
| | 0.25 | 5.74 | 6.92 | 7.69 | 8.13 | 7.88 | 8.18 | 9.46 | 7.45 | 8.65 | 7.45 | 1.97 |
| | Median | 6.25 | 7.34 | 8.09 | 8.42 | 8.26 | 8.45 | 9.67 | 8.17 | 9.11 | 7.76 | 2.86 |
| | 0.75 | 6.76 | 8.34 | 9.06 | 8.87 | 8.58 | 8.65 | 9.84 | 8.83 | 9.40 | 7.91 | 3.78 |
| | Max. | 8.02 | 9.74 | 10.11 | 9.77 | 9.47 | 9.05 | 10.58 | 10.34 | 10.77 | 8.13 | 5.13 |
| | Data Quality Objective: +/- 0.5 mg/l | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% |
| | QA Qualifier | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | rejected | rejected | Valid, 1 | Valid, 1 |
| Specific Conductance (µS/cm) | Min. | 489 | 0 | 1239 | 1404 | 510 | 1290 | 686 | 424 | 578 | 771 | 631 |
| | 0.25 | 502 | 230 | 1352 | 1581 | 737 | 1356 | 718 | 527 | 589 | 783 | 644 |
| | Median | 510 | 260 | 1361 | 1621 | 751 | 1372 | 733 | 592 | 591 | 793 | 675 |
| | 0.75 | 520 | 340 | 1367 | 1645 | 759 | 1394 | 744 | 600 | 595 | 803 | 685 |
| | Max. | 548 | 400 | 1401 | 1668 | 794 | 1416 | 760 | 669 | 601 | 817 | 696 |
| | Data Quality Objective: +/- 50 uS/cm | 23 | 7 | 7 | 2 | 1 | 1 | 10 | 4 | 3 | 2 | 2 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 3 |

Appendix D-5: Summary of continuous field measures in Suisun Creek Watershed

| Station | Winter Wet Season | | | Spring Runoff | | | | | | | |
|------------------------------|---|-----------|-----------|---------------|-----------|-----------|-----------|-----------|-----------|-----------|---------|
| | SUI020 | SUI110 | SUI125 | SUI020 | SUI050 | SUI090 | SUI110 | SUI125 | SUI180 | SUI185 | |
| Start Date | 1/21/2002 | 1/4/2002 | 2/15/2002 | 4/25/2002 | 4/25/2002 | 4/25/2002 | 3/26/2002 | 3/26/2002 | 3/26/2002 | 3/26/2002 | |
| End Date | 2/5/2002 | 1/21/2002 | 3/1/2002 | 5/7/2002 | 5/7/2002 | 5/7/2002 | 4/9/2002 | 4/9/2002 | 4/9/2002 | 3/29/2002 | |
| Data points | 1425 | 1631 | 1335 | 1141 | 1135 | 1141 | 1325 | 1325 | 1325 | 314 | |
| Temperature (°C) | Min. | 6.76 | 5.73 | 8.22 | 12.51 | 11.89 | 11.71 | 9.65 | 11.40 | 8.42 | 9.26 |
| | 0.25 | 8.28 | 7.87 | 10.48 | 14.17 | 13.70 | 13.86 | 12.90 | 13.63 | 12.63 | 11.45 |
| | Median | 8.79 | 10.19 | 11.09 | 15.12 | 14.87 | 15.07 | 14.68 | 14.56 | 13.90 | 13.18 |
| | 0.75 | 9.37 | 11.26 | 12.11 | 16.15 | 16.04 | 16.52 | 17.72 | 16.42 | 15.76 | 14.58 |
| | Max. | 10.81 | 13.31 | 14.71 | 18.78 | 19.85 | 20.07 | 22.48 | 18.66 | 18.33 | 17.40 |
| | HWAT | 9.1 | 11.5 | 12.0 | 15.7 | 15.2 | 16.0 | 16.2 | 15.5 | 14.8 | |
| QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | |
| pH | Min. | 7.62 | 7.80 | 7.84 | 7.40 | 7.85 | 7.58 | 8.00 | 7.70 | 7.72 | 8.15 |
| | 0.25 | 7.66 | 7.98 | 7.93 | 7.44 | 7.95 | 7.70 | 8.04 | 7.78 | 7.98 | 8.21 |
| | Median | 7.68 | 8.05 | 8.00 | 7.47 | 8.07 | 7.79 | 8.11 | 7.90 | 8.03 | 8.23 |
| | 0.75 | 7.72 | 8.10 | 8.19 | 7.51 | 8.34 | 8.03 | 8.29 | 8.16 | 8.26 | 8.35 |
| | Max. | 7.92 | 8.63 | 8.51 | 7.59 | 8.64 | 8.28 | 8.49 | 8.62 | 8.49 | 8.46 |
| | Data Quality Objective: +/- 0.50 pH units | 0.03 | 0.14 | 0.08 | 0.08 | 0.01 | 0.03 | 0.04 | 0.03 | 0.05 | 0.04 |
| QA Qualifier | Valid | Valid, 4 | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid | |
| Dissolved Oxygen (%) | Min. | 93.70 | Reject | 90.90 | 78.90 | 79.70 | 79.90 | 92.10 | 83.70 | 93.30 | 94.20 |
| | 0.25 | 95.30 | Reject | 94.00 | 83.40 | 86.20 | 85.10 | 95.30 | 88.40 | 98.20 | 95.30 |
| | Median | 96.40 | Reject | 96.50 | 86.70 | 92.50 | 90.00 | 97.60 | 91.50 | 101.30 | 96.30 |
| | 0.75 | 100.80 | Reject | 104.90 | 97.30 | 114.65 | 109.20 | 106.60 | 107.20 | 107.20 | 101.78 |
| | Max. | 124.50 | Reject | 122.90 | 107.30 | 135.30 | 128.80 | 119.20 | 124.30 | 114.50 | 110.10 |
| | Data Quality Objective: +/- 5.0% | 1.00 | Reject | 1.50 | 1.10 | 1.40 | 2.50 | 3.20 | 1.30 | 5.00 | 2.20 |
| QA Qualifier | Valid | Reject | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid | |
| Dissolved Oxygen (mg/L) | Min. | 10.58 | Reject | 9.54 | 7.70 | 7.76 | 7.58 | 8.50 | 8.17 | 8.86 | 9.33 |
| | 0.25 | 11.07 | Reject | 10.39 | 8.40 | 8.81 | 8.60 | 9.58 | 8.96 | 9.93 | 9.92 |
| | Median | 11.29 | Reject | 10.62 | 8.83 | 9.44 | 9.26 | 10.08 | 9.45 | 10.58 | 10.35 |
| | 0.75 | 11.71 | Reject | 11.44 | 9.64 | 11.45 | 10.86 | 10.58 | 10.81 | 11.19 | 10.78 |
| | Max. | 14.08 | Reject | 13.22 | 10.86 | 12.95 | 12.45 | 11.67 | 12.22 | 12.34 | 11.69 |
| | Data Quality Objective: +/- 0.5 mg/l | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% |
| QA Qualifier | Valid, 1 | Reject | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | |
| Specific Conductance (µS/cm) | Min. | 393 | 162 | 386 | 456 | 475 | 459 | 396 | 422 | 208 | 569 |
| | 0.25 | 412 | 285 | 398 | 477 | 482 | 478 | 427 | 437 | 222 | 573 |
| | Median | 419 | 340 | 412 | 481 | 486 | 482 | 444 | 445 | 230 | 575 |
| | 0.75 | 423 | 373 | 420 | 485 | 488 | 484 | 448 | 452 | 234 | 577 |
| | Max. | 435 | 390 | 434 | 493 | 493 | 490 | 464 | 458 | 251 | 580 |
| | Data Quality Objective: +/- 50 µS/cm | 24 | 14 | 16 | 2 | 6 | 6 | 1 | 2 | 5 | 2 |
| QA Qualifier | Valid | Valid, 4 | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid | |

Appendix D-5: Summary of continuous field measures in Suisun Creek Watershed

| Station | Summer Dry Season | | | | | | Late Summer | | | | | |
|------------------------------|---|-----------|-----------|-----------|-----------|-----------|-------------|--------------|--------------|--------------|------------|----------|
| | SUI020 | SUI050 | SUI090 | SUI110 | SUI125 | SUI185 | SUI020 | SUI050 | SUI090 | SUI125 | SUI185 | |
| Start Date | 7/3/2002 | 7/3/2002 | 7/3/2002 | 6/20/2002 | 6/20/2002 | 6/20/2002 | 10/24/2002 | 10/24/2002 | 10/23/2002 | 10/24/2002 | 10/24/2002 | |
| End Date | 7/16/2002 | 7/16/2002 | 7/16/2002 | 7/1/2002 | 7/1/2002 | 7/1/2002 | 10/30/2002 | 10/30/2002 | 10/29/2002 | 10/30/2002 | 10/30/2002 | |
| Data points | 1248 | 1246 | 1245 | 1050 | 1048 | 1050 | 567 | 570 | 564 | 562 | 561 | |
| Temperature (°C) | Min. | 17.44 | 17.68 | 17.84 | 17.02 | 15.29 | 15.30 | 12.67 | 11.11 | 11.81 | 12.29 | 9.64 |
| | 0.25 | 18.92 | 20.14 | 20.19 | 19.77 | 17.72 | 17.39 | 13.55 | 12.20 | 12.98 | 13.50 | 10.62 |
| | Median | 20.11 | 21.74 | 21.16 | 22.03 | 19.07 | 19.11 | 13.85 | 12.76 | 13.48 | 14.25 | 11.42 |
| | 0.75 | 21.03 | 23.09 | 21.99 | 24.80 | 20.23 | 20.47 | 14.20 | 13.35 | 13.92 | 14.76 | 12.01 |
| | Max. | 24.21 | 26.50 | 23.89 | 29.32 | 22.40 | 24.03 | 14.74 | 14.35 | 14.67 | 15.56 | 12.82 |
| | HWAT | 20.9 | 22.7 | 21.7 | 22.9 | 19.5 | 19.8 | | | | | |
| | QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 7.06 | 7.59 | 6.68 | 7.45 | 7.75 | 7.99 | 6.80 | 7.43 | 7.35 | 7.64 | 7.99 |
| | 0.25 | 7.48 | 7.76 | 7.46 | 7.54 | 7.79 | 8.07 | 7.00 | 7.85 | 7.46 | 7.71 | 8.02 |
| | Median | 7.61 | 7.85 | 7.54 | 7.67 | 7.86 | 8.16 | 7.37 | 7.88 | 7.48 | 7.74 | 8.03 |
| | 0.75 | 7.68 | 7.98 | 7.59 | 7.95 | 8.08 | 8.40 | 7.43 | 7.92 | 7.53 | 7.80 | 8.06 |
| | Max. | 7.93 | 8.15 | 7.72 | 8.41 | 8.23 | 8.62 | 7.50 | 8.06 | 7.62 | 7.93 | 8.10 |
| | Data Quality Objective: +/- 0.50 pH units | 0.03 | 0.03 | 0.07 | 0.00 | 0.06 | 0.07 | 0.05 | 0.06 | 0.10 | 0.06 | 0.04 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Valid | Estimated, 2 | Estimated, 2 | Estimated, 2 | Valid | Valid |
| Dissolved Oxygen (%) | Min. | 65.20 | 55.80 | 50.30 | 43.80 | 82.00 | 73.00 | 47.80 | 85.10 | 85.20 | 78.30 | 86.00 |
| | 0.25 | 78.50 | 68.30 | 76.80 | 58.93 | 85.20 | 82.70 | 70.20 | 88.40 | 86.90 | 86.70 | 88.00 |
| | Median | 83.65 | 79.55 | 80.30 | 83.45 | 90.75 | 87.85 | 78.60 | 92.00 | 89.15 | 88.20 | 89.20 |
| | 0.75 | 101.80 | 96.15 | 91.70 | 148.78 | 109.00 | 100.60 | 86.40 | 96.40 | 94.33 | 94.93 | 92.40 |
| | Max. | 117.40 | 112.10 | 100.70 | 205.10 | 121.40 | 112.60 | 104.40 | 106.10 | 101.40 | 103.30 | 96.00 |
| | Data Quality Objective: +/- 5.0% | 1.00 | 2.30 | 1.00 | 1.30 | 1.80 | 3.30 | 4.30 | 4.50 | 2.10 | 0.30 | 0.20 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid |
| Dissolved Oxygen (mg/L) | Min. | 6.06 | 4.95 | 4.42 | 3.90 | 7.36 | 6.62 | 5.06 | 9.13 | 8.83 | 8.25 | 9.37 |
| | 0.25 | 7.17 | 6.15 | 6.86 | 5.39 | 7.93 | 7.77 | 7.26 | 9.49 | 9.12 | 8.87 | 9.67 |
| | Median | 7.76 | 7.06 | 7.22 | 7.27 | 8.60 | 8.29 | 8.11 | 9.75 | 9.35 | 9.15 | 9.83 |
| | 0.75 | 9.03 | 8.18 | 8.12 | 12.47 | 10.14 | 9.17 | 8.87 | 10.06 | 9.78 | 9.79 | 10.05 |
| | Max. | 10.30 | 9.43 | 8.88 | 16.26 | 11.14 | 10.43 | 10.61 | 10.93 | 10.43 | 10.35 | 10.41 |
| | Data Quality Objective: +/- 0.5 mg/l | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% |
| | QA Qualifier | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 |
| Specific Conductance (µS/cm) | Min. | 363 | 482 | 442 | 558 | 421 | 486 | 364 | 471 | 430 | 414 | 522 |
| | 0.25 | 379 | 485 | 445 | 580 | 423 | 492 | 378 | 472 | 432 | 416 | 526 |
| | Median | 393 | 487 | 447 | 588 | 424 | 496 | 385 | 473 | 433 | 417 | 527 |
| | 0.75 | 412 | 487 | 451 | 593 | 425 | 500 | 393 | 475 | 433 | 417 | 529 |
| | Max. | 449 | 492 | 711 | 607 | 426 | 507 | 417 | 476 | 435 | 427 | 533 |
| | Data Quality Objective: +/- 50 uS/cm | 5 | 4 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Valid |

Appendix D-6: Summary of continuous field measures in Arroyo Las Positas Creek Watershed

| | Station | Winter Wet | | | Spring Wet | | | | |
|------------------------------|---|-----------------|-----------------|-----------------|--------------|--------------|--------------|--------------|--------------|
| | | ALP010 | ALP040 | ALP150 | ALP010 | ALP040 | ALP080 | ALP105 | ALP150 |
| | Start Date | 1/4/2002 | 1/18/2002 | 2/15/2002 | 4/25/2002 | 4/25/2002 | 4/25/2002 | 3/26/2002 | 3/26/2002 |
| | End date | 1/18/2002 | 2/5/2002 | 3/1/2002 | 5/7/02 | 5/7/2002 | 5/7/2002 | 4/9/2002 | 4/9/2002 |
| | Data points | 1340 | 1720 | 1337 | 1146 | 1146 | 1140 | 1343 | 1345 |
| Temperature (°C) | Min. | 5.35 | 3.89 | 7.93 | 13.16 | 11.54 | 12.50 | 11.23 | 11.34 |
| | 0.25 | 8.39 | 6.37 | 11.04 | 14.48 | 13.92 | 14.71 | 13.93 | 15.00 |
| | Median | 10.35 | 7.62 | 12.60 | 15.59 | 15.34 | 15.91 | 15.41 | 16.72 |
| | 0.75 | 11.61 | 8.69 | 14.58 | 17.19 | 18.13 | 17.74 | 19.01 | 19.27 |
| | Max. | 13.38 | 11.44 | 18.34 | 19.39 | 24.96 | 21.06 | 24.98 | 24.53 |
| | HWAT | 11.5 | 8.0 | 13.9 | 16.4 | 17.0 | 16.9 | 17.2 | 18.2 |
| | QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 8.13 | 8.08 | 7.60 | 8.24 | 8.08 | 7.89 | 7.88 | 7.93 |
| | 0.25 | 8.31 | 8.44 | 8.01 | 8.42 | 8.33 | 8.03 | 8.06 | 8.42 |
| | Median | 8.40 | 8.53 | 8.26 | 8.47 | 8.46 | 8.17 | 8.18 | 8.64 |
| | 0.75 | 8.48 | 8.61 | 8.46 | 8.50 | 8.62 | 8.30 | 8.42 | 8.77 |
| | Max. | 8.61 | 8.79 | 9.24 | 8.56 | 8.80 | 8.41 | 8.68 | 8.95 |
| | Data Quality Objective: +/- 0.50 pH units | 0.14 | 0.14 | 0.24 | 0.03 | 0.03 | 0.04 | 0.02 | 0.06 |
| | QA Qualifier | Estimated, 2, 3 | Estimated, 2, 3 | Estimated, 2, 3 | Estimated, 3 | Estimated, 3 | Estimated, 3 | Estimated, 3 | Estimated, 3 |
| Dissolved Oxygen (%) | Min. | 96.60 | 84.60 | 49.20 | 88.50 | 57.50 | 70.40 | 32.30 | 20.20 |
| | 0.25 | 98.50 | 97.00 | 81.10 | 96.80 | 80.40 | 80.00 | 60.25 | 93.50 |
| | Median | 101.60 | 106.10 | 96.80 | 102.90 | 106.25 | 89.90 | 90.90 | 124.80 |
| | 0.75 | 105.30 | 146.03 | 140.40 | 107.68 | 166.05 | 124.30 | 182.65 | 185.25 |
| | Max. | 115.20 | 212.10 | 218.10 | 119.20 | 234.90 | 174.20 | 395.10 | 266.50 |
| | Data Quality Objective: +/- 5.0% | 5.90 | 5.90 | 2.00 | 0.30 | 0.70 | 2.30 | 0.60 | 47.50 |
| | QA Qualifier | Rejected | Rejected | Estimated, 2 | Valid | Valid | Valid | Valid | Rejected |
| Dissolved Oxygen (mg/L) | Min. | 10.21 | 10.06 | 5.62 | 8.49 | 5.71 | 6.44 | 3.27 | 1.73 |
| | 0.25 | 10.69 | 11.89 | 8.78 | 9.54 | 8.35 | 7.97 | 6.18 | 9.34 |
| | Median | 11.38 | 12.77 | 10.55 | 10.11 | 10.69 | 9.06 | 8.98 | 12.02 |
| | 0.75 | 12.11 | 17.11 | 14.58 | 10.61 | 15.83 | 12.08 | 16.82 | 17.43 |
| | Max. | 14.06 | 23.20 | 20.78 | 11.75 | 19.75 | 16.51 | 36.42 | 23.92 |
| | Data Quality Objective: +/- 0.5 mg/l | NA | NA | NA | NA | NA | NA | NA | NA |
| | QA Qualifier | Rejected, 1 | Rejected, 1 | Estimated, 1, 2 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Rejected, 1 |
| Specific Conductance (µS/cm) | Min. | 1351 | 857 | 157 | 1137 | 1138 | 1013 | 2315 | 813 |
| | 0.25 | 1884 | 1554 | 829 | 1379 | 1490 | 1491 | 2652 | 949 |
| | Median | 2015 | 1705 | 928 | 1550 | 1560 | 1571 | 2803 | 990 |
| | 0.75 | 2035 | 1826 | 953 | 1599 | 1626 | 1681 | 2937 | 1060 |
| | Max. | 2078 | 2052 | 1033 | 1862 | 1983 | 2092 | 3354 | 1158 |
| | Data Quality Objective: +/- 50 uS/cm | 22 | 22 | 13 | 7 | 11 | 3 | 0 | 24 |
| | QA Qualifier | Estimated, 2 | Estimated, 2 | Estimated, 2 | Valid | Valid | Valid | Valid | Valid |

Appendix D-6: Summary of continuous field measures in Arroyo Las Positas Creek Watershed

| | | Summer Dry | | | | |
|------------------------------|---|------------|-----------|-----------------|-------------|-------------|
| | | ALP010 | ALP040 | ALP080 | ALP105 | ALP150 |
| | Station | | | | | |
| | Start Date | 7/18/2002 | 7/18/2002 | 7/18/2002 | 6/20/2002 | 6/20/2002 |
| | End date | 7/30/2002 | 7/30/2002 | 7/30/2002 | 7/1/2002 | 7/1/2002 |
| | Data points | 1149 | 1149 | 1150 | 1048 | 1049 |
| Temperature (°C) | Min. | 19.79 | 17.91 | 18.53 | 15.15 | 17.41 |
| | 0.25 | 20.71 | 19.89 | 19.88 | 18.27 | 19.89 |
| | Median | 21.55 | 21.44 | 21.43 | 20.78 | 21.85 |
| | 0.75 | 22.47 | 24.25 | 23.33 | 24.39 | 24.43 |
| | Max. | 24.14 | 27.35 | 25.22 | 32.10 | 29.27 |
| | HWAT | 21.5 | 21.9 | 21.5 | 22.2 | 23.1 |
| | QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 8.00 | 7.77 | 7.63 | 7.58 | 7.62 |
| | 0.25 | 8.19 | 7.89 | 7.87 | 7.97 | 7.79 |
| | Median | 8.26 | 8.02 | 8.11 | 8.17 | 8.19 |
| | 0.75 | 8.39 | 8.17 | 8.45 | 8.39 | 8.54 |
| | Max. | 8.52 | 8.75 | 8.69 | 8.87 | 8.99 |
| | Data Quality Objective: +/- 0.50 pH units | 0.01 | 0.03 | 0.06 | 0.09 | 0.04 |
| | QA Qualifier | Valid | Valid | Estimated, 2 | Valid | Valid |
| Dissolved Oxygen (%) | Min. | 81.60 | 42.60 | 56.80 | 6.00 | 6.30 |
| | 0.25 | 90.70 | 66.00 | 77.40 | 30.60 | 29.75 |
| | Median | 97.60 | 102.20 | 88.40 | 74.80 | 77.90 |
| | 0.75 | 104.80 | 169.60 | 122.30 | 148.50 | 158.78 |
| | Max. | 112.90 | 221.70 | 151.10 | 256.70 | 190.10 |
| | Data Quality Objective: +/- 5.0% | 1.60 | 3.90 | 0.80 | 0.40 | 5.40 |
| | QA Qualifier | Valid | Valid | Estimated, 2 | Valid | Rejected |
| Dissolved Oxygen (mg/L) | Min. | 7.06 | 3.98 | 4.96 | 0.56 | 0.59 |
| | 0.25 | 8.09 | 6.02 | 6.91 | 2.85 | 2.81 |
| | Median | 8.59 | 9.13 | 7.78 | 6.79 | 7.13 |
| | 0.75 | 9.09 | 14.25 | 10.63 | 12.35 | 13.47 |
| | Max. | 9.75 | 18.23 | 12.92 | 20.59 | 16.22 |
| | Data Quality Objective: +/- 0.5 mg/l | NA | NA | NA | NA | NA |
| | QA Qualifier | Valid, 1 | Valid, 1 | Estimated, 1, 2 | Valid, 1 | Rejected, 1 |
| Specific Conductance (µS/cm) | Min. | 984 | 1013 | 1050 | 835 | 824 |
| | 0.25 | 1087 | 1078 | 1147 | 2222 | 873 |
| | Median | 1134 | 1138 | 1204 | 2469 | 902 |
| | 0.75 | 1212 | 1209 | 1268 | 2559 | 929 |
| | Max. | 1586 | 1403 | 1515 | 3873 | 1048 |
| | Data Quality Objective: +/- 50 uS/cm | 7 | 5 | 5 | UNK | UNK |
| | QA Qualifier | Valid | Valid | Estimated, 2 | Rejected, 8 | Rejected, 8 |

Appendix D-7: Summary of continuous field measures in Pescadero and Butano Creek Watersheds

| | Station | Summer Dry Season | | | | | Late Summer | | | | |
|------------------------------|---|-------------------|-----------|-----------|-----------|-----------|--------------|--------------|--------------|--------------|--------------|
| | | PES060 | PES100 | PES140 | PES150 | PES170 | BUT010 | BUT050 | PES180 | PES190 | PES240 |
| | Start Date | 8/26/2002 | 8/26/2002 | 8/26/2002 | 8/26/2002 | 8/26/2002 | 10/9/2002 | 10/9/2002 | 10/9/2002 | 10/8/2002 | 10/9/2002 |
| | End date | 09/12/02 | 9/12/2002 | 9/12/2002 | 9/12/2002 | 9/12/2002 | 10/15/2002 | 10/15/2002 | 10/15/2002 | 10/14/2002 | 10/15/2002 |
| | Data points | 1630 | 1616 | 1627 | 1627 | 1627 | 572 | 571 | 574 | 571 | 573 |
| Temperature (°C) | Min. | 12.74 | 13.64 | 12.66 | 11.28 | 11.66 | 9.46 | 10.75 | 10.70 | 11.06 | 10.95 |
| | 0.25 | 15.50 | 15.92 | 14.54 | 12.79 | 12.97 | 11.88 | 11.48 | 11.60 | 11.84 | 11.77 |
| | Median | 16.45 | 16.97 | 15.31 | 13.27 | 13.48 | 12.94 | 12.01 | 12.37 | 12.56 | 12.75 |
| | 0.75 | 17.82 | 18.30 | 16.16 | 13.73 | 14.03 | 13.92 | 12.83 | 13.05 | 13.22 | 13.31 |
| | Max. | 19.85 | 20.68 | 17.58 | 14.88 | 15.07 | 16.18 | 13.31 | 13.73 | 13.88 | 14.08 |
| | HWAT | 17.1 | 17.8 | 16.1 | 13.8 | 14.1 | | | | | |
| | QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 7.56 | 7.82 | 8.08 | 7.45 | 7.86 | 7.90 | 7.49 | 7.62 | 7.43 | 7.59 |
| | 0.25 | 7.66 | 7.93 | 8.24 | 7.82 | 7.89 | 7.94 | 7.52 | 8.34 | 7.76 | 7.69 |
| | Median | 7.69 | 7.97 | 8.26 | 7.99 | 7.91 | 7.95 | 7.53 | 8.35 | 7.84 | 7.70 |
| | 0.75 | 7.73 | 8.04 | 8.28 | 8.10 | 7.92 | 8.00 | 7.54 | 8.36 | 7.91 | 7.70 |
| | Max. | 7.86 | 8.29 | 8.32 | 8.13 | 8.00 | 8.07 | 7.56 | 8.42 | 8.00 | 7.73 |
| | Data Quality Objective: +/- 0.50 pH units | 0.03 | 0.04 | 0.02 | 0.06 | 0.03 | 0.00 | 0.06 | 0.06 | 0.10 | 0.05 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 |
| Dissolved Oxygen (%) | Min. | 71.90 | 73.10 | 88.50 | 86.60 | 73.90 | 8.70 | 88.70 | 91.90 | 83.50 | 75.50 |
| | 0.25 | 78.70 | 77.60 | 92.40 | 89.85 | 77.50 | 88.20 | 89.70 | 94.10 | 87.40 | 84.70 |
| | Median | 81.70 | 84.30 | 94.70 | 90.60 | 79.30 | 92.30 | 90.60 | 95.50 | 90.80 | 85.60 |
| | 0.75 | 88.38 | 101.60 | 97.50 | 91.70 | 81.90 | 94.30 | 91.10 | 98.20 | 97.30 | 86.60 |
| | Max. | 101.80 | 135.00 | 104.90 | 94.10 | 94.00 | 138.00 | 93.80 | 103.40 | 106.10 | 87.60 |
| | Data Quality Objective: +/- 5.0% | 0.70 | 0.80 | 0.60 | 0.40 | 1.80 | 93.00 | 1.90 | 0.60 | 0.90 | 1.50 |
| | QA Qualifier | Valid, 4 | Valid, 4 | Valid, 4 | Valid, 4 | Valid, 4 | Rejected | Valid | Valid | Valid | Valid |
| Dissolved Oxygen (mg/L) | Min. | 6.85 | 7.13 | 8.76 | 8.90 | 7.70 | 0.91 | 9.32 | 9.62 | 8.76 | 7.76 |
| | 0.25 | 7.68 | 7.66 | 9.17 | 9.30 | 8.05 | 9.16 | 9.50 | 9.96 | 9.41 | 8.87 |
| | Median | 8.05 | 8.15 | 9.54 | 9.48 | 8.26 | 9.63 | 9.79 | 10.30 | 9.65 | 9.05 |
| | 0.75 | 8.61 | 9.66 | 9.74 | 9.67 | 8.47 | 10.05 | 9.91 | 10.52 | 10.26 | 9.33 |
| | Max. | 9.62 | 12.33 | 10.52 | 10.24 | 9.57 | 15.26 | 10.12 | 11.08 | 11.00 | 9.60 |
| | Data Quality Objective: +/- 0.5 mg/l | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% |
| | QA Qualifier | Valid, 4 | Valid, 4 | Valid, 4 | Valid, 4 | Valid, 4 | Rejected | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 |
| Specific Conductance (µS/cm) | Min. | 780 | 800 | 745 | 731 | 1414 | 446 | 365 | 847 | 701 | 817 |
| | 0.25 | 790 | 818 | 749 | 740 | 1456 | 465 | 366 | 849 | 705 | 819 |
| | Median | 794 | 824 | 758 | 743 | 1553 | 469 | 366 | 850 | 707 | 820 |
| | 0.75 | 799 | 836 | 769 | 748 | 1572 | 473 | 367 | 852 | 709 | 821 |
| | Max. | 808 | 849 | 777 | 755 | 1589 | 498 | 368 | 854 | 717 | 822 |
| | Data Quality Objective: +/- 50 uS/cm | 0 | 2 | 0 | 2 | 0 | 2 | 1 | 4 | 10 | 3 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 |

Appendix D-7: Summary of continuous field measures in Pescadero and Butano Creek Watersheds

| | | Winter Wet Season | | | | | Spring Runoff Season | | | | |
|------------------------------|---|-------------------|-----------|-----------|-----------|-----------|----------------------|--------------|--------------|-------------|--------------|
| Station | | BUT010 | BUT030 | PES050 | PES140 | PES150 | PES105 | BUT010 | BUT030 | PES050 | |
| Start Date | | 2/13/2003 | 2/13/2003 | 2/13/2003 | 2/13/2003 | 2/13/2003 | 6/5/2003 | 6/4/2003 | 6/5/2003 | 06/05/03 | |
| End date | | 02/27/03 | 2/27/2003 | 2/27/2003 | 2/27/2003 | 2/27/2003 | 6/12/2003 | 6/11/2003 | 6/12/2003 | 06/07/03 | |
| Data points | | 721 | 1345 | 1350 | 1328 | 1327 | 661 | 661 | 663 | 210.00 | |
| Temperature (°C) | Min. | 9.09 | 8.17 | 8.08 | 7.50 | 8.27 | 13.51 | 13.09 | 12.44 | 16.27 | |
| | 0.25 | 10.41 | 9.33 | 9.66 | 8.57 | 9.37 | 15.28 | 14.20 | 13.82 | 16.72 | |
| | Median | 11.00 | 9.77 | 10.35 | 9.32 | 9.72 | 16.08 | 14.69 | 14.35 | 17.33 | |
| | 0.75 | 11.56 | 10.25 | 10.90 | 9.95 | 9.98 | 16.85 | 15.49 | 15.02 | 17.94 | |
| | Max. | 12.92 | 10.96 | 11.69 | 10.83 | 10.69 | 18.40 | 16.52 | 16.24 | 18.41 | |
| | HWAT | | 9.9 | 9.9 | 10.2 | 9.4 | 9.7 | 15.9 | 14.8 | 14.3 | |
| | QA Qualifier | | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 7.02 | 7.44 | 7.65 | 7.88 | 7.22 | 8.15 | 7.29 | 0.09 | 7.73 | |
| | 0.25 | 7.25 | 7.45 | 7.84 | 8.06 | 7.82 | 8.17 | 7.41 | 0.11 | 7.75 | |
| | Median | 7.37 | 7.47 | 7.85 | 8.10 | 7.97 | 8.20 | 7.50 | 0.14 | 7.80 | |
| | 0.75 | 7.48 | 7.53 | 7.87 | 8.14 | 8.00 | 8.24 | 7.54 | 0.14 | 7.93 | |
| | Max. | 7.57 | 7.68 | 8.05 | 8.26 | 8.08 | 8.30 | 7.60 | 0.15 | 8.01 | |
| | Data Quality Objective: +/- 0.50 pH units | | 0.07 | 0.04 | 0.12 | 0.03 | 0.18 | 0.08 | 0.12 | NA | 0.08 |
| | QA Qualifier | | Valid | Valid | Valid | Valid | Valid | Estimated, 2 | Estimated, 2 | Rejected, 9 | Estimated, 2 |
| Dissolved Oxygen (%) | Min. | 65.30 | 95.40 | 88.70 | 99.10 | 97.50 | 96.90 | 57.00 | 7.66 | 83.70 | |
| | 0.25 | 93.70 | 96.50 | 95.00 | 100.50 | 100.40 | 97.70 | 75.10 | 7.69 | 84.53 | |
| | Median | 95.30 | 97.70 | 96.05 | 101.40 | 101.90 | 98.90 | 89.10 | 7.71 | 89.55 | |
| | 0.75 | 96.60 | 98.90 | 97.90 | 103.30 | 103.10 | 101.80 | 90.00 | 7.74 | 104.90 | |
| | Max. | 100.50 | 101.00 | 110.40 | 108.60 | 106.50 | 108.70 | 94.70 | 7.80 | 113.70 | |
| | Data Quality Objective: +/- 5.0% | | 1.20 | 0.00 | 0.80 | 1.10 | 0.40 | 2.00 | 3.60 | NA | NA |
| | QA Qualifier | | Valid | Valid | Valid | Valid | Valid | Valid | Valid | Rejected, 9 | Rejected, 9 |
| Dissolved Oxygen (mg/L) | Min. | 7.30 | 10.62 | 9.89 | 11.06 | 11.09 | 9.22 | 5.86 | 9.31 | 7.94 | |
| | 0.25 | 10.24 | 10.89 | 10.63 | 11.41 | 11.39 | 9.56 | 7.84 | 9.48 | 8.14 | |
| | Median | 10.48 | 11.10 | 10.80 | 11.71 | 11.58 | 9.87 | 8.88 | 9.71 | 8.65 | |
| | 0.75 | 10.65 | 11.29 | 11.04 | 11.93 | 11.75 | 10.16 | 9.08 | 9.97 | 10.01 | |
| | Max. | 11.28 | 11.73 | 12.13 | 12.61 | 12.49 | 10.61 | 9.31 | 10.31 | 10.77 | |
| | Data Quality Objective: +/- 0.5 mg/l | | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% | see DO% |
| | QA Qualifier | | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Valid, 1 | Rejected, 9 | Rejected, 9 |
| Specific Conductance (µS/cm) | Min. | 190 | 186 | 333 | 381 | 132 | 593 | 212 | 338 | 608 | |
| | 0.25 | 253 | 192 | 403 | 479 | 375 | 595 | 225 | 343 | 614 | |
| | Median | 257 | 196 | 435 | 512 | 407 | 596 | 282 | 344 | 616 | |
| | 0.75 | 332 | 219 | 444 | 521 | 429 | 597 | 305 | 345 | 618 | |
| | Max. | 373 | 340 | 583 | 572 | 506 | 599 | 352 | 346 | 620 | |
| | Data Quality Objective: +/- 50 uS/cm | | 1 | 4 | 18 | 7 | 3 | 15 | 0 | NA | 4 |
| | QA Qualifier | | Valid | Valid | Valid | Valid | Valid | Estimated, 2 | Estimated, 2 | Rejected, 9 | Estimated, 2 |

Appendix D-8: Summary of continuous field measures in San Gregorio Creek Watershed

| | Station | Late Summer | | | | | Winter Wet Season | | | | |
|------------------------------|---|--------------|--------------|--------------|--------------|--------------|-------------------|--------------|--------------|--------------|--------------|
| | | SGR010 | SGR040 | SGR090 | SGR100 | SGR120 | SGR010 | SGR020 | SGR090 | SGR100 | SGR120 |
| | Start Date | 9/26/2002 | 9/26/2002 | 9/26/2002 | 9/26/2002 | 9/26/2002 | 1/24/2003 | 1/24/2003 | 1/24/2003 | 1/24/2003 | 1/24/2003 |
| | End date | 10/4/2002 | 10/4/2002 | 10/4/2002 | 10/4/2002 | 10/4/2002 | 2/7/2003 | 2/7/2003 | 2/7/2003 | 2/7/2003 | 2/7/2003 |
| | Data points | 753 | 758 | 761 | 768 | 775 | 1351 | 1916 | 1331 | 1334 | 1326 |
| Temperature (°C) | Min. | 9.95 | 9.43 | 8.85 | 9.24 | 10.01 | 5.40 | 3.18 | 4.95 | 4.92 | 7.36 |
| | 0.25 | 12.69 | 12.41 | 11.26 | 11.81 | 10.87 | 8.44 | 6.90 | 7.68 | 7.99 | 8.85 |
| | Median | 14.31 | 14.21 | 12.72 | 13.36 | 11.85 | 10.69 | 9.13 | 9.86 | 10.24 | 10.21 |
| | 0.75 | 15.16 | 15.14 | 13.77 | 14.34 | 12.68 | 12.26 | 11.22 | 11.32 | 11.63 | 10.81 |
| | Max. | 16.09 | 17.12 | 14.72 | 15.36 | 13.17 | 13.89 | 14.26 | 12.26 | 12.94 | 11.33 |
| | HWAT | 14.2 | 14.0 | 12.6 | 13.2 | 11.9 | 11.8 | 11.3 | 10.8 | 11.2 | 10.6 |
| | QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 7.99 | 8.09 | 7.81 | 7.73 | 7.15 | 7.53 | 7.54 | 8.10 | 7.86 | 7.80 |
| | 0.25 | 8.05 | 8.15 | 8.12 | 8.23 | 7.31 | 7.86 | 8.29 | 8.19 | 8.11 | 7.88 |
| | Median | 8.08 | 8.19 | 8.18 | 8.25 | 7.38 | 7.89 | 8.31 | 8.20 | 8.14 | 7.89 |
| | 0.75 | 8.13 | 8.21 | 8.21 | 8.28 | 7.42 | 7.92 | 8.39 | 8.22 | 8.17 | 7.90 |
| | Max. | 8.19 | 8.30 | 8.32 | 8.36 | 7.46 | 7.98 | 8.50 | 8.27 | 8.23 | 7.91 |
| | Data Quality Objective: +/- 0.50 pH units | 0.10 | 0.06 | 0.08 | 0.02 | 0.10 | 0.06 | 0.18 | 0.02 | 0.05 | 0.05 |
| | QA Qualifier | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 |
| Dissolved Oxygen (%) | Min. | 82.70 | 86.00 | 78.60 | 91.10 | 25.50 | 96.30 | 93.70 | 98.60 | 98.10 | 96.20 |
| | 0.25 | 86.70 | 90.10 | 86.20 | 94.68 | 64.35 | 97.70 | 98.50 | 100.00 | 98.80 | 98.00 |
| | Median | 89.20 | 91.60 | 87.30 | 96.00 | 70.10 | 99.20 | 100.00 | 101.00 | 99.50 | 99.80 |
| | 0.75 | 97.30 | 95.18 | 89.10 | 97.60 | 72.70 | 101.10 | 102.30 | 102.30 | 100.20 | 101.60 |
| | Max. | 110.40 | 103.50 | 95.70 | 101.50 | 76.90 | 110.10 | 116.80 | 105.10 | 102.80 | 103.70 |
| | Data Quality Objective: +/- 5.0% | 0.70 | 0.00 | 1.10 | 2.40 | 1.20 | 0.50 | 4.80 | 2.70 | 0.00 | 1.20 |
| | QA Qualifier | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Valid | Estimated, 2 | Estimated, 2 | Estimated, 2 |
| Dissolved Oxygen (mg/L) | Min. | 8.19 | 8.47 | 8.14 | 9.18 | 2.73 | 9.99 | 10.15 | 10.58 | 10.38 | 10.61 |
| | 0.25 | 8.92 | 9.18 | 9.06 | 9.67 | 6.95 | 10.53 | 10.90 | 10.93 | 10.75 | 10.92 |
| | Median | 9.37 | 9.52 | 9.34 | 10.09 | 7.58 | 11.03 | 11.54 | 11.40 | 11.16 | 11.18 |
| | 0.75 | 10.02 | 9.96 | 9.67 | 10.47 | 7.98 | 11.82 | 12.27 | 12.20 | 11.84 | 11.77 |
| | Max. | 11.27 | 10.84 | 10.52 | 11.27 | 8.42 | 13.28 | 14.23 | 13.26 | 12.95 | 12.45 |
| | Data Quality Objective: +/- 0.5 mg/l | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| | QA Qualifier | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Valid, 1 | Estimated, 2 | Estimated, 2 | Estimated, 2 |
| Specific Conductance (µS/cm) | Min. | 1000 | 996 | 867 | 1060 | 696 | 749 | 495 | 778 | 776 | 384 |
| | 0.25 | 1010 | 1000 | 870 | 1066 | 710 | 762 | 508 | 792 | 787 | 410 |
| | Median | 1018 | 1001 | 873 | 1072 | 716 | 776 | 530 | 802 | 802 | 421 |
| | 0.75 | 1032 | 1003 | 875 | 1078 | 731 | 790 | 552 | 812 | 817 | 432 |
| | Max. | 1072 | 1007 | 884 | 1088 | 754 | 806 | 566 | 824 | 834 | 444 |
| | Data Quality Objective: +/- 0.05 mS/cm | 5 | 1 | 2 | 4 | 1 | 14 | 3 | 11 | 7 | 20 |
| | QA Qualifier | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 |

Appendix D-8: Summary of continuous field measures in San Gregorio Creek Watershed

| | | Summer Dry Season | | | | |
|------------------------------|---|-------------------|----------------|-----------------|----------------|----------------|
| | | SGR010 | SGR020 | SGR040 | SGR090 | SGR100 |
| | Station | SGR010 | SGR020 | SGR040 | SGR090 | SGR100 |
| | Start Date | 6/11/2003 | 6/12/2003 | 6/12/2003 | 6/12/2003 | 6/12/2003 |
| | End date | 6/18/2003 | 6/20/2003 | 06/20/03 | 6/20/2003 | 6/19/2003 |
| | Data points | 688 | 761 | 760 | 761 | 663 |
| Temperature (°C) | Min. | 13.14 | 11.08 | 12.01 | 11.39 | 11.30 |
| | 0.25 | 14.20 | 12.81 | 13.32 | 12.47 | 12.58 |
| | Median | 15.04 | 13.45 | 14.15 | 13.03 | 13.28 |
| | 0.75 | 16.29 | 14.11 | 15.19 | 13.97 | 14.60 |
| | Max. | 19.16 | 16.99 | 19.55 | 17.59 | 17.39 |
| | HWAT | 15.3 | 13.7 | 14.6 | 13.4 | 13.6 |
| | QA Qualifier | Valid, 5 | Rejected, 6, 9 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 7.98 | 8.11 | 8.26 | 8.24 | 8.10 |
| | 0.25 | 8.00 | 8.15 | 8.29 | 8.28 | 8.14 |
| | Median | 8.03 | 8.17 | 8.32 | 8.31 | 8.17 |
| | 0.75 | 8.06 | 8.21 | 8.37 | 8.37 | 8.23 |
| | Max. | 8.22 | 8.25 | 8.43 | 8.47 | 8.31 |
| | Data Quality Objective: +/- 0.50 pH units | 0.12 | no record | 0.09 | 0.08 | 0.08 |
| | QA Qualifier | Estimated, 2 | Rejected, 6, 9 | Estimated, 2 | Estimated, 2 | Estimated, 2 |
| Dissolved Oxygen (%) | Min. | 89.10 | 94.00 | 89.70 | 93.40 | 93.60 |
| | 0.25 | 91.30 | 96.50 | 91.40 | 94.50 | 95.00 |
| | Median | 92.40 | 97.30 | 92.35 | 95.10 | 95.60 |
| | 0.75 | 96.30 | 98.40 | 95.90 | 98.10 | 97.40 |
| | Max. | 103.70 | 101.20 | 100.10 | 102.00 | 100.50 |
| | Data Quality Objective: +/- 5.0% | 0.09 | no record | 1.40 | 0.60 | 1.10 |
| | QA Qualifier | Estimated, 2 | Rejected, 6, 9 | Estimated, 2 | Estimated, 2 | Estimated, 2 |
| Dissolved Oxygen (mg/L) | Min. | 8.43 | 9.08 | 8.53 | 9.31 | 9.10 |
| | 0.25 | 9.13 | 9.98 | 9.30 | 9.81 | 9.73 |
| | Median | 9.41 | 10.14 | 9.59 | 10.07 | 10.05 |
| | 0.75 | 9.68 | 10.36 | 9.78 | 10.24 | 10.22 |
| | Max. | 10.37 | 10.83 | 10.24 | 10.69 | 10.65 |
| | Data Quality Objective: +/- 0.5 mg/l | see DO% | no record | see DO% | see DO% | see DO% |
| | QA Qualifier | Estimated, 1,2 | Rejected, 6, 9 | Estimated, 1, 2 | Estimated, 1,2 | Estimated, 1,2 |
| Specific Conductance (µS/cm) | Min. | 796 | 568 | 836 | 812 | 843 |
| | 0.25 | 804 | 571 | 843 | 819 | 848 |
| | Median | 810 | 576 | 845 | 822 | 849 |
| | 0.75 | 815 | 582 | 847 | 824 | 852 |
| | Max. | 819 | 587 | 850 | 827 | 860 |
| | Data Quality Objective: +/- 0.05 mS/cm | 0 | no record | 5 | 15 | 4 |
| | QA Qualifier | Estimated, 2 | Rejected, 6, 9 | Estimated, 2 | Estimated, 2 | Estimated, 2 |

Appendix D-9: Summary of continuous field measures in Steven's and Permanente Creek Watersheds

| | Station | Late Summer | | | | | Winter Wet Season | | | |
|------------------------------|---|-------------|-----------|-----------|-----------|-----------|-------------------|--------------|--------------|--------------|
| | | PER010 | STE020 | STE060 | STE070 | STE100 | PER010 | STE020 | STE060 | STE070 |
| | Start Date | 9/17/2002 | 9/17/2002 | 9/17/2002 | 9/17/2002 | 9/17/2002 | 11/7/2002 | 11/7/2002 | 11/6/2002 | 11/7/2002 |
| | End Date | 9/24/2002 | 9/24/2002 | 9/24/2002 | 9/24/2002 | 09/24/02 | 12/6/2002 | 12/6/2002 | 12/5/2002 | 12/6/2002 |
| | Data points | 667 | 668 | 670 | 677 | 690 | 2803 | 2801 | 2791 | 2788 |
| Temperature (°C) | Min. | 16.80 | 17.66 | 16.94 | 20.91 | 13.14 | 11.47 | 14.33 | 9.28 | 11.91 |
| | 0.25 | 18.05 | 18.30 | 18.03 | 21.17 | 13.90 | 13.83 | 15.80 | 11.10 | 12.80 |
| | Median | 19.81 | 19.06 | 18.97 | 21.39 | 14.80 | 15.17 | 16.50 | 12.21 | 13.83 |
| | 0.75 | 22.58 | 20.81 | 19.85 | 21.82 | 15.54 | 16.56 | 17.28 | 13.31 | 14.40 |
| | Max. | 25.04 | 23.58 | 20.71 | 22.12 | 16.62 | 21.14 | 19.68 | 17.63 | 15.92 |
| | HWAT | 20.4 | 19.7 | 18.9 | 21.5 | 14.7 | 16.4 | 17.2 | 14.3 | 14.9 |
| | QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 7.45 | 7.52 | 8.16 | 7.66 | 7.83 | 6.95 | 7.13 | 7.55 | 7.44 |
| | 0.25 | 7.54 | 7.57 | 8.18 | 7.79 | 7.93 | 7.28 | 7.37 | 7.82 | 7.53 |
| | Median | 7.57 | 7.60 | 8.22 | 7.80 | 7.95 | 7.38 | 7.43 | 8.12 | 7.62 |
| | 0.75 | 7.64 | 7.71 | 8.30 | 7.82 | 7.99 | 7.49 | 7.52 | 8.15 | 7.70 |
| | Max. | 8.00 | 7.87 | 8.40 | 7.90 | 8.06 | 7.85 | 8.11 | 8.26 | 7.84 |
| | Data Quality Objective: +/- 0.50 pH units | 0.10 | 0.07 | 0.06 | 0.02 | 0.08 | 0.12 | 0.09 | 0.12 | 0.02 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 |
| Dissolved Oxygen (%) | Min. | 7.00 | 53.80 | 88.50 | 66.60 | 88.90 | 19.20 | 0.60 | 70.50 | 67.70 |
| | 0.25 | 17.40 | 59.90 | 90.20 | 72.00 | 90.30 | 62.70 | 2.60 | 91.20 | 75.50 |
| | Median | 28.60 | 64.10 | 91.50 | 73.10 | 91.30 | 67.70 | 2.90 | 92.80 | 84.10 |
| | 0.75 | 79.85 | 86.25 | 96.48 | 74.20 | 95.20 | 87.25 | 3.30 | 98.10 | 86.30 |
| | Max. | 153.40 | 114.60 | 103.90 | 78.10 | 132.30 | 206.70 | 100.40 | 104.10 | 90.40 |
| | Data Quality Objective: +/- 5.0% | 0.60 | 0.70 | 0.80 | 0.40 | 1.80 | 4.10 | 9.00 | 9.00 | 4.90 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Valid | Rejected | Rejected | Valid |
| Dissolved Oxygen (mg/L) | Min. | 0.65 | 4.94 | 7.97 | 5.92 | 8.85 | 1.86 | 0.05 | 7.21 | 6.92 |
| | 0.25 | 1.65 | 5.59 | 8.30 | 6.35 | 9.13 | 6.37 | 0.26 | 9.79 | 7.71 |
| | Median | 2.62 | 5.96 | 8.58 | 6.45 | 9.32 | 6.88 | 0.28 | 10.17 | 8.65 |
| | 0.75 | 6.86 | 7.78 | 9.02 | 6.53 | 9.65 | 8.54 | 0.32 | 10.50 | 9.09 |
| | Max. | 12.66 | 9.87 | 9.67 | 6.88 | 12.91 | 19.91 | 9.89 | 11.51 | 9.62 |
| | Data Quality Objective: +/- 0.5 mg/l | n/a | n/a | n/a | n/a | n/a | see DO% | see DO% | see DO% | see DO% |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Valid, 1 | Rejected | Rejected | Valid, 1 |
| Specific Conductance (µS/cm) | Min. | 1043 | 1001 | 529 | 512 | 659 | 155 | 111 | 76 | 318 |
| | 0.25 | 1135 | 1034 | 534 | 516 | 726 | 1206 | 882 | 403 | 530 |
| | Median | 1189 | 1051 | 537 | 517 | 739 | 1355 | 917 | 430 | 534 |
| | 0.75 | 1240 | 1062 | 540 | 518 | 748 | 1535 | 929 | 451 | 536 |
| | Max. | 1297 | 1073 | 548 | 520 | 794 | 14541 | 960 | 534 | 545 |
| | Data Quality Objective: +/- 50 uS/cm | 5 | 0 | 0 | 4 | 0 | 7 | 9 | 5 | 2 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Valid | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 |

Appendix D-9: Summary of continuous field measures in Steven's and Permanente Creek Watersheds

| | | Spring Runoff Season | | | Summer Dry Season | | | |
|------------------------------|---|----------------------|--------------|--------------|-------------------|--------------|--------------|--------------|
| Station | | PER010 | PER050 | STE060 | PER010 | STE020 | STE060 | STE070 |
| Start Date | | 3/3/2003 | 3/3/2003 | 3/3/2003 | 6/26/2003 | 6/25/2003 | 6/26/2003 | 6/26/2003 |
| End Date | | 3/13/2003 | 3/13/2003 | 3/13/2003 | 7/8/2003 | 07/07/03 | 7/8/2003 | 7/8/2003 |
| Data points | | 964 | 941 | 960 | 1165 | 1149 | 1160 | 1149 |
| Temperature (°C) | Min. | 10.59 | 9.60 | 8.65 | 15.86 | 16.78 | 13.94 | 12.36 |
| | 0.25 | 12.17 | 10.98 | 10.20 | 17.97 | 17.99 | 15.92 | 12.66 |
| | Median | 14.33 | 12.49 | 11.53 | 20.07 | 19.04 | 17.33 | 12.83 |
| | 0.75 | 16.11 | 13.55 | 12.42 | 23.99 | 21.69 | 18.43 | 13.22 |
| | Max. | 21.57 | 16.17 | 14.67 | 28.06 | 25.49 | 22.24 | 13.93 |
| | HWAT | 15.1 | 12.5 | 11.8 | 21.3 | 20.1 | 18.0 | 13.0 |
| | QA Qualifier | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 | Valid, 5 |
| pH | Min. | 7.60 | 8.09 | 8.03 | 7.31 | 7.31 | 7.95 | 7.44 |
| | 0.25 | 7.73 | 8.11 | 8.09 | 7.43 | 7.40 | 8.01 | 7.49 |
| | Median | 7.86 | 8.13 | 8.18 | 7.52 | 7.43 | 8.09 | 7.50 |
| | 0.75 | 8.16 | 8.18 | 8.42 | 7.89 | 7.55 | 8.29 | 7.51 |
| | Max. | 8.70 | 8.27 | 8.65 | 8.37 | 7.83 | 8.45 | 7.57 |
| | Data Quality Objective: +/- 0.50 pH units | 0.10 | 0.07 | 0.04 | 0.07 | 0.05 | 0.04 | 0.08 |
| | QA Qualifier | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 |
| Dissolved Oxygen (%) | Min. | 61.80 | 97.70 | 96.20 | 5.40 | 64.40 | 86.20 | 75.60 |
| | 0.25 | 74.40 | 99.60 | 97.90 | 25.60 | 74.20 | 93.40 | 78.40 |
| | Median | 83.10 | 100.20 | 99.50 | 59.20 | 86.30 | 96.50 | 79.40 |
| | 0.75 | 119.05 | 102.70 | 110.10 | 176.90 | 131.10 | 108.03 | 80.50 |
| | Max. | 239.70 | 108.90 | 127.90 | 321.00 | 164.30 | 121.30 | 84.20 |
| | Data Quality Objective: +/- 5.0% | 0.10 | 0.00 | 2.20 | 4.30 | 6.70 | 1.60 | 3.60 |
| | QA Qualifier | Valid | Valid | Valid | Valid | Rejected | Valid | Valid |
| Dissolved Oxygen (mg/L) | Min. | 6.17 | 9.90 | 9.92 | 0.49 | 5.91 | 7.76 | 7.99 |
| | 0.25 | 7.96 | 10.51 | 10.63 | 2.39 | 7.01 | 8.95 | 8.28 |
| | Median | 8.80 | 10.82 | 11.14 | 5.50 | 8.06 | 9.44 | 8.38 |
| | 0.75 | 11.77 | 11.10 | 12.05 | 14.92 | 11.54 | 10.36 | 8.47 |
| | Max. | 22.19 | 11.48 | 13.56 | 25.92 | 14.02 | 11.81 | 8.73 |
| | Data Quality Objective: +/- 0.5 mg/l | n/a | n/a | n/a | n/a | n/a | n/a | n/a |
| | QA Qualifier | Valid | Valid | Valid | Valid | Rejected | Valid | Valid |
| Specific Conductance (µS/cm) | Min. | 940 | 981 | 455 | 401 | 972 | 6 | 446 |
| | 0.25 | 1469 | 1006 | 484 | 1154 | 1004 | 524 | 459 |
| | Median | 1522 | 1030 | 491 | 1216 | 1013 | 530 | 460 |
| | 0.75 | 1597 | 1035 | 498 | 1260 | 1028 | 543 | 460 |
| | Max. | 1882 | 1045 | 517 | 1474 | 1048 | 561 | 463 |
| | Data Quality Objective: +/- 50 uS/cm | 1 | 7 | 3 | 4 | 12 | 10 | 6 |
| | QA Qualifier | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 | Estimated, 2 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | WLK-030 | WLK-050 | WLK-120 | WLK-100 | WLK-130 | WLK-140 | WLK-160 | WLK-170 | WLK-180 | WLK-190 | WLK-200 | WLK-230 | WLK-240 | SUI-010 | SUI-020 | SUI-050 | |
|----------------------|------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | |
| Dryopidae | | | | | | | | | | | | | | | | | | | | | |
| | | -- | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 5 | sh | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Dytiscidae | | | | | | | | | | | | | | | | | | | | | |
| | | -- | 5 | p | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 5 | p | | | | | | | | | | | | | | | | | |
| | | -- | 8 | p | 5 | 6 | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | -- | 5 | p | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | -- | 5 | p | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 5 | p | | | | | | | | | | | | | | | | | |
| | | Adults | 5 | p | | | | | | | | | | | | | | | | | |
| | | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 5 | p | | | | | | | | | | | | | | | | | |
| | | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Elmidae | | | | | | | | | | | | | | | | | | | | | |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | -- | 4 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | -- | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | -- | 4 | sc | 0 | 0 | 0 | 2 | 0 | 92 | 31 | 1 | 16 | 44 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| | | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Elmidae | | | | | | | | | | | | | | | | | | | | | |
| | | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 41 | 9 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gyrinidae | | | | | | | | | | | | | | | | | | | | | |
| | | | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 |
| Halplidae | | | | | | | | | | | | | | | | | | | | | |
| | | Adults | 5 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | -- | 5 | mh | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 5 | mh | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Helophoridae | | | | | | | | | | | | | | | | | | | | | |
| | | Larvae | | sh | | | | | | | | | | | | | | | | | |
| | | Larvae | | sh | | | | | | | | | | | | | | | | | |
| Hydraenidae | | | | | | | | | | | | | | | | | | | | | |
| | | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrophilidae | | | | | | | | | | | | | | | | | | | | | |
| | | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Larvae | | | | | | | | | | | | | | | | | | | |
| | | Larvae | 5 | p | | | | | | | | | | | | | | | | | |
| | | Larvae | 5 | mh | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 1 | 0 | 0 | 0 | 0 |
| | | -- | 5 | p | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 5 | cg | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Psephenidae | | | | | | | | | | | | | | | | | | | | | |
| | | | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 5 | 4 | 0 | 7 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | SUI-060 | SUI-110 | SUI-130 | SUI-180 | SUI-210 | SUI-260 | SPA-020 | SPA-050 | SPA-070 | SPA-100 | SPA-110 | SPA-130 | SPA-140 | SPA-150 | SPA-160 | SPA-170 | SPA-200 | |
|-------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | | |
| Dryopidae | | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helichus sp.</i> | -- | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helichus sp.</i> | Adults | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Postelichus sp.</i> | Adults | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Dytiscidae | | 5 | p | 0 | 1 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 1 | |
| <i>Agabinus</i> | | 5 | p | | | | | | | | | | | | | | | | | | | |
| <i>Agabus sp.</i> | -- | 8 | p | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 |
| <i>Sanfilippodytes sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stictotarsus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agabinus sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | | | |
| <i>Agabus sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | | | |
| <i>Oreodytes sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sanfilippodytes sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | | | |
| <i>Stictotarsus sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Elmidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ampumixis dispar</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cleptelmis addenda sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lara sp.</i> | -- | 4 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Narpus sp.</i> | -- | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Optioservus sp.</i> | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ordobrevia nubrifera</i> | -- | 4 | sc | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Zaitzevia sp.</i> | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Elmidae | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dubiraphia sp.</i> | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Narpus sp.</i> | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Optioservus sp.</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ordobrevia nubrifera</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Zaitzevia sp.</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gyrinidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gyrinus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Haliphidae | | 5 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Apteraliplus parvulus</i> | Adults | | | | | | | | | | | | | | | | | | | | | |
| <i>Brychius sp.</i> | | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Peltodytes sp.</i> | -- | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Peltodytes sp.</i> | Adults | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Helophoridae | | | sh | | | | | | | | | | | | | | | | | | | |
| <i>Helophorus sp.</i> | Larvae | | sh | | | | | | | | | | | | | | | | | | | |
| Hydraenidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydraena sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| <i>Hydraena sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrophilidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ametor sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helochares sp.</i> | Larvae | | | | | | | | | | | | | | | | | | | | | |
| <i>Hydrochus sp.</i> | Larvae | 5 | p | | | | | | | | | | | | | | | | | | | |
| <i>Laccobius sp.</i> | Larvae | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tropisternus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cymbiodyta sp.</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Paracymus sp.</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ametor scabrosus</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Psephenidae | | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Acneus sp.</i> | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Eubrianax edwardsi</i> | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | | | | | | | | | | | | | | | | |
|-------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---|
| | | | SPA-220 | SPA-240 | LAG-170 | LAG-130 | LAG-150 | LAG-160 | LAG-165 | LAG-180 | LAG-190 | LAG-210 | LAG-220 | LAG-240 | LAG-270 | LAG-290 | LAG-300 | LAG-320 | |
| | | | Year: | | | | | | | | | | | | | | | | |
| | | | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | |
| Dryopidae | | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helichus sp.</i> | -- | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helichus sp.</i> | Adults | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Postelichus sp.</i> | Adults | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Dytiscidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | |
| <i>Agabinus</i> | | 5 | p | | | | | | | | | | | | | | | | |
| <i>Agabus sp.</i> | -- | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Sanfilippodytes sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Stictotarsus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Agabinus sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | |
| <i>Agabus sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | |
| <i>Oreodytes sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Sanfilippodytes sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | |
| <i>Stictotarsus sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Elmidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Ampumixis dispar</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Cleptelmis addenda sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Lara sp.</i> | -- | 4 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Narpus sp.</i> | -- | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 20 | 4 | 1 | 1 | 2 | 0 | |
| <i>Optioservus sp.</i> | -- | 4 | sc | 0 | 0 | 110 | 254 | 2 | 0 | 44 | 100 | 20 | 59 | 104 | 49 | 59 | 76 | 95 | |
| <i>Ordobrevia nubrifera</i> | -- | 4 | sc | 0 | 0 | 0 | 2 | 0 | 0 | 4 | 73 | 6 | 0 | 4 | 29 | 9 | 1 | 0 | |
| <i>Zaitzevia sp.</i> | -- | 4 | sc | 0 | 0 | 11 | 13 | 0 | 0 | 36 | 4 | 5 | 2 | 1 | 14 | 1 | 2 | 1 | |
| Elmidae | Adults | 4 | cg | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Dubiraphia sp.</i> | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Narpus sp.</i> | Adults | 4 | cg | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | |
| <i>Optioservus sp.</i> | Adults | 4 | sc | 0 | 0 | 77 | 15 | 1 | 0 | 54 | 25 | 38 | 22 | 25 | 11 | 7 | 1 | 12 | |
| <i>Ordobrevia nubrifera</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Zaitzevia sp.</i> | Adults | 4 | sc | 0 | 0 | 4 | 3 | 0 | 0 | 8 | 6 | 4 | 2 | 0 | 0 | 1 | 5 | 1 | |
| Gyrinidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Gyrinus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Halplidae | | 5 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Apteraltiplus parvulus</i> | Adults | | | | | | | | | | | | | | | | | | |
| <i>Brychius sp.</i> | | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Pelodytes sp.</i> | -- | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Pelodytes sp.</i> | Adults | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Helophoridae | | | sh | | | | | | | | | | | | | | | | |
| <i>Helophorus sp.</i> | Larvae | | sh | | | | | | | | | | | | | | | | |
| Hydraenidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Hydraena sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Hydraena sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Hydrophilidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Ametor sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Helochaeres sp.</i> | Larvae | | | | | | | | | | | | | | | | | | |
| <i>Hydrochus sp.</i> | Larvae | 5 | p | | | | | | | | | | | | | | | | |
| <i>Laccobius sp.</i> | Larvae | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Tropisternus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Cymbiodyta sp.</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Paracymus sp.</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Ametor scabrosus</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Psephenidae | | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Acneus sp.</i> | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Eubrianax edwardsi</i> | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 36 | 15 | 0 | 5 | 12 | 3 | 1 | 9 | |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | | | | | | | | | | | | | | | | | |
|-------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | LAG-330 | LAG-335 | LAG-380 | LAG-390 | ALP-010 | ALP-040 | ALP-070 | ALP-080 | ALP-100 | ALP-110 | ALP-140 | SLE-030 | SLE-170 | SLE-180 | SLE-210 | SLE-220 | SLE-230 | |
| Year: | | | | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | |
| Dryopidae | | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helichus sp.</i> | -- | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helichus sp.</i> | Adults | 5 | sh | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 12 |
| <i>Postelichus sp.</i> | Adults | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Dytiscidae | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agabinus</i> | | 5 | p | | | | | | | | | | | | | | | | | | |
| <i>Agabus sp.</i> | -- | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sanfilippodytes sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stictotarsus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agabinus sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | | |
| <i>Agabus sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | | |
| <i>Oreodytes sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sanfilippodytes sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | | |
| <i>Stictotarsus sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Elmidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ampumixis dispar</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cleptelmis addenda sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lara sp.</i> | -- | 4 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Narpus sp.</i> | -- | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Optoservus sp.</i> | -- | 4 | sc | 0 | 2 | 17 | 11 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 54 | 33 | 6 | 8 | 43 | |
| <i>Ordobrevia nubrifera</i> | -- | 4 | sc | 29 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Zaitzevia sp.</i> | -- | 4 | sc | 10 | 31 | 94 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10 |
| Elmidae | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dubiraphia sp.</i> | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Narpus sp.</i> | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Optoservus sp.</i> | Adults | 4 | sc | 2 | 0 | 6 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 4 |
| <i>Ordobrevia nubrifera</i> | Adults | 4 | sc | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Zaitzevia sp.</i> | Adults | 4 | sc | 6 | 14 | 10 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 3 |
| Gyrinidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gyrinus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Haliplidae | | 5 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Apteraltipus parvulus</i> | Adults | | | | | | | | | | | | | | | | | | | | |
| <i>Brychius sp.</i> | | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pelodytes sp.</i> | -- | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pelodytes sp.</i> | Adults | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Helophoridae | | | sh | | | | | | | | | | | | | | | | | | |
| <i>Helophorus sp.</i> | Larvae | | sh | | | | | | | | | | | | | | | | | | |
| Hydraenidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydraena sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydraena sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrophilidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Amator sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helochaeres sp.</i> | Larvae | | | | | | | | | | | | | | | | | | | | |
| <i>Hydrochus sp.</i> | Larvae | 5 | p | | | | | | | | | | | | | | | | | | |
| <i>Laccobius sp.</i> | Larvae | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tropisternus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cymbiodyta sp.</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Paracymus sp.</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Amator scabrosus</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 |
| Psephenidae | | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Acneus sp.</i> | -- | 4 | sc | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| <i>Eubrianax edwardsi</i> | -- | 4 | sc | 7 | 66 | 27 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | WIL-100 | WIL-170 | WIL-190 | BUT-010 | BUT-020 | BUT-030 | BUT-040 | BUT-050 | PER-010 | PER-020 | PER-030 | PER-040 | PER-050 | PER-070 | PER-080 | PES-050 | PES-060 |
|-------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y1 | Y1 | Y1 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | |
| Dryopidae | | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helichus sp.</i> | -- | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helichus sp.</i> | Adults | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| <i>Postelichus sp.</i> | Adults | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Dytiscidae | | 5 | p | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 0 | 0 | 0 |
| <i>Agabius</i> | | 5 | p | | | | | | | | | | | | | | | | | |
| <i>Agabus sp.</i> | -- | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 5 | 0 | 0 | 0 |
| <i>Sanfilippodytes sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stictotarsus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agabius sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | |
| <i>Agabus sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | |
| <i>Oreodytes sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sanfilippodytes sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | |
| <i>Stictotarsus sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Elmidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ampumixis dispar</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cleptelmis addenda sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lara sp.</i> | -- | 4 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Narpus sp.</i> | -- | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Optioservus sp.</i> | -- | 4 | sc | 3 | 10 | 0 | 11 | 30 | 37 | 91 | 57 | 0 | 0 | 0 | 0 | 2 | 0 | 2 | 21 | 47 |
| <i>Ordobrevia nubrifera</i> | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Zaitzevia sp.</i> | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Elmidae | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dubiraphia sp.</i> | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Narpus sp.</i> | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Optioservus sp.</i> | Adults | 4 | sc | 0 | 0 | 1 | 8 | 41 | 16 | 22 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 20 | 38 |
| <i>Ordobrevia nubrifera</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Zaitzevia sp.</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gyrinidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gyrinus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Halipidae | | 5 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Apteraliplus parvulus</i> | Adults | | | | | | | | | | | | | | | | | | | |
| <i>Brychius sp.</i> | | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| <i>Peltodytes sp.</i> | -- | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Peltodytes sp.</i> | Adults | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Helophoridae | | | sh | | | | | | | | | | | | | | | | | |
| <i>Helophorus sp.</i> | Larvae | | sh | | | | | | | | | | | | | | | | | |
| Hydraenidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydraena sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydraena sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrophilidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ametor sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helochares sp.</i> | Larvae | | | | | | | | | | | | | | | | | | | |
| <i>Hydrochus sp.</i> | Larvae | 5 | p | | | | | | | | | | | | | | | | | |
| <i>Laccobius sp.</i> | Larvae | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| <i>Tropisternus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cymbiodyta sp.</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Paracymus sp.</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ametor scabrosus</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Psephenidae | | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Acneus sp.</i> | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Eubrianax edwardsi</i> | -- | 4 | sc | 0 | 3 | 0 | 0 | 0 | 2 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | | | | | | | | | | | | | | | | |
|-------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | | PES-070 | PES-080 | PES-095 | PES-100 | PES-120 | PES-140 | PES-150 | PES-160 | PES-170 | PES-180 | PES-190 | PES-200 | PES-210 | PES-230 | PES-240 | SGR-010 | SGR-030 |
| | | | | Year: | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | |
| Dryopidae | | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helichus sp.</i> | -- | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 8 | 0 | 0 | 19 | 1 | 0 |
| <i>Helichus sp.</i> | Adults | 5 | sh | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 5 | 1 | 1 | 0 | 0 |
| <i>Postelichus sp.</i> | Adults | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Dytiscidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agabinus</i> | | 5 | p | | | | | | | | | | | | | | | | | |
| <i>Agabus sp.</i> | -- | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sanfilippodytes sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stictotarsus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agabinus sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | |
| <i>Agabus sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | |
| <i>Oreodytes sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| <i>Sanfilippodytes sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | |
| <i>Stictotarsus sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Elmidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ampumixis dispar</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| <i>Cleptelmis addenda sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lara sp.</i> | -- | 4 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Narpus sp.</i> | -- | 4 | cg | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 10 | 2 | 11 | 4 | 5 | 4 | 0 | 5 | 0 | 2 |
| <i>Optioservus sp.</i> | -- | 4 | sc | 27 | 26 | 14 | 19 | 23 | 12 | 15 | 51 | 34 | 18 | 27 | 56 | 61 | 109 | 66 | 105 | 69 |
| <i>Ordobrevia nubrifera</i> | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 95 | 0 | 0 | 0 | 0 |
| <i>Zaitzevia sp.</i> | -- | 4 | sc | 0 | 0 | 1 | 2 | 3 | 1 | 1 | 21 | 41 | 20 | 25 | 20 | 40 | 1 | 1 | 1 | 2 |
| Elmidae | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dubiraphia sp.</i> | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Narpus sp.</i> | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| <i>Optioservus sp.</i> | Adults | 4 | sc | 37 | 8 | 33 | 54 | 25 | 50 | 18 | 30 | 22 | 1 | 6 | 80 | 43 | 21 | 42 | 52 | 23 |
| <i>Ordobrevia nubrifera</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| <i>Zaitzevia sp.</i> | Adults | 4 | sc | 0 | 0 | 2 | 3 | 4 | 1 | 1 | 2 | 17 | 12 | 16 | 28 | 19 | 5 | 3 | 0 | 1 |
| Gyrinidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gyrinus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Halplidae | | 5 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Apteraliplus parvulus</i> | Adults | | | | | | | | | | | | | | | | | | | |
| <i>Brychius sp.</i> | | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 5 | 0 |
| <i>Peltodytes sp.</i> | -- | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Peltodytes sp.</i> | Adults | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Helophoridae | | | sh | | | | | | | | | | | | | | | | | |
| <i>Helophorus sp.</i> | Larvae | | sh | | | | | | | | | | | | | | | | | |
| Hydraenidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydraena sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydraena sp.</i> | | 5 | p | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrophilidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ametor sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helochares sp.</i> | Larvae | | | | | | | | | | | | | | | | | | | |
| <i>Hydrochus sp.</i> | Larvae | 5 | p | | | | | | | | | | | | | | | | | |
| <i>Laccobius sp.</i> | Larvae | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tropisternus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cymbiodyta sp.</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Paracymus sp.</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ametor scabrosus</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Psephenidae | | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Acneus sp.</i> | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| <i>Eubrianax edwardsi</i> | -- | 4 | sc | 2 | 0 | 0 | 0 | 8 | 1 | 0 | 35 | 23 | 5 | 33 | 94 | 43 | 4 | 4 | 1 | 1 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | SGR-040 | SGR-060 | SGR-075 | SGR-080 | SGR-090 | SGR-110 | SGR-120 | SGR-130 | SGR-150 | STE-020 | STE-030 | STE-040 | STE-060 | STE-070 | STE-100 | STE-110 | STE-120 | |
|-------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | | |
| Dryopidae | | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helichus sp.</i> | -- | 5 | | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 |
| <i>Helichus sp.</i> | Adults | 5 | sh | 4 | 4 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | |
| <i>Postelichus sp.</i> | Adults | 5 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Dytiscidae | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agabinus</i> | | 5 | p | | | | | | | | | | | | | | | | | | | |
| <i>Agabus sp.</i> | -- | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sanfilippodytes sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stictotarsus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agabinus sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | | | |
| <i>Agabus sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | | | |
| <i>Oreodytes sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sanfilippodytes sp.</i> | Adults | 5 | p | | | | | | | | | | | | | | | | | | | |
| <i>Stictotarsus sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Elmidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ampunixis dispar</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cleptelmis addenda sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lara sp.</i> | -- | 4 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Narpus sp.</i> | -- | 4 | cg | 1 | 5 | 1 | 1 | 2 | 9 | 2 | 10 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 0 | 0 |
| <i>Optioservus sp.</i> | -- | 4 | sc | 21 | 14 | 44 | 19 | 10 | 46 | 39 | 20 | 31 | 0 | 0 | 0 | 22 | 0 | 14 | 46 | 27 | | |
| <i>Ordobrevia nubrifera</i> | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 10 | 0 | | |
| <i>Zaitzevia sp.</i> | -- | 4 | sc | 0 | 4 | 1 | 1 | 2 | 0 | 5 | 7 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 27 | 2 | |
| Elmidae | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dubiraphia sp.</i> | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Narpus sp.</i> | Adults | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Optioservus sp.</i> | Adults | 4 | sc | 30 | 11 | 31 | 16 | 7 | 13 | 19 | 9 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 21 | 14 | |
| <i>Ordobrevia nubrifera</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | |
| <i>Zaitzevia sp.</i> | Adults | 4 | sc | 0 | 2 | 0 | 0 | 2 | 1 | 0 | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14 | 5 | |
| Gyrinidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gyrinus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Halplidae | | 5 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Apteralipus parvulus</i> | Adults | | | | | | | | | | | | | | | | | | | | | |
| <i>Brychius sp.</i> | | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pelodytes sp.</i> | -- | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pelodytes sp.</i> | Adults | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Helophoridae | | | sh | | | | | | | | | | | | | | | | | | | |
| <i>Helophorus sp.</i> | Larvae | | sh | | | | | | | | | | | | | | | | | | | |
| Hydraenidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydraena sp.</i> | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| <i>Hydraena sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrophilidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| <i>Ametor sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| <i>Helochares sp.</i> | Larvae | | | | | | | | | | | | | | | | | | | | | |
| <i>Hydrochus sp.</i> | Larvae | 5 | p | | | | | | | | | | | | | | | | | | | |
| <i>Laccobius sp.</i> | Larvae | 5 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tropisternus sp.</i> | -- | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cymbiodyta sp.</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Paracymus sp.</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ametor scabrosus</i> | Adults | 5 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Psephenidae | | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Acneus sp.</i> | -- | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| <i>Eubrianax edwardsi</i> | -- | 4 | sc | 1 | 7 | 3 | 1 | 2 | 22 | 0 | 12 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 21 | 15 | |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | WLK-030 | WLK-050 | WLK-120 | WLK-100 | WLK-130 | WLK-140 | WLK-160 | WLK-170 | WLK-180 | WLK-190 | WLK-200 | WLK-230 | WLK-240 | SUI-010 | SUI-020 | SUI-050 | |
|----------------------|-----------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | |
| | <i>Eubrianax edwardsi</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Psephenus</i> | Larvae | 4 | sc | | | | | | | | | | | | | | | | | |
| Scirtidae | | | | | | | | | | | | | | | | | | | | | |
| | <i>Elodes sp.</i> | -- | 5 | sc | | | | | | | | | | | | | | | | | |
| | Staphylinidae | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Diptera | | | | | | | | | | | | | | | | | | | | | |
| | Blephariceridae | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Agathon sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Blepharicera sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Philorus sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Brachycera | | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Brachycera sp.</i> | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cecidomyiidae | | | | | | | | | | | | | | | | | | | | | |
| | Ceratopogonidae | | 6 | p | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Bezzia/ Palpomyia</i> | | 6 | p | 0 | 0 | 5 | 5 | 0 | 5 | 0 | 9 | 0 | 28 | 1 | 20 | 5 | 0 | 1 | 0 | 0 |
| | <i>Atrichopogon sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Ceratopogon sp.</i> | | 6 | p | | | | | | | | | | | | | | | | | |
| | <i>Culicoides sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| | <i>Dasyhelea sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Forcipomyiinae | | 6 | cg | | | | | | | | | | | | | | | | | |
| | <i>Probezzia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Stilobezzia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Ceratopogonidae | Pupae | 6 | p | | | | | | | | | | | | | | | | | |
| | Chaoboridae | | 7 | p | | | | | | | | | | | | | | | | | |
| | <i>Chaoborus</i> | | 7 | p | | | | | | | | | | | | | | | | | |
| | Chironomidae | | 6 | | 411 | 291 | 208 | 135 | 514 | 83 | 436 | 278 | 85 | 164 | 286 | 579 | 565 | 491 | 374 | 334 | |
| Culicidae | | | | | | | | | | | | | | | | | | | | | |
| | <i>Aedes sp.</i> | | 8 | cg | | | | | | | | | | | | | | | | | |
| | <i>Culiseta sp.</i> | | | | | | | | | | | | | | | | | | | | |
| | Dixidae | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Dixa sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Dixella sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Meringodixa sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Dolichopodidae | | 4 | p | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Empididae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Chelifera/ Metachela</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Clinocera sp.</i> | | 6 | p | 13 | 3 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Hemerodromia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Neoplasta sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Oreogeton sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Trichoclinocera sp.</i> | | 6 | p | | | | | | | | | | | | | | | | | |
| | <i>Wiedemannia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Ephydriidae | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Hemerodromia sp.</i> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Hydrellia sp.</i> | | 6 | sh | | | | | | | | | | | | | | | | | |
| | <i>Scatella sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Ephydriidae | Pupae | 6 | | | | | | | | | | | | | | | | | | |
| | Muscidae | | 6 | p | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 |
| | Pelecorynchidae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | <i>Glutops sp.</i> | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Phoridae | | 5 | cg | | | | | | | | | | | | | | | | | |
| | Psychodidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | SUI-060 | SUI-110 | SUI-130 | SUI-180 | SUI-210 | SUI-260 | SPA-020 | SPA-050 | SPA-070 | SPA-100 | SPA-110 | SPA-130 | SPA-140 | SPA-150 | SPA-160 | SPA-170 | SPA-200 | |
|-----------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | | |
| <i>Eubrianax edwardsi</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Psephenus</i> | Larvae | 4 | sc | | | | | | | | | | | | | | | | | | | |
| Scirtidae | | 5 | sc | | | | | | | | | | | | | | | | | | | |
| <i>Elodes sp.</i> | -- | | sc | | | | | | | | | | | | | | | | | | | |
| Staphylinidae | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| Diptera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Blephariceridae | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agathon sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Blepharicera sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Philorus sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Brachycera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Brachycera sp.</i> | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cecidomyiidae | | 5 | | | | | | | | | | | | | | | | | | | | |
| Ceratopogonidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Bezzia/ Palpomyia</i> | | 6 | p | 0 | 0 | 0 | 0 | 3 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 10 | 0 | 1 | 0 | 6 | 0 | 0 |
| <i>Atrichopogon sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ceratopogon sp.</i> | | 6 | p | | | | | | | | | | | | | | | | | | | |
| <i>Culicoides sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 |
| <i>Dasyhelea sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Forcipomyiinae | | 6 | cg | | | | | | | | | | | | | | | | | | | |
| <i>Probezzia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| <i>Sitobezzia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ceratopogonidae | Pupae | 6 | p | | | | | | | | | | | | | | | | | | | |
| Chaoboridae | | 7 | p | | | | | | | | | | | | | | | | | | | |
| <i>Chaoborus</i> | | 7 | p | | | | | | | | | | | | | | | | | | | |
| Chironomidae | | 6 | | 328 | 162 | 207 | 213 | 307 | 445 | 289 | 234 | 418 | 442 | 523 | 585 | 237 | 114 | 103 | 321 | 183 | | |
| Culicidae | | 8 | cg | | | | | | | | | | | | | | | | | | | |
| <i>Aedes sp.</i> | | | | | | | | | | | | | | | | | | | | | | |
| <i>Culiseta sp.</i> | | | | | | | | | | | | | | | | | | | | | | |
| Dixidae | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dixa sp.</i> | | 2 | cg | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dixella sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Meringodixa sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Dolichopodidae | | 4 | p | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Empididae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| <i>Chelifera/ Metachela</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Clinocera sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hemerodromia sp.</i> | | 6 | p | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 9 | 1 | | |
| <i>Neoplasta sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| <i>Oreogeton sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Trichoclinocera sp.</i> | | 6 | p | | | | | | | | | | | | | | | | | | | |
| <i>Wiedemannia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ephydriidae | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hemerodromia sp.</i> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydrellia sp.</i> | | 6 | sh | | | | | | | | | | | | | | | | | | | |
| <i>Scatella sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ephydriidae | Pupae | 6 | | | | | | | | | | | | | | | | | | | | |
| Muscidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pelecorynchidae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Glutops sp.</i> | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 0 |
| Phoridae | | 5 | cg | | | | | | | | | | | | | | | | | | | |
| Psychodidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | SPA-220 | SPA-240 | LAG-170 | LAG-130 | LAG-150 | LAG-160 | LAG-165 | LAG-180 | LAG-190 | LAG-210 | LAG-220 | LAG-240 | LAG-270 | LAG-290 | LAG-300 | LAG-320 |
|-----------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | |
| <i>Eubrianax edwardsi</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Psephenus</i> | Larvae | 4 | sc | | | | | | | | | | | | | | | | |
| Scirtidae | | 5 | sc | | | | | | | | | | | | | | | | |
| <i>Elodes sp.</i> | -- | | sc | | | | | | | | | | | | | | | | |
| Staphylinidae | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Diptera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Blephariceridae | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agathon sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Blepharicera sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Phylorus sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Brachycera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Brachycera sp.</i> | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cecidomyiidae | | 5 | | | | | | | | | | | | | | | | | |
| Ceratopogonidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Bezzia/ Palpomyia</i> | | 6 | p | 0 | 1 | 3 | 0 | 6 | 5 | 4 | 18 | 6 | 2 | 8 | 33 | 23 | 7 | 8 | 2 |
| <i>Atrichopogon sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ceratopogon sp.</i> | | 6 | p | | | | | | | | | | | | | | | | |
| <i>Culicoides sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dasyhelea sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Forcipomyiinae | | 6 | cg | | | | | | | | | | | | | | | | |
| <i>Probezzia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stilobezzia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ceratopogonidae | Pupae | 6 | p | | | | | | | | | | | | | | | | |
| Chaoboridae | | 7 | p | | | | | | | | | | | | | | | | |
| <i>Chaoborus</i> | | 7 | p | | | | | | | | | | | | | | | | |
| Chironomidae | | 6 | | 192 | 271 | 183 | 147 | 341 | 719 | 180 | 173 | 267 | 191 | 218 | 145 | 350 | 461 | 527 | 203 |
| Culicidae | | 8 | cg | | | | | | | | | | | | | | | | |
| <i>Aedes sp.</i> | | | | | | | | | | | | | | | | | | | |
| <i>Culiseta sp.</i> | | | | | | | | | | | | | | | | | | | |
| Dixidae | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dixa sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dixella sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Meringodixa sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Dolichopodidae | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Empididae | | 6 | p | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Chelifera/ Metachela</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| <i>Clinocera sp.</i> | | 6 | p | 0 | 0 | 5 | 0 | 0 | 0 | 2 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hemerodromia sp.</i> | | 6 | p | 0 | 1 | 4 | 1 | 0 | 0 | 6 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
| <i>Neoplasta sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Oreogeton sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Trichoclinocera sp.</i> | | 6 | p | | | | | | | | | | | | | | | | |
| <i>Wiedemannia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ephydriidae | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hemerodromia sp.</i> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydrellia sp.</i> | | 6 | sh | | | | | | | | | | | | | | | | |
| <i>Scatella sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ephydriidae | Pupae | 6 | | | | | | | | | | | | | | | | | |
| Muscidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pelecorynchidae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Glutops sp.</i> | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phoridae | | 5 | cg | | | | | | | | | | | | | | | | |
| Psychodidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | | | | | | | | | | | | | | | | | |
|-----------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | LAG-330 | LAG-335 | LAG-380 | LAG-390 | ALP-010 | ALP-040 | ALP-070 | ALP-080 | ALP-100 | ALP-110 | ALP-140 | SLE-030 | SLE-170 | SLE-180 | SLE-210 | SLE-220 | SLE-230 | |
| Year: | | | | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | |
| <i>Eubrianax edwardsi</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Psephenus</i> | Larvae | 4 | sc | | | | | | | | | | | | | | | | | | |
| Scirtidae | | 5 | sc | | | | | | | | | | | | | | | | | | |
| <i>Elodes sp.</i> | -- | | sc | | | | | | | | | | | | | | | | | | |
| Staphylinidae | Adults | 5 | p | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Diptera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Blephariceridae | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| <i>Agathon sp.</i> | | 0 | sc | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Blepharicera sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Philorus sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Brachycera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Brachycera sp.</i> | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cecidomyiidae | | 5 | | | | | | | | | | | | | | | | | | | |
| Ceratopogonidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Bezzia/ Palpomyia</i> | | 6 | p | 12 | 8 | 8 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | 30 | 1 | 22 | 0 | 0 |
| <i>Atrichopogon sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ceratopogon sp.</i> | | 6 | p | | | | | | | | | | | | | | | | | | |
| <i>Culicoides sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dasyhelea sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Forcipomyiinae | | 6 | cg | | | | | | | | | | | | | | | | | | |
| <i>Probezzia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stilobezzia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ceratopogonidae | Pupae | 6 | p | | | | | | | | | | | | | | | | | | |
| Chaoboridae | | 7 | p | | | | | | | | | | | | | | | | | | |
| <i>Chaoborus</i> | | 7 | p | | | | | | | | | | | | | | | | | | |
| Chironomidae | | 6 | | 171 | 209 | 196 | 98 | 688 | 657 | 348 | 670 | 685 | 512 | 119 | 428 | 594 | 261 | 544 | 252 | 189 | |
| Culicidae | | 8 | cg | | | | | | | | | | | | | | | | | | |
| <i>Aedes sp.</i> | | | | | | | | | | | | | | | | | | | | | |
| <i>Culiseta sp.</i> | | | | | | | | | | | | | | | | | | | | | |
| Dixidae | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dixa sp.</i> | | 2 | cg | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dixella sp.</i> | | 2 | cg | 0 | 4 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| <i>Meringodixa sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Dolichopodidae | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Empididae | | 6 | p | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Chelifera/ Metachela</i> | | 6 | p | 0 | 2 | 5 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Clinocera sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hemerodromia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Neoplasta sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Oreogeton sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Trichoclinocera sp.</i> | | 6 | p | | | | | | | | | | | | | | | | | | |
| <i>Wiedemannia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ephydriidae | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hemerodromia sp.</i> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydrellia sp.</i> | | 6 | sh | | | | | | | | | | | | | | | | | | |
| <i>Scatella sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ephydriidae | Pupae | 6 | | | | | | | | | | | | | | | | | | | |
| Muscidae | | 6 | p | 0 | 0 | 0 | 0 | 3 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pelecorynchidae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Glutops sp.</i> | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 |
| Phoridae | | 5 | cg | | | | | | | | | | | | | | | | | | |
| Psychodidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | WIL-100 | WIL-170 | WIL-190 | BUT-010 | BUT-020 | BUT-030 | BUT-040 | BUT-050 | PER-010 | PER-020 | PER-030 | PER-040 | PER-050 | PER-070 | PER-080 | PES-050 | PES-060 |
|-----------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y1 | Y1 | Y1 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | |
| <i>Eubrianax edwardsi</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Psephenus</i> | Larvae | 4 | sc | | | | | | | | | | | | | | | | | |
| Scirtidae | | 5 | sc | | | | | | | | | | | | | | | | | |
| <i>Elodes sp.</i> | -- | | sc | | | | | | | | | | | | | | | | | |
| Staphylinidae | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Diptera</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Blephariceridae | | 0 | sc | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agathon sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Blepharicera sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Philorus sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Brachycera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Brachycera sp.</i> | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cecidomyiidae | | 5 | | | | | | | | | | | | | | | | | | |
| Ceratopogonidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 |
| <i>Bezzia/ Palpomyia</i> | | 6 | p | 1 | 0 | 4 | 0 | 0 | 8 | 12 | 3 | 0 | 0 | 5 | 0 | 0 | 0 | 4 | 0 | 1 |
| <i>Atrichopogon sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ceratopogon sp.</i> | | 6 | p | | | | | | | | | | | | | | | | | |
| <i>Culicoides sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dasyhelea sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Forcipomyiinae | | 6 | cg | | | | | | | | | | | | | | | | | |
| <i>Probezzia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stilobezzia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ceratopogonidae | Pupae | 6 | p | | | | | | | | | | | | | | | | | |
| Chaoboridae | | 7 | p | | | | | | | | | | | | | | | | | |
| <i>Chaoborus</i> | | 7 | p | | | | | | | | | | | | | | | | | |
| Chironomidae | | 6 | | 261 | 344 | 86 | 64 | 336 | 101 | 77 | 51 | 136 | 808 | 640 | 839 | 317 | 551 | 86 | 128 | 338 |
| Culicidae | | 8 | cg | | | | | | | | | | | | | | | | | |
| <i>Aedes sp.</i> | | | | | | | | | | | | | | | | | | | | |
| <i>Culiseta sp.</i> | | | | | | | | | | | | | | | | | | | | |
| Dixidae | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dixa sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| <i>Dixella sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Meringodixa sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| Dolichopodidae | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Empididae | | 6 | p | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 0 | 0 | 0 |
| <i>Chelifera/ Metachela</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Clinocera sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| <i>Hemerodromia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 4 |
| <i>Neoplasta sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 2 | 4 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 4 | 0 | 0 |
| <i>Oreogeton sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Trichoclinocera sp.</i> | | 6 | p | | | | | | | | | | | | | | | | | |
| <i>Wiedemannia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ephydriidae | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hemerodromia sp.</i> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydrellia sp.</i> | | 6 | sh | | | | | | | | | | | | | | | | | |
| <i>Scatella sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| Ephydriidae | Pupae | 6 | | | | | | | | | | | | | | | | | | |
| Muscidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pelecorynchidae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Glutops sp.</i> | | 3 | p | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phoridae | | 5 | cg | | | | | | | | | | | | | | | | | |
| Psychodidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | PES-070 | PES-080 | PES-095 | PES-100 | PES-120 | PES-140 | PES-150 | PES-160 | PES-170 | PES-180 | PES-190 | PES-200 | PES-210 | PES-230 | PES-240 | SGR-010 | SGR-030 |
|-----------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | |
| <i>Eubrianax edwardsi</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Psephenus</i> | Larvae | 4 | sc | | | | | | | | | | | | | | | | | |
| Scirtidae | | 5 | sc | | | | | | | | | | | | | | | | | |
| <i>Elodes sp.</i> | -- | | sc | | | | | | | | | | | | | | | | | |
| Staphylinidae | Adults | 5 | p | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Diptera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Blephariceridae | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agathon sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Blepharicera sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Philorus sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Brachycera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Brachycera sp.</i> | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cecidomyiidae | | 5 | | | | | | | | | | | | | | | | | | |
| Ceratopogonidae | | 6 | p | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Bezzia/ Palpomyia</i> | | 6 | p | 0 | 2 | 2 | 6 | 0 | 1 | 4 | 6 | 13 | 4 | 5 | 9 | 4 | 16 | 16 | 8 | 5 |
| <i>Atrichopogon sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 6 |
| <i>Ceratopogon sp.</i> | | 6 | p | | | | | | | | | | | | | | | | | |
| <i>Culicoides sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| <i>Dasyhelea sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 |
| Forcipomyiinae | | 6 | cg | | | | | | | | | | | | | | | | | |
| <i>Probezzia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stilobezzia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| Ceratopogonidae | Pupae | 6 | p | | | | | | | | | | | | | | | | | |
| Chaoboridae | | 7 | p | | | | | | | | | | | | | | | | | |
| <i>Chaoborus</i> | | 7 | p | | | | | | | | | | | | | | | | | |
| Chironomidae | | 6 | | 262 | 416 | 261 | 40 | 16 | 54 | 68 | 47 | 31 | 104 | 41 | 64 | 45 | 200 | 68 | 157 | 43 |
| Culicidae | | 8 | cg | | | | | | | | | | | | | | | | | |
| <i>Aedes sp.</i> | | | | | | | | | | | | | | | | | | | | |
| <i>Culiseta sp.</i> | | | | | | | | | | | | | | | | | | | | |
| Dixidae | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dixa sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| <i>Dixella sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Meringodixa sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Dolichopodidae | | 4 | p | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Empididae | | 6 | p | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| <i>Chelifera/ Metachela</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| <i>Clinocera sp.</i> | | 6 | p | 2 | 0 | 3 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 10 | 0 |
| <i>Hemerodromia sp.</i> | | 6 | p | 5 | 0 | 2 | 0 | 1 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 31 | 0 |
| <i>Neoplasta sp.</i> | | 6 | p | 1 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 5 | 4 | 0 | 0 | 2 | 0 | 5 | 0 |
| <i>Oreogeton sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |
| <i>Trichoclinocera sp.</i> | | 6 | p | | | | | | | | | | | | | | | | | |
| <i>Wiedemannia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |
| Ephydriidae | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| <i>Hemerodromia sp.</i> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydrellia sp.</i> | | 6 | sh | | | | | | | | | | | | | | | | | |
| <i>Scatella sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ephydriidae | Pupae | 6 | | | | | | | | | | | | | | | | | | |
| Muscidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| Pelecorynchidae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Glutops sp.</i> | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
| Phoridae | | 5 | cg | | | | | | | | | | | | | | | | | |
| Psychodidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | SGR-040 | SGR-060 | SGR-075 | SGR-080 | SGR-090 | SGR-110 | SGR-120 | SGR-130 | SGR-150 | STE-020 | STE-030 | STE-040 | STE-060 | STE-070 | STE-100 | STE-110 | STE-120 |
|-----------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | |
| <i>Eubrianax edwardsi</i> | Adults | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Psephenus</i> | Larvae | 4 | sc | | | | | | | | | | | | | | | | | |
| Scirtidae | | 5 | sc | | | | | | | | | | | | | | | | | |
| <i>Elodes sp.</i> | -- | | sc | | | | | | | | | | | | | | | | | |
| Staphylinidae | Adults | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Diptera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Blephariceridae | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| <i>Agathon sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 7 |
| <i>Blepharicera sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| <i>Philorus sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Brachycera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Brachycera sp.</i> | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 17 | 0 | 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cecidomyiidae | | 5 | | | | | | | | | | | | | | | | | | |
| Ceratopogonidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Bezzia/ Palpomyia</i> | | 6 | p | 2 | 9 | 13 | 5 | 7 | 2 | 13 | 7 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 13 | 28 |
| <i>Atrichopogon sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ceratopogon sp.</i> | | 6 | p | | | | | | | | | | | | | | | | | |
| <i>Culicoides sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| <i>Dasyhelea sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Forcipomyiinae | | 6 | cg | | | | | | | | | | | | | | | | | |
| <i>Probezzia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stilobezzia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ceratopogonidae | Pupae | 6 | p | | | | | | | | | | | | | | | | | |
| Chaoboridae | | 7 | p | | | | | | | | | | | | | | | | | |
| <i>Chaoborus</i> | | 7 | p | | | | | | | | | | | | | | | | | |
| Chironomidae | | 6 | | 13 | 65 | 34 | 59 | 45 | 18 | 115 | 57 | 61 | 742 | 625 | 561 | 610 | 471 | 269 | 123 | 91 |
| Culicidae | | 8 | cg | | | | | | | | | | | | | | | | | |
| <i>Aedes sp.</i> | | | | | | | | | | | | | | | | | | | | |
| <i>Culiseta sp.</i> | | | | | | | | | | | | | | | | | | | | |
| Dixidae | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dixa sp.</i> | | 2 | cg | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| <i>Dixella sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Meringodixa sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Dolichopodidae | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| Empididae | | 6 | p | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| <i>Chelifera/ Metachela</i> | | 6 | p | 0 | 5 | 2 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| <i>Clinocera sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 0 |
| <i>Hemerodromia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 |
| <i>Neoplasta sp.</i> | | 6 | p | 0 | 0 | 3 | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 3 |
| <i>Oreogeton sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Trichoclinocera sp.</i> | | 6 | p | | | | | | | | | | | | | | | | | |
| <i>Wiedemannia sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ephydriidae | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hemerodromia sp.</i> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydrellia sp.</i> | | 6 | sh | | | | | | | | | | | | | | | | | |
| <i>Scatella sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ephydriidae | Pupae | 6 | | | | | | | | | | | | | | | | | | |
| Muscidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 2 | 0 | 0 | 0 | 0 |
| Pelecynchidae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Glutops sp.</i> | | 3 | p | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| Phoridae | | 5 | cg | | | | | | | | | | | | | | | | | |
| Psychodidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | WLK-030 | WLK-050 | WLK-120 | WLK-100 | WLK-130 | WLK-140 | WLK-160 | WLK-170 | WLK-180 | WLK-190 | WLK-200 | WLK-230 | WLK-240 | SUI-010 | SUI-020 | SUI-050 | |
|---------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | |
| <i>Maruina lanceolata sp.</i> | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 16 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pericoma/Telmatoscopus</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Psychoda sp.</i> | | 10 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Psychodidae | Pupae | 4 | cg | | | | | | | | | | | | | | | | | | |
| Scathophagidae | | 5 | | | | | | | | | | | | | | | | | | | |
| Sciomyzidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Antichaeta sp.</i> | | 6 | | | | | | | | | | | | | | | | | | | |
| <i>Sepeidon sp.</i> | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Simuliidae | | 6 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 |
| <i>Prosimulium sp.</i> | | 3 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Simulium sp.</i> | | 6 | cf | 226 | 73 | 4 | 335 | 59 | 19 | 181 | 13 | 60 | 3 | 18 | 169 | 203 | 39 | 35 | 199 | | |
| Stratiomyidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Caloparyphus/ Euparyphus</i> | | 7 | cg | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hedriodiscus/Odontomyia</i> | | 8 | cg | | | | | | | | | | | | | | | | | | |
| <i>Nemotelus</i> | | 8 | cg | | | | | | | | | | | | | | | | | | |
| Syrphidae | | 10 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Tabanidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tabanus - Atylotus sp.</i> | | 5 | p | | | | | | | | | | | | | | | | | | |
| <i>Chrysops sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Silvius sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Tipulidae | | 3 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Antocha sp.</i> | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 6 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cryptolabis</i> | | 3 | sh | | | | | | | | | | | | | | | | | | |
| <i>Dicranota sp.</i> | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Erioptera sp.</i> | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gonomyia sp.</i> | | 3 | cg | | | | | | | | | | | | | | | | | | |
| <i>Hesperocanopa sp.</i> | | 1 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hexatoma sp.</i> | | 2 | p | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Limnophila sp.</i> | | 3 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Limonia sp.</i> | | 6 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lipsothrix sp.</i> | | 5 | | | | | | | | | | | | | | | | | | | |
| <i>Molophilus sp.</i> | | 4 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ormosia sp.</i> | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pilaria sp.</i> | | 7 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Rhabdomastix sp.</i> | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tipula sp.</i> | | 4 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ulomorpha sp.</i> | | 5 | | | | | | | | | | | | | | | | | | | |
| <i>Tipulidae sp A.</i> | | 5 | | | | | | | | | | | | | | | | | | | |
| <u>Ephemeroptera</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ameletidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ameletus sp.</i> | | 0 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Baetidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 17 | 2 | 0 | 0 | 0 |
| <i>Centroptilum/ Procloeon</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Acentrella sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Baetis sp.</i> | | 5 | cg | 45 | 205 | 75 | 261 | 3 | 191 | 55 | 205 | 132 | 32 | 98 | 23 | 45 | 151 | 210 | 216 | | |
| <i>Callibaetis sp.</i> | | 9 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Diphotor hageni</i> | | 5 | cg | 0 | 0 | 0 | 1 | 0 | 73 | 0 | 5 | 49 | 3 | 52 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| <i>Fallceon quilleri</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 7 | 1 | |
| Caenidae | | 7 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Caenis sp.</i> | | 7 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ephemerelellidae | | 1 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 28 | 12 | |
| <i>Drunella sp.</i> | | 0 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 2 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | SUI-060 | SUI-110 | SUI-130 | SUI-180 | SUI-210 | SUI-260 | SPA-020 | SPA-050 | SPA-070 | SPA-100 | SPA-110 | SPA-130 | SPA-140 | SPA-150 | SPA-160 | SPA-170 | SPA-200 |
|--------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | |
| <i>Maruina lanceolata sp.</i> | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pericoma/Telmatoscopus</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 |
| <i>Psychoda sp.</i> | | 10 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Psychodidae | Pupae | 4 | cg | | | | | | | | | | | | | | | | | |
| Scathophagidae | | 5 | | | | | | | | | | | | | | | | | | |
| Sciomyzidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Antichaeta sp.</i> | | 6 | | | | | | | | | | | | | | | | | | |
| <i>Sepedon sp.</i> | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Simuliidae | | 6 | cf | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Prosimulium sp.</i> | | 3 | cf | 0 | 0 | 0 | 0 | 21 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Simulium sp.</i> | | 6 | cf | 139 | 191 | 145 | 6 | 322 | 37 | 1 | 58 | 285 | 30 | 7 | 260 | 283 | 14 | 1 | 1 | 225 |
| Stratiomyidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Caloparyphus/Euparyphus</i> | | 7 | cg | 0 | 1 | 1 | 0 | 0 | 11 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 6 | 2 | 3 | 0 |
| <i>Hedriodiscus/Odontomyia</i> | | 8 | cg | | | | | | | | | | | | | | | | | |
| <i>Nemotelus</i> | | 8 | cg | | | | | | | | | | | | | | | | | |
| Syrphidae | | 10 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Tabanidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tabanus - Arylotus sp.</i> | | 5 | p | | | | | | | | | | | | | | | | | |
| <i>Chrysops sp.</i> | | 8 | p | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Silvius sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Tipulidae | | 3 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Antocha sp.</i> | | 3 | cg | 0 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cryptolabis</i> | | 3 | sh | | | | | | | | | | | | | | | | | |
| <i>Dicranota sp.</i> | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 8 | 9 | 0 | 0 |
| <i>Erioptera sp.</i> | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gonomyia sp.</i> | | 3 | cg | | | | | | | | | | | | | | | | | |
| <i>Hesperocanopa sp.</i> | | 1 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hexatoma sp.</i> | | 2 | p | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Limnophila sp.</i> | | 3 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| <i>Limonia sp.</i> | | 6 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lipsothrix sp.</i> | | 5 | | | | | | | | | | | | | | | | | | |
| <i>Molophilus sp.</i> | | 4 | sh | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ormosia sp.</i> | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pilaria sp.</i> | | 7 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Rhabdomastix sp.</i> | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tipula sp.</i> | | 4 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 6 | 7 | 0 | 0 | 0 | 0 | 0 |
| <i>Ulomorpha sp.</i> | | 5 | | | | | | | | | | | | | | | | | | |
| <i>Tipulidae sp A.</i> | | 5 | | | | | | | | | | | | | | | | | | |
| Ephemeroptera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ameletidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ameletus sp.</i> | | 0 | cg | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Baetidae | | | | 4 | 0 | 2 | 0 | 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Centroptilum/Proclon</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| <i>Acentrella sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Baetis sp.</i> | | 5 | cg | 273 | 381 | 328 | 144 | 39 | 2 | 158 | 286 | 111 | 6 | 7 | 1 | 23 | 6 | 10 | 495 | 431 |
| <i>Callibaetis sp.</i> | | 9 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dipheter hageni</i> | | 5 | cg | 0 | 0 | 0 | 2 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 34 | 96 | 130 | 0 | 0 |
| <i>Fallceon quilleri</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Caenidae | | | | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Caenis sp.</i> | | 7 | cg | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ephemereillidae | | | | 1 | 31 | 89 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Drunella sp.</i> | | 0 | cg | 0 | 14 | 0 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | SPA-220 | SPA-240 | LAG-170 | LAG-130 | LAG-150 | LAG-160 | LAG-165 | LAG-180 | LAG-190 | LAG-210 | LAG-220 | LAG-240 | LAG-270 | LAG-290 | LAG-300 | LAG-320 |
|--------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | |
| <i>Maruina lanceolata sp.</i> | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 3 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| <i>Pericoma/Telmatoscopus</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Psychoda sp.</i> | | 10 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Psychodidae | Pupae | 4 | cg | | | | | | | | | | | | | | | | |
| Scathophagidae | | 5 | | | | | | | | | | | | | | | | | |
| Sciomyzidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Antichaeta sp.</i> | | 6 | | | | | | | | | | | | | | | | | |
| <i>Sepedon sp.</i> | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Simuliidae | | 6 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Prosimulium sp.</i> | | 3 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 9 | 40 | 10 | 13 |
| <i>Simulium sp.</i> | | 6 | cf | 81 | 54 | 176 | 16 | 4 | 13 | 57 | 15 | 6 | 156 | 11 | 2 | 22 | 33 | 28 | 1 |
| Stratiomyidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Caloparyphus/Euparyphus</i> | | 7 | cg | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hedriodiscus/Odontomyia</i> | | 8 | cg | | | | | | | | | | | | | | | | |
| <i>Nemotelus</i> | | 8 | cg | | | | | | | | | | | | | | | | |
| Syrphidae | | 10 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Tabanidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tabanus - Atylotus sp.</i> | | 5 | p | | | | | | | | | | | | | | | | |
| <i>Chrysops sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Silvius sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Tipulidae | | 3 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Antocha sp.</i> | | 3 | cg | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 5 | 0 | 0 | 0 | 1 | 3 |
| <i>Cryptolabis</i> | | 3 | sh | | | | | | | | | | | | | | | | |
| <i>Dicranota sp.</i> | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 1 | 0 | 0 | 0 |
| <i>Erioptera sp.</i> | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gonomyia sp.</i> | | 3 | cg | | | | | | | | | | | | | | | | |
| <i>Hesperocanopa sp.</i> | | 1 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hexatoma sp.</i> | | 2 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 11 | 11 | 1 | 2 | 6 | 0 | 0 | 4 | 1 |
| <i>Limnophila sp.</i> | | 3 | s | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 124 | 44 | 0 | 11 | 2 | |
| <i>Limonia sp.</i> | | 6 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lipsothrix sp.</i> | | 5 | | | | | | | | | | | | | | | | | |
| <i>Molophilus sp.</i> | | 4 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ormosia sp.</i> | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pilaria sp.</i> | | 7 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Rhabdomastix sp.</i> | | 3 | p | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tipula sp.</i> | | 4 | om | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| <i>Ulomorpha sp.</i> | | 5 | | | | | | | | | | | | | | | | | |
| <i>Tipulidae sp A.</i> | | 5 | | | | | | | | | | | | | | | | | |
| Ephemeroptera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ameletidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ameletus sp.</i> | | 0 | cg | 0 | 0 | 0 | 0 | 1 | 1 | 4 | 1 | 1 | 0 | 0 | 7 | 5 | 0 | 3 | 0 |
| Baetidae | | 4 | cg | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Centropilum/Procloeon</i> | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Acentrella sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Baetis sp.</i> | | 5 | cg | 405 | 434 | 61 | 122 | 129 | 67 | 55 | 64 | 39 | 75 | 17 | 81 | 104 | 100 | 65 | 44 |
| <i>Callibaetis sp.</i> | | 9 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dipheter hageni</i> | | 5 | cg | 0 | 0 | 16 | 24 | 1 | 0 | 76 | 44 | 73 | 35 | 41 | 8 | 8 | 12 | 40 | 3 |
| <i>Fallceon quilleri</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Caenidae | | 7 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Caenis sp.</i> | | 7 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ephemerellidae | | 1 | cg | 0 | 0 | 12 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Drunella sp.</i> | | 0 | cg | 0 | 0 | 9 | 2 | 45 | 0 | 6 | 13 | 68 | 24 | 21 | 14 | 16 | 4 | 1 | 49 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | | | | | | | | | | | | | | | | |
|-------------------|------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | | LAG-330 | LAG-335 | LAG-380 | LAG-390 | ALP-010 | ALP-040 | ALP-070 | ALP-080 | ALP-100 | ALP-110 | ALP-140 | SLE-030 | SLE-170 | SLE-180 | SLE-210 | SLE-220 | SLE-230 |
| | | | | Year: | | | | | | | | | | | | | | | | |
| | | | | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | |
| | | | | | | | | | | | | | | | | | | | | |
| | | | 2 | sc | 2 | 8 | 19 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 0 | 0 |
| | | | 10 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | Pupae | 4 | cg | | | | | | | | | | | | | | | | |
| | | | 5 | | | | | | | | | | | | | | | | | |
| | | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 6 | | | | | | | | | | | | | | | | | |
| | | | 6 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | cf | 8 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 6 | cf | 1 | 2 | 0 | 19 | 14 | 5 | 14 | 92 | 62 | 22 | 10 | 178 | 102 | 6 | 122 | 110 |
| | | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 7 | cg | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 0 | 8 | 3 |
| | | | 8 | cg | | | | | | | | | | | | | | | | |
| | | | 8 | cg | | | | | | | | | | | | | | | | |
| | | | 10 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 5 | p | | | | | | | | | | | | | | | | |
| | | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | sh | | | | | | | | | | | | | | | | |
| | | | 3 | p | 0 | 1 | 5 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 6 | 0 |
| | | | 3 | cg | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | cg | | | | | | | | | | | | | | | | |
| | | | 1 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 2 | p | 1 | 5 | 2 | 19 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 2 |
| | | | 3 | s | 39 | 69 | 25 | 94 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 6 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 5 | | | | | | | | | | | | | | | | | |
| | | | 4 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 7 | p | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | p | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 7 | 0 |
| | | | 4 | om | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 4 | 0 |
| | | | 5 | | | | | | | | | | | | | | | | | |
| | | | 5 | | | | | | | | | | | | | | | | | |
| | | | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 0 | cg | 2 | 1 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 10 | 0 | 46 | 0 |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 5 | cg | 158 | 110 | 28 | 38 | 5 | 0 | 4 | 37 | 0 | 0 | 631 | 4 | 83 | 104 | 127 | 123 |
| | | | 9 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 5 | cg | 20 | 10 | 77 | 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 107 |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 7 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 7 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 1 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 0 | cg | 3 | 3 | 3 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 11 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | | WIL-100 | WIL-170 | WIL-190 | BUT-010 | BUT-020 | BUT-030 | BUT-040 | BUT-050 | PER-010 | PER-020 | PER-030 | PER-040 | PER-050 | PER-070 | PER-080 | PES-050 | PES-060 | |
|--------|----------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | Year: | | Y1 | Y1 | Y1 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | |
| Phylum | Insecta | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| | <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | |
| | | | 2 | sc | 0 | 5 | 0 | 0 | 0 | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| | | | 10 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Psychodidae | Pupae | 4 | cg | | | | | | | | | | | | | | | | | | |
| | Scathophagidae | | 5 | | | | | | | | | | | | | | | | | | | |
| | Sciomyzidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 6 | | | | | | | | | | | | | | | | | | | |
| | | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Simuliidae | | 6 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 6 | cf | 379 | 170 | 27 | 1 | 91 | 104 | 68 | 80 | 0 | 0 | 184 | 1 | 47 | 40 | 28 | 9 | 50 | |
| | Stratiomyidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 7 | cg | 0 | 7 | 2 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 1 |
| | | | 8 | cg | | | | | | | | | | | | | | | | | | |
| | | | 8 | cg | | | | | | | | | | | | | | | | | | |
| | Syrphidae | | 10 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Tabanidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 5 | p | | | | | | | | | | | | | | | | | | |
| | | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Tipulidae | | 3 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | sh | | | | | | | | | | | | | | | | | | |
| | | | 3 | p | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | cg | | | | | | | | | | | | | | | | | | |
| | | | 1 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 2 | p | 0 | 0 | 1 | 0 | 2 | 2 | 2 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14 | 12 |
| | | | 3 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 6 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 5 | | | | | | | | | | | | | | | | | | | |
| | | | 4 | sh | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 7 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | p | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | om | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 5 | | | | | | | | | | | | | | | | | | | |
| | <u>Ephemeroptera</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Ameletidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 0 | cg | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 0 |
| | Baetidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 5 | cg | 178 | 44 | 191 | 12 | 63 | 128 | 169 | 25 | 0 | 0 | 8 | 10 | 102 | 168 | 71 | 30 | 53 | |
| | | | 9 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| | | | 5 | cg | 0 | 69 | 133 | 5 | 60 | 24 | 19 | 49 | 0 | 0 | 0 | 0 | 0 | 0 | 23 | 41 | 42 | |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Caenidae | | 7 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 7 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | Ephemerellidae | | 1 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 0 | cg | 0 | 2 | 0 | 0 | 9 | 35 | 19 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 4 |

Appendix E-1: Bioassessment master taxa list

| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | Site Code: | PES-070 | PES-080 | PES-095 | PES-100 | PES-120 | PES-140 | PES-150 | PES-160 | PES-170 | PES-180 | PES-190 | PES-200 | PES-210 | PES-230 | PES-240 | SGR-010 | SGR-030 | |
|---------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | | |
| <i>Maruina lanceolata</i> sp. | | 2 | sc | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 1 | 5 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pericoma/ Telmatoscopus</i> | | 4 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 0 | 1 | 0 | |
| <i>Psychoda</i> sp. | | 10 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Psychodidae | Pupae | 4 | cg | | | | | | | | | | | | | | | | | | | |
| Scathophagidae | | 5 | | | | | | | | | | | | | | | | | | | | |
| Sciomyzidae | | 6 | p | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Antichaeta</i> sp. | | 6 | | | | | | | | | | | | | | | | | | | | |
| <i>Sepedon</i> sp. | | 6 | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Simuliidae | | 6 | cf | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Prosimulium</i> sp. | | 3 | cf | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Simulium</i> sp. | | 6 | cf | | 201 | 98 | 66 | 75 | 89 | 2 | 29 | 37 | 8 | 52 | 49 | 2 | 26 | 11 | 26 | 23 | 14 | |
| Stratiomyidae | | 8 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Catoparyphus/ Euparyphus</i> | | 7 | cg | | 0 | 0 | 2 | 2 | 0 | 1 | 0 | 1 | 2 | 5 | 1 | 7 | 1 | 3 | 0 | 2 | 4 | |
| <i>Hedriodiscus/Odontomyia</i> | | 8 | cg | | | | | | | | | | | | | | | | | | | |
| <i>Nemotelus</i> | | 8 | cg | | | | | | | | | | | | | | | | | | | |
| Syrphidae | | 10 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | |
| Tabanidae | | 8 | p | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Tabanus - Atrylotus</i> sp. | | 5 | p | | | | | | | | | | | | | | | | | | | |
| <i>Chrysops</i> sp. | | 8 | p | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Silvius</i> sp. | | 8 | p | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | |
| Tipulidae | | 3 | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 2 | 0 | 1 | 0 | 1 | 0 | |
| <i>Antocha</i> sp. | | 3 | cg | | 0 | 0 | 0 | 2 | 2 | 1 | 0 | 2 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | |
| <i>Cryptolabis</i> | | 3 | sh | | | | | | | | | | | | | | | | | | | |
| <i>Dicranota</i> sp. | | 3 | p | | 0 | 2 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 3 | 0 | 0 | |
| <i>Erioptera</i> sp. | | 3 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Gonomyia</i> sp. | | 3 | cg | | | | | | | | | | | | | | | | | | | |
| <i>Hesperocanopa</i> sp. | | 1 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | |
| <i>Hexatoma</i> sp. | | 2 | p | | 3 | 0 | 3 | 0 | 4 | 4 | 0 | 6 | 4 | 3 | 2 | 6 | 1 | 3 | 5 | 1 | 2 | |
| <i>Limnophila</i> sp. | | 3 | s | | 0 | 0 | 2 | 2 | 3 | 3 | 3 | 21 | 0 | 2 | 6 | 27 | 33 | 1 | 7 | 1 | 0 | |
| <i>Limonia</i> sp. | | 6 | sh | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Lipsothrix</i> sp. | | 5 | | | | | | | | | | | | | | | | | | | | |
| <i>Molophilus</i> sp. | | 4 | sh | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Ormosia</i> sp. | | 3 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 1 | 0 | 0 | | |
| <i>Pilaria</i> sp. | | 7 | p | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Rhabdomastix</i> sp. | | 3 | p | | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | |
| <i>Tipula</i> sp. | | 4 | om | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | |
| <i>Ulomorpha</i> sp. | | 5 | | | | | | | | | | | | | | | | | | | | |
| <i>Tipulidae</i> sp A. | | 5 | | | | | | | | | | | | | | | | | | | | |
| <u>Ephemeroptera</u> | | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <u>Ameletidae</u> | | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Ameletus</i> sp. | | 0 | cg | | 0 | 0 | 0 | 3 | 0 | 16 | 22 | 0 | 33 | 1 | 0 | 8 | 2 | 0 | 5 | 1 | 2 | |
| Baetidae | | 4 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Centroptilum/ Procloeon</i> | | 2 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Acentrella</i> sp. | | 4 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Baetis</i> sp. | | 5 | cg | | 81 | 56 | 109 | 107 | 98 | 33 | 3 | 25 | 10 | 108 | 28 | 48 | 63 | 25 | 22 | 5 | 331 | |
| <i>Callibaetis</i> sp. | | 9 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Dipheter hageni</i> | | 5 | cg | | 18 | 23 | 16 | 96 | 35 | 25 | 94 | 11 | 17 | 9 | 22 | 17 | 18 | 1 | 20 | 58 | 11 | |
| <i>Fallceon quilleri</i> | | 4 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Caenidae | | 7 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Caenis</i> sp. | | 7 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Ephemerellidae | | 1 | cg | | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| <i>Drunella</i> sp. | | 0 | cg | | 13 | 0 | 20 | 11 | 43 | 60 | 0 | 45 | 5 | 4 | 13 | 17 | 1 | 5 | 4 | 5 | 23 | |

Appendix E-1: Bioassessment master taxa list

| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | Site Code: | SGR-040 | SGR-060 | SGR-075 | SGR-080 | SGR-090 | SGR-110 | SGR-120 | SGR-130 | SGR-150 | STE-020 | STE-030 | STE-040 | STE-060 | STE-070 | STE-100 | STE-110 | STE-120 | |
|---------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | | |
| <i>Maruina lanceolata sp.</i> | | 2 | sc | | 0 | 5 | 1 | 0 | 1 | 3 | 0 | 14 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 0 |
| <i>Pericoma/ Telmatoscopus</i> | | 4 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Psychoda sp.</i> | | 10 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Psychodidae | Pupae | 4 | cg | | | | | | | | | | | | | | | | | | | |
| Scathophagidae | | 5 | | | | | | | | | | | | | | | | | | | | |
| Sciomyzidae | | 6 | p | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Antichaeta sp.</i> | | 6 | | | | | | | | | | | | | | | | | | | | |
| <i>Sepdon sp.</i> | | 6 | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Simuliidae | | 6 | cf | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Prosimulium sp.</i> | | 3 | cf | | 0 | 0 | 0 | 0 | 0 | 0 | 32 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 56 | |
| <i>Simulium sp.</i> | | 6 | cf | | 36 | 118 | 6 | 149 | 39 | 9 | 7 | 61 | 5 | 60 | 10 | 18 | 102 | 21 | 19 | 5 | 1 | |
| Stratiomyidae | | 8 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Caloparyphus/ Euparyphus</i> | | 7 | cg | | 5 | 9 | 4 | 1 | 0 | 0 | 3 | 2 | 3 | 0 | 0 | 5 | 0 | 0 | 1 | 4 | 1 | |
| <i>Hedriodiscus/Odontomyia</i> | | 8 | cg | | | | | | | | | | | | | | | | | | | |
| <i>Nemotelus</i> | | 8 | cg | | | | | | | | | | | | | | | | | | | |
| Syrphidae | | 10 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Tabanidae | | 8 | p | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tabanus - Atylotus sp.</i> | | 5 | p | | | | | | | | | | | | | | | | | | | |
| <i>Chrysops sp.</i> | | 8 | p | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Silvius sp.</i> | | 8 | p | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Tipulidae | | 3 | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Antocha sp.</i> | | 3 | cg | | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 3 | 0 | 0 | 1 | 0 | |
| <i>Cryptolabis</i> | | 3 | sh | | | | | | | | | | | | | | | | | | | |
| <i>Dicranota sp.</i> | | 3 | p | | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 0 | 2 | 1 |
| <i>Erioptera sp.</i> | | 3 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gonomyia sp.</i> | | 3 | cg | | | | | | | | | | | | | | | | | | | |
| <i>Hesperocanopa sp.</i> | | 1 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hexatoma sp.</i> | | 2 | p | | 0 | 0 | 2 | 1 | 4 | 2 | 3 | 1 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 4 |
| <i>Limnophila sp.</i> | | 3 | s | | 45 | 1 | 26 | 12 | 18 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 17 | 24 | 0 | |
| <i>Limonia sp.</i> | | 6 | sh | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lipsothrix sp.</i> | | 5 | | | | | | | | | | | | | | | | | | | | |
| <i>Molophilus sp.</i> | | 4 | sh | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ormosia sp.</i> | | 3 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 1 | |
| <i>Pilaria sp.</i> | | 7 | p | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Rhabdomastix sp.</i> | | 3 | p | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tipula sp.</i> | | 4 | om | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ulmomorpha sp.</i> | | 5 | | | | | | | | | | | | | | | | | | | | |
| <i>Tipulidae sp A.</i> | | 5 | | | | | | | | | | | | | | | | | | | | |
| Ephemeroptera | | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ameletidae | | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ameletus sp.</i> | | 0 | cg | | 0 | 3 | 1 | 0 | 3 | 4 | 12 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 4 | 15 | |
| Baetidae | | 4 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Centroptilum/ Procloeon</i> | | 2 | cg | | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Acentrella sp.</i> | | 4 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Baetis sp.</i> | | 5 | cg | | 10 | 120 | 14 | 56 | 69 | 76 | 456 | 81 | 30 | 1 | 26 | 24 | 8 | 235 | 55 | 33 | 160 | |
| <i>Callibaetis sp.</i> | | 9 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dipheter hageni</i> | | 5 | cg | | 6 | 35 | 6 | 10 | 25 | 19 | 1 | 4 | 9 | 0 | 0 | 0 | 0 | 0 | 35 | 14 | 8 | |
| <i>Fallceon quilleri</i> | | 4 | cg | | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Caenidae | | 7 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Caenis sp.</i> | | 7 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| EphemereIIDae | | 1 | cg | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Drunella sp.</i> | | 0 | cg | | 24 | 10 | 44 | 32 | 14 | 32 | 18 | 19 | 11 | 0 | 0 | 0 | 0 | 0 | 20 | 27 | 20 | |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | WLK-030 | WLK-050 | WLK-120 | WLK-100 | WLK-130 | WLK-140 | WLK-160 | WLK-170 | WLK-180 | WLK-190 | WLK-200 | WLK-230 | WLK-240 | SUI-010 | SUI-020 | SUI-050 | |
|--------------------|------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | |
| | | | 1 | cg | 0 | 0 | 105 | 5 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| | | | 2 | cg | 0 | 0 | 0 | 37 | 0 | 12 | 1 | 5 | 0 | 4 | 0 | 1 | 1 | 3 | 6 | 9 | |
| | | | 4 | sc | 0 | 0 | 0 | 2 | 0 | 0 | 1 | 7 | 0 | 9 | 0 | 0 | 0 | 0 | 9 | 5 | |
| | | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | sc | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 37 | 2 | 0 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 3 | 6 | 0 | 2 | 0 | 0 | 0 | 0 | 2 | 1 | 6 | |
| | | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | cg | 0 | 1 | 1 | 18 | 0 | 0 | 0 | 2 | 5 | 6 | 5 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 7 | cg | | | | | | | | | | | | | | | | | |
| | | | 7 | cg | | | | | | | | | | | | | | | | | |
| Hemiptera | | | | | | | | | | | | | | | | | | | | | |
| | | | 8 | p | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 5 | | | | | | | | | | | | | | | | | | |
| | | Adults | 5 | | | | | | | | | | | | | | | | | | |
| Lepidoptera | | | | | | | | | | | | | | | | | | | | | |
| | | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 5 | sc | | | | | | | | | | | | | | | | | |
| | | | 5 | | | | | | | | | | | | | | | | | | |
| Megaloptera | | | | | | | | | | | | | | | | | | | | | |
| | | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Odonata | | | | | | | | | | | | | | | | | | | | | |
| | | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 7 | p | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 |
| | | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 1 | sh | | | | | | | | | | | | | | | | | |
| | | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | SUI-060 | SUI-110 | SUI-130 | SUI-180 | SUI-210 | SUI-260 | SPA-020 | SPA-050 | SPA-070 | SPA-100 | SPA-110 | SPA-130 | SPA-140 | SPA-150 | SPA-160 | SPA-170 | SPA-200 | |
|--------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | | |
| | | | cg | 0 | 16 | 0 | 238 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | cg | 9 | 33 | 5 | 40 | 0 | 62 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | sc | 2 | 15 | 0 | 11 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 59 | 29 | 50 | 0 | 0 | 0 |
| | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 94 | 157 | 184 | 0 | 0 | 0 |
| | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 3 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 3 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cg | 0 | 0 | 8 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cg | 1 | 9 | 0 | 51 | 137 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 12 | 35 | 15 | 0 | 0 | 0 | 0 |
| | | 7 | cg | | | | | | | | | | | | | | | | | | | |
| | | 7 | cg | | | | | | | | | | | | | | | | | | | |
| | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | | | | | | | | | | | | | | | | | | | | |
| | Adults | 5 | | | | | | | | | | | | | | | | | | | | |
| | | 5 | | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 0 | 3 | 0 | 0 | 0 | 1 |
| | | | | | | | | | | | | | | | | | | | | | | |
| | | | | | | | | | | | | | | | | | | | | | | |
| | | 5 | | | | | | | | | | | | | | | | | | | | |
| | | 5 | sc | | | | | | | | | | | | | | | | | | | |
| | | | | | | | | | | | | | | | | | | | | | | |
| | | 5 | | | | | | | | | | | | | | | | | | | | |
| | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 7 | p | 0 | 0 | 9 | 0 | 5 | 0 | 0 | 0 | 2 | 9 | 2 | 6 | 2 | 7 | 4 | 2 | 2 | 5 | 5 |
| | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 1 | 0 | 1 |
| | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 5 | 5 | 0 | 0 | 0 | 0 |
| | | 1 | sh | | | | | | | | | | | | | | | | | | | |
| | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 11 | 2 | 0 | 0 | 0 |
| | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | SPA-220 | SPA-240 | LAG-170 | LAG-130 | LAG-150 | LAG-160 | LAG-165 | LAG-180 | LAG-190 | LAG-210 | LAG-220 | LAG-240 | LAG-270 | LAG-290 | LAG-300 | LAG-320 |
|-----------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | |
| <i>Ephemera</i> sp. | | 1 | cg | 0 | 0 | 14 | 2 | 27 | 15 | 10 | 0 | 9 | 7 | 17 | 83 | 75 | 24 | 4 | 0 |
| <i>Serratella</i> sp. | | 2 | cg | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 2 | 74 |
| Heptageniidae | | 4 | sc | 0 | 0 | 3 | 5 | 3 | 0 | 0 | 0 | 0 | 17 | 34 | 36 | 9 | 7 | 4 | 25 |
| <i>Cinygma</i> sp. | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 69 | 28 | 46 | 53 | 0 | 0 | 0 | 0 | 0 | 37 |
| <i>Cinygmula</i> sp. | | 4 | sc | 0 | 0 | 36 | 32 | 0 | 0 | 0 | 0 | 0 | 47 | 78 | 0 | 0 | 2 | 4 | 76 |
| <i>Epeorus</i> sp. | | 0 | sc | 0 | 0 | 5 | 2 | 0 | 1 | 2 | 0 | 4 | 18 | 5 | 0 | 0 | 0 | 0 | 9 |
| <i>Ironodes</i> sp. | | 3 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| <i>Leucrocota/Nixe/Heptagenia</i> | | 3 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Rhithrogena</i> sp. | | 0 | sc | 0 | 0 | 4 | 1 | 0 | 0 | 3 | 1 | 2 | 3 | 2 | 0 | 0 | 0 | 0 | 0 |
| Leptohiphidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tricorythodes</i> sp. | | 4 | cg | 0 | 0 | 2 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Leptophlebiidae | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Paraleptophlebia</i> sp. | | 4 | cg | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 10 | 0 | 0 | 28 | 7 | 0 | 1 | 0 |
| Siphonuridae | | 7 | cg | | | | | | | | | | | | | | | | |
| <i>Siphonurus</i> sp. | | 7 | cg | | | | | | | | | | | | | | | | |
| Hemiptera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corixidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sigara</i> sp. | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Veliidae | | | | | | | | | | | | | | | | | | | |
| <i>Microvelia</i> sp. | Adults | 5 | | | | | | | | | | | | | | | | | |
| Lepidoptera | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Noctuidae | | | | | | | | | | | | | | | | | | | |
| <i>Archana</i> sp. | | | | | | | | | | | | | | | | | | | |
| Pyralidae | | 5 | | | | | | | | | | | | | | | | | |
| <i>Petrophila</i> | | 5 | sc | | | | | | | | | | | | | | | | |
| Tortricidae | | | | | | | | | | | | | | | | | | | |
| <i>Archips</i> sp. | | 5 | | | | | | | | | | | | | | | | | |
| Megaloptera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corydalidae | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dysmicohermes</i> sp. | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Orohermes crepuscularis</i> | | 0 | p | | | | | | | | | | | | | | | | |
| <i>Neohermes</i> sp. | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sialidae | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sialis</i> sp. | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Odonata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aeshnidae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Anax</i> sp. | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Calopterygidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hetaerina</i> sp. | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Coenagrionidae | | | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Argia</i> sp. | | 7 | p | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cordulegastridae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cordulegaster dorsalis</i> | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gomphidae | | 4 | p | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 1 | 0 | 0 | 1 | 0 |
| <i>Erpetogomphus</i> sp. | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Octogomphus specularis</i> | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Plecoptera | | 1 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Capniidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Mesocapnia</i> sp. | | 1 | sh | | | | | | | | | | | | | | | | |
| Chloroperlidae | | 1 | p | 0 | 0 | 6 | 0 | 4 | 0 | 1 | 0 | 0 | 0 | 4 | 12 | 7 | 0 | 2 | 0 |
| <i>Bisancora</i> sp. | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Haploperla</i> sp. | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | | | | | | | | | | | | | | | | |
|-----------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | | WIL-100 | WIL-170 | WIL-190 | BUT-010 | BUT-020 | BUT-030 | BUT-040 | BUT-050 | PER-010 | PER-020 | PER-030 | PER-040 | PER-050 | PER-070 | PER-080 | PES-050 | PES-060 |
| | | | | Year: | | | | | | | | | | | | | | | | |
| | | | | Y1 | Y1 | Y1 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | |
| <i>Ephemerella sp.</i> | | 1 | cg | 0 | 0 | 0 | 0 | 37 | 92 | 22 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 24 | 30 |
| <i>Serratella sp.</i> | | 2 | cg | 0 | 0 | 0 | 0 | 13 | 62 | 39 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| Heptageniidae | | 4 | sc | 0 | 21 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 5 |
| <i>Cinygma sp.</i> | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cinygmula sp.</i> | | 4 | sc | 0 | 18 | 0 | 10 | 89 | 27 | 28 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 16 | 21 |
| <i>Epeorus sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 14 | 39 | 70 | 9 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 15 |
| <i>Ironodes sp.</i> | | 3 | sc | 0 | 3 | 0 | 0 | 0 | 7 | 14 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Leucrocuta/Nixe/Heptagenia</i> | | 3 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Rhithrogena sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 5 | 4 | 2 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 2 |
| Leptohyphidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tricorythodes sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 52 | 121 |
| Leptophlebiidae | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Paraleptophlebia sp.</i> | | 4 | cg | 4 | 73 | 1 | 1 | 6 | 3 | 5 | 70 | 1 | 0 | 0 | 0 | 0 | 2 | 338 | 0 | 1 |
| Siphonuridae | | 7 | cg | | | | | | | | | | | | | | | | | |
| <i>Siphonurus sp.</i> | | 7 | cg | | | | | | | | | | | | | | | | | |
| Hemiptera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corixidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sigara sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 |
| Veliidae | | | | | | | | | | | | | | | | | | | | |
| <i>Microvelia sp.</i> | Adults | 5 | | | | | | | | | | | | | | | | | | |
| Lepidoptera | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Noctuidae | | | | | | | | | | | | | | | | | | | | |
| <i>Archana sp.</i> | | | | | | | | | | | | | | | | | | | | |
| Pyralidae | | 5 | | | | | | | | | | | | | | | | | | |
| <i>Petrophila</i> | | 5 | sc | | | | | | | | | | | | | | | | | |
| Tortricidae | | | | | | | | | | | | | | | | | | | | |
| <i>Archips sp.</i> | | 5 | | | | | | | | | | | | | | | | | | |
| Megaloptera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corydalidae | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dysmicohermes sp.</i> | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Orohermes crepuscularis</i> | | 0 | p | | | | | | | | | | | | | | | | | |
| <i>Neohermes sp.</i> | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sialidae | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sialis sp.</i> | | 4 | p | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |
| Odonata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aeshnidae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| <i>Anax sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Calopterygidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hetaerina sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Coenagrionidae | | | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Argia sp.</i> | | 7 | p | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 3 | 8 | 0 | 0 |
| Cordulegastridae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cordulegaster dorsalis</i> | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |
| Gomphidae | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Erpetogomphus sp.</i> | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Octogomphus specularis</i> | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Plecoptera | | 1 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Capniidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Mesocapnia sp.</i> | | 1 | sh | | | | | | | | | | | | | | | | | |
| Chloroperlidae | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Bisancora sp.</i> | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Haploperla sp.</i> | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | PES-070 | PES-080 | PES-095 | PES-100 | PES-120 | PES-140 | PES-150 | PES-160 | PES-170 | PES-180 | PES-190 | PES-200 | PES-210 | PES-230 | PES-240 | SGR-010 | SGR-030 | |
|-------------------|-----------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | | |
| | <i>Ephemerella sp.</i> | | 1 | cg | 22 | 0 | 121 | 220 | 31 | 305 | 2 | 144 | 25 | 25 | 6 | 13 | 25 | 118 | 42 | 74 | 5 | |
| | <i>Serratella sp.</i> | | 2 | cg | 4 | 0 | 19 | 10 | 153 | 42 | 0 | 36 | 1 | 86 | 174 | 0 | 32 | 39 | 16 | 2 | 85 | |
| | Heptageniidae | | 4 | sc | 3 | 1 | 0 | 4 | 2 | 12 | 2 | 22 | 0 | 0 | 9 | 1 | 0 | 0 | 0 | 2 | 0 | |
| | <i>Cinygma sp.</i> | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Cinygmula sp.</i> | | 4 | sc | 21 | 73 | 12 | 12 | 5 | 62 | 0 | 61 | 275 | 95 | 140 | 72 | 72 | 18 | 94 | 19 | 20 | |
| | <i>Epeorus sp.</i> | | 0 | sc | 15 | 1 | 26 | 28 | 23 | 36 | 0 | 82 | 2 | 68 | 66 | 7 | 9 | 0 | 22 | 22 | 23 | |
| | <i>Ironodes sp.</i> | | 3 | sc | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 5 | 7 | 10 | 21 | 0 | 4 | 0 | 6 | |
| | <i>Leucrocuta/Nixe/Heptagenia</i> | | 3 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Rhithrogena sp.</i> | | 0 | sc | 3 | 3 | 5 | 4 | 4 | 16 | 5 | 14 | 1 | 2 | 3 | 4 | 8 | 0 | 18 | 2 | 7 | |
| | Leptohiphidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Tricorythodes sp.</i> | | 4 | cg | 38 | 0 | 34 | 4 | 85 | 19 | 0 | 13 | 0 | 3 | 3 | 0 | 0 | 0 | 0 | 71 | 0 | |
| | Leptophlebiidae | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Paraleptophlebia sp.</i> | | 4 | cg | 0 | 5 | 1 | 3 | 13 | 1 | 8 | 0 | 0 | 7 | 5 | 5 | 1 | 6 | 7 | 3 | 0 | |
| | Siphonuridae | | 7 | cg | | | | | | | | | | | | | | | | | | |
| | <i>Siphonurus sp.</i> | | 7 | cg | | | | | | | | | | | | | | | | | | |
| | Hemiptera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Corixidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Sigara sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Veliidae | | | | | | | | | | | | | | | | | | | | | |
| | <i>Microvelia sp.</i> | Adults | 5 | | | | | | | | | | | | | | | | | | | |
| | Lepidoptera | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Noctuidae | | | | | | | | | | | | | | | | | | | | | |
| | <i>Archana sp.</i> | | | | | | | | | | | | | | | | | | | | | |
| | Pyralidae | | 5 | | | | | | | | | | | | | | | | | | | |
| | <i>Petrophila</i> | | 5 | sc | | | | | | | | | | | | | | | | | | |
| | Tortricidae | | | | | | | | | | | | | | | | | | | | | |
| | <i>Archips sp.</i> | | 5 | | | | | | | | | | | | | | | | | | | |
| | Megaloptera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Corydalidae | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Dysmicohermes sp.</i> | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Orohermes crepuscularis</i> | | 0 | p | | | | | | | | | | | | | | | | | | |
| | <i>Neohermes sp.</i> | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Sialidae | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Sialis sp.</i> | | 4 | p | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Odonata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Aeshnidae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Anax sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Calopterygidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Hetaerina sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Coenagrionidae | | | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Argia sp.</i> | | 7 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | |
| | Cordulegastridae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Cordulegaster dorsalis</i> | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 0 | 2 | 4 | 0 | 9 | 0 | 2 | 8 | 0 | 2 | |
| | Gomphidae | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Erpetogomphus sp.</i> | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Octogomphus specularis</i> | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 5 | 2 | 1 | 6 | 0 | 2 | 0 | |
| | Plecoptera | | 1 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Capniidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Mesocapnia sp.</i> | | 1 | sh | | | | | | | | | | | | | | | | | | |
| | Chloroperlidae | | 1 | p | 0 | 0 | 0 | 1 | 0 | 8 | 0 | 13 | 17 | 0 | 2 | 2 | 0 | 1 | 6 | 1 | 0 | |
| | <i>Bisancora sp.</i> | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Haploperla sp.</i> | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | SGR-040 | SGR-060 | SGR-075 | SGR-080 | SGR-090 | SGR-110 | SGR-120 | SGR-130 | SGR-150 | STE-020 | STE-030 | STE-040 | STE-060 | STE-070 | STE-100 | STE-110 | STE-120 | |
|--------------------------|-----------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Insecta | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | | |
| | <i>Ephemera</i> sp. | | 1 | cg | 125 | 48 | 187 | 59 | 66 | 11 | 0 | 26 | 141 | 0 | 0 | 0 | 0 | 0 | 182 | 78 | 25 | |
| | <i>Serratella</i> sp. | | 2 | cg | 30 | 99 | 23 | 57 | 5 | 0 | 4 | 3 | 8 | 0 | 0 | 0 | 0 | 0 | 22 | 18 | 0 | |
| | Heptageniidae | | 4 | sc | 1 | 3 | 0 | 9 | 4 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | |
| | <i>Cinygma</i> sp. | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Cinygmula</i> sp. | | 4 | sc | 84 | 117 | 116 | 83 | 165 | 38 | 1 | 97 | 137 | 0 | 0 | 0 | 0 | 0 | 17 | 36 | 1 | |
| | <i>Epeorus</i> sp. | | 0 | sc | 120 | 85 | 35 | 63 | 93 | 68 | 10 | 89 | 69 | 0 | 0 | 0 | 0 | 0 | 10 | 27 | 1 | |
| | <i>Ironodes</i> sp. | | 3 | sc | 1 | 0 | 1 | 0 | 3 | 7 | 8 | 5 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 4 | |
| | <i>Leucrocota/Nixe/Heptagenia</i> | | 3 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Rhithrogena</i> sp. | | 0 | sc | 5 | 2 | 1 | 0 | 3 | 4 | 5 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 20 | |
| | Leptohiphidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Tricorythodes</i> sp. | | 4 | cg | 2 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Leptophlebiidae | | 2 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Paraleptophlebia</i> sp. | | 4 | cg | 9 | 3 | 13 | 10 | 15 | 5 | 2 | 2 | 4 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 57 | |
| | Siphonuridae | | 7 | cg | | | | | | | | | | | | | | | | | | |
| | <i>Siphonurus</i> sp. | | 7 | cg | | | | | | | | | | | | | | | | | | |
| | <u>Hemiptera</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Corixidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Sigara</i> sp. | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Veliidae | | | | | | | | | | | | | | | | | | | | | |
| | <i>Microvelia</i> sp. | Adults | 5 | | | | | | | | | | | | | | | | | | | |
| | <u>Lepidoptera</u> | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <u>Noctuidae</u> | | | | | | | | | | | | | | | | | | | | | |
| | <i>Archana</i> sp. | | | | | | | | | | | | | | | | | | | | | |
| | Pyrilidae | | 5 | | | | | | | | | | | | | | | | | | | |
| | <i>Petrophila</i> | | 5 | sc | | | | | | | | | | | | | | | | | | |
| | Tortricidae | | | | | | | | | | | | | | | | | | | | | |
| | <i>Archips</i> sp. | | 5 | | | | | | | | | | | | | | | | | | | |
| | <u>Megaloptera</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Corydalidae | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Dysmicohermes</i> sp. | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Orohermes crepuscularis</i> | | 0 | p | | | | | | | | | | | | | | | | | | |
| | <i>Neohermes</i> sp. | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | |
| | Sialidae | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Sialis</i> sp. | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <u>Odonata</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Aeshnidae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Anax</i> sp. | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Calopterygidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Hetaerina</i> sp. | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Coenagrionidae | | | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Argia</i> sp. | | 7 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | |
| | Cordulegastridae | | 3 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Cordulegaster dorsalis</i> | | 3 | p | 0 | 1 | 0 | 0 | 0 | 0 | 2 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | |
| | Gomphidae | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Erpetogomphus</i> sp. | | 4 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Octogomphus specularis</i> | | 4 | p | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 0 | |
| | <u>Plecoptera</u> | | 1 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | Capniidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Mesocapnia</i> sp. | | 1 | sh | | | | | | | | | | | | | | | | | | |
| | Chloroperlidae | | 1 | p | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 12 | |
| | <i>Bisancora</i> sp. | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | <i>Haploperla</i> sp. | | 1 | p | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | | | | | | | | | | | | | SUI-010 | | SUI-020 | SUI-050 |
|-------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|---------|---------|
| | | | | WLK-030 | WLK-050 | WLK-120 | WLK-100 | WLK-130 | WLK-140 | WLK-160 | WLK-170 | WLK-180 | WLK-190 | WLK-200 | WLK-230 | WLK-240 | Y1 | Y1 | Y1 | |
| | | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | |
| Phylum | Arthropoda | | | | | | | | | | | | | | | | | | | |
| Insecta | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | |
| | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | 1 | p | 0 | 0 | 0 | 1 | 0 | 96 | 24 | 5 | 5 | 54 | 0 | 0 | 0 | 0 | 0 | 1 | |
| | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | |
| Leuctridae | | | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Nemouridae | | | | 2 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 2 | sh | 0 | 0 | 0 | 40 | 0 | 18 | 0 | 83 | 173 | 48 | 108 | 0 | 2 | 0 | 0 | 2 | |
| | | 2 | sh | | | | | | | | | | | | | | | | | |
| | | 2 | sh | | | | | | | | | | | | | | | | | |
| Peltoperlidae | | | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Perlidae | | | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 2 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | 2 | p | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Perlodidae | | | | 2 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 2 | | | | | | | | | | | | | | | | | | |
| | | 2 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | 2 | p | 0 | 0 | 139 | 2 | 0 | 2 | 1 | 45 | 4 | 19 | 7 | 5 | 4 | 1 | 1 | 0 | |
| | | 2 | p | 0 | 0 | 0 | 3 | 0 | 2 | 0 | 4 | 1 | 3 | 0 | 1 | 2 | 0 | 0 | 0 | |
| | | 2 | p | | | | | | | | | | | | | | | | | |
| | | 2 | p | | | | | | | | | | | | | | | | | |
| Pteronarcyidae | | | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Taeniopterygidae | | | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Trichoptera | | | | 3 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Apataniidae | | | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Brachycentridae | | | | 1 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | 1 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 4 | 1 | 1 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Calamoceratidae | | | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Glossosomatidae | | | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | |
| | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 2 | 7 | 4 | 77 | 28 | 23 | 0 | 6 | 1 | 26 | 17 | |
| | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 1 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | |
| | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 40 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | 3 | sc | | | | | | | | | | | | | | | | | |
| Hydropsychidae | | | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 20 | 3 | 16 | |
| | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | 5 | cf | 0 | 0 | 0 | 0 | 0 | 4 | 3 | 0 | 0 | 0 | 0 | 0 | 4 | 1 | 0 | 0 | |
| | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 10 | 1 | 1 | 41 | 3 | 30 | 0 | 0 | 7 | 3 | 2 | |
| | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Hydropsychidae | | | | Pupae | | | | | | | | | | | | | | | | |
| Hydroptilidae | | | | 4 | ph | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | SUI-060 | SUI-110 | SUI-130 | SUI-180 | SUI-210 | SUI-260 | SPA-020 | SPA-050 | SPA-070 | SPA-100 | SPA-110 | SPA-130 | SPA-140 | SPA-150 | SPA-160 | SPA-170 | SPA-200 | |
|-------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | | |
| <i>Paraperla sp.</i> | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Suwallia sp.</i> | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 |
| <i>Sweltsa sp.</i> | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 8 | 2 | 0 | 0 |
| Leuctridae | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Despaxia augusta</i> | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Paraleuctra sp.</i> | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Perlomyia sp.</i> | | 0 | sh | | | | | | | | | | | | | | | | | | | |
| Nemouridae | | 2 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Malenka sp.</i> | | 2 | sh | 1 | 0 | 1 | 1 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 108 | 141 | 191 | 0 | 0 | 0 |
| <i>Soyedina</i> | | 2 | sh | | | | | | | | | | | | | | | | | | | |
| <i>Zapada</i> | | 2 | sh | | | | | | | | | | | | | | | | | | | |
| Peltoperlidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Soliperla sp.</i> | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Perlidae | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hesperoperla sp.</i> | | 2 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Calineuria californica</i> | | 2 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Perlodidae | | 2 | p | 0 | 0 | 1 | 19 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Baumannella alameda</i> | | 2 | | | | | | | | | | | | | | | | | | | | |
| <i>Cultus sp.</i> | | 2 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Isoperla sp.</i> | | 2 | p | 0 | 3 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Kogotus sp.</i> | | 2 | p | 0 | 0 | 0 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 1 | 0 | 0 | 0 | 0 |
| <i>Osobenus yakimae</i> | | 2 | p | | | | | | | | | | | | | | | | | | | |
| <i>Rickera sorpta</i> | | 2 | p | | | | | | | | | | | | | | | | | | | |
| Pteronarcyidae | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pteronarcella sp.</i> | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pteronarcys sp.</i> | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Taeniopterygidae | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Taenionema sp.</i> | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Trichoptera | | 3 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Apataniidae | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Apatania sp.</i> | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Brachycentridae | | 1 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Brachycentrus sp.</i> | | 1 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Micrasema sp.</i> | | 1 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Amiocentrus aspilus</i> | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Calamoceratidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Heteroplectron</i> | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Glossosomatidae | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agapetus sp.</i> | | 0 | sc | 40 | 1 | 1 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Anagapetus sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Culoptila sp.</i> | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Glossosoma sp.</i> | | 1 | sc | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Protoptila sp.</i> | | 1 | sc | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helicopsyche borealis</i> | | 3 | sc | | | | | | | | | | | | | | | | | | | |
| Hydropsychidae | | 4 | cf | 8 | 3 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Arctopsyche sp.</i> | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cheumatopsyche sp.</i> | | 5 | cf | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Homoplectra</i> | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydropsyche sp.</i> | | 4 | cf | 3 | 1 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 16 | 33 | 0 | 0 | 0 | 0 |
| <i>Parapsyche sp.</i> | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 |
| Hydropsychidae | Pupae | | | | | | | | | | | | | | | | | | | | | |
| Hydroptilidae | | 4 | ph | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | SPA-220 | SPA-240 | LAG-170 | LAG-130 | LAG-150 | LAG-160 | LAG-165 | LAG-180 | LAG-190 | LAG-210 | LAG-220 | LAG-240 | LAG-270 | LAG-290 | LAG-300 | LAG-320 |
|-------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | |
| <i>Paraperla sp.</i> | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 2 | 0 | 0 | 0 |
| <i>Suwallia sp.</i> | | 1 | p | 0 | 0 | 18 | 5 | 118 | 44 | 51 | 14 | 1 | 13 | 5 | 60 | 41 | 5 | 10 | 2 |
| <i>Sweltsa sp.</i> | | 1 | p | 0 | 0 | 32 | 1 | 2 | 0 | 38 | 0 | 2 | 7 | 3 | 5 | 3 | 2 | 2 | 12 |
| Leuctridae | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Despaxia augusta</i> | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Paraleuctra sp.</i> | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Perlomyia sp.</i> | | 0 | sh | | | | | | | | | | | | | | | | |
| Nemouridae | | 2 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Malenka sp.</i> | | 2 | sh | 0 | 0 | 22 | 87 | 5 | 0 | 82 | 126 | 43 | 45 | 41 | 17 | 46 | 12 | 8 | 136 |
| <i>Soyedina</i> | | 2 | sh | | | | | | | | | | | | | | | | |
| <i>Zapada</i> | | 2 | sh | | | | | | | | | | | | | | | | |
| Peltoperlidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Soliperla sp.</i> | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Perlidae | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | 38 | 10 | 0 | 0 | 0 | 1 |
| <i>Hesperoperla sp.</i> | | 2 | p | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 10 | 4 | 0 | 0 | 0 | 0 | 0 | 5 |
| <i>Calineuria californica</i> | | 2 | p | 0 | 0 | 1 | 1 | 0 | 0 | 1 | 8 | 32 | 6 | 10 | 2 | 3 | 0 | 0 | 2 |
| Perlodidae | | 2 | p | 0 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 |
| <i>Baumannella alameda</i> | | 2 | | | | | | | | | | | | | | | | | |
| <i>Cultus sp.</i> | | 2 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Isoperla sp.</i> | | 2 | p | 0 | 0 | 10 | 4 | 111 | 22 | 9 | 4 | 0 | 1 | 0 | 1 | 5 | 10 | 3 | 2 |
| <i>Kogotus sp.</i> | | 2 | p | 1 | 0 | 0 | 0 | 1 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 |
| <i>Osobenus yakimae</i> | | 2 | p | | | | | | | | | | | | | | | | |
| <i>Rickera sorpta</i> | | 2 | p | | | | | | | | | | | | | | | | |
| Pteronarcyidae | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pteronarcella sp.</i> | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pteronarcys sp.</i> | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Taeniopterygidae | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Taenionema sp.</i> | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Trichoptera | | 3 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Apataniidae | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Apatania sp.</i> | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Brachycentridae | | 1 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Brachycentrus sp.</i> | | 1 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Micrasema sp.</i> | | 1 | mh | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 6 |
| <i>Amiocentrus aspilus</i> | | 3 | cg | 0 | 0 | 3 | 1 | 1 | 0 | 1 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| Calamoceratidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Heteroplectron</i> | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Glossosomatidae | | 0 | sc | 0 | 0 | 5 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agapetus sp.</i> | | 0 | sc | 0 | 0 | 19 | 46 | 84 | 1 | 6 | 3 | 0 | 5 | 25 | 0 | 2 | 0 | 0 | 0 |
| <i>Anagapetus sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Culoptila sp.</i> | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Glossosoma sp.</i> | | 1 | sc | 0 | 0 | 0 | 1 | 0 | 0 | 5 | 1 | 1 | 16 | 15 | 0 | 0 | 0 | 0 | 19 |
| <i>Protoptila sp.</i> | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helicopsyche borealis</i> | | 3 | sc | | | | | | | | | | | | | | | | |
| Hydropsychidae | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Arctopsyche sp.</i> | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cheumatopsyche sp.</i> | | 5 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Homoplectra</i> | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydropsyche sp.</i> | | 4 | cf | 0 | 0 | 15 | 34 | 0 | 0 | 2 | 5 | 40 | 12 | 8 | 1 | 2 | 3 | 1 | 17 |
| <i>Parapsyche sp.</i> | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydropsychidae | Pupae | | | | | | | | | | | | | | | | | | |
| Hydroptilidae | | 4 | ph | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | | | | | | | | | | | | | | | | |
|-------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | | LAG-330 | LAG-335 | LAG-380 | LAG-390 | ALP-010 | ALP-040 | ALP-070 | ALP-080 | ALP-100 | ALP-110 | ALP-140 | SLE-030 | SLE-170 | SLE-180 | SLE-210 | SLE-220 | SLE-230 |
| Year: | | | | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | |
| <i>Paraperla sp.</i> | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Suwallia sp.</i> | | 1 | p | 0 | 0 | 6 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 13 | 0 | 0 | 0 |
| <i>Sweltsa sp.</i> | | 1 | p | 18 | 4 | 4 | 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 6 |
| Leuctridae | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Despaxia augusta</i> | | 0 | sh | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Paraleuctra sp.</i> | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Perlomyia sp.</i> | | 0 | sh | | | | | | | | | | | | | | | | | |
| Nemouridae | | 2 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Malenka sp.</i> | | 2 | sh | 59 | 24 | 61 | 106 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 24 | 1 | 51 | 9 |
| <i>Soyedina</i> | | 2 | sh | | | | | | | | | | | | | | | | | |
| <i>Zapada</i> | | 2 | sh | | | | | | | | | | | | | | | | | |
| Peltoperlidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Soliperla sp.</i> | | 1 | sh | 0 | 0 | 3 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Perlidae | | 1 | p | 0 | 0 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| <i>Hesperoperla sp.</i> | | 2 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Calineuria californica</i> | | 2 | p | 3 | 0 | 6 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 |
| Perlodidae | | 2 | p | 0 | 0 | 0 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 |
| <i>Baumannella alameda</i> | | 2 | | | | | | | | | | | | | | | | | | |
| <i>Cultus sp.</i> | | 2 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Isoptera sp.</i> | | 2 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 57 | 0 | 1 | 0 | 0 |
| <i>Kogotus sp.</i> | | 2 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 26 | 0 | 2 | 2 | 2 |
| <i>Osobenus yakimae</i> | | 2 | p | | | | | | | | | | | | | | | | | |
| <i>Rickera sorpta</i> | | 2 | p | | | | | | | | | | | | | | | | | |
| Pteronarcyidae | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pteronarcella sp.</i> | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pteronarcys sp.</i> | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Taeniopterygidae | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Taenionema sp.</i> | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Trichoptera | | 3 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Apataniidae | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Apatania sp.</i> | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Brachycentridae | | 1 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Brachycentrus sp.</i> | | 1 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Micrasema sp.</i> | | 1 | mh | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 18 |
| <i>Amiocentrus aspilus</i> | | 3 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Calamoceratidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Heteroplectron</i> | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| Glossosomatidae | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agapetus sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 13 | 0 |
| <i>Anagapetus sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Culoptila sp.</i> | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Glossosoma sp.</i> | | 1 | sc | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Protoptila sp.</i> | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helicopsyche borealis</i> | | 3 | sc | | | | | | | | | | | | | | | | | |
| Hydropsychidae | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Arctopsyche sp.</i> | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cheumatopsyche sp.</i> | | 5 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| <i>Homoptera</i> | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydropsyche sp.</i> | | 4 | cf | 5 | 1 | 6 | 16 | 0 | 0 | 0 | 0 | 1 | 86 | 0 | 0 | 0 | 0 | 0 | 0 | 20 |
| <i>Parapsyche sp.</i> | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydropsychidae | Pupae | | | | | | | | | | | | | | | | | | | |
| Hydroptilidae | | 4 | ph | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | | | | | | | | | | | | | | | | |
|--------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | WIL-100 | WIL-170 | WIL-190 | BUT-010 | BUT-020 | BUT-030 | BUT-040 | BUT-050 | PER-010 | PER-020 | PER-030 | PER-040 | PER-050 | PER-070 | PER-080 | PES-050 | PES-060 |
| | | | Year: | Y1 | Y1 | Y1 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | |
| | | | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | p | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 49 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | p | 0 | 0 | 119 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Leuctridae | | | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Nemouridae | | | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | sh | 22 | 35 | 79 | 0 | 19 | 89 | 65 | 74 | 0 | 0 | 0 | 0 | 0 | 29 | 3 | 6 |
| | | | sh | | | | | | | | | | | | | | | | |
| | | | sh | | | | | | | | | | | | | | | | |
| Peltoperlidae | | | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Perlidae | | | p | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | p | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | p | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 70 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Perlodidae | | | p | 0 | 0 | 0 | 0 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| | | | | 2 | | | | | | | | | | | | | | | |
| | | | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| | | | p | 0 | 0 | 0 | 0 | 11 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 13 |
| | | | p | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | p | 2 | | | | | | | | | | | | | | | |
| | | | p | 2 | | | | | | | | | | | | | | | |
| Pteronarcyidae | | | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Taeniopterygidae | | | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |
| Trichoptera | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Apataniidae | | | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Brachycentridae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | om | 0 | 0 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | mh | 0 | 0 | 0 | 0 | 0 | 10 | 58 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | eg | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 6 |
| Calamoceratidae | | | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Glossosomatidae | | | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | sc | 0 | 0 | 0 | 0 | 12 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 3 | 1 |
| | | | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| | | | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydropsychidae | | | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| | | | p | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | cf | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 16 |
| | | | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | | cf | 0 | 3 | 13 | 0 | 15 | 9 | 10 | 5 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 24 |
| | | | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydropsychidae | Pupae | | | | | | | | | | | | | | | | | | |
| Hydroptilidae | | | ph | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | PES-070 | PES-080 | PES-095 | PES-100 | PES-120 | PES-140 | PES-150 | PES-160 | PES-170 | PES-180 | PES-190 | PES-200 | PES-210 | PES-230 | PES-240 | SGR-010 | SGR-030 |
|-------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | |
| <i>Paraperla sp.</i> | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Suwallia sp.</i> | | 1 | p | 0 | 0 | 1 | 0 | 0 | 2 | 0 | 2 | 0 | 3 | 3 | 3 | 4 | 9 | 8 | 0 | 0 |
| <i>Sweltsa sp.</i> | | 1 | p | 0 | 0 | 0 | 1 | 2 | 1 | 10 | 0 | 57 | 0 | 1 | 14 | 8 | 1 | 20 | 0 | 1 |
| Leuctridae | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Despaxia augusta</i> | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| <i>Paraleuctra sp.</i> | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 |
| <i>Perlomyia sp.</i> | | 0 | sh | | | | | | | | | | | | | | | | | |
| Nemouridae | | 2 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Malenka sp.</i> | | 2 | sh | 1 | 117 | 7 | 33 | 34 | 41 | 17 | 49 | 43 | 27 | 43 | 16 | 3 | 111 | 39 | 17 | 29 |
| <i>Soyedina</i> | | 2 | sh | | | | | | | | | | | | | | | | | |
| <i>Zapada</i> | | 2 | sh | | | | | | | | | | | | | | | | | |
| Peltoperlidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Soliperla sp.</i> | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| Perlidae | | 1 | p | 0 | 0 | 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hesperoperla sp.</i> | | 2 | p | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 13 | 11 | 3 | 6 | 0 | 5 | 0 | 6 |
| <i>Calineuria californica</i> | | 2 | p | 0 | 0 | 1 | 3 | 2 | 17 | 74 | 26 | 56 | 33 | 25 | 18 | 24 | 10 | 100 | 2 | 0 |
| Perlodidae | | 2 | p | 2 | 0 | 1 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Baumannella alameda</i> | | 2 | | | | | | | | | | | | | | | | | | |
| <i>Cultus sp.</i> | | 2 | p | 2 | 0 | 1 | 2 | 1 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Isoperla sp.</i> | | 2 | p | 16 | 0 | 23 | 48 | 18 | 3 | 0 | 2 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 13 | 0 |
| <i>Kogotus sp.</i> | | 2 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Osobenus yakimae</i> | | 2 | p | | | | | | | | | | | | | | | | | |
| <i>Rickera sorpta</i> | | 2 | p | | | | | | | | | | | | | | | | | |
| Pteronarcyidae | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pteronarcella sp.</i> | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 0 | 0 | 0 | 0 |
| <i>Pteronarcys sp.</i> | | 0 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 |
| Taeniopterygidae | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Taenionema sp.</i> | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Trichoptera | | 3 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Apataniidae | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Apatania sp.</i> | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| Brachycentridae | | 1 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Brachycentrus sp.</i> | | 1 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Micrasema sp.</i> | | 1 | mh | 4 | 0 | 8 | 2 | 77 | 11 | 0 | 12 | 0 | 45 | 23 | 3 | 7 | 0 | 1 | 1 | 77 |
| <i>Amiocentrus aspilus</i> | | 3 | cg | 4 | 1 | 7 | 3 | 27 | 2 | 0 | 6 | 0 | 2 | 7 | 0 | 1 | 0 | 0 | 4 | 5 |
| Calamoceratidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Heteroplectron</i> | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Glossosomatidae | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Agapetus sp.</i> | | 0 | sc | 1 | 3 | 9 | 8 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 11 | 0 |
| <i>Anagapetus sp.</i> | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Culoptila sp.</i> | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Glossosoma sp.</i> | | 1 | sc | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| <i>Protoptila sp.</i> | | 1 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helicopsyche borealis</i> | | 3 | sc | | | | | | | | | | | | | | | | | |
| Hydropsychidae | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Arctopsyche sp.</i> | | 1 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cheumatopsyche sp.</i> | | 5 | cf | 8 | 0 | 4 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| <i>Homoplectra</i> | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydropsyche sp.</i> | | 4 | cf | 25 | 0 | 24 | 35 | 17 | 4 | 0 | 8 | 1 | 40 | 16 | 11 | 4 | 1 | 1 | 9 | 3 |
| <i>Parapsyche sp.</i> | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydropsychidae | Pupae | | | | | | | | | | | | | | | | | | | |
| Hydroptilidae | | 4 | ph | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | SGR-040 | SGR-060 | SGR-075 | SGR-080 | SGR-090 | SGR-110 | SGR-120 | SGR-130 | SGR-150 | STE-020 | STE-030 | STE-040 | STE-060 | STE-070 | STE-100 | STE-110 | STE-120 | |
|------------------|-------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | | |
| | <i>Paraperla sp.</i> | | 0 | p | | | | | | | | | | | | | | | | | | |
| | <i>Suwallia sp.</i> | | 1 | p | | | | | | | | | | | | | | | | | | |
| | <i>Sweltsa sp.</i> | | 1 | p | | | | | | | | | | | | | | | | | | |
| Leuctridae | | | 0 | sh | | | | | | | | | | | | | | | | | | |
| | <i>Despaxia augusta</i> | | 0 | sh | | | | | | | | | | | | | | | | | | |
| | <i>Paraleuctra sp.</i> | | 0 | sh | | | | | | | | | | | | | | | | | | |
| | <i>Perlomyia sp.</i> | | 0 | sh | | | | | | | | | | | | | | | | | | |
| Nemouridae | | | 2 | sh | | | | | | | | | | | | | | | | | | |
| | <i>Malenka sp.</i> | | 2 | sh | | | | | | | | | | | | | | | | | | |
| | <i>Soyedina</i> | | 2 | sh | | | | | | | | | | | | | | | | | | |
| | <i>Zapada</i> | | 2 | sh | | | | | | | | | | | | | | | | | | |
| Peltoperlidae | | | 1 | sh | | | | | | | | | | | | | | | | | | |
| | <i>Soliperla sp.</i> | | 1 | sh | | | | | | | | | | | | | | | | | | |
| Perlidae | | | 1 | p | | | | | | | | | | | | | | | | | | |
| | <i>Hesperoperla sp.</i> | | 2 | p | | | | | | | | | | | | | | | | | | |
| | <i>Calineuria californica</i> | | 2 | p | | | | | | | | | | | | | | | | | | |
| Perlodidae | | | 2 | p | | | | | | | | | | | | | | | | | | |
| | <i>Baumannella alameda</i> | | 2 | | | | | | | | | | | | | | | | | | | |
| | <i>Cultus sp.</i> | | 2 | p | | | | | | | | | | | | | | | | | | |
| | <i>Isoperla sp.</i> | | 2 | p | | | | | | | | | | | | | | | | | | |
| | <i>Kogotus sp.</i> | | 2 | p | | | | | | | | | | | | | | | | | | |
| | <i>Osobenus yakimae</i> | | 2 | p | | | | | | | | | | | | | | | | | | |
| | <i>Rickera sorpta</i> | | 2 | p | | | | | | | | | | | | | | | | | | |
| Pteronarcyidae | | | 0 | om | | | | | | | | | | | | | | | | | | |
| | <i>Pteronarcella sp.</i> | | 0 | om | | | | | | | | | | | | | | | | | | |
| | <i>Pteronarcys sp.</i> | | 0 | om | | | | | | | | | | | | | | | | | | |
| Taeniopterygidae | | | 2 | om | | | | | | | | | | | | | | | | | | |
| | <i>Taenionema sp.</i> | | 2 | om | | | | | | | | | | | | | | | | | | |
| Trichoptera | | | 3 | | | | | | | | | | | | | | | | | | | |
| Apataniidae | | | 1 | sc | | | | | | | | | | | | | | | | | | |
| | <i>Apatania sp.</i> | | 1 | sc | | | | | | | | | | | | | | | | | | |
| Brachycentridae | | | 1 | | | | | | | | | | | | | | | | | | | |
| | <i>Brachycentrus sp.</i> | | 1 | om | | | | | | | | | | | | | | | | | | |
| | <i>Micrasema sp.</i> | | 1 | mh | | | | | | | | | | | | | | | | | | |
| | <i>Amiocentrus aspilus</i> | | 3 | cg | | | | | | | | | | | | | | | | | | |
| Calamoceratidae | | | 1 | sh | | | | | | | | | | | | | | | | | | |
| | <i>Heteroplectron</i> | | 1 | sh | | | | | | | | | | | | | | | | | | |
| Glossosomatidae | | | 0 | sc | | | | | | | | | | | | | | | | | | |
| | <i>Agapetus sp.</i> | | 0 | sc | | | | | | | | | | | | | | | | | | |
| | <i>Anagapetus sp.</i> | | 0 | sc | | | | | | | | | | | | | | | | | | |
| | <i>Culoptila sp.</i> | | 2 | sc | | | | | | | | | | | | | | | | | | |
| | <i>Glossosoma sp.</i> | | 1 | sc | | | | | | | | | | | | | | | | | | |
| | <i>Protoptila sp.</i> | | 1 | sc | | | | | | | | | | | | | | | | | | |
| | <i>Helicopsyche borealis</i> | | 3 | sc | | | | | | | | | | | | | | | | | | |
| Hydropsychidae | | | 4 | cf | | | | | | | | | | | | | | | | | | |
| | <i>Arctopsyche sp.</i> | | 1 | p | | | | | | | | | | | | | | | | | | |
| | <i>Cheumatopsyche sp.</i> | | 5 | cf | | | | | | | | | | | | | | | | | | |
| | <i>Homoplectra</i> | | 4 | cf | | | | | | | | | | | | | | | | | | |
| | <i>Hydropsyche sp.</i> | | 4 | cf | | | | | | | | | | | | | | | | | | |
| | <i>Parapsyche sp.</i> | | 0 | p | | | | | | | | | | | | | | | | | | |
| Hydropsychidae | | Pupae | | | | | | | | | | | | | | | | | | | | |
| Hydroptilidae | | | 4 | ph | | | | | | | | | | | | | | | | | | |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | WLK-030 | WLK-050 | WLK-120 | WLK-100 | WLK-130 | WLK-140 | WLK-160 | WLK-170 | WLK-180 | WLK-190 | WLK-200 | WLK-230 | WLK-240 | SUI-010 | SUI-020 | SUI-050 | |
|----------------------------|------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|--|
| | | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | |
| Phylum | Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | |
| | | | 6 | ph | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 5 | |
| | | | 4 | ph | 0 | 0 | 0 | 0 | 0 | 1 | 22 | 8 | 1 | 0 | 0 | 0 | 0 | 13 | 4 | 11 | |
| | | | 3 | ph | 1 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 1 | sh | 0 | 0 | 1 | 0 | 0 | 6 | 13 | 21 | 1 | 2 | 3 | 0 | 0 | 0 | 0 | 0 | |
| | | Pupae | 1 | sh | | | | | | | | | | | | | | | | | |
| | | | 4 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 4 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 1 | sh | | | | | | | | | | | | | | | | | |
| | | | 1 | om | | | | | | | | | | | | | | | | | |
| | | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 1 | sh | | | | | | | | | | | | | | | | | |
| | | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 2 | sh | | | | | | | | | | | | | | | | | |
| | | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 3 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | |
| | | | 2 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 3 | cf | 0 | 0 | 0 | 29 | 0 | 80 | 0 | 0 | 1 | 4 | 0 | 0 | 0 | 26 | 20 | 4 | |
| | | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 0 | p | 1 | 0 | 0 | 2 | 0 | 4 | 6 | 5 | 10 | 22 | 8 | 1 | 0 | 0 | 0 | 0 | |
| | | Pupae | 0 | p | | | | | | | | | | | | | | | | | |
| | | | 3 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 3 | sh | 0 | 0 | 1 | 2 | 0 | 44 | 43 | 34 | 118 | 264 | 33 | 0 | 0 | 0 | 0 | 0 | |
| | | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 3 | ng | 0 | 0 | 0 | 0 | 0 | 1 | 7 | 1 | 32 | 18 | 41 | 0 | 0 | 0 | 0 | 0 | |
| | | | 0 | ng | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| Subphylum Crustacea | | | | | | | | | | | | | | | | | | | | | |
| Malacostraca | | | | | | | | | | | | | | | | | | | | | |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 16 | 72 | 0 | |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 4 | cg | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 6 | cg | | | | | | | | | | | | | | | | | |
| | | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 8 | cg | 8 | 141 | 215 | 0 | 265 | 1 | 2 | 7 | 4 | 0 | 4 | 0 | 0 | 0 | 0 | 1 | |
| | | | 8 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 6 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| | | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | SUI-060 | SUI-110 | SUI-130 | SUI-180 | SUI-210 | SUI-260 | SPA-020 | SPA-050 | SPA-070 | SPA-100 | SPA-110 | SPA-130 | SPA-140 | SPA-150 | SPA-160 | SPA-170 | SPA-200 |
|--------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | |
| <i>Hydroptila sp.</i> | | 6 | ph | 0 | 0 | 4 | 0 | 0 | 3 | 4 | 1 | 1 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 1 |
| <i>Ochrotrichia sp.</i> | | 4 | ph | 11 | 1 | 3 | 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Oxyethira sp.</i> | | 3 | ph | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepidostomatidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lepidostoma sp.</i> | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 5 | 206 | 88 | 4 | 0 |
| <i>Lepidostoma sp.</i> | Pupae | 1 | sh | | | | | | | | | | | | | | | | | |
| Leptoceridae | | 4 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Oecetis disjuncta</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Limnephilidae | | 4 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Amphicomoeus canax</i> | | 1 | sh | | | | | | | | | | | | | | | | | |
| <i>Dicosmoecus gilvipes</i> | | 1 | om | | | | | | | | | | | | | | | | | |
| <i>Ecclisomyia sp.</i> | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydatophylax hesperus</i> | | 1 | sh | | | | | | | | | | | | | | | | | |
| <i>Onocosmoecus sp.</i> | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Psychoglypha</i> | | 2 | sh | | | | | | | | | | | | | | | | | |
| Odontoceridae | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Parthina sp.</i> | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Philopotamidae | | 3 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Chimarra sp.</i> | | 4 | cf | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dolophilodes sp.</i> | | 2 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Wormaldia sp.</i> | | 3 | cf | 12 | 5 | 1 | 5 | 8 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Polycentropodidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Polycentropus sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Psychomyiidae | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Timodes sp.</i> | | 2 | sc | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Rhyacophilidae | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Rhyacophila sp.</i> | | 0 | p | 0 | 1 | 0 | 0 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 |
| <i>Rhyacophila sp.</i> | Pupae | 0 | p | | | | | | | | | | | | | | | | | |
| Sericostomatidae | | 3 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gumaga sp.</i> | | 3 | sh | 0 | 0 | 1 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 |
| Uenoidae | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Neophylax sp.</i> | | 3 | g | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Neothremma sp.</i> | | 0 | g | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Subphylum Crustacea | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Malacostraca | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Amphipoda</u> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corophiidae | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Corophium sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Crangonyctidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Crangonyx sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| <i>Stygobromus sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Synurella sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 13 | 2 | 0 |
| Gammaridae | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Eogammarus sp.</i> | | 6 | cg | | | | | | | | | | | | | | | | | |
| <i>Gammarus sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 57 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hyalellidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hyalella</i> | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Decapoda | | 8 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Astacidae | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pacifasticus lenisculus</i> | | 6 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Atyidae | | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Syncaris</i> | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | SPA-220 | SPA-240 | LAG-170 | LAG-130 | LAG-150 | LAG-160 | LAG-165 | LAG-180 | LAG-190 | LAG-210 | LAG-220 | LAG-240 | LAG-270 | LAG-290 | LAG-300 | LAG-320 |
|--------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | |
| <i>Hydroptila sp.</i> | | 6 | ph | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 8 | 0 |
| <i>Ochrotrichia sp.</i> | | 4 | ph | 0 | 0 | 0 | 10 | 2 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Oxyethira sp.</i> | | 3 | ph | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepidostomatidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lepidostoma sp.</i> | | 1 | sh | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 2 | 0 | 1 | 0 |
| <i>Lepidostoma sp.</i> | Pupae | 1 | sh | | | | | | | | | | | | | | | | |
| Leptoceridae | | 4 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Oecetis disjuncta</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Limnephilidae | | 4 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Amphicmoecus canax</i> | | 1 | sh | | | | | | | | | | | | | | | | |
| <i>Dicosmoecus gilvipes</i> | | 1 | om | | | | | | | | | | | | | | | | |
| <i>Ecclisomyia sp.</i> | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydatophylax hesperus</i> | | 1 | sh | | | | | | | | | | | | | | | | |
| <i>Onocosmoecus sp.</i> | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Psychoglypha</i> | | 2 | sh | | | | | | | | | | | | | | | | |
| Odontoceridae | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Parthina sp.</i> | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 11 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 |
| Philopotamidae | | 3 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Chimarra sp.</i> | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dolophilodes sp.</i> | | 2 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Wormaldia sp.</i> | | 3 | cf | 0 | 0 | 3 | 13 | 0 | 0 | 28 | 25 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 |
| Polycentropodidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Polycentropus sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| Psychomyiidae | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tinodes sp.</i> | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Rhyacophilidae | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Rhyacophila sp.</i> | | 0 | p | 0 | 0 | 4 | 5 | 0 | 0 | 11 | 16 | 44 | 13 | 21 | 27 | 22 | 16 | 8 | 10 |
| <i>Rhyacophila sp.</i> | Pupae | 0 | p | | | | | | | | | | | | | | | | |
| Sericostomatidae | | 3 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gumaga sp.</i> | | 3 | sh | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 31 | 2 | 0 | 22 | 18 | 11 | 1 | 15 | 0 |
| Uenoidae | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Neophylax sp.</i> | | 3 | g | 0 | 0 | 0 | 3 | 0 | 0 | 2 | 4 | 6 | 3 | 21 | 4 | 4 | 0 | 1 | 2 |
| <i>Neothrenna sp.</i> | | 0 | g | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Subphylum Crustacea | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Malacostraca | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Amphipoda | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corophiidae | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Corophium sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Crangonyctidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Crangonyx sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stygobromus sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Synurella sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gammaridae | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Eogammarus sp.</i> | | 6 | cg | | | | | | | | | | | | | | | | |
| <i>Gammarus sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hyalellidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hyalella</i> | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Decapoda | | 8 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Astacidae | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pacifasticus lenisculus</i> | | 6 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Atyidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Syncaris</i> | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | | | | | | | | | | | | | | | | | |
|----------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | LAG-330 | LAG-335 | LAG-380 | LAG-390 | ALP-010 | ALP-040 | ALP-070 | ALP-080 | ALP-100 | ALP-110 | ALP-140 | SLE-030 | SLE-170 | SLE-180 | SLE-210 | SLE-220 | SLE-230 | |
| Year: | | | | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | |
| | | | | | | | | | | | | | | | | | | | | | |
| | | 6 | ph | 0 | 0 | 1 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | ph | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 3 | ph | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | sh | 21 | 0 | 7 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 6 | 41 |
| | Pupae | 1 | sh | | | | | | | | | | | | | | | | | | |
| | | 4 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | sh | | | | | | | | | | | | | | | | | | |
| | | 1 | om | | | | | | | | | | | | | | | | | | |
| | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | sh | | | | | | | | | | | | | | | | | | |
| | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| | | 2 | sh | | | | | | | | | | | | | | | | | | |
| | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | sh | 90 | 21 | 2 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 13 | 27 | 0 | 4 | 91 | |
| | | 3 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 2 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 3 | cf | 2 | 1 | 3 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 6 | p | 7 | 42 | 3 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | p | 74 | 51 | 75 | 26 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 35 | 0 | 108 | 43 | |
| | Pupae | 0 | p | | | | | | | | | | | | | | | | | | |
| | | 3 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 3 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 3 | g | 0 | 1 | 2 | 31 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 15 | 0 | 0 | 15 | |
| | | 0 | g | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 67 |
| Subphylum Crustacea | | | | | | | | | | | | | | | | | | | | | |
| Malacostraca | | | | | | | | | | | | | | | | | | | | | |
| | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cg | 2 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 |
| | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 6 | cg | | | | | | | | | | | | | | | | | | |
| | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 8 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 6 | om | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | | | | | | | | | | | | | | | | | |
|----------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | WIL-100 | WIL-170 | WIL-190 | BUT-010 | BUT-020 | BUT-030 | BUT-040 | BUT-050 | PER-010 | PER-020 | PER-030 | PER-040 | PER-050 | PER-070 | PER-080 | PES-050 | PES-060 | |
| Year: | | | | Y1 | Y1 | Y1 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | |
| | | | | | | | | | | | | | | | | | | | | | |
| | | 6 | ph | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| | | 4 | ph | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 |
| | | 3 | ph | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | sh | 0 | 0 | 2 | 0 | 1 | 0 | 2 | 11 | 0 | 0 | 0 | 0 | 0 | 9 | 72 | 0 | 4 | 0 |
| | Pupae | 1 | sh | | | | | | | | | | | | | | | | | | |
| | | 4 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | sh | | | | | | | | | | | | | | | | | | |
| | | 1 | om | | | | | | | | | | | | | | | | | | |
| | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 1 | sh | | | | | | | | | | | | | | | | | | |
| | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 2 | sh | | | | | | | | | | | | | | | | | | |
| | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | sh | 0 | 0 | 160 | 0 | 0 | 0 | 3 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 3 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 2 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 3 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 0 | 0 | 0 |
| | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 6 | p | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | p | 2 | 18 | 40 | 0 | 3 | 28 | 7 | 12 | 0 | 0 | 1 | 0 | 1 | 1 | 43 | 0 | 1 | 0 |
| | Pupae | 0 | p | | | | | | | | | | | | | | | | | | |
| | | 3 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 3 | sh | 2 | 0 | 0 | 9 | 6 | 2 | 1 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 23 | 0 |
| | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 3 | g | 0 | 1 | 0 | 0 | 2 | 5 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | g | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Subphylum Crustacea | | | | | | | | | | | | | | | | | | | | | |
| Malacostraca | | | | | | | | | | | | | | | | | | | | | |
| | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |
| | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cg | 0 | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 6 | cg | | | | | | | | | | | | | | | | | | |
| | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 8 | cg | 3 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 8 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 6 | | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 6 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | PES-070 | PES-080 | PES-095 | PES-100 | PES-120 | PES-140 | PES-150 | PES-160 | PES-170 | PES-180 | PES-190 | PES-200 | PES-210 | PES-230 | PES-240 | SGR-010 | SGR-030 |
|--------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | |
| <i>Hydroptila sp.</i> | | 6 | ph | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 2 | 2 |
| <i>Ochrotrichia sp.</i> | | 4 | ph | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| <i>Oxyethira sp.</i> | | 3 | ph | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepidostomatidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lepidostoma sp.</i> | | 1 | sh | 3 | 0 | 22 | 27 | 2 | 1 | 21 | 0 | 19 | 4 | 0 | 21 | 0 | 0 | 6 | 35 | 11 |
| <i>Lepidostoma sp.</i> | Pupae | 1 | sh | | | | | | | | | | | | | | | | | |
| Leptoceridae | | 4 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Oecetis disjuncta</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| Limnephilidae | | 4 | s | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 3 | 0 | 1 | 1 | 0 | 0 | 0 | 0 |
| <i>Amphicomoeus canax</i> | | 1 | sh | | | | | | | | | | | | | | | | | |
| <i>Dicosmoecus gilvipes</i> | | 1 | om | | | | | | | | | | | | | | | | | |
| <i>Ecclisomyia sp.</i> | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydatophylax hesperus</i> | | 1 | sh | | | | | | | | | | | | | | | | | |
| <i>Onocosmoecus sp.</i> | | 1 | sh | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Psychoglypha</i> | | 2 | sh | | | | | | | | | | | | | | | | | |
| Odontoceridae | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Parthina sp.</i> | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 1 | 8 | 0 | 26 | 4 | 2 | 70 | 20 | 3 | 14 | 0 | 0 |
| Philopotamidae | | 3 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| <i>Chimarra sp.</i> | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dolophilodes sp.</i> | | 2 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| <i>Wormaldia sp.</i> | | 3 | cf | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 18 | 0 | 0 |
| Polycentropodidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Polycentropus sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 |
| Psychomyiidae | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tinodes sp.</i> | | 2 | sc | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| Rhyacophilidae | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Rhyacophila sp.</i> | | 0 | p | 4 | 8 | 7 | 7 | 17 | 28 | 10 | 15 | 27 | 30 | 30 | 24 | 18 | 1 | 20 | 6 | 13 |
| <i>Rhyacophila sp.</i> | Pupae | 0 | p | | | | | | | | | | | | | | | | | |
| Sericostomatidae | | 3 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gumaga sp.</i> | | 3 | sh | 50 | 0 | 21 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 2 | 29 | 0 |
| Uenoidae | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Neophylax sp.</i> | | 3 | g | 2 | 0 | 2 | 8 | 9 | 8 | 0 | 19 | 0 | 4 | 2 | 30 | 12 | 14 | 8 | 0 | 21 |
| <i>Neothremma sp.</i> | | 0 | g | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Subphylum Crustacea | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Malacostraca | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Amphipoda | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corophiidae | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Corophium sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| Crangonyctidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Crangonyx sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stygobromus sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Synurella sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gammaridae | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Eogammarus sp.</i> | | 6 | cg | | | | | | | | | | | | | | | | | |
| <i>Gammarus sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hyalellidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hyalella</i> | | 8 | cg | 0 | 2 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Decapoda | | 8 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Astacidae | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pacifasticus lenisculus</i> | | 6 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Atyidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Syncaris</i> | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | SGR-040 | SGR-060 | SGR-075 | SGR-080 | SGR-090 | SGR-110 | SGR-120 | SGR-130 | SGR-150 | STE-020 | STE-030 | STE-040 | STE-060 | STE-070 | STE-100 | STE-110 | STE-120 | |
|--------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | | |
| <i>Hydropila sp.</i> | | 6 | ph | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 18 | 0 | 0 | 0 | 0 |
| <i>Ochrotichia sp.</i> | | 4 | ph | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Oxyethira sp.</i> | | 3 | ph | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepidostomatidae | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lepidostoma sp.</i> | | 1 | sh | 138 | 0 | 43 | 57 | 0 | 1 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 2 | |
| <i>Lepidostoma sp.</i> | Pupae | 1 | sh | | | | | | | | | | | | | | | | | | | |
| Leptoceridae | | 4 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Oecetis disjuncta</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Limnephilidae | | 4 | s | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Amphicomoeus canax</i> | | 1 | sh | | | | | | | | | | | | | | | | | | | |
| <i>Dicosmoecus gilvipes</i> | | 1 | om | | | | | | | | | | | | | | | | | | | |
| <i>Ecclisomyia sp.</i> | | 2 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | |
| <i>Hydatophylax hesperus</i> | | 1 | sh | | | | | | | | | | | | | | | | | | | |
| <i>Onocosmoecus sp.</i> | | 1 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Psychoglypha</i> | | 2 | sh | | | | | | | | | | | | | | | | | | | |
| Odontoceridae | | 0 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Parthina sp.</i> | | 0 | sh | 0 | 14 | 0 | 0 | 1 | 6 | 4 | 2 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 16 | 3 | |
| Philopotamidae | | 3 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Chimarra sp.</i> | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Dolophilodes sp.</i> | | 2 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Wormaldia sp.</i> | | 3 | cf | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Polycentropodidae | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Polycentropus sp.</i> | | 6 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | |
| Psychomyiidae | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Tinodes sp.</i> | | 2 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Rhyacophilidae | | 0 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Rhyacophila sp.</i> | | 0 | p | 22 | 32 | 33 | 28 | 59 | 15 | 17 | 17 | 45 | 0 | 0 | 0 | 0 | 0 | 0 | 33 | 68 | 18 | |
| <i>Rhyacophila sp.</i> | Pupae | 0 | p | | | | | | | | | | | | | | | | | | | |
| Sericostomatidae | | 3 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gumaga sp.</i> | | 3 | sh | 2 | 0 | 1 | 0 | 1 | 0 | 1 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Uenoidae | | 0 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Neophylax sp.</i> | | 3 | g | 2 | 6 | 17 | 3 | 6 | 4 | 10 | 0 | 33 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 6 | 1 | |
| <i>Neothremma sp.</i> | | 0 | g | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Subphylum Crustacea | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Malacostraca | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Amphipoda | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corophiidae | | 4 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Corophium sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Crangonyctidae | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Crangonyx sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 22 | 0 | 0 | 0 | 0 |
| <i>Stygobromus sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 3 | |
| <i>Synurella sp.</i> | | 4 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gammaridae | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Eogammarus sp.</i> | | 6 | cg | | | | | | | | | | | | | | | | | | | |
| <i>Gammarus sp.</i> | | 6 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hyalellidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hyalella</i> | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Decapoda | | 8 | sh | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Astacidae | | 6 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pacifasticus lenisculus</i> | | 6 | om | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Atyidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Syncaris</i> | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | WLK-030 | WLK-050 | WLK-120 | WLK-100 | WLK-130 | WLK-140 | WLK-160 | WLK-170 | WLK-180 | WLK-190 | WLK-200 | WLK-230 | WLK-240 | SUI-010 | SUI-020 | SUI-050 |
|------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | |
| Isopoda | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Asellidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Caecidota sp. | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sphaeromatidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gnorimosphaeroma | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ostracoda | | 8 | cg | 0 | 2 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 1 | 1 | 0 |
| Subphylum Chelicerata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Arachnida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Acari | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Oribatei sp. | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | <i>Site Code:</i> | | | | | | | | | | | | | | | | | |
|------------------------------|------------|----------------------|--------------------------------|-------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | SUI-060 | SUI-110 | SUI-130 | SUI-180 | SUI-210 | SUI-260 | SPA-020 | SPA-050 | SPA-070 | SPA-100 | SPA-110 | SPA-130 | SPA-140 | SPA-150 | SPA-160 | SPA-170 | SPA-200 | |
| | | | | <i>Year:</i> | | | | | | | | | | | | | | | | | |
| | | | | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | |
| <u>Isopoda</u> | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Asellidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Caecidota sp.</u> | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sphaeromatidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Gnorimosphaeroma</u> | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ostracoda | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 11 | 1 | 0 | 0 | 4 | 8 | 1 | 0 | 0 | 0 | 0 | 0 | 4 |
| Subphylum Chelicerata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Arachnida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Acari</u> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Oribatei sp.</u> | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | <i>Site Code:</i> | SPA-220 | SPA-240 | LAG-170 | LAG-130 | LAG-150 | LAG-160 | LAG-165 | LAG-180 | LAG-190 | LAG-210 | LAG-220 | LAG-240 | LAG-270 | LAG-290 | LAG-300 | LAG-320 |
|------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | <i>Year:</i> | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | |
| Isopoda | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Asellidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Caecidota sp.</i> | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sphaeromatidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gnorimosphaeroma</i> | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ostracoda | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 3 | 6 | 0 | 0 | 1 | 3 | 20 |
| Subphylum Chelicerata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Arachnida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Acari | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Oribatei sp.</i> | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | <i>Site Code:</i> | | | | | | | | | | | | | | | | | |
|------------------------------|------------|----------------------|--------------------------------|-------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---|
| | | | | LAG-330 | LAG-335 | LAG-380 | LAG-390 | ALP-010 | ALP-040 | ALP-070 | ALP-080 | ALP-100 | ALP-110 | ALP-140 | SLE-030 | SLE-170 | SLE-180 | SLE-210 | SLE-220 | SLE-230 | |
| | | | | <i>Year:</i> | | | | | | | | | | | | | | | | | |
| | | | | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | |
| Isopoda | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Asellidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Caecidota sp.</i> | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sphaeromatidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gnorimosphaeroma</i> | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ostracoda | | 8 | cg | 9 | 0 | 0 | 5 | 2 | 0 | 1 | 1 | 4 | 10 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| Subphylum Chelicerata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Arachnida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Acari | | 5 | p | 5 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 |
| <i>Oribatei sp.</i> | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | | | | | | | | | | | | | | | | | |
|------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | WIL-100 | WIL-170 | WIL-190 | BUT-010 | BUT-020 | BUT-030 | BUT-040 | BUT-050 | PER-010 | PER-020 | PER-030 | PER-040 | PER-050 | PER-070 | PER-080 | PES-050 | PES-060 | |
| | | | | Year: | | | | | | | | | | | | | | | | | |
| | | | | Y1 | Y1 | Y1 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | |
| Isopoda | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Asellidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Caecidota sp.</i> | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sphaeromatidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gnoringosphaeroma</i> | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 488 | 0 |
| Ostracoda | | 8 | cg | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 59 | 1 | 3 | 3 | 0 | 7 | 4 | 0 | 0 | 0 |
| Subphylum Chelicerata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Arachnida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Acari | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Oribatei sp.</i> | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | <i>Site Code:</i> | PES-070 | PES-080 | PES-095 | PES-100 | PES-120 | PES-140 | PES-150 | PES-160 | PES-170 | PES-180 | PES-190 | PES-200 | PES-210 | PES-230 | PES-240 | SGR-010 | SGR-030 |
|------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | <i>Year:</i> | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum Arthropoda | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | |
| Isopoda | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Asellidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Caecidota sp. | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sphaeromatidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gnoringosphaeroma | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 272 | 0 | 29 | 8 | 0 | 0 | 100 | 1 | 15 | 0 | 0 |
| Ostracoda | | 8 | cg | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Subphylum Chelicerata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Arachnida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Acari | | 5 | p | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| Oribatei sp. | | 5 | | 0 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | <i>Site Code:</i> | | | | | | | | | | | | | | | | | |
|------------------------------|------------|----------------------|--------------------------------|-------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | SGR-040 | SGR-060 | SGR-075 | SGR-080 | SGR-090 | SGR-110 | SGR-120 | SGR-130 | SGR-150 | STE-020 | STE-030 | STE-040 | STE-060 | STE-070 | STE-100 | STE-110 | STE-120 | |
| <i>Year:</i> | | | | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | |
| Isopoda | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Asellidae | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Caecidota sp.</i> | | 8 | cg | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 |
| Sphaeromatidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gnorimosphaeroma</i> | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ostracoda | | 8 | cg | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 12 | 0 | 3 | 1 | 0 | 0 | 0 | 0 | 0 |
| Subphylum Chelicerata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Arachnida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Acari | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Oribatei sp.</i> | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | WLK-030 | WLK-050 | WLK-120 | WLK-100 | WLK-130 | WLK-140 | WLK-160 | WLK-170 | WLK-180 | WLK-190 | WLK-200 | WLK-230 | WLK-240 | SUI-010 | SUI-020 | SUI-050 | |
|-------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | |
| Anisitsiellidae / Mideopsidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydriphantiidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Protzia sp.</u> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hygrobatidae | | 8 | p | 0 | 0 | 0 | 1 | 0 | 3 | 2 | 4 | 0 | 4 | 1 | 65 | 3 | 4 | 5 | 4 | | |
| <u>Atractides sp.</u> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Corticacarus sp.</u> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Hygrobates sp.</u> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Mesobates sp.</u> | | 8 | p | | | | | | | | | | | | | | | | | | |
| Lebertiidae | | 8 | p | 2 | 13 | 0 | 0 | 0 | 6 | 0 | 11 | 5 | 10 | 4 | 0 | 1 | 7 | 0 | 0 | | |
| <u>Lebertia sp.</u> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sperchontidae | | 8 | p | 0 | 1 | 1 | 0 | 0 | 2 | 2 | 6 | 1 | 15 | 0 | 0 | 32 | 22 | 22 | | | |
| <u>Sperchon sp.</u> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Sperchonopsis sp.</u> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Stygothrombiidae | | 5 | p | | | | | | | | | | | | | | | | | | |
| <u>Stygothrombium sp.</u> | | 5 | p | | | | | | | | | | | | | | | | | | |
| Torrenticolidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 5 | 2 | 0 | 1 | 70 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Testudacarus sp.</u> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Torrenticola sp.</u> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Unidentified #1 | | 5 | p | | | | | | | | | | | | | | | | | | |
| Phylum Annelida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aclitellata | | | | | | | | | | | | | | | | | | | | | |
| Polychaeta | | | | | | | | | | | | | | | | | | | | | |
| <u>Aciculata</u> | | | | | | | | | | | | | | | | | | | | | |
| Nereididae | | | cf | | | | | | | | | | | | | | | | | | |
| Subphylum Clitellata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hirudinea | | 10 | pa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pharyngobdellida | | | | | | | | | | | | | | | | | | | | | |
| Erpobdellidae | | 8 | p | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Rhynchobdellida</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Glossiphoniidae | | 8 | pa | 0 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Helobdella</u> | | 6 | pa | | | | | | | | | | | | | | | | | | |
| <u>Placobdella sp.</u> | | 6 | pa | | | | | | | | | | | | | | | | | | |
| Oligochaeta | | 5 | cg | 194 | 157 | 85 | 23 | 43 | 0 | 3 | 37 | 12 | 3 | 23 | 36 | 27 | 104 | 47 | 18 | | |
| <u>Megadrili</u> | | 5 | | 0 | 0 | 2 | 0 | 0 | 2 | 0 | 1 | 4 | 0 | 3 | 0 | 0 | 0 | 3 | 0 | | |
| Phylum Coelenterata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrozoa | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydroida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydridae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Hydra sp.</u> | | 5 | p | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Mollusca | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Bivalvia | | | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 2 | 1 | 0 | | |
| <u>Veneroidea</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corbiculidae | | 10 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Corbicula sp.</u> | | 10 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sphaeriidae | | 8 | cf | 20 | 1 | 4 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Pisidium sp.</u> | | 8 | cf | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Sphaerium sp.</u> | | 8 | gc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gastropoda | | 8 | g | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 0 | | |
| Ancylidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Ferrissia sp.</u> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrobiidae | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lymnaeidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | SUI-060 | SUI-110 | SUI-130 | SUI-180 | SUI-210 | SUI-260 | SPA-020 | SPA-050 | SPA-070 | SPA-100 | SPA-110 | SPA-130 | SPA-140 | SPA-150 | SPA-160 | SPA-170 | SPA-200 | |
|-------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | | |
| Anisitsiellidae / Mideopsidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydryphantidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Protzia sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hygrobatidae | | 8 | p | 9 | 2 | 1 | 11 | 3 | 32 | 0 | 0 | 1 | 0 | 1 | 3 | 3 | 2 | 2 | 2 | 3 | 0 | 0 |
| <i>Atractides sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Corticacarus sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hygrobates sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Mesobates sp.</i> | | 8 | p | | | | | | | | | | | | | | | | | | | |
| Lebertiidae | | 8 | p | 4 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| <i>Lebertia sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sperchontidae | | 8 | p | 11 | 2 | 14 | 0 | 4 | 14 | 2 | 2 | 14 | 4 | 0 | 2 | 20 | 43 | 34 | 9 | 3 | | |
| <i>Sperchon sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sperchonopsis sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Stygothrombidiidae | | 5 | p | | | | | | | | | | | | | | | | | | | |
| <i>Stygothrombium sp.</i> | | 5 | p | | | | | | | | | | | | | | | | | | | |
| Torrenticolidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 67 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Testudacarus sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Torrenticola sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Unidentified #1 | | 5 | p | | | | | | | | | | | | | | | | | | | |
| Phylum Annelida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aclitellata | | | | | | | | | | | | | | | | | | | | | | |
| Polychaeta | | | | | | | | | | | | | | | | | | | | | | |
| <u>Aciculata</u> | | | | | | | | | | | | | | | | | | | | | | |
| Nereididae | | | cf | | | | | | | | | | | | | | | | | | | |
| Subphylum Clitellata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hirudinea | | 10 | pa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pharyngobdellida | | | | | | | | | | | | | | | | | | | | | | |
| Erpobdellidae | | 8 | p | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Rhynchobdellida</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Glossiphoniidae | | 8 | pa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helobdella</i> | | 6 | pa | | | | | | | | | | | | | | | | | | | |
| <i>Placobdella sp.</i> | | 6 | pa | | | | | | | | | | | | | | | | | | | |
| Oligochaeta | | 5 | cg | 3 | 8 | 27 | 24 | 0 | 130 | 366 | 330 | 85 | 335 | 324 | 28 | 3 | 1 | 22 | 28 | 31 | | |
| <i>Megadrili</i> | | 5 | | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 5 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | |
| Phylum Coelenterata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrozoa | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Hydroida</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydridae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydra sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Mollusca | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Bivalvia | | | cf | 1 | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Veneroidea</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corbiculidae | | 10 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Corbicula sp.</i> | | 10 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sphaeriidae | | 8 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pisidium sp.</i> | | 8 | cf | 0 | 0 | 1 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sphaerium sp.</i> | | 8 | gc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gastropoda | | 8 | g | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ancylidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ferrissia sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrobiidae | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lymnaeidae | | 6 | sc | 0 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | SPA-220 | SPA-240 | LAG-170 | LAG-130 | LAG-150 | LAG-160 | LAG-165 | LAG-180 | LAG-190 | LAG-210 | LAG-220 | LAG-240 | LAG-270 | LAG-290 | LAG-300 | LAG-320 |
|-------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | |
| Anisitsiellidae / Mideopsidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydryphantidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Protzia sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hygrobatidae | | 8 | p | 0 | 3 | 0 | 0 | 5 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| <i>Atracides sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Corticacarus sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hygrobates sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Mesobates sp.</i> | | 8 | p | | | | | | | | | | | | | | | | |
| Lebertiidae | | 8 | p | 0 | 0 | 5 | 1 | 0 | 1 | 2 | 3 | 0 | 3 | 0 | 2 | 0 | 2 | 2 | 2 |
| <i>Lebertia sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sperchontidae | | 8 | p | 5 | 4 | 1 | 1 | 0 | 0 | 0 | 1 | 0 | 1 | 1 | 0 | 1 | 0 | 0 | 1 |
| <i>Sperchon sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sperchonopsis sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Stygothrombidiidae | | 5 | p | | | | | | | | | | | | | | | | |
| <i>Stygothrombium sp.</i> | | 5 | p | | | | | | | | | | | | | | | | |
| Torrenticolidae | | 5 | p | 0 | 0 | 3 | 0 | 11 | 0 | 0 | 14 | 11 | 0 | 4 | 3 | 1 | 1 | 3 | 3 |
| <i>Testudacarus sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Torrenticola sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Unidentified #1 | | 5 | p | | | | | | | | | | | | | | | | |
| Phylum Annelida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aclitellata | | | | | | | | | | | | | | | | | | | |
| Polychaeta | | | | | | | | | | | | | | | | | | | |
| <u>Aciculata</u> | | | | | | | | | | | | | | | | | | | |
| Nereididae | | | cf | | | | | | | | | | | | | | | | |
| Subphylum Clitellata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hirudinea | 10 | | pa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Pharyngobdellida</u> | | | | | | | | | | | | | | | | | | | |
| Erpobdellidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Rhynchobdellida</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Glossiphoniidae | | 8 | pa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helobdella</i> | | 6 | pa | | | | | | | | | | | | | | | | |
| <i>Placobdella sp.</i> | | 6 | pa | | | | | | | | | | | | | | | | |
| Oligochaeta | | 5 | cg | 197 | 71 | 4 | 0 | 10 | 0 | 35 | 3 | 14 | 0 | 1 | 2 | 11 | 12 | 2 | 0 |
| <i>Megadrili</i> | | 5 | | 3 | 10 | 0 | 0 | 1 | 1 | 6 | 1 | 0 | 16 | 1 | 0 | 1 | 2 | 2 | 5 |
| Phylum Coelenterata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrozoa | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Hydroida</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydridae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydra sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Mollusca | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Bivalvia | | | cf | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Veneroidea</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corbiculidae | | 10 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Corbicula sp.</i> | | 10 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sphaeriidae | | 8 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 5 | 1 | 0 | 0 | 0 | 0 |
| <i>Pisidium sp.</i> | | 8 | cf | 0 | 0 | 3 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sphaerium sp.</i> | | 8 | gc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| Gastropoda | | 8 | g | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ancylidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ferrissia sp.</i> | | 6 | sc | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrobiidae | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lymnaeidae | | 6 | sc | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | | | | | | | | | | | | | | | | | |
|-------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | LAG-330 | LAG-335 | LAG-380 | LAG-390 | ALP-010 | ALP-040 | ALP-070 | ALP-080 | ALP-100 | ALP-110 | ALP-140 | SLE-030 | SLE-170 | SLE-180 | SLE-210 | SLE-220 | SLE-230 | |
| Year: | | | | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | |
| Anisitsiellidae / Mideopsidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydryphantidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Protzia sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hygrobatidae | | 8 | p | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 4 | 0 | 7 | 0 | 0 |
| <i>Atracides sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Corticacarus sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hygrobates sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Mesobates sp.</i> | | 8 | p | | | | | | | | | | | | | | | | | | |
| Lebertiidae | | 8 | p | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 11 | 1 | 3 | 4 | 1 | 0 | 0 |
| <i>Lebertia sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sperchontidae | | 8 | p | 0 | 1 | 8 | 4 | 3 | 1 | 2 | 9 | 0 | 17 | 3 | 55 | 1 | 1 | 2 | 3 | 8 | 0 |
| <i>Sperchon sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sperchonopsis sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Stygothrombiidae | | 5 | p | | | | | | | | | | | | | | | | | | |
| <i>Stygothrombium sp.</i> | | 5 | p | | | | | | | | | | | | | | | | | | |
| Torrenticolidae | | 5 | p | 10 | 10 | 12 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Testudacarus sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Torrenticola sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Unidentified #1 | | 5 | p | | | | | | | | | | | | | | | | | | |
| Phylum Annelida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aclitellata | | | | | | | | | | | | | | | | | | | | | |
| Polychaeta | | | | | | | | | | | | | | | | | | | | | |
| <u>Aciculata</u> | | | | | | | | | | | | | | | | | | | | | |
| Nereididae | | | cf | | | | | | | | | | | | | | | | | | |
| Subphylum Clitellata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hirudinea | | 10 | pa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pharyngobdellida | | | | | | | | | | | | | | | | | | | | | |
| Erpobdellidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Rhynchobdellida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Glossiphoniidae | | 8 | pa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helobdella</i> | | 6 | pa | | | | | | | | | | | | | | | | | | |
| <i>Placobdella sp.</i> | | 6 | pa | | | | | | | | | | | | | | | | | | |
| Oligochaeta | | 5 | cg | 5 | 2 | 3 | 0 | 24 | 127 | 375 | 85 | 16 | 314 | 1 | 12 | 17 | 2 | 86 | 30 | 0 | 0 |
| <i>Megadrili</i> | | 5 | | 0 | 0 | 0 | 4 | 0 | 11 | 4 | 3 | 0 | 2 | 0 | 0 | 0 | 20 | 2 | 9 | 1 | 0 |
| Phylum Coelenterata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrozoa | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Hydroida</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydriidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydra sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Phylum Mollusca | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Bivalvia | | | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Veneroida</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corbiculidae | | 10 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Corbicula sp.</i> | | 10 | cf | 0 | 0 | 0 | 0 | 10 | 13 | 16 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sphaeriidae | | 8 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| <i>Pisidium sp.</i> | | 8 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sphaerium sp.</i> | | 8 | gc | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 2 | 1 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 1 | 0 |
| Gastropoda | | 8 | g | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ancylidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ferrissia sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrobiidae | | 8 | sc | 0 | 0 | 0 | 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 |
| Lymnaeidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | WIL-100 | WIL-170 | WIL-190 | BUT-010 | BUT-020 | BUT-030 | BUT-040 | BUT-050 | PER-010 | PER-020 | PER-030 | PER-040 | PER-050 | PER-070 | PER-080 | PES-050 | PES-060 |
|-------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y1 | Y1 | Y1 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | |
| Anisitsiellidae / Mideopsidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydraphantidae | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Protzia sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hygrobatidae | | 8 | p | 0 | 2 | 0 | 1 | 1 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 3 | 4 | 1 | 3 |
| <i>Atractides sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 8 | 0 | 0 |
| <i>Corticacarus sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hygrobates sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Mesobates sp.</i> | | 8 | p | | | | | | | | | | | | | | | | | |
| Lebertiidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lebertia sp.</i> | | 8 | p | 0 | 0 | 0 | 9 | 19 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 1 | 0 | 0 | 15 | 24 |
| Sperchontidae | | 8 | p | 2 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sperchon sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 5 | 0 | 8 | 3 | 10 | 3 | 4 |
| <i>Sperchonopsis sp.</i> | | 8 | p | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Stygothrombiidae | | 5 | p | | | | | | | | | | | | | | | | | |
| <i>Stygothrombium sp.</i> | | 5 | p | | | | | | | | | | | | | | | | | |
| Torrenticolidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Testudacarus sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Torrenticola sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| Unidentified #1 | | 5 | p | | | | | | | | | | | | | | | | | |
| Phylum Annelida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aclitellata | | | | | | | | | | | | | | | | | | | | |
| Polychaeta | | | | | | | | | | | | | | | | | | | | |
| <u>Aciculata</u> | | | | | | | | | | | | | | | | | | | | |
| Nereididae | | | cf | | | | | | | | | | | | | | | | | |
| Subphylum Clitellata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hirudinea | | 10 | pa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Pharyngobdellida</u> | | | | | | | | | | | | | | | | | | | | |
| Erpobdellidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Rhynchobdellida</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Glossiphoniidae | | 8 | pa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helobdella</i> | | 6 | pa | | | | | | | | | | | | | | | | | |
| <i>Placobdella sp.</i> | | 6 | pa | | | | | | | | | | | | | | | | | |
| Oligochaeta | | 5 | cg | 38 | 0 | 0 | 19 | 0 | 1 | 9 | 9 | 629 | 99 | 25 | 60 | 131 | 49 | 29 | 2 | 9 |
| <i>Megadrili</i> | | 5 | | 0 | 1 | 7 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 12 | 3 | 4 | 0 | 16 | 0 | 0 |
| Phylum Coelenterata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrozoa | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Hydroida</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydriadae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydra sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| Phylum Mollusca | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Bivalvia | | | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <u>Veneroidea</u> | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corbiculidae | | 10 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Corbicula sp.</i> | | 10 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sphaeriidae | | 8 | cf | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 2 | 0 | 0 |
| <i>Pisidium sp.</i> | | 8 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sphaerium sp.</i> | | 8 | gc | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gastropoda | | 8 | g | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ancylidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ferrissia sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrobiidae | | 8 | sc | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lymnaeidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | PES-070 | PES-080 | PES-095 | PES-100 | PES-120 | PES-140 | PES-150 | PES-160 | PES-170 | PES-180 | PES-190 | PES-200 | PES-210 | PES-230 | PES-240 | SGR-010 | SGR-030 |
|-------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | |
| Anisitsiellidae / Mideopsidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydriphantiidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Protzia sp. | | 8 | p | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| Hygrobatidae | | 8 | p | 2 | 0 | 4 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 11 | 0 |
| Atractides sp. | | 8 | p | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corticacarus sp. | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 11 | 0 |
| Hygrobates sp. | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 |
| Mesobates sp. | | 8 | p | | | | | | | | | | | | | | | | | |
| Lebertiidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lebertia sp. | | 8 | p | 3 | 3 | 2 | 3 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 44 | 2 |
| Sperchontidae | | 8 | p | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sperchon sp. | | 8 | p | 1 | 0 | 1 | 0 | 2 | 0 | 0 | 0 | 1 | 0 | 2 | 1 | 0 | 0 | 0 | 10 | 0 |
| Sperchonopsis sp. | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 2 | 0 | 0 | 1 | 5 | |
| Stygothrombiidae | | 5 | p | | | | | | | | | | | | | | | | | |
| Stygothrombium sp. | | 5 | p | | | | | | | | | | | | | | | | | |
| Torrenticolidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Testudacarus sp. | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Torrenticola sp. | | 5 | p | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 8 | 2 | 4 | 3 | 1 | 0 | 0 | 0 | 1 |
| Unidentified #1 | | 5 | p | | | | | | | | | | | | | | | | | |
| Phylum Annelida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aclitellata | | | | | | | | | | | | | | | | | | | | |
| Polychaeta | | | | | | | | | | | | | | | | | | | | |
| Aciculata | | | | | | | | | | | | | | | | | | | | |
| Nereididae | | | cf | | | | | | | | | | | | | | | | | |
| Subphylum Clitellata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hirudinea | | 10 | pa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pharyngobdellida | | | | | | | | | | | | | | | | | | | | |
| Erpobdellidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Rhynchobdellida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Glossiphoniidae | | 8 | pa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Helobdella | | 6 | pa | | | | | | | | | | | | | | | | | |
| Placobdella sp. | | 6 | pa | | | | | | | | | | | | | | | | | |
| Oligochaeta | | 5 | cg | 1 | 25 | 1 | 0 | 4 | 3 | 2 | 3 | 40 | 1 | 8 | 7 | 1 | 11 | 2 | 2 | 27 |
| Megadrili | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| Phylum Coelenterata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrozoa | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydroida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydriidae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydra sp. | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Mollusca | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Bivalvia | | | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Veneroida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corbiculidae | | 10 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corbicula sp. | | 10 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sphaeriidae | | 8 | cf | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pisidium sp. | | 8 | cf | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 |
| Sphaerium sp. | | 8 | gc | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gastropoda | | 8 | g | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ancylidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ferrissia sp. | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrobiidae | | 8 | sc | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| Lymnaeidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | <i>Site Code:</i> | | | | | | | | | | | | | | | | | |
|-------------------------------|------------|----------------------|--------------------------------|-------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | SGR-040 | SGR-060 | SGR-075 | SGR-080 | SGR-090 | SGR-110 | SGR-120 | SGR-130 | SGR-150 | STE-020 | STE-030 | STE-040 | STE-060 | STE-070 | STE-100 | STE-110 | STE-120 | |
| <i>Year:</i> | | | | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | | |
| Anisitsiellidae / Mideopsidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydryphantidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Protzia sp.</i> | | 8 | p | 2 | 5 | 6 | 0 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hygrobatidae | | 8 | p | 0 | 1 | 0 | 1 | 0 | 0 | 2 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Atractides sp.</i> | | 8 | p | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| <i>Corticacarus sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hygrobates sp.</i> | | 8 | p | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Mesobates sp.</i> | | 8 | p | | | | | | | | | | | | | | | | | | |
| Lebertiidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lebertia sp.</i> | | 8 | p | 7 | 6 | 3 | 2 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 10 | 5 | 8 | 0 | 3 | |
| Sperchontidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Sperchon sp.</i> | | 8 | p | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 1 | 11 | 31 | 3 | 1 | 0 | |
| <i>Sperchonopsis sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 0 |
| Stygothrombidiidae | | 5 | p | | | | | | | | | | | | | | | | | | |
| <i>Stygothrombium sp.</i> | | 5 | p | | | | | | | | | | | | | | | | | | |
| Torrenticolidae | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Testudacarus sp.</i> | | 5 | p | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Torrenticola sp.</i> | | 5 | p | 1 | 1 | 0 | 0 | 0 | 1 | 1 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | 1 | |
| Unidentified #1 | | 5 | p | | | | | | | | | | | | | | | | | | |
| Phylum Annelida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aclitellata | | | | | | | | | | | | | | | | | | | | | |
| Polychaeta | | | | | | | | | | | | | | | | | | | | | |
| Aciculata | | | | | | | | | | | | | | | | | | | | | |
| Nereididae | | | cf | | | | | | | | | | | | | | | | | | |
| Subphylum Clitellata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hirudinea | | 10 | pa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pharyngobdellida | | | | | | | | | | | | | | | | | | | | | |
| Erpobdellidae | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Rhynchobdellida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Glossiphoniidae | | 8 | pa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helobdella</i> | | 6 | pa | | | | | | | | | | | | | | | | | | |
| <i>Placobdella sp.</i> | | 6 | pa | | | | | | | | | | | | | | | | | | |
| Oligochaeta | | 5 | eg | 2 | 6 | 5 | 24 | 1 | 0 | 4 | 10 | 7 | 103 | 217 | 297 | 125 | 20 | 48 | 9 | 4 | |
| <i>Megadrili</i> | | 5 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 1 | 0 | 1 | 0 | 0 | 1 | 1 | 0 | |
| Phylum Coelenterata | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrozoa | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydroida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydridae | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Hydra sp.</i> | | 5 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10 | 0 | 0 | 0 | 0 |
| Phylum Mollusca | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Bivalvia | | | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Veneroida | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Corbiculidae | | 10 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Corbicula sp.</i> | | 10 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| Sphaeriidae | | 8 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| <i>Pisidium sp.</i> | | 8 | cf | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 |
| <i>Sphaerium sp.</i> | | 8 | gc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gastropoda | | 8 | g | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ancylidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Ferrissia sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydrobiidae | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lymnaeidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | WLK-030 | WLK-050 | WLK-120 | WLK-100 | WLK-130 | WLK-140 | WLK-160 | WLK-170 | WLK-180 | WLK-190 | WLK-200 | WLK-230 | WLK-240 | SUI-010 | SUI-020 | SUI-050 |
|-------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| <i>Coleoptera</i> | | | | | | | | | | | | | | | | | | | | |
| <i>Fossaria sp.</i> | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lanx sp.</i> | | 6 | gc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stagnicola</i> | | 10 | sc | | | | | | | | | | | | | | | | | |
| Physidae | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Physa/ Physella</i> | | 8 | sc | 0 | 0 | 2 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Planorbidae | | 6 | sc | 0 | 0 | 4 | 0 | 1 | 0 | 0 | 0 | 9 | 0 | 51 | 4 | 0 | 0 | 0 | 0 | 0 |
| <i>Gyraulus sp.</i> | | 8 | sc | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Menetus sp.</i> | | 6 | sc | 1 | 0 | 7 | 0 | 0 | 0 | 0 | 0 | 14 | 0 | 45 | 3 | 0 | 0 | 0 | 0 | 0 |
| <i>Planorbella sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Nematomorpha | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Nemertea | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Prostoma sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Platyhelminthes | | 4 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Turbellaria | | 4 | | 1 | 1 | 2 | 0 | 0 | 1 | 1 | 8 | 2 | 3 | 0 | 0 | 0 | 0 | 4 | 10 | 5 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | | | | | | | | | | | | | | | | | |
|-------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | SUI-060 | SUI-110 | SUI-130 | SUI-180 | SUI-210 | SUI-260 | SPA-020 | SPA-050 | SPA-070 | SPA-100 | SPA-110 | SPA-130 | SPA-140 | SPA-150 | SPA-160 | SPA-170 | SPA-200 | |
| | | | | Year: | | | | | | | | | | | | | | | | | |
| | | | | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | |
| <i>Fossaria sp.</i> | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lanx sp.</i> | | 6 | gc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stagnicola</i> | | 10 | sc | | | | | | | | | | | | | | | | | | |
| Physidae | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Physa/ Physella</i> | | 8 | sc | 0 | 13 | 0 | 3 | 1 | 5 | 0 | 0 | 0 | 0 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 10 |
| Planorbidae | | 6 | sc | 0 | 6 | 0 | 86 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gyraulus sp.</i> | | 8 | sc | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Menetus sp.</i> | | 6 | sc | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Planorbella sp.</i> | | 6 | sc | 0 | 2 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Nematomorpha | | 8 | | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Nemertea | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Prostoma sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Platyhelminthes | | 4 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Turbellaria | | 4 | | 2 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 10 | 2 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | SPA-220 | SPA-240 | LAG-170 | LAG-130 | LAG-150 | LAG-160 | LAG-165 | LAG-180 | LAG-190 | LAG-210 | LAG-220 | LAG-240 | LAG-270 | LAG-290 | LAG-300 | LAG-320 |
|-------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | |
| <i>Fossaria sp.</i> | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lanx sp.</i> | | 6 | gc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stagnicola</i> | | 10 | sc | | | | | | | | | | | | | | | | |
| Physidae | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Physa/ Physella</i> | | 8 | sc | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Planorbidae | | 6 | sc | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 21 | 2 | 0 |
| <i>Gyraulus sp.</i> | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Menetus sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Planorbella sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Nematomorpha | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| Phylum Nemertea | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Prostoma sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Platyhelminthes | | 4 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Turbellaria | | 4 | | 3 | 55 | 3 | 81 | 0 | 0 | 2 | 6 | 15 | 2 | 16 | 4 | 2 | 5 | 1 | 8 |

Appendix E-1: Bioassessment master taxa list

| | | | | <i>Site Code:</i> | | | | | | | | | | | | | | | | | |
|-------------------------------|------------|----------------------|--------------------------------|-------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | LAG-330 | LAG-335 | LAG-380 | LAG-390 | ALP-010 | ALP-040 | ALP-070 | ALP-080 | ALP-100 | ALP-110 | ALP-140 | SLE-030 | SLE-170 | SLE-180 | SLE-210 | SLE-220 | SLE-230 | |
| | | | | <i>Year:</i> | | | | | | | | | | | | | | | | | |
| | | | | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 | Y1 |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | |
| <i>Fossaria sp.</i> | | 8 | sc | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lanx sp.</i> | | 6 | gc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stagnicola</i> | | 10 | sc | | | | | | | | | | | | | | | | | | |
| Physidae | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Physa/ Physella</i> | | 8 | sc | 0 | 0 | 0 | 0 | 47 | 7 | 11 | 3 | 3 | 9 | 1 | 1 | 0 | 0 | 1 | 0 | 3 | |
| Planorbidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 45 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gyraulus sp.</i> | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Menetus sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Planorbella sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Nematomorpha | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Nemertea | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Prostoma sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Platyhelminthes | | 4 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Turbellaria | | 4 | | 13 | 0 | 27 | 34 | 37 | 4 | 14 | 8 | 61 | 5 | 0 | 150 | 0 | 0 | 7 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | | | | | | | | | | | | | | | | | |
|-------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | WIL-100 | WIL-170 | WIL-190 | BUT-010 | BUT-020 | BUT-030 | BUT-040 | BUT-050 | PER-010 | PER-020 | PER-030 | PER-040 | PER-050 | PER-070 | PER-080 | PES-050 | PES-060 | |
| | | | | Year: | | | | | | | | | | | | | | | | | |
| | | | | Y1 | Y1 | Y1 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | |
| <i>Fossaria sp.</i> | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lanx sp.</i> | | 6 | gc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stagnicola</i> | | 10 | sc | | | | | | | | | | | | | | | | | | |
| Physidae | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Physa/ Physella</i> | | 8 | sc | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 1 | 2 | 0 | 9 | 40 | 1 | 0 | 0 | 0 |
| Planorbidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gyraulus sp.</i> | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Menetus sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Planorbella sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Nematomorpha | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Nemertea | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Prostoma sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Phylum Platyhelminthes | | 4 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Turbellaria | | 4 | | 18 | 1 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 42 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | Site Code: | PES-070 | PES-080 | PES-095 | PES-100 | PES-120 | PES-140 | PES-150 | PES-160 | PES-170 | PES-180 | PES-190 | PES-200 | PES-210 | PES-230 | PES-240 | SGR-010 | SGR-030 |
|-------------------------------|------------|----------------------|--------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | Year: | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | |
| Coleoptera | | | | | | | | | | | | | | | | | | | | |
| <i>Fossaria sp.</i> | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lanx sp.</i> | | 6 | gc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stagnicola</i> | | 10 | sc | | | | | | | | | | | | | | | | | |
| Physidae | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Physa/ Physella</i> | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Planorbidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gyraulus sp.</i> | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Menetus sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Planorbella sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Nematomorpha | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Nemertea | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Prostoma sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Platyhelminthes | | 4 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Turbellaria | | 4 | | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E-1: Bioassessment master taxa list

| | | | | Site Code: | | | | | | | | | | | | | | | | | |
|-------------------------------|------------|----------------------|--------------------------------|------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----|
| | | | | SGR-040 | SGR-060 | SGR-075 | SGR-080 | SGR-090 | SGR-110 | SGR-120 | SGR-130 | SGR-150 | STE-020 | STE-030 | STE-040 | STE-060 | STE-070 | STE-100 | STE-110 | STE-120 | |
| | | | | Year: | | | | | | | | | | | | | | | | | |
| | | | | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 | Y2 |
| Phylum | Life Stage | TV (Tolerance Value) | FFG (Functional Feeding Group) | | | | | | | | | | | | | | | | | | |
| Phylum Arthropoda | | | | | | | | | | | | | | | | | | | | | |
| Insecta | | | | | | | | | | | | | | | | | | | | | |
| <u>Coleoptera</u> | | | | | | | | | | | | | | | | | | | | | |
| <i>Fossaria sp.</i> | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Lanx sp.</i> | | 6 | gc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Stagnicola</i> | | 10 | sc | | | | | | | | | | | | | | | | | | |
| Physidae | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Physa/ Physella</i> | | 8 | sc | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 0 | 12 | 5 | 0 | 0 | 0 | 0 | 0 |
| Planorbidae | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Gyraulus sp.</i> | | 8 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Menetus sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 |
| <i>Planorbella sp.</i> | | 6 | sc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Nematomorpha | | 8 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Nemertea | | | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Prostoma sp.</i> | | 8 | p | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Phylum Platyhelminthes | | 4 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Turbellaria | | 4 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 7 | 0 | 1 | 5 | 0 | 0 |

Appendix E-2: Bioassessment physical habitat scores

| Site ID | Collection Date | EpifaunalSubstrate | Embeddedness | Velocity/ Depth Regimes |
|---------|-----------------|--------------------|--------------|-------------------------|
| LAG-130 | 36997 | 15 | 10 | |
| LAG-150 | 36997 | 15 | 15 | 10 |
| LAG-160 | 36997 | 6 | 10 | 13 |
| LAG-180 | 36998 | 9 | 8 | 12 |
| LAG-170 | 36998 | 12 | 15 | 16 |
| LAG-190 | 36998 | 15 | 12 | 15 |
| LAG-210 | 36998 | 16 | 13 | 16 |
| LAG-220 | 36998 | 7 | 11 | 12 |
| LAG-165 | 36999 | 15 | 15 | 18 |
| LAG-240 | 36999 | 10 | 18 | 16 |
| LAG-270 | 36999 | 9 | 15 | 16 |
| LAG-320 | 36999 | 17 | 18 | 14 |
| LAG-300 | 37000 | 13 | 12 | 15 |
| LAG-330 | 37000 | 15 | 13 | 14 |
| LAG-335 | 37000 | 15 | 14 | 14 |
| LAG-380 | 37000 | 17 | 14 | 15 |
| LAG-390 | 37000 | 19 | 20 | 15 |
| WLK-180 | 36997 | 15 | 12 | 17 |
| WLK-190 | 36997 | 14 | 10 | 12 |
| WLK-230 | 36997 | 13 | 11 | 11 |
| WLK-240 | 36997 | 10 | 12 | 12 |
| WLK-170 | 36999 | 9 | 8 | 7 |
| WLK-200 | 36999 | 12 | 9 | 14 |
| SUI-010 | 37000 | 10 | 11 | 12 |
| SUI-020 | 37000 | 9 | 9 | 10 |
| SUI-050 | 37000 | 12 | 13 | 11 |
| SUI-060 | 37000 | 11 | 9 | 10 |
| SUI-110 | 37000 | 4 | 1 | 12 |
| SUI-130 | 37000 | 11 | 9 | 12 |
| SUI-180 | 37001 | 13 | 14 | 13 |
| SUI-210 | 37001 | 7 | 8 | 12 |
| SUI-260 | 37013 | 5 | 5 | 6 |
| ALP-010 | 37001 | 13 | 16 | 18 |
| ALP-040 | 37007 | 2 | 2 | 13 |
| ALP-070 | 37007 | 14 | 5 | 15 |
| ALP-080 | 37007 | 6 | 8 | 17 |
| ALP-100 | 37007 | 2 | 4 | 16 |
| ALP-110 | 37007 | 3 | 3 | 16 |
| ALP-140 | 37007 | 8 | 3 | 14 |
| SLE-030 | 37013 | 6 | 13 | 11 |
| SLE-170 | 37013 | 15 | 16 | 11 |
| SLE-180 | 37013 | 15 | 18 | 10 |
| SLE-220 | 37013 | 6 | 15 | 6 |
| SLE-210 | 37014 | 16 | 15 | 15 |
| SLE-230 | 37014 | 16 | 16 | 12 |
| SPA-020 | 37013 | 2 | 1 | 11 |
| SPA-050 | 37013 | 3 | 3 | 14 |
| SPA-070 | 37013 | 5 | 5 | 11 |
| SPA-100 | 37013 | 3 | 5 | 12 |
| SPA-110 | 37013 | 6 | 7 | 14 |
| SPA-130 | 37013 | 8 | 8 | 13 |
| SPA-140 | 37014 | 7 | 8 | 10 |
| SPA-150 | 37014 | 4 | 6 | 10 |
| SPA-160 | 37014 | 6 | 5 | 10 |
| SPA-170 | 37015 | 13 | 7 | 16 |

Appendix E-2: Bioassessment physical habitat scores

| Site ID | Collection Date | EpifaunalSubstrate | Embeddedness | Velocity/ Depth Regimes |
|---------|-----------------|--------------------|--------------|-------------------------|
| SPA-200 | 37015 | 5 | 3 | 12 |
| SPA-220 | 37015 | 5 | 5 | 11 |
| SPA-240 | 37015 | 5 | 5 | 11 |
| WIL-100 | 37014 | 17 | 15 | 13 |
| WIL_170 | 37015 | 13 | 18 | 13 |
| WIL_190 | 37015 | 9 | | 8 |
| BUT-010 | 37355 | 3 | 1 | 9 |
| BUT-020 | 37355 | 10 | 12 | 17 |
| BUT-030 | 37355 | 12 | 8 | 16 |
| BUT-040 | 37355 | 18 | 8 | 18 |
| BUT-050 | 37355 | 17 | 16 | 15 |
| PES-050 | 37355 | 9 | 12 | 17 |
| PES-140 | 37355 | 14 | 15 | 17 |
| PES-200 | 37355 | 17 | 14 | 16 |
| PES-210 | 37355 | 18 | 13 | 17 |
| PES-230 | 37355 | 14 | 4 | 15 |
| PES-240 | 37355 | 14 | 12 | 14 |
| PES-060 | 37356 | 14 | 12 | 16 |
| PES-070 | 37356 | 15 | 14 | 17 |
| PES-080 | 37356 | 11 | 16 | 10 |
| PES-095 | 37356 | 15 | 17 | 17 |
| PES-100 | 37356 | 15 | 16 | 18 |
| PES-120 | 37356 | 18 | 10 | 19 |
| PES-150 | 37356 | 4 | 4 | 6 |
| PES-160 | 37356 | 15 | 13 | 14 |
| PES-170 | 37356 | 16 | 13 | 14 |
| PES-180 | 37356 | 14 | 14 | 10 |
| PES-190 | 37356 | 17 | 14 | 17 |
| SGR-010 | 37357 | 16 | 12 | 18 |
| SGR-030 | 37357 | 16 | 10 | 17 |
| SGR-040 | 37357 | 14 | 13 | 15 |
| SGR-060 | 37357 | 17 | 10 | 18 |
| SGR-075 | 37357 | 18 | 14 | 18 |
| SGR-080 | 37357 | 14 | 13 | 15 |
| SGR-090 | 37357 | 16 | 16 | 17 |
| SGR-110 | 37357 | 16 | 14 | 18 |
| SGR-120 | 37357 | 15 | 15 | 12 |
| SGR-130 | 37357 | 15 | 10 | 10 |
| SGR-150 | 37357 | 16 | 14 | 17 |
| STE-020 | 37354 | 10 | 10 | 10 |
| STE-030 | 37354 | 5 | 7 | 6 |
| STE-040 | 37354 | 13 | 9 | 10 |
| STE-060 | 37354 | 12 | 8 | 15 |
| STE-070 | 37354 | 15 | 10 | 16 |
| STE-100 | 37354 | 14 | 11 | 16 |
| STE-110 | 37358 | 16 | 15 | 15 |
| STE-120 | 37358 | 12 | 10 | 11 |
| PER-030 | 37357 | 5 | 13 | 15 |
| PER-040 | 37357 | 2 | 16 | 6 |
| PER-010 | 37358 | 5 | 2 | 3 |
| PER-020 | 37358 | 3 | 10 | 6 |
| PER-050 | 37358 | 5 | 6 | 10 |
| PER-070 | 37358 | 12 | 5 | 15 |
| PER-080 | 37358 | 7 | 15 | 15 |

Appendix E-2: I

| Site ID | SedimentDepth | ChannelFlow | ChannelAlteration | RiffleFrequency | BankStability-Left |
|---------|---------------|-------------|-------------------|-----------------|--------------------|
| LAG-130 | 10 | 10 | 20 | 11 | 7 |
| LAG-150 | 16 | 10 | 18 | 15 | 5 |
| LAG-160 | 14 | 4 | 20 | 18 | 6 |
| LAG-180 | 18 | 12 | 20 | 16 | 7 |
| LAG-170 | 15 | 18 | 20 | 14 | 9 |
| LAG-190 | 18 | 18 | 20 | 18 | 9 |
| LAG-210 | 18 | 20 | 20 | 15 | 10 |
| LAG-220 | 18 | 19 | 20 | 15 | 9 |
| LAG-165 | 15 | 19 | 20 | 17 | 9 |
| LAG-240 | 16 | 12 | 14 | 10 | 8 |
| LAG-270 | 15 | 15 | 16 | 16 | 7 |
| LAG-320 | 18 | 20 | 20 | 20 | 9 |
| LAG-300 | 13 | 16 | 14 | 12 | 7 |
| LAG-330 | 18 | 11 | 20 | 18 | 7 |
| LAG-335 | 18 | 15 | 20 | 20 | 8 |
| LAG-380 | 18 | 18 | 20 | 20 | 9 |
| LAG-390 | 20 | 20 | 20 | 20 | 10 |
| WLK-180 | 8 | 6 | 17 | 16 | 4 |
| WLK-190 | 14 | 8 | 19 | 15 | 4 |
| WLK-230 | 7 | 7 | 19 | 13 | 2 |
| WLK-240 | 4 | 6 | 19 | 15 | 1 |
| WLK-170 | 6 | 5 | 17 | 8 | 3 |
| WLK-200 | 14 | 10 | 18 | 15 | 6 |
| SUI-010 | 13 | 10 | 12 | 12 | 7 |
| SUI-020 | 10 | 18 | 7 | 19 | 8 |
| SUI-050 | 14 | 14 | 9 | 14 | 6 |
| SUI-060 | 11 | 9 | 12 | 18 | 4 |
| SUI-110 | 15 | 13 | 18 | 18 | 9 |
| SUI-130 | 10 | 13 | 20 | 16 | 4 |
| SUI-180 | 16 | 17 | 20 | 15 | 8 |
| SUI-210 | 8 | 15 | 18 | 16 | 1 |
| SUI-260 | 5 | 14 | 20 | 14 | 6 |
| ALP-010 | 14 | 18 | 9 | 15 | 3 |
| ALP-040 | 4 | 17 | 12 | 11 | 7 |
| ALP-070 | 2 | 16 | 8 | 14 | 1 |
| ALP-080 | 4 | 19 | 7 | 11 | 5 |
| ALP-100 | 2 | 17 | 3 | 2 | 0 |
| ALP-110 | 5 | 19 | 3 | 1 | 4 |
| ALP-140 | 8 | 16 | 14 | 8 | 5 |
| SLE-030 | 13 | 14 | 0 | 12 | 10 |
| SLE-170 | 17 | 13 | 9 | 15 | 8 |
| SLE-180 | 18 | 9 | 18 | 18 | 8 |
| SLE-220 | 4 | 8 | 20 | 5 | 3 |
| SLE-210 | 14 | 10 | 8 | 6 | 3 |
| SLE-230 | 17 | 20 | 20 | 10 | 9 |
| SPA-020 | 8 | 17 | 6 | 3 | 9 |
| SPA-050 | 4 | 18 | 3 | 17 | 4 |
| SPA-070 | 6 | 18 | 2 | 9 | 9 |
| SPA-100 | 4 | 17 | 3 | 11 | 5 |
| SPA-110 | 10 | 18 | 5 | 16 | 7 |
| SPA-130 | 15 | 18 | 19 | 16 | 7 |
| SPA-140 | 7 | 18 | 18 | 18 | 4 |
| SPA-150 | 4 | 18 | 20 | 18 | 3 |
| SPA-160 | 4 | 18 | 20 | 16 | 6 |
| SPA-170 | 7 | 15 | 9 | 18 | 7 |

Appendix E-2: I

| Site ID | SedimentDepth | ChannelFlow | ChannelAlteration | RiffleFrequency | BankStability-Left |
|---------|---------------|-------------|-------------------|-----------------|--------------------|
| SPA-200 | 8 | 18 | 3 | 17 | 7 |
| SPA-220 | 5 | 18 | 7 | 17 | 8 |
| SPA-240 | 12 | 18 | 6 | 17 | 4 |
| WIL-100 | 16 | 8 | 20 | 9 | 8 |
| WIL_170 | 20 | 12 | 8 | 17 | 4 |
| WIL_190 | 19 | 5 | 19 | 18 | 3 |
| BUT-010 | 1 | 18 | 16 | 10 | 4 |
| BUT-020 | 7 | 18 | 17 | 10 | 7 |
| BUT-030 | 5 | 18 | 18 | 18 | 6 |
| BUT-040 | 11 | 18 | 18 | 19 | 8 |
| BUT-050 | 13 | 19 | 17 | 18 | 6 |
| PES-050 | 5 | 17 | 17 | 10 | 2 |
| PES-140 | 12 | 20 | 18 | 14 | 9 |
| PES-200 | 13 | 20 | 20 | 18 | 5 |
| PES-210 | 13 | 20 | 20 | 18 | 10 |
| PES-230 | 5 | 20 | 2 | 12 | 9 |
| PES-240 | 12 | 20 | 20 | 18 | 4 |
| PES-060 | 4 | 19 | 8 | 18 | 7 |
| PES-070 | 13 | 18 | 17 | 16 | 6 |
| PES-080 | 16 | 18 | 18 | 19 | 5 |
| PES-095 | 12 | 18 | 18 | 17 | 6 |
| PES-100 | 16 | 18 | 11 | 18 | 7 |
| PES-120 | 12 | 18 | 18 | 18 | 5 |
| PES-150 | 5 | 20 | 20 | 14 | 6 |
| PES-160 | 14 | 20 | 20 | 18 | 7 |
| PES-170 | 12 | 20 | 20 | 18 | 9 |
| PES-180 | 17 | 20 | 20 | 18 | 4 |
| PES-190 | 14 | 20 | 20 | 15 | 7 |
| SGR-010 | 12 | 20 | 20 | 18 | 8 |
| SGR-030 | 10 | 20 | 20 | 18 | 7 |
| SGR-040 | 14 | 20 | 18 | 18 | 7 |
| SGR-060 | 14 | 20 | 20 | 18 | 5 |
| SGR-075 | 15 | 20 | 19 | 18 | 9 |
| SGR-080 | 16 | 18 | 15 | 18 | 8 |
| SGR-090 | 16 | 19 | 16 | 16 | 5 |
| SGR-110 | 16 | 19 | 18 | 16 | 7 |
| SGR-120 | 16 | 18 | 16 | 18 | 7 |
| SGR-130 | 13 | 18 | 17 | 19 | 7 |
| SGR-150 | 10 | 18 | 16 | 19 | 7 |
| STE-020 | 3 | 7 | 7 | 14 | 3 |
| STE-030 | 5 | 16 | 1 | 10 | 7 |
| STE-040 | 9 | 18 | 8 | 17 | 9 |
| STE-060 | 13 | 18 | 10 | 18 | 8 |
| STE-070 | 15 | 20 | 20 | 14 | 9 |
| STE-100 | 8 | 20 | 18 | 17 | 9 |
| STE-110 | 16 | 20 | 20 | 17 | 8 |
| STE-120 | 10 | 20 | 20 | 16 | 7 |
| PER-030 | 15 | 15 | 15 | 15 | 5 |
| PER-040 | 10 | 18 | 0 | 2 | 10 |
| PER-010 | 2 | 15 | 5 | 1 | 5 |
| PER-020 | 10 | 10 | 0 | 5 | 10 |
| PER-050 | 10 | 20 | 17 | 19 | 8 |
| PER-070 | 5 | 20 | 20 | 15 | 9 |
| PER-080 | 16 | 20 | 18 | 16 | 7 |

Appendix E-2: I

| Site ID | BankStability-Right | BankVegetation-Left | BankVegetation-Right | RiparianZone-Left |
|---------|---------------------|---------------------|----------------------|-------------------|
| LAG-130 | 8 | 8 | 9 | 6 |
| LAG-150 | 6 | 6 | 6 | 6 |
| LAG-160 | 6 | 8 | 8 | 7 |
| LAG-180 | 6 | 7 | 7 | 7 |
| LAG-170 | 8 | 10 | 9 | 9 |
| LAG-190 | 9 | 7 | 6 | 9 |
| LAG-210 | 9 | 9 | 8 | 9 |
| LAG-220 | 9 | 8 | 9 | 9 |
| LAG-165 | 8 | 9 | 8 | 9 |
| LAG-240 | 7 | 7 | 6 | 3 |
| LAG-270 | 8 | 8 | 7 | 7 |
| LAG-320 | 9 | 9 | 9 | 9 |
| LAG-300 | 8 | 7 | 7 | 3 |
| LAG-330 | 7 | 7 | 7 | 10 |
| LAG-335 | 9 | 8 | 9 | 10 |
| LAG-380 | 9 | 9 | 9 | 10 |
| LAG-390 | 10 | 9 | 9 | 10 |
| WLK-180 | 4 | 5 | 6 | 7 |
| WLK-190 | 4 | 7 | 6 | 7 |
| WLK-230 | 4 | 6 | 7 | 5 |
| WLK-240 | 1 | 1 | 2 | 1 |
| WLK-170 | 3 | 4 | 3 | 2 |
| WLK-200 | 7 | 5 | 7 | 7 |
| SUI-010 | 8 | 7 | 6 | 7 |
| SUI-020 | 6 | 5 | 8 | 4 |
| SUI-050 | 6 | 7 | 6 | 5 |
| SUI-060 | 4 | 4 | 4 | 5 |
| SUI-110 | 7 | 2 | 7 | 9 |
| SUI-130 | 3 | 4 | 4 | 4 |
| SUI-180 | 8 | 8 | 8 | 9 |
| SUI-210 | 2 | 2 | 2 | 9 |
| SUI-260 | 6 | 4 | 4 | 9 |
| ALP-010 | 3 | 3 | 3 | 4 |
| ALP-040 | 7 | 6 | 6 | 3 |
| ALP-070 | 1 | 2 | 1 | 2 |
| ALP-080 | 5 | 3 | 3 | 4 |
| ALP-100 | 1 | 0 | 0 | 0 |
| ALP-110 | 4 | 2 | 2 | 2 |
| ALP-140 | 5 | 5 | 4 | 5 |
| SLE-030 | 10 | 2 | 2 | 2 |
| SLE-170 | 9 | 9 | 7 | 6 |
| SLE-180 | 8 | 9 | 9 | 6 |
| SLE-220 | 5 | 6 | 6 | 10 |
| SLE-210 | 7 | 6 | 6 | 4 |
| SLE-230 | 8 | 9 | 8 | 8 |
| SPA-020 | 9 | 8 | 6 | 4 |
| SPA-050 | 4 | 5 | 5 | 2 |
| SPA-070 | 9 | 3 | 3 | 2 |
| SPA-100 | 5 | 4 | 5 | 2 |
| SPA-110 | 8 | 5 | 4 | 3 |
| SPA-130 | 7 | 7 | 7 | 9 |
| SPA-140 | 4 | 5 | 5 | 9 |
| SPA-150 | 2 | 3 | 1 | 10 |
| SPA-160 | 6 | 4 | 4 | 10 |
| SPA-170 | 7 | 6 | 6 | 5 |

Appendix E-2: I

| Site ID | BankStability-Right | BankVegetation-Left | BankVegetation-Right | RiparianZone-Left |
|---------|---------------------|---------------------|----------------------|-------------------|
| SPA-200 | 6 | 4 | 5 | 1 |
| SPA-220 | 8 | 2 | 2 | 3 |
| SPA-240 | 2 | 3 | 1 | 3 |
| WIL-100 | 8 | 9 | 9 | 10 |
| WIL_170 | 5 | 6 | 4 | 9 |
| WIL_190 | 3 | 4 | 4 | 8 |
| BUT-010 | 4 | 8 | 8 | 8 |
| BUT-020 | 7 | 6 | 8 | 4 |
| BUT-030 | 7 | 7 | 7 | 6 |
| BUT-040 | 9 | 8 | 9 | 7 |
| BUT-050 | 7 | 7 | 8 | 9 |
| PES-050 | 2 | 6 | 8 | 4 |
| PES-140 | 9 | 7 | 7 | 10 |
| PES-200 | 6 | 5 | 6 | 10 |
| PES-210 | 8 | 8 | 7 | 10 |
| PES-230 | 8 | 9 | 9 | 7 |
| PES-240 | 4 | 5 | 7 | 10 |
| PES-060 | 7 | 8 | 8 | 3 |
| PES-070 | 7 | 7 | 7 | 4 |
| PES-080 | 5 | 6 | 6 | 4 |
| PES-095 | 4 | 7 | 7 | 8 |
| PES-100 | 4 | 6 | 6 | 2 |
| PES-120 | 5 | 7 | 8 | 3 |
| PES-150 | 6 | 6 | 6 | 10 |
| PES-160 | 6 | 8 | 7 | 10 |
| PES-170 | 8 | 9 | 8 | 10 |
| PES-180 | 7 | 5 | 8 | 10 |
| PES-190 | 7 | 6 | 6 | 10 |
| SGR-010 | 8 | 9 | 9 | 10 |
| SGR-030 | 7 | 9 | 8 | 10 |
| SGR-040 | 6 | 6 | 7 | 9 |
| SGR-060 | 6 | 6 | 7 | 7 |
| SGR-075 | 8 | 8 | 9 | 8 |
| SGR-080 | 7 | 3 | 3 | 2 |
| SGR-090 | 4 | 4 | 7 | 3 |
| SGR-110 | 8 | 8 | 7 | 8 |
| SGR-120 | 5 | 8 | 8 | 5 |
| SGR-130 | 7 | 7 | 7 | 9 |
| SGR-150 | 3 | 8 | 4 | 9 |
| STE-020 | 3 | 3 | 1 | 3 |
| STE-030 | 9 | 6 | 5 | 2 |
| STE-040 | 9 | 6 | 4 | 3 |
| STE-060 | 8 | 6 | 7 | 5 |
| STE-070 | 9 | 8 | 9 | 9 |
| STE-100 | 10 | 5 | 7 | 9 |
| STE-110 | 8 | 6 | 7 | 10 |
| STE-120 | 5 | 6 | 6 | 10 |
| PER-030 | 5 | 4 | 4 | 2 |
| PER-040 | 10 | 0 | 0 | 0 |
| PER-010 | 5 | 5 | 5 | 0 |
| PER-020 | 10 | 0 | 0 | 0 |
| PER-050 | 8 | 6 | 7 | 6 |
| PER-070 | 9 | 10 | 10 | 10 |
| PER-080 | 7 | 5 | 5 | 10 |

Appendix E-2: I

| Site ID | RiparianZone-Right | Canopy Cover T1 | Canopy Cover T2 | Canopy Cover T3 |
|---------|--------------------|-----------------|-----------------|-----------------|
| LAG-130 | 6 | 50 | 50 | 30 |
| LAG-150 | 6 | 50 | 0 | 50 |
| LAG-160 | 7 | 75 | 75 | 90 |
| LAG-180 | 7 | 80 | 90 | 90 |
| LAG-170 | 9 | 20 | 20 | 50 |
| LAG-190 | 9 | 90 | 90 | 90 |
| LAG-210 | 8 | 50 | 90 | 90 |
| LAG-220 | 10 | 80 | 90 | 80 |
| LAG-165 | 9 | 50 | 80 | 80 |
| LAG-240 | 3 | 70 | 75 | 50 |
| LAG-270 | 7 | 30 | 80 | 60 |
| LAG-320 | 8 | 70 | 80 | 70 |
| LAG-300 | 3 | 80 | 70 | 20 |
| LAG-330 | 10 | 100 | 80 | 80 |
| LAG-335 | 10 | 80 | 100 | 100 |
| LAG-380 | 10 | 100 | 100 | 100 |
| LAG-390 | 9 | 90 | 90 | 100 |
| WLK-180 | 7 | 15 | 5 | 25 |
| WLK-190 | 8 | 85 | 35 | 95 |
| WLK-230 | 6 | 0 | 0 | 0 |
| WLK-240 | 1 | 0 | 0 | 0 |
| WLK-170 | 2 | 45 | 0 | 90 |
| WLK-200 | 7 | 90 | 95 | 75 |
| SUI-010 | 6 | 30 | 85 | 95 |
| SUI-020 | 5 | 90 | 70 | 95 |
| SUI-050 | 4 | 50 | 70 | 70 |
| SUI-060 | 5 | 100 | 60 | 70 |
| SUI-110 | 6 | 90 | 30 | 10 |
| SUI-130 | 2 | 20 | 90 | 20 |
| SUI-180 | 7 | 100 | 40 | 5 |
| SUI-210 | 5 | 95 | 95 | 95 |
| SUI-260 | 10 | 5 | 5 | 60 |
| ALP-010 | 4 | 5 | 5 | 5 |
| ALP-040 | 3 | 5 | 5 | 5 |
| ALP-070 | 1 | 80 | 80 | 85 |
| ALP-080 | 4 | 10 | 5 | 40 |
| ALP-100 | 1 | 5 | 5 | 5 |
| ALP-110 | 2 | 5 | 5 | 5 |
| ALP-140 | 5 | 85 | 85 | 50 |
| SLE-030 | 2 | 10 | 15 | 30 |
| SLE-170 | 5 | 80 | 80 | 90 |
| SLE-180 | 10 | 100 | 100 | 90 |
| SLE-220 | 7 | 100 | 100 | 90 |
| SLE-210 | 5 | 100 | 100 | |
| SLE-230 | 6 | 100 | 100 | 100 |
| SPA-020 | 4 | 5 | 10 | 50 |
| SPA-050 | 2 | 90 | 100 | 35 |
| SPA-070 | 2 | 85 | 65 | 60 |
| SPA-100 | 3 | 100 | 100 | 50 |
| SPA-110 | 2 | 20 | 85 | 95 |
| SPA-130 | 8 | 70 | 65 | 80 |
| SPA-140 | 8 | 80 | 90 | 50 |
| SPA-150 | 9 | 20 | 85 | 95 |
| SPA-160 | 10 | 95 | 60 | 90 |
| SPA-170 | 5 | 30 | 30 | 45 |

Appendix E-2: I

| Site ID | RiparianZone-Right | Canopy Cover T1 | Canopy Cover T2 | Canopy Cover T3 |
|---------|--------------------|-----------------|-----------------|-----------------|
| SPA-200 | 1 | 100 | 95 | 90 |
| SPA-220 | 3 | 80 | 65 | 80 |
| SPA-240 | 1 | 80 | 90 | 90 |
| WIL-100 | 9 | 85 | 75 | 50 |
| WIL_170 | 3 | 100 | 100 | 75 |
| WIL_190 | 3 | 100 | 100 | 100 |
| BUT-010 | 8 | 100 | 94 | 94 |
| BUT-020 | 3 | 94 | 94 | 59 |
| BUT-030 | 9 | 59 | 94 | 53 |
| BUT-040 | 10 | 100 | 88 | 71 |
| BUT-050 | 10 | 100 | 88 | 100 |
| PES-050 | 6 | 18 | 53 | 88 |
| PES-140 | 10 | 71 | 72 | 87 |
| PES-200 | 10 | 99 | 96 | 99 |
| PES-210 | 10 | 84 | 85 | 93 |
| PES-230 | 10 | 85 | 43 | 54 |
| PES-240 | 10 | 99 | 100 | 99 |
| PES-060 | 2 | 47 | 82 | |
| PES-070 | 5 | 6 | 24 | 94 |
| PES-080 | 5 | 53 | 82 | 76 |
| PES-095 | 3 | 65 | 53 | 24 |
| PES-100 | 2 | 59 | 18 | |
| PES-120 | 9 | 71 | 82 | 82 |
| PES-150 | 10 | 100 | 100 | 100 |
| PES-160 | 10 | 84 | 84 | 90 |
| PES-170 | 10 | 97 | 100 | 94 |
| PES-180 | 10 | 96 | 90 | 82 |
| PES-190 | 10 | 91 | 99 | 93 |
| SGR-010 | 10 | 35 | 41 | 21 |
| SGR-030 | 10 | 81 | 76 | 93 |
| SGR-040 | 7 | 34 | 78 | 53 |
| SGR-060 | 10 | 90 | 91 | 93 |
| SGR-075 | 10 | 94 | 79 | 85 |
| SGR-080 | 5 | 94 | 94 | 82 |
| SGR-090 | 2 | 94 | 100 | 94 |
| SGR-110 | 9 | 100 | 94 | 100 |
| SGR-120 | 7 | 100 | 94 | 88 |
| SGR-130 | 9 | 100 | 94 | 94 |
| SGR-150 | 7 | 94 | 100 | 94 |
| STE-020 | 1 | 0 | 6 | 12 |
| STE-030 | 2 | 94 | 47 | 82 |
| STE-040 | 3 | 29 | 18 | 35 |
| STE-060 | 4 | 100 | 88 | 71 |
| STE-070 | 10 | 99 | 100 | 85 |
| STE-100 | 10 | 97 | 97 | 94 |
| STE-110 | 10 | 100 | 100 | 99 |
| STE-120 | 10 | 94 | 100 | 97 |
| PER-030 | 2 | 100 | 100 | 100 |
| PER-040 | 0 | 24 | 35 | 35 |
| PER-010 | 0 | 0 | 0 | 0 |
| PER-020 | 0 | 12 | 12 | 12 |
| PER-050 | 6 | 96 | 94 | 99 |
| PER-070 | 10 | 100 | 100 | 100 |
| PER-080 | 9 | 100 | 100 | 100 |

Appendix E-2: I

| Site ID | Substrate Complexity T1 | Substrate Complexity T2 | Substrate Complexity T3 |
|---------|-------------------------|-------------------------|-------------------------|
| LAG-130 | 10 | 10 | |
| LAG-150 | 12 | 12 | 5 |
| LAG-160 | 7 | 9 | 8 |
| LAG-180 | 17 | 15 | 15 |
| LAG-170 | 8 | 11 | 15 |
| LAG-190 | 17 | 16 | 18 |
| LAG-210 | 15 | 20 | 15 |
| LAG-220 | 8 | 8 | 16 |
| LAG-165 | 13 | 13 | 14 |
| LAG-240 | 12 | 10 | 9 |
| LAG-270 | 10 | 17 | |
| LAG-320 | 17 | 16 | 16 |
| LAG-300 | 15 | 12 | 12 |
| LAG-330 | 17 | 18 | 16 |
| LAG-335 | 16 | 16 | 16 |
| LAG-380 | 18 | 16 | 18 |
| LAG-390 | 16 | 18 | 19 |
| WLK-180 | 14 | 14 | 13 |
| WLK-190 | 16 | 13 | 13 |
| WLK-230 | 8 | 6 | 6 |
| WLK-240 | 8 | 8 | 7 |
| WLK-170 | 16 | 5 | 5 |
| WLK-200 | 17 | 15 | 14 |
| SUI-010 | 10 | 11 | 9 |
| SUI-020 | 10 | 12 | 12 |
| SUI-050 | 12 | 10 | 10 |
| SUI-060 | 10 | 11 | 11 |
| SUI-110 | 5 | 3 | 3 |
| SUI-130 | 12 | 11 | 13 |
| SUI-180 | 11 | 14 | 15 |
| SUI-210 | 8 | 5 | 10 |
| SUI-260 | 5 | 5 | 5 |
| ALP-010 | 10 | 10 | 10 |
| ALP-040 | 2 | 2 | 2 |
| ALP-070 | 4 | 6 | 6 |
| ALP-080 | 12 | 10 | 12 |
| ALP-100 | 7 | 8 | 2 |
| ALP-110 | 4 | 5 | 7 |
| ALP-140 | 6 | 8 | 10 |
| SLE-030 | 7 | 13 | 7 |
| SLE-170 | 15 | 16 | 15 |
| SLE-180 | 14 | 12 | 16 |
| SLE-220 | 3 | 5 | 11 |
| SLE-210 | 16 | 15 | |
| SLE-230 | 15 | 15 | 15 |
| SPA-020 | 7 | 1 | 1 |
| SPA-050 | 6 | 6 | 6 |
| SPA-070 | 13 | 13 | 3 |
| SPA-100 | 10 | 8 | 6 |
| SPA-110 | 13 | 11 | 8 |
| SPA-130 | 5 | 9 | 5 |
| SPA-140 | 5 | 5 | 5 |
| SPA-150 | 12 | 12 | 12 |
| SPA-160 | 11 | 11 | 10 |
| SPA-170 | 12 | 11 | 12 |

Appendix E-2: I

| Site ID | Substrate Complexity T1 | Substrate Complexity T2 | Substrate Complexity T3 |
|---------|-------------------------|-------------------------|-------------------------|
| SPA-200 | 8 | 9 | 8 |
| SPA-220 | 13 | 11 | 11 |
| SPA-240 | 14 | 14 | 14 |
| WIL-100 | 16 | 18 | 16 |
| WIL_170 | 18 | 18 | 17 |
| WIL_190 | 17 | 16 | 16 |
| BUT-010 | 2 | 6 | 4 |
| BUT-020 | 7 | 8 | 6 |
| BUT-030 | 14 | 12 | 13 |
| BUT-040 | 17 | 14 | 12 |
| BUT-050 | 12 | 11 | 11 |
| PES-050 | 6 | 8 | 7 |
| PES-140 | 15 | 15 | 16 |
| PES-200 | 18 | 16 | 17 |
| PES-210 | 17 | 17 | 18 |
| PES-230 | 14 | 5 | 8 |
| PES-240 | 14 | 14 | 13 |
| PES-060 | 12 | 13 | 11 |
| PES-070 | 12 | 12 | 12 |
| PES-080 | 12 | 14 | 13 |
| PES-095 | 12 | 12 | 11 |
| PES-100 | 14 | 13 | 11 |
| PES-120 | 17 | 16 | 15 |
| PES-150 | 8 | 10 | 8 |
| PES-160 | 15 | 14 | 15 |
| PES-170 | 12 | 11 | 16 |
| PES-180 | 17 | 16 | 16 |
| PES-190 | 17 | 17 | 16 |
| SGR-010 | 14 | 15 | 15 |
| SGR-030 | 13 | 17 | 15 |
| SGR-040 | 15 | 12 | 13 |
| SGR-060 | 13 | 14 | 15 |
| SGR-075 | 14 | 14 | 12 |
| SGR-080 | 13 | 13 | 9 |
| SGR-090 | 8 | 9 | 15 |
| SGR-110 | 12 | 13 | 12 |
| SGR-120 | 10 | 10 | 11 |
| SGR-130 | 16 | 14 | 15 |
| SGR-150 | 16 | 16 | 16 |
| STE-020 | 10 | 8 | 10 |
| STE-030 | 10 | 11 | 16 |
| STE-040 | 15 | 12 | 15 |
| STE-060 | 4 | 10 | 16 |
| STE-070 | 16 | 14 | 14 |
| STE-100 | 15 | 15 | 14 |
| STE-110 | 17 | 14 | 17 |
| STE-120 | 13 | 14 | 15 |
| PER-030 | 8 | 10 | 8 |
| PER-040 | 5 | 3 | 5 |
| PER-010 | 2 | 2 | 2 |
| PER-020 | 1 | 1 | 1 |
| PER-050 | 12 | 11 | 11 |
| PER-070 | 10 | 10 | 10 |
| PER-080 | 11 | 11 | 11 |

Appendix E-2: I

| Site ID | Rifle Embeddedness T1 | Rifle Embeddedness T2 | Rifle Embeddedness T3 |
|---------|-----------------------|-----------------------|-----------------------|
| LAG-130 | 15 | 15 | 16 |
| LAG-150 | 15 | 15 | 15 |
| LAG-160 | 16 | 8 | 8 |
| LAG-180 | 9 | 7 | 8 |
| LAG-170 | 16 | 16 | 12 |
| LAG-190 | 10 | 16 | 12 |
| LAG-210 | 13 | 14 | 11 |
| LAG-220 | 10 | 10 | 14 |
| LAG-165 | 15 | 15 | 15 |
| LAG-240 | 18 | 18 | 18 |
| LAG-270 | 15 | 15 | 16 |
| LAG-320 | 18 | 18 | 18 |
| LAG-300 | 10 | 14 | 12 |
| LAG-330 | 13 | 14 | 12 |
| LAG-335 | 14 | 14 | 14 |
| LAG-380 | 13 | 14 | 14 |
| LAG-390 | 20 | | |
| WLK-180 | 12 | 12 | 12 |
| WLK-190 | 8 | 14 | 8 |
| WLK-230 | 11 | 11 | 11 |
| WLK-240 | 12 | 12 | 12 |
| WLK-170 | 10 | 7 | 7 |
| WLK-200 | 6 | 13 | 8 |
| SUI-010 | 10 | 14 | 8 |
| SUI-020 | 7 | 7 | 12 |
| SUI-050 | 12 | 14 | 12 |
| SUI-060 | 8 | 9 | 9 |
| SUI-110 | 2 | 1 | 1 |
| SUI-130 | 8 | 7 | 12 |
| SUI-180 | 12 | 14 | 15 |
| SUI-210 | 8 | 5 | 12 |
| SUI-260 | 8 | 8 | 8 |
| ALP-010 | 16 | 16 | 16 |
| ALP-040 | 4 | 4 | 4 |
| ALP-070 | 2 | 10 | 10 |
| ALP-080 | 9 | 9 | 9 |
| ALP-100 | 6 | 6 | 1 |
| ALP-110 | 4 | 4 | 8 |
| ALP-140 | 3 | 3 | 3 |
| SLE-030 | 13 | 9 | 13 |
| SLE-170 | 16 | 16 | 18 |
| SLE-180 | 18 | 18 | 15 |
| SLE-220 | 15 | 15 | 15 |
| SLE-210 | 15 | 15 | 15 |
| SLE-230 | 17 | 15 | 15 |
| SPA-020 | 8 | 0 | 0 |
| SPA-050 | 10 | 10 | 10 |
| SPA-070 | 7 | 14 | 13 |
| SPA-100 | 8 | 7 | 7 |
| SPA-110 | 10 | 14 | 12 |
| SPA-130 | 7 | 12 | 7 |
| SPA-140 | 10 | 10 | 10 |
| SPA-150 | 9 | 9 | 9 |
| SPA-160 | 8 | 8 | 8 |
| SPA-170 | 12 | 8 | 10 |

Appendix E-2: I

| Site ID | Rifle Embeddedness T1 | Rifle Embeddedness T2 | Rifle Embeddedness T3 |
|---------|-----------------------|-----------------------|-----------------------|
| SPA-200 | 4 | 4 | 4 |
| SPA-220 | 6 | 6 | 5 |
| SPA-240 | 6 | 6 | 6 |
| WIL-100 | 16 | 15 | 15 |
| WIL_170 | 18 | 18 | 18 |
| WIL_190 | 16 | 16 | 17 |
| BUT-010 | 1 | 1 | 1 |
| BUT-020 | 10 | 16 | 12 |
| BUT-030 | 8 | 10 | 15 |
| BUT-040 | 6 | 10 | 8 |
| BUT-050 | 17 | 17 | 17 |
| PES-050 | 12 | 14 | 12 |
| PES-140 | 17 | 17 | 17 |
| PES-200 | 17 | 14 | 15 |
| PES-210 | 16 | 16 | 16 |
| PES-230 | 6 | 10 | 8 |
| PES-240 | 16 | 16 | 16 |
| PES-060 | 10 | 16 | 13 |
| PES-070 | 18 | 16 | 10 |
| PES-080 | 17 | 16 | 15 |
| PES-095 | 17 | 18 | 16 |
| PES-100 | 12 | 17 | 18 |
| PES-120 | 10 | 10 | 10 |
| PES-150 | 6 | 6 | 5 |
| PES-160 | 14 | 16 | 16 |
| PES-170 | 16 | 16 | 16 |
| PES-180 | 15 | 17 | 17 |
| PES-190 | 16 | 16 | 16 |
| SGR-010 | 14 | 14 | 14 |
| SGR-030 | 13 | 14 | 14 |
| SGR-040 | 17 | 17 | 16 |
| SGR-060 | 10 | 12 | 13 |
| SGR-075 | 16 | 16 | 16 |
| SGR-080 | 17 | 8 | 18 |
| SGR-090 | 17 | 16 | 17 |
| SGR-110 | 15 | 16 | 17 |
| SGR-120 | 12 | 17 | 18 |
| SGR-130 | 12 | 10 | 12 |
| SGR-150 | 14 | 14 | 18 |
| STE-020 | 10 | 10 | 13 |
| STE-030 | 10 | 10 | 8 |
| STE-040 | 4 | 10 | 16 |
| STE-060 | 2 | 8 | 9 |
| STE-070 | 12 | 12 | 12 |
| STE-100 | 12 | 12 | 10 |
| STE-110 | 17 | 14 | 17 |
| STE-120 | 12 | 12 | 13 |
| PER-030 | 14 | 14 | 16 |
| PER-040 | 10 | 10 | 10 |
| PER-010 | 1 | 1 | 1 |
| PER-020 | 10 | 10 | 10 |
| PER-050 | 8 | 8 | 8 |
| PER-070 | 8 | 8 | 8 |
| PER-080 | 16 | 16 | 16 |

Appendix E-2: I

| Site ID | Substrate Consolidation T1 | Substrate Consolidation T2 | Substrate Consolidation T3 | Fines T1 |
|---------|----------------------------|----------------------------|----------------------------|----------|
| LAG-130 | Low | Low | Low | 45 |
| LAG-150 | Low | Low | Low | 30 |
| LAG-160 | Low | Med | Med | 45 |
| LAG-180 | High | Med | Med | 10 |
| LAG-170 | Low | Low | Low | 40 |
| LAG-190 | Med | Low | Med | 10 |
| LAG-210 | Med | Low | Med | 10 |
| LAG-220 | Med | Med | Low | 30 |
| LAG-165 | Low | Low | Low | 50 |
| LAG-240 | Low | Low | Low | 20 |
| LAG-270 | Low | Med | Low | 25 |
| LAG-320 | Low | Low | Med | 10 |
| LAG-300 | Low | Low | Low | 20 |
| LAG-330 | Med | Med | Med | 10 |
| LAG-335 | Med | Med | Med | 10 |
| LAG-380 | Med | Med | Med | 15 |
| LAG-390 | Low | Low | Low | 5 |
| WLK-180 | Med | Med | Med | 25 |
| WLK-190 | Med | Med | Med | 15 |
| WLK-230 | Low | Low | Low | 45 |
| WLK-240 | Low | Low | Low | 35 |
| WLK-170 | High | Low | Low | 20 |
| WLK-200 | High | Med | Med | 20 |
| SUI-010 | Med | Low | Low | 40 |
| SUI-020 | Med | Med | Med | 50 |
| SUI-050 | Med | Med | Med | 25 |
| SUI-060 | Med | Med | Med | 30 |
| SUI-110 | High | High | High | 10 |
| SUI-130 | Med | Med | Med | 40 |
| SUI-180 | Med | Med | Med | 25 |
| SUI-210 | Low | Low | Low | 50 |
| SUI-260 | Low | Low | Low | 35 |
| ALP-010 | Low | Low | Low | 10 |
| ALP-040 | Low | Low | Low | 50 |
| ALP-070 | Low | Low | Low | 70 |
| ALP-080 | Med | Med | Low | 35 |
| ALP-100 | Med | Med | High | 60 |
| ALP-110 | Med | Med | Med | 80 |
| ALP-140 | High | High | High | 60 |
| SLE-030 | Low | Low | Low | 40 |
| SLE-170 | Low | Low | Low | 20 |
| SLE-180 | Low | Low | Med | 5 |
| SLE-220 | Low | Low | Low | 50 |
| SLE-210 | Low | Low | Low | 35 |
| SLE-230 | Low | Low | Low | 10 |
| SPA-020 | High | High | High | 10 |
| SPA-050 | Low | Low | Low | 30 |
| SPA-070 | High | Low | High | 20 |
| SPA-100 | Med | Low | Low | 30 |
| SPA-110 | Med | Med | High | 20 |
| SPA-130 | Low | Low | Low | 60 |
| SPA-140 | Low | Low | Low | 40 |
| SPA-150 | Med | Low | Low | 40 |
| SPA-160 | Low | Low | Low | 50 |
| SPA-170 | Low | Med | Med | 20 |

Appendix E-2: I

| Site ID | Substrate Consolidation T1 | Substrate Consolidation T2 | Substrate Consolidation T3 | Fines T1 |
|---------|----------------------------|----------------------------|----------------------------|----------|
| SPA-200 | High | Med | Med | 40 |
| SPA-220 | Med | Med | High | 25 |
| SPA-240 | Med | High | High | 30 |
| WIL-100 | Med | Med | Med | 5 |
| WIL_170 | Low | Low | Low | <5 |
| WIL_190 | Low | Low | Low | 10 |
| BUT-010 | Low | Low | Low | 95 |
| BUT-020 | Med | Low | Low | 40 |
| BUT-030 | Low | Low | Low | 35 |
| BUT-040 | Med | Med | High | 20 |
| BUT-050 | Low | Low | Low | 20 |
| PES-050 | Low | Low | Low | 50 |
| PES-140 | Low | Low | Low | 30 |
| PES-200 | Low | Med | Low | 15 |
| PES-210 | Low | Low | Low | 10 |
| PES-230 | Low | Low | Low | 50 |
| PES-240 | Low | Low | Low | 20 |
| PES-060 | Low | Low | Low | 40 |
| PES-070 | Low | Low | Low | 20 |
| PES-080 | Low | Low | Low | 20 |
| PES-095 | Low | Low | Low | 30 |
| PES-100 | Low | Low | Low | 25 |
| PES-120 | Med | Med | High | 20 |
| PES-150 | Med | High | Low | 60 |
| PES-160 | Med | Low | Med | 30 |
| PES-170 | Low | Low | Low | 30 |
| PES-180 | Low | Low | Low | 15 |
| PES-190 | Low | Low | Low | 15 |
| SGR-010 | Med | Med | Med | 20 |
| SGR-030 | Low | Med | Med | 30 |
| SGR-040 | Low | Low | Med | 15 |
| SGR-060 | Low | Med | High | 20 |
| SGR-075 | Med | Low | Med | 25 |
| SGR-080 | Med | High | Med | 25 |
| SGR-090 | Med | Med | Med | 15 |
| SGR-110 | High | Med | Med | 35 |
| SGR-120 | Med | Med | Med | 30 |
| SGR-130 | High | High | High | 20 |
| SGR-150 | High | High | Low | 10 |
| STE-020 | Low | Low | Low | 25 |
| STE-030 | Low | Low | Hogh | 30 |
| STE-040 | Med | Med | Low | 10 |
| STE-060 | High | Low | Med | 5 |
| STE-070 | Low | Low | Low | 30 |
| STE-100 | Low | Low | Low | 40 |
| STE-110 | Low | Low | High | 15 |
| STE-120 | Low | Low | Med | 30 |
| PER-030 | Med | Low | Low | 40 |
| PER-040 | Low | Low | Low | 50 |
| PER-010 | Low | Low | Low | 90 |
| PER-020 | High | High | High | 15 |
| PER-050 | Low | Low | Low | 40 |
| PER-070 | Low | Low | Low | 40 |
| PER-080 | Low | Low | Low | 25 |

Appendix E-2: I

| Site ID | Fines T2 | Fines T3 | Gravel T1 | Gravel T2 | Gravel T3 | Coble T1 | Coble T2 | Coble T3 |
|---------|----------|----------|-----------|-----------|-----------|----------|----------|----------|
| LAG-130 | 45 | 40 | 45 | 45 | 30 | 10 | 10 | 30 |
| LAG-150 | 30 | 65 | 60 | 60 | 30 | 10 | 10 | 5 |
| LAG-160 | 30 | 50 | 50 | 50 | 40 | 5 | 20 | 10 |
| LAG-180 | 20 | 20 | 30 | 40 | 40 | 60 | 40 | 40 |
| LAG-170 | 30 | 25 | 50 | 40 | 30 | 10 | 30 | 40 |
| LAG-190 | 10 | 10 | 25 | 20 | 20 | 25 | 20 | 25 |
| LAG-210 | 10 | 10 | 20 | 20 | 25 | 65 | 20 | 60 |
| LAG-220 | 30 | 10 | 35 | 35 | 30 | 30 | 30 | 50 |
| LAG-165 | 50 | 50 | 50 | 45 | 40 | 0 | 5 | 10 |
| LAG-240 | 25 | 13 | 20 | 30 | 35 | 55 | 40 | 50 |
| LAG-270 | 20 | 25 | 65 | 30 | 65 | 10 | 20 | 10 |
| LAG-320 | 10 | 10 | 20 | 25 | 20 | 60 | 60 | 60 |
| LAG-300 | 20 | 10 | 50 | 60 | 60 | 25 | 20 | 30 |
| LAG-330 | 15 | 10 | 20 | 15 | 40 | 30 | 30 | 10 |
| LAG-335 | 10 | 10 | 10 | 10 | 10 | 30 | 30 | 30 |
| LAG-380 | 10 | 10 | 20 | 10 | 10 | 40 | 65 | 55 |
| LAG-390 | 5 | 3 | 15 | 15 | 7 | 20 | 30 | 30 |
| WLK-180 | 20 | 50 | 30 | 25 | 45 | 40 | 50 | 5 |
| WLK-190 | 15 | 15 | 30 | 45 | 35 | 40 | 35 | 50 |
| WLK-230 | 60 | 60 | 50 | 40 | 35 | 5 | 0 | 5 |
| WLK-240 | 55 | 55 | 50 | 40 | 40 | 15 | 5 | 5 |
| WLK-170 | 65 | 65 | 30 | 30 | 30 | 30 | 5 | 5 |
| WLK-200 | 30 | 30 | 20 | 30 | 35 | 35 | 35 | 30 |
| SUI-010 | 25 | 40 | 30 | 45 | 50 | 30 | 25 | 5 |
| SUI-020 | 40 | 30 | 30 | 30 | 30 | 20 | 25 | 35 |
| SUI-050 | 20 | 30 | 40 | 50 | 40 | 30 | 30 | 30 |
| SUI-060 | 20 | 25 | 40 | 50 | 45 | 30 | 30 | 30 |
| SUI-110 | 5 | 5 | 5 | 5 | 5 | 5 | 0 | 0 |
| SUI-130 | 40 | 30 | 25 | 30 | 30 | 30 | 30 | 40 |
| SUI-180 | 15 | 15 | 25 | 35 | 30 | 30 | 35 | 40 |
| SUI-210 | 70 | 40 | 30 | 20 | 40 | 20 | 10 | 20 |
| SUI-260 | 39 | 39 | 60 | 60 | 60 | 5 | 1 | 1 |
| ALP-010 | 5 | 0 | 0 | 0 | 0 | 10 | 15 | 0 |
| ALP-040 | 40 | 60 | 40 | 60 | 35 | 10 | 0 | 5 |
| ALP-070 | 40 | 40 | 30 | 50 | 50 | 0 | 10 | 10 |
| ALP-080 | 30 | 30 | 45 | 50 | 40 | 20 | 20 | 30 |
| ALP-100 | 70 | 95 | 30 | 15 | 5 | 10 | 10 | 0 |
| ALP-110 | 60 | 15 | 15 | 35 | 15 | 5 | 10 | 50 |
| ALP-140 | 50 | 10 | 30 | 20 | 15 | 5 | 20 | 50 |
| SLE-030 | 30 | 40 | 60 | 30 | 60 | 0 | 30 | 0 |
| SLE-170 | 15 | 15 | 20 | 25 | 25 | 30 | 40 | 50 |
| SLE-180 | 10 | 15 | 15 | 50 | 15 | 60 | 30 | 30 |
| SLE-220 | 45 | 35 | 50 | 45 | 40 | 0 | 10 | 15 |
| SLE-210 | 35 | 20 | 30 | 35 | 35 | 25 | 20 | 25 |
| SLE-230 | 10 | 10 | 5 | 5 | 5 | 25 | 35 | 30 |
| SPA-020 | 0 | 0 | 10 | 0 | 0 | 30 | 0 | 0 |
| SPA-050 | 40 | 40 | 55 | 60 | 60 | 10 | <5 | <5 |
| SPA-070 | 20 | 10 | 20 | 40 | 10 | 30 | 20 | 10 |
| SPA-100 | 40 | 40 | 40 | 30 | 30 | 20 | 20 | <5 |
| SPA-110 | 10 | 20 | 50 | 30 | 20 | 10 | 30 | 10 |
| SPA-130 | 30 | 60 | 40 | 70 | 40 | 0 | <5 | 0 |
| SPA-140 | 50 | 50 | 60 | 50 | 50 | 0 | 0 | 0 |
| SPA-150 | 40 | 40 | 40 | 40 | 50 | 20 | 20 | 10 |
| SPA-160 | 50 | 45 | 50 | 50 | 55 | <5 | <5 | <5 |
| SPA-170 | 30 | 10 | 70 | 50 | 70 | 10 | 20 | 20 |

Appendix E-2: I

| Site ID | Fines T2 | Fines T3 | Gravel T1 | Gravel T2 | Gravel T3 | Coble T1 | Coble T2 | Coble T3 |
|---------|----------|----------|-----------|-----------|-----------|----------|----------|----------|
| SPA-200 | 35 | 40 | 50 | 40 | 50 | 10 | 10 | 10 |
| SPA-220 | 30 | 30 | 35 | 30 | 20 | 15 | 20 | 20 |
| SPA-240 | 30 | 30 | 25 | 25 | 25 | 40 | 40 | 40 |
| WIL-100 | 10 | 10 | 20 | 15 | 10 | 65 | 25 | 30 |
| WIL_170 | <5 | <2 | 15 | 20 | 5 | 25 | 30 | 30 |
| WIL_190 | 5 | 5 | 25 | 20 | 25 | 35 | 25 | 30 |
| BUT-010 | 75 | 85 | 5 | 25 | 15 | 0 | 0 | 0 |
| BUT-020 | 40 | 50 | 30 | 50 | 30 | 15 | 10 | 20 |
| BUT-030 | 30 | 35 | 30 | 40 | 40 | 35 | 30 | 25 |
| BUT-040 | 30 | 25 | 30 | 10 | 15 | 25 | 20 | 50 |
| BUT-050 | 20 | 40 | 5 | 30 | 30 | 75 | 50 | 30 |
| PES-050 | 30 | 40 | 40 | 60 | 55 | 10 | 10 | 5 |
| PES-140 | 30 | 30 | 40 | 40 | 35 | 30 | 30 | 30 |
| PES-200 | 25 | 20 | 20 | 25 | 20 | 40 | 35 | 40 |
| PES-210 | 10 | 5 | 30 | 40 | 30 | 40 | 30 | 40 |
| PES-230 | 50 | 40 | 20 | 40 | 40 | 15 | 10 | 10 |
| PES-240 | 20 | 20 | 30 | 50 | 50 | 40 | 30 | 25 |
| PES-060 | 50 | 45 | 40 | 25 | 40 | 20 | 25 | 15 |
| PES-070 | 30 | 20 | 30 | 50 | 30 | 50 | 20 | 50 |
| PES-080 | 35 | 25 | 50 | 30 | 50 | 30 | 35 | 25 |
| PES-095 | 35 | 20 | 40 | 35 | 40 | 30 | 30 | 40 |
| PES-100 | 20 | 20 | 25 | 40 | 30 | 50 | 40 | 50 |
| PES-120 | 20 | 30 | 20 | 20 | 10 | 35 | 40 | 15 |
| PES-150 | 60 | 70 | 30 | 20 | 20 | 10 | 10 | 10 |
| PES-160 | 20 | 20 | 20 | 20 | 10 | 30 | 40 | 40 |
| PES-170 | 40 | 25 | 30 | 45 | 30 | 30 | 15 | 30 |
| PES-180 | 10 | 10 | 20 | 15 | 15 | 30 | 40 | 40 |
| PES-190 | 15 | 10 | 30 | 35 | 20 | 35 | 35 | 30 |
| SGR-010 | 25 | 25 | 50 | 35 | 40 | 30 | 30 | 30 |
| SGR-030 | 15 | 25 | 30 | 15 | 15 | 10 | 30 | 25 |
| SGR-040 | 25 | 20 | 30 | 45 | 40 | 45 | 30 | 40 |
| SGR-060 | 20 | 15 | 30 | 30 | 30 | 40 | 40 | 35 |
| SGR-075 | 25 | 20 | 40 | 40 | 35 | 35 | 35 | 30 |
| SGR-080 | 25 | 15 | 40 | 40 | 25 | 35 | 35 | 60 |
| SGR-090 | 15 | 20 | 15 | 25 | 40 | 70 | 60 | 30 |
| SGR-110 | 30 | 25 | 20 | 25 | 20 | 40 | 30 | 50 |
| SGR-120 | 30 | 30 | 20 | 20 | 40 | 50 | 50 | 30 |
| SGR-130 | 30 | 10 | 10 | 25 | 10 | 30 | 30 | 30 |
| SGR-150 | 15 | 20 | 15 | 20 | 20 | 60 | 60 | 60 |
| STE-020 | 15 | 35 | 30 | 30 | 25 | 40 | 50 | 40 |
| STE-030 | 5 | 15 | 60 | 60 | 35 | 10 | 30 | 15 |
| STE-040 | 25 | 40 | 15 | 10 | 20 | 5 | 5 | 20 |
| STE-060 | 20 | 25 | 10 | 55 | 30 | 0 | 25 | 40 |
| STE-070 | 30 | 35 | 20 | 25 | 25 | 40 | 35 | 35 |
| STE-100 | 45 | 45 | 30 | 35 | 35 | 20 | 10 | 5 |
| STE-110 | 30 | 15 | 30 | 45 | 30 | 30 | 15 | 35 |
| STE-120 | 20 | 20 | 25 | 35 | 30 | 20 | 30 | 40 |
| PER-030 | 28 | 40 | 20 | 20 | 40 | 40 | 50 | 20 |
| PER-040 | 40 | 40 | 50 | 60 | 60 | 0 | 0 | 0 |
| PER-010 | 0 | 0 | 5 | 0 | 0 | 5 | 0 | 0 |
| PER-020 | 15 | 15 | 0 | 0 | 0 | 0 | 0 | 0 |
| PER-050 | 40 | 40 | 40 | 45 | 45 | 10 | 15 | 15 |
| PER-070 | 50 | 50 | 40 | 40 | 40 | 20 | 10 | 10 |
| PER-080 | 25 | 25 | 35 | 35 | 35 | 30 | 30 | 30 |

Appendix E-2: I

| Site ID | Boulder T1 | Boulder T2 | Boulder T3 | Bedrock T1 | Bedrock T2 | Bedrock T3 | Gradient T1 |
|---------|------------|------------|------------|------------|------------|------------|-------------|
| LAG-130 | 0 | 0 | 0 | 0 | 0 | 0 | 1.3 |
| LAG-150 | 0 | 0 | 0 | 0 | 0 | 0 | 7 |
| LAG-160 | 0 | 0 | 0 | 0 | 0 | 0 | 2.2 |
| LAG-180 | 0 | 0 | 0 | 0 | 0 | 0 | 1.8 |
| LAG-170 | 0 | 0 | 5 | 0 | 0 | 0 | 1.3 |
| LAG-190 | 15 | 20 | 30 | 30 | 30 | 15 | 1.2 |
| LAG-210 | 5 | 30 | 5 | 0 | 20 | 0 | 1 |
| LAG-220 | 0 | 0 | 10 | 0 | 0 | 0 | 0.5 |
| LAG-165 | 0 | 0 | 0 | 0 | 0 | 0 | 1.5 |
| LAG-240 | 5 | 5 | 2 | 0 | 0 | 0 | 1 |
| LAG-270 | 0 | 30 | 0 | 0 | 0 | 0 | 1 |
| LAG-320 | 5 | 5 | 10 | 5 | 0 | 0 | 2 |
| LAG-300 | 5 | 0 | 0 | 0 | 0 | 0 | 1 |
| LAG-330 | 40 | 40 | 40 | 0 | 0 | 0 | 3 |
| LAG-335 | 50 | 50 | 50 | 0 | 0 | 0 | 1 |
| LAG-380 | 25 | 15 | 25 | 0 | 0 | 0 | 2 |
| LAG-390 | 0 | 40 | 60 | 60 | 10 | 0 | 7 |
| WLK-180 | 5 | 5 | 0 | 0 | 0 | 0 | 4 |
| WLK-190 | 15 | 5 | 0 | 0 | 0 | 0 | 2 |
| WLK-230 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| WLK-240 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| WLK-170 | 20 | 0 | 0 | 0 | 0 | 0 | 7 |
| WLK-200 | 25 | 5 | 5 | 0 | 0 | 0 | 5 |
| SUI-010 | 0 | 5 | 5 | 0 | 0 | 0 | 1 |
| SUI-020 | 0 | 5 | 5 | 0 | 0 | 0 | 2 |
| SUI-050 | 5 | 0 | 0 | 0 | 0 | 0 | 2 |
| SUI-060 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| SUI-110 | 0 | 0 | 0 | 80 | 90 | 90 | 2 |
| SUI-130 | 5 | 0 | 0 | 0 | 0 | 0 | 3 |
| SUI-180 | 20 | 15 | 15 | 0 | 0 | 0 | 2 |
| SUI-210 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| SUI-260 | 0 | 0 | 0 | | | | |
| ALP-010 | 20 | 30 | 20 | 60 | 60 | 80 | 1.8 |
| ALP-040 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| ALP-070 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| ALP-080 | 0 | 0 | 0 | 0 | 0 | 0 | 0.5 |
| ALP-100 | 0 | 5 | 0 | 0 | 0 | 0 | 2 |
| ALP-110 | 10 | 5 | 20 | 0 | 0 | 0 | 1 |
| ALP-140 | 5 | 10 | 25 | 0 | 0 | 0 | 2 |
| SLE-030 | 0 | 10 | 0 | 0 | 0 | 0 | 1 |
| SLE-170 | 30 | 20 | 10 | 0 | 0 | 0 | 1.3 |
| SLE-180 | 20 | 10 | 40 | 0 | 0 | 0 | 4 |
| SLE-220 | 0 | 0 | 5 | 0 | 0 | 5 | 0.4 |
| SLE-210 | 10 | 10 | 20 | 0 | 0 | 0 | 1 |
| SLE-230 | 60 | 50 | 55 | 0 | 0 | 0 | 3.5 |
| SPA-020 | 50 | 0 | 0 | 0 | 100 | 100 | 8 |
| SPA-050 | 5 | 0 | 0 | 0 | 0 | 0 | |
| SPA-070 | 30 | 20 | 70 | 0 | 0 | 0 | 4.5 |
| SPA-100 | 10 | 10 | 30 | 0 | 0 | 0 | |
| SPA-110 | 20 | 30 | 50 | 0 | 0 | 0 | 11 |
| SPA-130 | 0 | 0 | 0 | 0 | 0 | 0 | 10 |
| SPA-140 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| SPA-150 | 0 | 0 | <5 | 0 | 0 | 0 | 3 |
| SPA-160 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| SPA-170 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |

Appendix E-2: I

| Site ID | Boulder T1 | Boulder T2 | Boulder T3 | Bedrock T1 | Bedrock T2 | Bedrock T3 | Gradient T1 |
|---------|------------|------------|------------|------------|------------|------------|-------------|
| SPA-200 | <5 | 15 | <5 | 0 | 0 | 0 | 0.5 |
| SPA-220 | 20 | 20 | 30 | 0 | 0 | 0 | 2 |
| SPA-240 | 5 | 5 | 5 | 0 | 0 | 0 | |
| WIL-100 | 10 | 50 | 50 | 0 | 0 | 0 | 4.5 |
| WIL_170 | 60 | 50 | 60 | 0 | 0 | 5 | 5 |
| WIL_190 | 30 | 50 | 40 | 0 | 0 | 0 | 7 |
| BUT-010 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| BUT-020 | 15 | 0 | 0 | 0 | 0 | 0 | 1 |
| BUT-030 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| BUT-040 | 25 | 40 | 10 | 0 | 0 | 0 | 1 |
| BUT-050 | 0 | 0 | 0 | 0 | 0 | 0 | 5 |
| PES-050 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| PES-140 | 0 | 0 | 5 | 0 | 0 | 0 | 0.5 |
| PES-200 | 25 | 20 | 20 | 0 | 0 | 0 | 5 |
| PES-210 | 20 | 20 | 20 | 0 | 0 | 5 | 3 |
| PES-230 | 15 | 0 | 10 | 0 | 0 | 0 | 2.5 |
| PES-240 | 10 | 0 | 5 | 0 | 0 | 0 | 2.5 |
| PES-060 | 0 | 0 | 0 | 0 | 0 | 0 | 1.5 |
| PES-070 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| PES-080 | 0 | 0 | 0 | 0 | 0 | 0 | 5 |
| PES-095 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| PES-100 | 0 | 0 | 0 | 0 | 0 | 0 | 1.5 |
| PES-120 | 25 | 20 | 45 | 0 | 0 | 0 | 1.5 |
| PES-150 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| PES-160 | 20 | 20 | 30 | 0 | 0 | 0 | 2.5 |
| PES-170 | 10 | 0 | 15 | 0 | 0 | 0 | 3.5 |
| PES-180 | 20 | 15 | 30 | 15 | 20 | 5 | 2.5 |
| PES-190 | 10 | 10 | 30 | 10 | 5 | 10 | 2.5 |
| SGR-010 | 0 | 10 | 5 | 0 | 0 | 0 | 1 |
| SGR-030 | 0 | 40 | 35 | 30 | 0 | 0 | 2.5 |
| SGR-040 | 10 | 0 | 0 | 0 | 0 | 0 | 4.5 |
| SGR-060 | 10 | 10 | 20 | 0 | 0 | 0 | 3 |
| SGR-075 | 0 | 0 | 15 | 0 | 0 | 0 | 3.5 |
| SGR-080 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| SGR-090 | 0 | 0 | 10 | 0 | 0 | 0 | 2 |
| SGR-110 | 5 | 15 | 15 | 0 | 0 | 0 | 3 |
| SGR-120 | 0 | 0 | 0 | 0 | 0 | 0 | 15 |
| SGR-130 | 40 | 15 | 50 | 0 | 0 | 0 | 5 |
| SGR-150 | 15 | 5 | 0 | 0 | 0 | 0 | 4 |
| STE-020 | 5 | 5 | 0 | 0 | 0 | 0 | 2 |
| STE-030 | 0 | 5 | 35 | 0 | 0 | 0 | 1 |
| STE-040 | 70 | 60 | 20 | 0 | 0 | 0 | 3 |
| STE-060 | 0 | 0 | 5 | 85 | 0 | 0 | 3 |
| STE-070 | 10 | 10 | 5 | 0 | 0 | 0 | 4 |
| STE-100 | 10 | 10 | 10 | 0 | 0 | 5 | 3.5 |
| STE-110 | 25 | 10 | 20 | 0 | 0 | 0 | 5 |
| STE-120 | 25 | 15 | 10 | 0 | 0 | 0 | 1.5 |
| PER-030 | 0 | 2 | 0 | 0 | 0 | 0 | 1 |
| PER-040 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| PER-010 | 0 | 0 | 0 | 0 | 100 | 100 | 1 |
| PER-020 | 0 | 0 | 0 | 85 | 85 | 85 | 1 |
| PER-050 | 10 | 0 | 0 | 0 | 0 | 0 | 1 |
| PER-070 | 0 | 0 | 0 | 0 | 0 | 0 | 2.5 |
| PER-080 | 0 | 0 | 0 | 0 | 0 | 0 | 1.5 |

Appendix E-2: I

| Site ID | Gradient T2 | Gradient T3 | Specific Conductance | Disolved Oxygen | pH | WaterTemp |
|---------|-------------|-------------|----------------------|-----------------|-----|-----------|
| LAG-130 | 1.3 | 1.5 | 200 | 11.5 | 8.1 | 11.5 |
| LAG-150 | 4 | 3 | 201 | 11 | 8 | 15.2 |
| LAG-160 | 1.5 | 1.3 | 150 | 5.8 | 8.1 | 11.6 |
| LAG-180 | 1 | 1.5 | 236 | 10.74 | 8.3 | 11.3 |
| LAG-170 | 1 | 1.8 | 139 | 11.23 | 8.2 | 11.1 |
| LAG-190 | 2.9 | 1.8 | 270 | 10.3 | 8.5 | 12.7 |
| LAG-210 | 1.8 | 1.5 | 131 | 11.75 | 8.4 | 11.6 |
| LAG-220 | 0.5 | 1 | 176 | 9.5 | 8.4 | 12.2 |
| LAG-165 | 1 | 1.25 | 139 | 11.37 | 8.2 | 10.5 |
| LAG-240 | 1 | 1 | 340 | 12 | 8.6 | 11.5 |
| LAG-270 | 1.5 | 1 | 394 | 11 | 8.4 | 11.6 |
| LAG-320 | 1.5 | 2 | 106 | 11.26 | 8.2 | 10.4 |
| LAG-300 | 1 | 2.5 | 383 | 11.2 | 8.3 | 11.3 |
| LAG-330 | 3 | 1.7 | 390 | 10.2 | 8.4 | 9.6 |
| LAG-335 | 5 | 3 | 560 | 10.34 | 8.4 | 10.8 |
| LAG-380 | 3.5 | 3 | 323 | 11 | 8 | 9.7 |
| LAG-390 | 9 | 4.5 | 179 | 13.24 | 8.1 | 9.1 |
| WLK-180 | 2.5 | | 152 | 10.37 | 8.7 | 12.4 |
| WLK-190 | 2 | | 219 | 98.3 | 8.5 | 15.5 |
| WLK-230 | 1 | 5 | 285 | 12.7 | 8.2 | 14.9 |
| WLK-240 | 1 | 1 | 126 | 10 | 7.6 | 16.2 |
| WLK-170 | 4 | 2 | 288 | 9.3 | 8.2 | 12.9 |
| WLK-200 | 2 | 6 | 152 | 12.63 | 8.5 | 11.6 |
| SUI-010 | 1 | 1 | 465 | 10.5 | | 13.4 |
| SUI-020 | 2 | 2 | 441 | 11.3 | | 13.9 |
| SUI-050 | 1 | 2 | 350 | 11.1 | | 14.4 |
| SUI-060 | 1 | 2 | 350 | 11.1 | | 14.4 |
| SUI-110 | 4 | 1 | 481 | 9.6 | | 18.8 |
| SUI-130 | 2 | 3 | 414 | 11.3 | | 14.3 |
| SUI-180 | 3 | 4 | 169 | 12.1 | | 10.8 |
| SUI-210 | 2 | 4 | 486 | 10.3 | | 10.7 |
| SUI-260 | | | 612 | 8.6 | 7.7 | 14.3 |
| ALP-010 | 2 | 2 | 1043 | 7.7 | 8.2 | 14 |
| ALP-040 | 1 | 1 | 1533 | 12.2 | 8.6 | 18.6 |
| ALP-070 | 3 | 0.5 | 1630 | 9.4 | 8.5 | 18.3 |
| ALP-080 | 1 | 0.5 | 1610 | 8.6 | 8.4 | 18.4 |
| ALP-100 | 2 | 2 | 2640 | 13.06 | 8.4 | 21.8 |
| ALP-110 | 2 | 4 | 1114 | 10 | 7.4 | 22.9 |
| ALP-140 | 5 | 6.5 | 1145 | 7.76 | 8.6 | 17.4 |
| SLE-030 | 0.5 | 0.3 | 548 | 12.76 | 8.1 | 18.4 |
| SLE-170 | 2 | 1.2 | 410 | 8.92 | 8.3 | 13 |
| SLE-180 | 1 | 1.5 | 497 | 8.4 | 7.8 | 13.5 |
| SLE-220 | 0.2 | 4 | 495 | 8.48 | 8.2 | 12.8 |
| SLE-210 | 1.5 | | 640 | 8 | 8.2 | 13.9 |
| SLE-230 | 4.5 | 2 | 498 | 8.12 | 8.2 | 10.9 |
| SPA-020 | 12 | 2 | 1238 | 11.7 | 8.3 | 16.2 |
| SPA-050 | | | 1119 | 9.9 | 8.1 | 16.7 |
| SPA-070 | 3 | 4 | 1.5 | 11.43 | 8.1 | 15.6 |
| SPA-100 | | | 1609 | 9.62 | 7.8 | 14 |
| SPA-110 | 2 | 7 | 1845 | 11.35 | 7.8 | 15.2 |
| SPA-130 | 3 | 4 | 1713 | 11.46 | 8.1 | 13.9 |
| SPA-140 | 3 | 1 | 813 | 12.69 | 8.1 | 14.7 |
| SPA-150 | 3 | 2 | 920 | 11.91 | 8.4 | 15 |
| SPA-160 | 1.5 | | 932 | 10.57 | 8.3 | 14.4 |
| SPA-170 | 1 | 2 | 319.2 | 13.09 | 7.6 | 11.9 |

Appendix E-2: I

| Site ID | Gradient T2 | Gradient T3 | Specific Conductance | Disolved Oxygen | pH | WaterTemp |
|---------|-------------|-------------|----------------------|-----------------|-----|-----------|
| SPA-200 | 2 | 2 | 1455 | 10.5 | 7.9 | 12.3 |
| SPA-220 | 2 | 1.5 | 737 | 11.35 | 8.3 | 13.6 |
| SPA-240 | | | 695 | 10.2 | 8.1 | 14.4 |
| WIL-100 | 1.5 | 1 | 477 | 7.81 | 8.4 | 16.2 |
| WIL_170 | 5 | 3 | 527 | 8.37 | 8.4 | 12 |
| WIL_190 | 5 | | 472 | 8.51 | 8.2 | 10.4 |
| BUT-010 | 1.5 | 1 | 295 | 10.6 | | 13.5 |
| BUT-020 | 3 | 1 | 283 | 10.1 | | 13.3 |
| BUT-030 | 1 | 1 | 281 | 11.1 | | 11.2 |
| BUT-040 | 1 | 0.5 | 274 | 11.8 | | 11 |
| BUT-050 | 2 | 4 | 149 | 10.1 | | 10.8 |
| PES-050 | 1 | 1 | 500 | 10.1 | | 13.9 |
| PES-140 | 3 | 1 | 543 | 10.8 | 8.5 | 12.2 |
| PES-200 | 4 | 3.5 | 664 | 10.5 | 8.4 | 10.7 |
| PES-210 | 2 | 4 | 606 | 10.5 | 8.5 | 10.9 |
| PES-230 | 0.5 | 1 | 672 | 9.8 | 8.2 | 11.8 |
| PES-240 | 2 | 3 | 465 | 10.3 | 8.3 | 11.2 |
| PES-060 | 1 | 2 | 493 | 10.1 | | 12.5 |
| PES-070 | 1 | 1 | 490 | 10.7 | | 13.1 |
| PES-080 | 2 | 2 | 535 | 10.4 | | 12.8 |
| PES-095 | 1 | 2 | 489 | 10.2 | | 15.9 |
| PES-100 | 2 | 1.5 | 491 | 11.2 | | 13.5 |
| PES-120 | 2 | | 479 | 9.08 | | 13.1 |
| PES-150 | 2.5 | 1 | 307 | 9.98 | 8.1 | 11.7 |
| PES-160 | 2.5 | 3.5 | 519 | 10.9 | 8.4 | 12 |
| PES-170 | 1 | 1.5 | 421 | 9.84 | 8.1 | 11.1 |
| PES-180 | 2.5 | 3.5 | 667 | 10.9 | 8.5 | 13 |
| PES-190 | 2.5 | 5 | 551 | 9.62 | 8.3 | 13.4 |
| SGR-010 | 4 | 2 | 726 | 11.6 | 8.5 | 13.6 |
| SGR-030 | 6.5 | 6 | 435 | 11.28 | 8.4 | 9.9 |
| SGR-040 | 2.5 | 4.5 | 791 | 10.9 | 8.5 | 13.9 |
| SGR-060 | 3 | 6 | 608 | 10.81 | 8.5 | 13 |
| SGR-075 | 1.5 | 5 | 645 | 10.5 | 8.5 | 14.7 |
| SGR-080 | 1 | 2 | 703 | 10.83 | | 11.2 |
| SGR-090 | 0.5 | 2 | 689 | 11.1 | | 11.5 |
| SGR-110 | 2 | 3 | 229 | 10.6 | | 11.8 |
| SGR-120 | 5 | 3 | 351 | 10.9 | | 9.8 |
| SGR-130 | 3 | | 649 | 11 | | 12.3 |
| SGR-150 | 3 | 5 | 820 | 10.7 | | 12.5 |
| STE-020 | 1 | 1 | 930 | 9.8 | | 18.1 |
| STE-030 | 8 | 14 | 235 | 8.5 | | 16.2 |
| STE-040 | 1 | 3 | 481 | 9.8 | | 16.7 |
| STE-060 | 1 | 4 | 429 | 83.8 | | 13.8 |
| STE-070 | 3 | 4 | 448 | 9.98 | 7.9 | 9.9 |
| STE-100 | 2 | 2 | 515 | 10.7 | 8.5 | 11.8 |
| STE-110 | 2.5 | 4 | 501 | 10.7 | 8.5 | 10.1 |
| STE-120 | 4 | 5 | 372 | 10 | 8.2 | 9.8 |
| PER-030 | 4 | 1 | 1802 | 9.5 | | 15.3 |
| PER-040 | 1 | 1 | 902 | 9.9 | | 16.4 |
| PER-010 | 1 | 1 | 1180 | 12.3 | | 16.6 |
| PER-020 | 1 | 1 | 1280 | 13.2 | | 15.3 |
| PER-050 | 2 | 1.5 | 980 | 10 | 8.5 | 15.5 |
| PER-070 | 3 | 2 | 914 | 9.85 | 8.3 | 15.4 |
| PER-080 | 1.5 | 0.5 | 849 | 8.53 | 7.9 | 14.6 |

Appendix E-2: I

| Site ID | Average Depth T1 | Average Depth T2 | Average Depth T3 | Average Velocity T1 |
|---------|------------------|------------------|------------------|---------------------|
| LAG-130 | 11 | 10 | 14 | 0.36 |
| LAG-150 | 8 | 10 | 7 | 0.67 |
| LAG-160 | 5 | 8 | 5 | 0.35 |
| LAG-180 | 5 | 5 | 5 | 0.18 |
| LAG-170 | 8 | 8 | 16 | 0.6 |
| LAG-190 | 11 | 12 | 13 | 0.62 |
| LAG-210 | 20 | 15 | 12 | 0.4 |
| LAG-220 | 13 | 10 | 15 | 0.35 |
| LAG-165 | 8 | 6 | 7 | 0.54 |
| LAG-240 | 5 | 4 | 7 | 0.38 |
| LAG-270 | 5 | 6 | 7 | 0.32 |
| LAG-320 | 9 | 8 | 13 | 0.44 |
| LAG-300 | 3 | 4 | 5 | 0.3 |
| LAG-330 | 3 | 3 | 3 | 0.24 |
| LAG-335 | 5 | 6 | 3 | 0.26 |
| LAG-380 | 4 | 3 | 5 | 0.25 |
| LAG-390 | 6 | 1 | 6 | 0.58 |
| WLK-180 | 18 | 13 | 17 | 0.85 |
| WLK-190 | 9 | 7 | 6 | 0.24 |
| WLK-230 | 5 | 5 | 3 | 0.32 |
| WLK-240 | 4 | 9 | 6 | 0.31 |
| WLK-170 | 6 | 5 | 6 | 0.27 |
| WLK-200 | 16 | 33 | 22 | 0.35 |
| SUI-010 | 10 | 19 | 14 | 0.38 |
| SUI-020 | 13 | 12 | 17 | 0.26 |
| SUI-050 | 9 | 10 | 9 | 0.3 |
| SUI-060 | 9 | 19 | 17 | 0.47 |
| SUI-110 | 7 | 13 | 10 | 0.5 |
| SUI-130 | 11 | 15 | 16 | 0.35 |
| SUI-180 | 12 | 8 | 9 | 0.27 |
| SUI-210 | 5 | 9 | 5 | 0.35 |
| SUI-260 | 9 | 5 | 4 | 0.51 |
| ALP-010 | 20 | 16 | 13 | 0.87 |
| ALP-040 | 5 | 3 | 8 | 0.4 |
| ALP-070 | 16 | 5 | 3 | 0.5 |
| ALP-080 | 9 | 7 | 11 | 0.65 |
| ALP-100 | 11 | 13 | 10 | 0.39 |
| ALP-110 | 16 | 11 | 8 | 0.45 |
| ALP-140 | 6 | 10 | 2 | 0.23 |
| SLE-030 | | | | |
| SLE-170 | | | | |
| SLE-180 | | | | |
| SLE-220 | | | | |
| SLE-210 | | | | |
| SLE-230 | | | | |
| SPA-020 | 15 | 9 | 6 | 0.33 |
| SPA-050 | 9 | 7 | 13 | 0.37 |
| SPA-070 | 17 | 7 | 11 | 0.24 |
| SPA-100 | 6 | 6 | 5 | 0.13 |
| SPA-110 | 13 | 9 | 5 | 0.45 |
| SPA-130 | 6 | 5 | 5 | 0.3 |
| SPA-140 | 4 | 4 | 3 | 0.21 |
| SPA-150 | 5 | 4 | 2 | 0.25 |
| SPA-160 | 5 | 3 | 3 | 0.17 |
| SPA-170 | 27 | 21 | 13 | 0.9 |

Appendix E-2: I

| Site ID | Average Depth T1 | Average Depth T2 | Average Depth T3 | Average Velocity T1 |
|---------|------------------|------------------|------------------|---------------------|
| SPA-200 | 6 | 10 | 7 | 0.21 |
| SPA-220 | 10 | 15 | 12 | 0.36 |
| SPA-240 | 8 | 8 | 9 | 0.13 |
| WIL-100 | . | | | |
| WIL_170 | | | | |
| WIL_190 | | | | |
| BUT-010 | 14 | 10 | 10 | 0.45 |
| BUT-020 | 13 | 9 | 10 | 0.25 |
| BUT-030 | 12 | 15 | 10 | 0.24 |
| BUT-040 | 13 | 15 | 11 | 0.5 |
| BUT-050 | 6 | 5 | 6 | 0.35 |
| PES-050 | 8 | 7 | 8 | 0.12 |
| PES-140 | 12 | 13 | 19 | 0.4 |
| PES-200 | 10 | 10 | 9 | 0.45 |
| PES-210 | 12 | 12 | 8 | 0.503 |
| PES-230 | 8 | 11 | 7 | 0.34 |
| PES-240 | 6 | 6 | 6 | 0.237 |
| PES-060 | 5 | 6 | 10 | 0.3 |
| PES-070 | 17 | 17 | 11 | 0.5 |
| PES-080 | 10 | 8 | 11 | 0.58 |
| PES-095 | 15 | 18 | 14 | 0.35 |
| PES-100 | 13 | 19 | 13 | 0.7 |
| PES-120 | 25 | 19 | 31 | 0.5 |
| PES-150 | 4 | 4 | 5 | 0.15 |
| PES-160 | 15 | 18 | 12 | 0.45 |
| PES-170 | 6 | 3 | 4 | 0.21 |
| PES-180 | 16 | 12 | 8 | 0.23 |
| PES-190 | 18 | 10 | 15 | 0.42 |
| SGR-010 | 9 | 21 | 19 | 0.17 |
| SGR-030 | 14 | 13 | 14 | 0.66 |
| SGR-040 | 17 | 8 | 11 | 0.8 |
| SGR-060 | 9 | 8 | 9 | 0.17 |
| SGR-075 | 10 | 12 | 13 | 0.42 |
| SGR-080 | 12 | 13 | 7 | 0.5 |
| SGR-090 | 12 | 5 | 12 | 0.75 |
| SGR-110 | 12 | 15 | 7 | 0.6 |
| SGR-120 | 12 | 11 | 4 | 0.65 |
| SGR-130 | 14 | 14 | 14 | 0.72 |
| SGR-150 | 7 | 10 | 7 | 0.33 |
| STE-020 | 10 | 9 | 13 | 0.4 |
| STE-030 | 2 | 3 | 5 | 0.08 |
| STE-040 | 17 | 20 | 13 | 0.3 |
| STE-060 | 20 | 15 | 13 | 1.35 |
| STE-070 | 16 | 14 | 17 | 0.38 |
| STE-100 | 8 | 13 | 16 | 0.4 |
| STE-110 | 8 | 12 | 8 | 0.45 |
| STE-120 | 7 | 6 | 5 | 0.22 |
| PER-030 | 5 | 5 | 4 | 0.2 |
| PER-040 | 3 | 3 | 3 | 0.19 |
| PER-010 | 16 | 17 | 5 | 0.05 |
| PER-020 | 4 | 4 | 4 | 0.2 |
| PER-050 | 8 | 6 | 6 | 0.28 |
| PER-070 | 13 | 16 | 10 | 0.43 |
| PER-080 | 5 | 5 | 5 | 0.2 |

Appendix E-2: I

| Site ID | Average Velocity T2 | Average Velocity T3 |
|---------|---------------------|---------------------|
| LAG-130 | 0.6 | 0.65 |
| LAG-150 | 0.55 | 0.48 |
| LAG-160 | 0.3 | 0.25 |
| LAG-180 | 0.32 | 0.3 |
| LAG-170 | 0.38 | 0.75 |
| LAG-190 | 0.4 | 0.25 |
| LAG-210 | 0.61 | 0.7 |
| LAG-220 | 0.28 | 0.5 |
| LAG-165 | 0.53 | 0.63 |
| LAG-240 | 0.33 | 0.26 |
| LAG-270 | 0.32 | 0.59 |
| LAG-320 | 0.64 | 0.34 |
| LAG-300 | 0.28 | 0.26 |
| LAG-330 | 0.28 | 0.19 |
| LAG-335 | 0.13 | 0.19 |
| LAG-380 | 0.27 | 0.33 |
| LAG-390 | 0.24 | 0.44 |
| WLK-180 | 0.67 | 0.65 |
| WLK-190 | 0.19 | 0.21 |
| WLK-230 | 0.18 | 0.3 |
| WLK-240 | 0.27 | 0.09 |
| WLK-170 | 0.33 | 0.42 |
| WLK-200 | 0.3 | 0.42 |
| SUI-010 | 0.37 | 0.3 |
| SUI-020 | 0.43 | 0.34 |
| SUI-050 | 0.43 | 0.39 |
| SUI-060 | 0.41 | 0.59 |
| SUI-110 | 0.43 | 0.47 |
| SUI-130 | 0.38 | 0.55 |
| SUI-180 | 0.3 | 0.23 |
| SUI-210 | 0.32 | 0.11 |
| SUI-260 | 0.26 | 0.11 |
| ALP-010 | 0.78 | 0.73 |
| ALP-040 | 0.26 | 0.33 |
| ALP-070 | 0.7 | 0.35 |
| ALP-080 | 0.45 | 0.75 |
| ALP-100 | 0.33 | 0.48 |
| ALP-110 | 0.45 | 0.78 |
| ALP-140 | 0.2 | 0.3 |
| SLE-030 | | |
| SLE-170 | | |
| SLE-180 | | |
| SLE-220 | | |
| SLE-210 | | |
| SLE-230 | | |
| SPA-020 | 0.24 | 0.4 |
| SPA-050 | 0.31 | 0.43 |
| SPA-070 | 0.4 | 0.43 |
| SPA-100 | 0.15 | 0.1 |
| SPA-110 | 0.14 | 0.13 |
| SPA-130 | 0.12 | 0.47 |
| SPA-140 | 0.26 | 0.2 |
| SPA-150 | 0.26 | 0.11 |
| SPA-160 | 0.15 | 0.19 |
| SPA-170 | 0.75 | 0.38 |

Appendix E-2: I

| Site ID | Average Velocity T2 | Average Velocity T3 |
|---------|---------------------|---------------------|
| SPA-200 | 0.27 | 0.16 |
| SPA-220 | 0.46 | 0.31 |
| SPA-240 | 0.34 | |
| WIL-100 | | |
| WIL_170 | | |
| WIL_190 | | |
| BUT-010 | 0.35 | 0.45 |
| BUT-020 | 0.35 | 0.54 |
| BUT-030 | 0.34 | 0.52 |
| BUT-040 | 0.29 | 0.6 |
| BUT-050 | 0.2 | 0.3 |
| PES-050 | 0.31 | 0.3 |
| PES-140 | 0.4 | 0.52 |
| PES-200 | 0.3 | 0.48 |
| PES-210 | 0.277 | 0.457 |
| PES-230 | 0.34 | 0.32 |
| PES-240 | 0.177 | 0.173 |
| PES-060 | 0.4 | 0.23 |
| PES-070 | 0.65 | 0.61 |
| PES-080 | 0.4 | 0.33 |
| PES-095 | 0.54 | 0.55 |
| PES-100 | 0.65 | 0.65 |
| PES-120 | 0.75 | 0.35 |
| PES-150 | 0.2 | 0.26 |
| PES-160 | 0.47 | 0.39 |
| PES-170 | 0.16 | 0.16 |
| PES-180 | 0.52 | 0.43 |
| PES-190 | 0.39 | 0.36 |
| SGR-010 | 0.46 | 0.57 |
| SGR-030 | 0.36 | 0.35 |
| SGR-040 | 0.38 | 0.4 |
| SGR-060 | 0.25 | 0.41 |
| SGR-075 | 0.52 | 0.34 |
| SGR-080 | 0.58 | 0.4 |
| SGR-090 | 0.32 | 0.5 |
| SGR-110 | 0.6 | 0.25 |
| SGR-120 | 0.5 | 0.2 |
| SGR-130 | 0.5 | 0.5 |
| SGR-150 | 0.5 | 0.63 |
| STE-020 | 0.3 | 0.35 |
| STE-030 | 0.05 | 0.03 |
| STE-040 | 0.38 | 0.56 |
| STE-060 | 0.83 | 0.6 |
| STE-070 | 0.45 | 0.41 |
| STE-100 | 0.3 | 0.4 |
| STE-110 | 0.43 | 0.37 |
| STE-120 | 0.2 | 0.18 |
| PER-030 | 0.12 | 0.35 |
| PER-040 | 0.19 | 0.2 |
| PER-010 | 0.05 | 0.01 |
| PER-020 | 0.2 | 0.2 |
| PER-050 | 0.31 | 0.2 |
| PER-070 | 0.63 | 0.32 |
| PER-080 | 0.2 | 0.15 |

Appendix E, E-3: Bioassessment metrics

| Site Code: | ALP-010 | ALP-040 | ALP-070 | ALP-080 | ALP-100 | ALP-110 | ALP-140 | BUT-010 | BUT-020 | BUT-030 | BUT-040 | BUT-050 | LAG-130 | LAG-150 |
|----------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| Coleoptera Taxa | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 5 | 8 | 7 | 6 | 2 |
| Diptera Taxa | 4 | 3 | 4 | 2 | 3 | 3 | 3 | 3 | 6 | 10 | 8 | 7 | 3 | 4 |
| Ephemeroptera Taxa | 1 | 0 | 1 | 1 | 0 | 0 | 1 | 4 | 10 | 10 | 10 | 12 | 8 | 8 |
| Hemiptera Taxa | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepidoptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Megaloptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Plecoptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 2 | 3 | 5 | 5 | 5 |
| Trichoptera Taxa | 1 | 0 | 0 | 0 | 1 | 2 | 3 | 2 | 7 | 7 | 8 | 7 | 8 | 3 |
| Non-Insect Taxa | 8 | 7 | 9 | 10 | 7 | 7 | 5 | 5 | 4 | 2 | 4 | 7 | 5 | 4 |
| Taxa Richness | 16 | 11 | 14 | 13 | 11 | 12 | 12 | 16 | 32 | 36 | 41 | 45 | 35 | 26 |
| EPT Taxa | 2 | 0 | 1 | 1 | 1 | 2 | 4 | 6 | 20 | 19 | 21 | 24 | 21 | 16 |
| | | | | | | | | | | | | | | |
| % EPT | 1 | 0 | 1 | 4 | 0 | 0 | 82 | 25 | 41 | 67 | 62 | 70 | 43 | 58 |
| % Sensitive EPT | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 10 | 22 | 47 | 37 | 41 | 21 | 44 |
| % Chironomidae | 81 | 79 | 44 | 73 | 81 | 57 | 14 | 42 | 37 | 12 | 9 | 6 | 16 | 37 |
| % Coleoptera | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 12 | 8 | 7 | 15 | 10 | 30 | 0 |
| % Oligochaeta | 3 | 15 | 47 | 9 | 2 | 35 | 0 | 12 | 0 | 0 | 1 | 1 | 0 | 1 |
| % Non-insect | 15 | 20 | 54 | 13 | 11 | 40 | 1 | 20 | 3 | 0 | 2 | 2 | 9 | 3 |
| | | | | | | | | | | | | | | |
| % Intolerant | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 15 | 44 | 36 | 35 | 17 | 44 |
| % Tolerant | 8 | 3 | 4 | 2 | 2 | 4 | 1 | 8 | 3 | 0 | 0 | 1 | 1 | 1 |
| Tolerance Value | 6.1 | 5.9 | 5.6 | 5.9 | 5.9 | 5.7 | 5.1 | 5.3 | 4.8 | 3.4 | 3.5 | 3.5 | 4.0 | 3.6 |
| | | | | | | | | | | | | | | |
| % Predator | 1 | 1 | 1 | 2 | 0 | 2 | 2 | 7 | 5 | 7 | 5 | 19 | 2 | 28 |
| % Collector-filterer | 3 | 2 | 4 | 10 | 8 | 3 | 11 | 1 | 12 | 13 | 9 | 10 | 8 | 0 |
| %Collector-gatherer | 89 | 96 | 94 | 87 | 90 | 94 | 86 | 67 | 59 | 52 | 41 | 26 | 35 | 61 |
| % Scraper | 6 | 1 | 1 | 1 | 0 | 1 | 1 | 19 | 21 | 16 | 29 | 31 | 44 | 10 |
| % Shredder | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 3 | 11 | 8 | 14 | 10 | 1 |
| % Other | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 1 | 7 | 0 | 1 | 0 |

Appendix E, E-3: Bioassessment metrics

| Site Code: | LAG-160 | LAG-165 | LAG-170 | LAG-180 | LAG-190 | LAG-210 | LAG-220 | LAG-240 | LAG-270 | LAG-290 | LAG-300 | LAG-320 | LAG-330 | LAG-335 |
|----------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| Coleoptera Taxa | 0 | 8 | 4 | 8 | 7 | 5 | 7 | 6 | 7 | 7 | 6 | 5 | 7 | 4 |
| Diptera Taxa | 3 | 5 | 7 | 6 | 11 | 8 | 7 | 7 | 7 | 4 | 7 | 9 | 8 | 13 |
| Ephemeroptera Taxa | 4 | 8 | 8 | 9 | 10 | 8 | 8 | 7 | 7 | 5 | 8 | 8 | 7 | 7 |
| Hemiptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepidoptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Megaloptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Plecoptera Taxa | 3 | 6 | 6 | 4 | 5 | 6 | 4 | 6 | 6 | 4 | 5 | 6 | 3 | 2 |
| Trichoptera Taxa | 1 | 8 | 7 | 8 | 10 | 7 | 7 | 6 | 7 | 4 | 7 | 6 | 6 | 8 |
| Non-Insect Taxa | 3 | 5 | 6 | 10 | 4 | 5 | 9 | 6 | 5 | 7 | 8 | 7 | 7 | 5 |
| Taxa Richness | 14 | 40 | 38 | 45 | 47 | 39 | 42 | 38 | 39 | 31 | 41 | 41 | 38 | 40 |
| EPT Taxa | 8 | 22 | 21 | 21 | 25 | 21 | 19 | 19 | 20 | 13 | 20 | 20 | 16 | 17 |
| | | | | | | | | | | | | | | |
| % EPT | 17 | 51 | 34 | 43 | 51 | 47 | 49 | 49 | 41 | 23 | 21 | 60 | 60 | 40 |
| % Sensitive EPT | 10 | 37 | 21 | 32 | 36 | 24 | 34 | 32 | 28 | 10 | 11 | 38 | 34 | 14 |
| % Chironomidae | 80 | 20 | 20 | 19 | 30 | 22 | 25 | 17 | 38 | 53 | 58 | 23 | 20 | 25 |
| % Coleoptera | 0 | 16 | 22 | 28 | 11 | 10 | 18 | 14 | 9 | 10 | 12 | 9 | 7 | 13 |
| % Oligochaeta | 0 | 4 | 0 | 0 | 2 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 1 | 0 |
| % Non-insect | 0 | 5 | 2 | 4 | 5 | 3 | 4 | 1 | 2 | 5 | 2 | 5 | 5 | 4 |
| | | | | | | | | | | | | | | |
| % Intolerant | 10 | 33 | 18 | 26 | 33 | 25 | 24 | 29 | 25 | 8 | 6 | 41 | 33 | 19 |
| % Tolerant | 0 | 0 | 1 | 1 | 0 | 1 | 1 | 0 | 0 | 0 | 1 | 3 | 1 | 0 |
| Tolerance Value | 5.5 | 3.9 | 4.4 | 4.0 | 3.8 | 4.2 | 3.9 | 3.5 | 4.2 | 5.1 | 5.2 | 3.6 | 3.6 | 4.3 |
| | | | | | | | | | | | | | | |
| % Predator | 9 | 14 | 10 | 10 | 14 | 7 | 12 | 20 | 12 | 5 | 5 | 5 | 16 | 21 |
| % Collector-filterer | 1 | 10 | 22 | 5 | 5 | 19 | 3 | 1 | 4 | 9 | 4 | 4 | 2 | 0 |
| %Collector-gatherer | 90 | 41 | 34 | 34 | 56 | 40 | 43 | 43 | 63 | 71 | 73 | 47 | 45 | 43 |
| % Scraper | 0 | 26 | 30 | 33 | 16 | 28 | 35 | 17 | 10 | 13 | 13 | 27 | 12 | 21 |
| % Shredder | 0 | 9 | 3 | 18 | 7 | 5 | 7 | 18 | 11 | 2 | 4 | 16 | 25 | 13 |
| % Other | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 1 |

Appendix E, E-3: Bioassessment metrics

| Site Code: | LAG-380 | LAG-390 | PER-010 | PER-020 | PER-030 | PER-040 | PER-050 | PER-070 | PER-080 | PES-050 | PES-060 | PES-070 | PES-080 | PES-095 | PES-100 |
|----------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| Coleoptera Taxa | 7 | 6 | 0 | 0 | 0 | 0 | 2 | 1 | 1 | 3 | 5 | 3 | 3 | 5 | 5 |
| Diptera Taxa | 10 | 9 | 4 | 2 | 5 | 2 | 6 | 6 | 6 | 5 | 6 | 6 | 5 | 9 | 6 |
| Ephemeroptera Taxa | 6 | 8 | 2 | 0 | 1 | 1 | 1 | 3 | 4 | 9 | 10 | 9 | 7 | 10 | 11 |
| Hemiptera Taxa | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Lepidoptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Megaloptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Plecoptera Taxa | 5 | 7 | 0 | 0 | 1 | 0 | 0 | 0 | 2 | 3 | 3 | 3 | 1 | 5 | 6 |
| Trichoptera Taxa | 9 | 8 | 0 | 0 | 1 | 0 | 1 | 6 | 6 | 5 | 8 | 10 | 4 | 10 | 9 |
| Non-Insect Taxa | 7 | 7 | 4 | 4 | 6 | 4 | 7 | 5 | 10 | 6 | 5 | 5 | 6 | 7 | 5 |
| Taxa Richness | 44 | 45 | 11 | 6 | 14 | 7 | 17 | 21 | 30 | 32 | 37 | 36 | 26 | 46 | 42 |
| EPT Taxa | 20 | 23 | 2 | 0 | 3 | 1 | 2 | 9 | 12 | 17 | 21 | 22 | 12 | 25 | 26 |
| | | | | | | | | | | | | | | | |
| % EPT | 43 | 58 | 0 | 0 | 1 | 1 | 16 | 21 | 70 | 22 | 42 | 38 | 33 | 55 | 76 |
| % Sensitive EPT | 30 | 35 | 0 | 0 | 0 | 0 | 0 | 1 | 21 | 11 | 16 | 19 | 18 | 35 | 58 |
| % Chironomidae | 23 | 11 | 16 | 89 | 72 | 91 | 50 | 61 | 10 | 14 | 36 | 29 | 46 | 29 | 4 |
| % Coleoptera | 18 | 5 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 5 | 9 | 7 | 4 | 6 | 9 |
| % Oligochaeta | 0 | 0 | 75 | 11 | 3 | 7 | 21 | 5 | 3 | 0 | 1 | 0 | 3 | 0 | 0 |
| % Non-insect | 7 | 9 | 83 | 11 | 5 | 8 | 25 | 11 | 14 | 56 | 5 | 1 | 5 | 1 | 1 |
| | | | | | | | | | | | | | | | |
| % Intolerant | 23 | 31 | 0 | 0 | 0 | 0 | 0 | 1 | 18 | 7 | 10 | 11 | 15 | 30 | 46 |
| % Tolerant | 1 | 3 | 8 | 0 | 1 | 1 | 3 | 6 | 4 | 56 | 3 | 1 | 1 | 1 | 1 |
| Tolerance Value | 3.9 | 3.5 | 5.4 | 5.9 | 6.0 | 5.9 | 5.7 | 5.8 | 4.1 | 6.4 | 4.8 | 4.8 | 5.0 | 4.0 | 3.2 |
| | | | | | | | | | | | | | | | |
| % Predator | 18 | 16 | 0 | 0 | 1 | 0 | 3 | 3 | 11 | 11 | 7 | 5 | 3 | 6 | 8 |
| % Collector-filterer | 2 | 5 | 0 | 0 | 21 | 0 | 7 | 5 | 4 | 4 | 10 | 26 | 11 | 11 | 13 |
| %Collector-gatherer | 38 | 25 | 99 | 100 | 77 | 99 | 87 | 86 | 70 | 68 | 65 | 50 | 59 | 65 | 57 |
| % Scraper | 30 | 24 | 1 | 0 | 0 | 1 | 2 | 4 | 1 | 16 | 14 | 12 | 13 | 11 | 15 |
| % Shredder | 12 | 30 | 0 | 0 | 0 | 0 | 0 | 1 | 13 | 1 | 4 | 6 | 13 | 6 | 7 |
| % Other | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |

Appendix E, E-3: Bioassessment metrics

| Site Code: | PES-120 | PES-140 | PES-150 | PES-160 | PES-170 | PES-180 | PES-190 | PES-200 | PES-210 | PES-230 | PES-240 | SGR-010 | SGR-030 | SGR-040 | SGR-060 |
|----------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| Coleoptera Taxa | 7 | 6 | 6 | 7 | 6 | 10 | 7 | 8 | 9 | 8 | 8 | 7 | 8 | 5 | 8 |
| Diptera Taxa | 8 | 9 | 8 | 8 | 6 | 11 | 9 | 11 | 9 | 13 | 14 | 13 | 8 | 7 | 8 |
| Ephemeroptera Taxa | 10 | 11 | 7 | 10 | 10 | 12 | 11 | 10 | 11 | 7 | 11 | 11 | 10 | 11 | 11 |
| Hemiptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepidoptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Megaloptera Taxa | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Plecoptera Taxa | 5 | 6 | 4 | 4 | 4 | 6 | 7 | 5 | 8 | 5 | 8 | 4 | 3 | 5 | 5 |
| Trichoptera Taxa | 8 | 9 | 5 | 9 | 6 | 8 | 7 | 9 | 9 | 5 | 12 | 10 | 8 | 10 | 5 |
| Non-Insect Taxa | 5 | 1 | 4 | 4 | 8 | 5 | 4 | 6 | 6 | 4 | 2 | 9 | 6 | 4 | 8 |
| Taxa Richness | 43 | 42 | 35 | 42 | 40 | 52 | 45 | 49 | 52 | 42 | 55 | 54 | 43 | 42 | 45 |
| EPT Taxa | 23 | 26 | 16 | 23 | 20 | 26 | 25 | 24 | 28 | 17 | 31 | 25 | 21 | 26 | 21 |
| | | | | | | | | | | | | | | | |
| % EPT | 78 | 84 | 39 | 69 | 69 | 69 | 73 | 48 | 41 | 47 | 62 | 44 | 76 | 81 | 69 |
| % Sensitive EPT | 54 | 70 | 38 | 54 | 37 | 41 | 49 | 31 | 23 | 40 | 47 | 32 | 35 | 65 | 41 |
| % Chironomidae | 2 | 6 | 10 | 5 | 3 | 12 | 5 | 7 | 5 | 26 | 8 | 18 | 5 | 1 | 7 |
| % Coleoptera | 7 | 7 | 5 | 17 | 15 | 8 | 13 | 34 | 35 | 18 | 17 | 19 | 11 | 6 | 5 |
| % Oligochaeta | 0 | 0 | 0 | 0 | 4 | 0 | 1 | 1 | 0 | 1 | 0 | 0 | 3 | 0 | 1 |
| % Non-insect | 2 | 0 | 39 | 1 | 9 | 1 | 2 | 2 | 12 | 2 | 2 | 10 | 4 | 1 | 3 |
| | | | | | | | | | | | | | | | |
| % Intolerant | 47 | 66 | 24 | 51 | 35 | 40 | 46 | 27 | 20 | 39 | 41 | 22 | 31 | 64 | 37 |
| % Tolerant | 1 | 0 | 39 | 0 | 3 | 1 | 0 | 1 | 12 | 1 | 2 | 9 | 1 | 1 | 2 |
| Tolerance Value | 3.1 | 2.1 | 5.3 | 2.6 | 3.3 | 3.4 | 3.1 | 3.3 | 4.1 | 3.7 | 3.2 | 4.3 | 3.7 | 2.1 | 3.6 |
| | | | | | | | | | | | | | | | |
| % Predator | 8 | 8 | 25 | 8 | 22 | 12 | 11 | 11 | 9 | 7 | 27 | 18 | 5 | 8 | 9 |
| % Collector-filterer | 12 | 1 | 7 | 5 | 1 | 10 | 7 | 2 | 4 | 2 | 6 | 4 | 2 | 7 | 14 |
| %Collector-gatherer | 57 | 64 | 46 | 41 | 19 | 42 | 36 | 26 | 26 | 55 | 26 | 44 | 63 | 26 | 45 |
| % Scraper | 11 | 21 | 10 | 37 | 48 | 27 | 38 | 44 | 52 | 20 | 33 | 24 | 17 | 30 | 27 |
| % Shredder | 4 | 5 | 12 | 8 | 11 | 5 | 6 | 17 | 8 | 15 | 9 | 9 | 4 | 26 | 3 |
| % Other | 9 | 1 | 0 | 1 | 0 | 5 | 3 | 0 | 1 | 1 | 0 | 1 | 9 | 3 | 1 |

Appendix E, E-3: Bioassessment metrics

| Site Code: | SGR-075 | SGR-080 | SGR-090 | SGR-110 | SGR-120 | SGR-130 | SGR-150 | SLE-030 | SLE-170 | SLE-180 | SLE-210 | SLE-220 | SLE-230 | SPA-020 | SPA-050 |
|----------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| Coleoptera Taxa | 5 | 5 | 6 | 6 | 5 | 7 | 9 | 0 | 2 | 2 | 2 | 2 | 7 | 0 | 0 |
| Diptera Taxa | 9 | 9 | 6 | 6 | 10 | 11 | 13 | 2 | 3 | 10 | 5 | 7 | 7 | 2 | 2 |
| Ephemeroptera Taxa | 12 | 9 | 11 | 10 | 10 | 11 | 9 | 1 | 2 | 5 | 1 | 5 | 6 | 1 | 1 |
| Hemiptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepidoptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Megaloptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 |
| Plecoptera Taxa | 4 | 5 | 6 | 5 | 4 | 4 | 6 | 0 | 2 | 7 | 1 | 3 | 4 | 0 | 0 |
| Trichoptera Taxa | 8 | 6 | 7 | 7 | 9 | 6 | 6 | 0 | 4 | 4 | 0 | 4 | 9 | 1 | 1 |
| Non-Insect Taxa | 5 | 3 | 2 | 2 | 6 | 2 | 8 | 9 | 3 | 6 | 8 | 6 | 5 | 6 | 4 |
| Taxa Richness | 43 | 37 | 38 | 36 | 44 | 41 | 51 | 12 | 16 | 34 | 17 | 27 | 39 | 10 | 9 |
| EPT Taxa | 24 | 20 | 24 | 22 | 23 | 21 | 21 | 1 | 8 | 16 | 2 | 12 | 19 | 2 | 2 |
| | | | | | | | | | | | | | | | |
| % EPT | 79 | 68 | 84 | 73 | 69 | 71 | 75 | 0 | 12 | 55 | 14 | 47 | 66 | 18 | 32 |
| % Sensitive EPT | 58 | 46 | 50 | 43 | 17 | 38 | 53 | 0 | 3 | 28 | 0 | 27 | 48 | 0 | 0 |
| % Chironomidae | 4 | 6 | 5 | 4 | 13 | 7 | 7 | 48 | 67 | 30 | 60 | 28 | 22 | 32 | 25 |
| % Coleoptera | 9 | 4 | 3 | 19 | 7 | 8 | 8 | 0 | 6 | 4 | 1 | 1 | 9 | 0 | 0 |
| % Oligochaeta | 1 | 3 | 0 | 0 | 0 | 1 | 1 | 1 | 2 | 0 | 9 | 3 | 0 | 40 | 36 |
| % Non-insect | 2 | 3 | 0 | 0 | 1 | 2 | 3 | 31 | 2 | 4 | 12 | 6 | 2 | 51 | 36 |
| | | | | | | | | | | | | | | | |
| % Intolerant | 56 | 45 | 47 | 39 | 16 | 40 | 49 | 0 | 3 | 27 | 0 | 26 | 35 | 0 | 0 |
| % Tolerant | 1 | 0 | 0 | 0 | 1 | 1 | 1 | 8 | 0 | 1 | 1 | 1 | 1 | 0 | 0 |
| Tolerance Value | 2.5 | 3.2 | 2.8 | 2.9 | 4.3 | 3.3 | 2.8 | 5.8 | 5.6 | 4.1 | 5.7 | 4.2 | 3.4 | 5.4 | 5.3 |
| | | | | | | | | | | | | | | | |
| % Predator | 11 | 7 | 18 | 10 | 7 | 17 | 22 | 9 | 2 | 23 | 1 | 18 | 7 | 0 | 0 |
| % Collector-filterer | 5 | 21 | 10 | 6 | 5 | 16 | 3 | 24 | 11 | 1 | 14 | 13 | 3 | 0 | 6 |
| %Collector-gatherer | 39 | 34 | 28 | 37 | 72 | 28 | 38 | 60 | 79 | 55 | 84 | 57 | 54 | 99 | 93 |
| % Scraper | 26 | 21 | 32 | 41 | 10 | 35 | 31 | 7 | 6 | 14 | 1 | 5 | 15 | 0 | 0 |
| % Shredder | 14 | 12 | 11 | 5 | 7 | 3 | 5 | 0 | 2 | 7 | 0 | 7 | 18 | 0 | 0 |
| % Other | 4 | 6 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |

Appendix E, E-3: Bioassessment metrics

| Site Code: | SPA-070 | SPA-100 | SPA-110 | SPA-130 | SPA-140 | SPA-150 | SPA-160 | SPA-170 | SPA-200 | SPA-220 | SPA-240 | STE-020 | STE-030 | STE-040 | STE-060 |
|----------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| Coleoptera Taxa | 0 | 0 | 0 | 1 | 0 | 2 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
| Diptera Taxa | 5 | 5 | 5 | 7 | 9 | 6 | 8 | 6 | 3 | 4 | 4 | 3 | 4 | 4 | 5 |
| Ephemeroptera Taxa | 1 | 1 | 1 | 1 | 4 | 4 | 5 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| Hemiptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepidoptera Taxa | 0 | 1 | 0 | 1 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Megaloptera Taxa | 1 | 1 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Plecoptera Taxa | 0 | 0 | 0 | 0 | 4 | 4 | 5 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Trichoptera Taxa | 1 | 0 | 1 | 1 | 4 | 4 | 4 | 1 | 1 | 2 | 0 | 1 | 0 | 0 | 2 |
| Non-Insect Taxa | 4 | 6 | 6 | 5 | 3 | 3 | 4 | 7 | 7 | 7 | 7 | 7 | 2 | 5 | 9 |
| Taxa Richness | 12 | 14 | 13 | 17 | 24 | 25 | 26 | 18 | 14 | 15 | 12 | 12 | 7 | 10 | 18 |
| EPT Taxa | 2 | 1 | 2 | 2 | 12 | 12 | 14 | 3 | 2 | 4 | 1 | 2 | 1 | 1 | 3 |
| | | | | | | | | | | | | | | | |
| % EPT | 12 | 1 | 1 | 0 | 39 | 78 | 79 | 56 | 48 | 46 | 48 | 0 | 3 | 3 | 1 |
| % Sensitive EPT | 0 | 0 | 0 | 0 | 18 | 51 | 47 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| % Chironomidae | 45 | 52 | 59 | 64 | 26 | 12 | 11 | 36 | 20 | 21 | 30 | 80 | 71 | 61 | 67 |
| % Coleoptera | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| % Oligochaeta | 9 | 39 | 36 | 3 | 0 | 0 | 2 | 3 | 3 | 22 | 8 | 11 | 25 | 32 | 14 |
| % Non-insect | 11 | 42 | 38 | 4 | 3 | 5 | 8 | 5 | 6 | 23 | 16 | 13 | 25 | 34 | 18 |
| | | | | | | | | | | | | | | | |
| % Intolerant | 0 | 0 | 0 | 0 | 14 | 40 | 33 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| % Tolerant | 2 | 1 | 1 | 1 | 3 | 5 | 4 | 1 | 2 | 1 | 1 | 2 | 0 | 2 | 3 |
| Tolerance Value | 5.8 | 5.5 | 5.6 | 6.0 | 5.0 | 3.6 | 3.8 | 5.4 | 5.5 | 5.3 | 5.3 | 5.9 | 5.7 | 5.7 | 5.9 |
| | | | | | | | | | | | | | | | |
| % Predator | 2 | 2 | 1 | 3 | 4 | 9 | 7 | 4 | 1 | 1 | 1 | 1 | 0 | 0 | 3 |
| % Collector-filterer | 31 | 5 | 1 | 28 | 31 | 3 | 4 | 0 | 25 | 9 | 6 | 6 | 1 | 2 | 11 |
| %Collector-gatherer | 67 | 93 | 97 | 67 | 34 | 28 | 32 | 96 | 72 | 89 | 92 | 93 | 99 | 97 | 82 |
| % Scraper | 0 | 0 | 0 | 0 | 17 | 20 | 26 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 3 |
| % Shredder | 0 | 0 | 0 | 0 | 13 | 39 | 31 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| % Other | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix E, E-3: Bioassessment metrics

| Site Code: | STE-070 | STE-100 | STE-110 | STE-120 | SUI-010 | SUI-020 | SUI-050 | SUI-060 | SUI-110 | SUI-130 | SUI-180 | SUI-210 | SUI-260 | WIL-100 | WIL-170 | WIL-190 |
|----------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| Coleoptera Taxa | 0 | 5 | 9 | 9 | 2 | 1 | 1 | 0 | 1 | 1 | 0 | 1 | 1 | 2 | 2 | 1 |
| Diptera Taxa | 5 | 14 | 15 | 11 | 2 | 3 | 2 | 2 | 6 | 5 | 3 | 4 | 9 | 3 | 5 | 9 |
| Ephemeroptera Taxa | 1 | 10 | 11 | 10 | 4 | 5 | 6 | 4 | 7 | 3 | 8 | 3 | 6 | 2 | 6 | 5 |
| Hemiptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepidoptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| Megaloptera Taxa | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| Plecoptera Taxa | 1 | 4 | 8 | 6 | 1 | 1 | 2 | 1 | 1 | 2 | 2 | 2 | 1 | 1 | 1 | 3 |
| Trichoptera Taxa | 2 | 6 | 8 | 5 | 6 | 6 | 6 | 6 | 5 | 9 | 2 | 3 | 4 | 5 | 4 | 5 |
| Non-Insect Taxa | 8 | 5 | 7 | 4 | 9 | 9 | 5 | 7 | 7 | 6 | 10 | 5 | 7 | 6 | 8 | 3 |
| Taxa Richness | 17 | 44 | 58 | 46 | 24 | 25 | 22 | 20 | 27 | 26 | 25 | 19 | 28 | 19 | 27 | 26 |
| EPT Taxa | 4 | 20 | 27 | 21 | 11 | 12 | 14 | 11 | 13 | 14 | 12 | 8 | 11 | 8 | 11 | 13 |
| | | | | | | | | | | | | | | | | |
| % EPT | 33 | 53 | 54 | 65 | 25 | 36 | 35 | 43 | 57 | 52 | 61 | 25 | 11 | 23 | 34 | 84 |
| % Sensitive EPT | 4 | 44 | 43 | 32 | 3 | 9 | 6 | 8 | 11 | 11 | 37 | 4 | 8 | 3 | 15 | 61 |
| % Chironomidae | 53 | 30 | 14 | 13 | 52 | 42 | 37 | 37 | 18 | 23 | 23 | 34 | 51 | 28 | 41 | 10 |
| % Coleoptera | 0 | 2 | 17 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 0 |
| % Oligochaeta | 2 | 5 | 1 | 1 | 11 | 5 | 2 | 0 | 1 | 3 | 3 | 0 | 15 | 4 | 0 | 0 |
| % Non-insect | 10 | 7 | 4 | 2 | 18 | 18 | 6 | 4 | 4 | 6 | 15 | 1 | 30 | 7 | 2 | 2 |
| | | | | | | | | | | | | | | | | |
| % Intolerant | 4 | 40 | 42 | 32 | 1 | 7 | 5 | 7 | 11 | 11 | 36 | 3 | 9 | 3 | 7 | 46 |
| % Tolerant | 4 | 1 | 1 | 0 | 5 | 3 | 3 | 3 | 2 | 2 | 2 | 1 | 8 | 1 | 2 | 0 |
| Tolerance Value | 5.5 | 3.7 | 3.1 | 3.5 | 5.6 | 5.1 | 5.4 | 5.2 | 5.0 | 4.9 | 3.9 | 5.4 | 5.5 | 5.6 | 5.2 | 3.1 |
| | | | | | | | | | | | | | | | | |
| % Predator | 6 | 9 | 23 | 17 | 5 | 3 | 3 | 3 | 1 | 3 | 5 | 4 | 16 | 1 | 4 | 20 |
| % Collector-filterer | 3 | 3 | 4 | 8 | 10 | 7 | 25 | 19 | 22 | 20 | 1 | 39 | 4 | 42 | 20 | 5 |
| %Collector-gatherer | 84 | 73 | 38 | 56 | 83 | 84 | 67 | 72 | 73 | 75 | 80 | 53 | 78 | 53 | 64 | 48 |
| % Scraper | 0 | 7 | 26 | 14 | 0 | 4 | 3 | 5 | 4 | 1 | 12 | 2 | 1 | 0 | 8 | 0 |
| % Shredder | 4 | 8 | 6 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 3 | 4 | 28 |
| % Other | 2 | 1 | 3 | 1 | 2 | 1 | 2 | 1 | 0 | 1 | 2 | 0 | 0 | 1 | 0 | 0 |

Appendix E, E-3: Bioassessment metrics

| Site Code: | WLK-030 | WLK-050 | WLK-100 | WLK-120 | WLK-130 | WLK-140 | WLK-160 | WLK-170 | WLK-180 | WLK-190 | WLK-200 | WLK-230 | WLK-240 |
|----------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| Coleoptera Taxa | 1 | 9 | 2 | 1 | 0 | 6 | 4 | 3 | 3 | 4 | 0 | 2 | 1 |
| Diptera Taxa | 4 | 4 | 4 | 5 | 2 | 5 | 3 | 5 | 4 | 7 | 7 | 4 | 4 |
| Ephemeroptera Taxa | 1 | 2 | 6 | 3 | 1 | 7 | 5 | 5 | 6 | 5 | 3 | 2 | 3 |
| Hemiptera Taxa | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepidoptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Megaloptera Taxa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Plecoptera Taxa | 0 | 0 | 4 | 1 | 0 | 5 | 3 | 4 | 4 | 4 | 2 | 3 | 3 |
| Trichoptera Taxa | 2 | 1 | 4 | 2 | 0 | 12 | 12 | 9 | 10 | 9 | 9 | 1 | 1 |
| Non-Insect Taxa | 7 | 10 | 2 | 13 | 3 | 6 | 9 | 10 | 12 | 7 | 10 | 4 | 3 |
| Taxa Richness | 15 | 27 | 22 | 25 | 6 | 41 | 36 | 36 | 39 | 36 | 31 | 16 | 15 |
| EPT Taxa | 3 | 3 | 14 | 6 | 1 | 24 | 20 | 18 | 20 | 18 | 14 | 6 | 7 |
| | | | | | | | | | | | | | |
| % EPT | 5 | 23 | 45 | 37 | 0 | 70 | 23 | 52 | 74 | 59 | 47 | 4 | 9 |
| % Sensitive EPT | 0 | 1 | 13 | 28 | 0 | 47 | 13 | 26 | 54 | 53 | 32 | 1 | 2 |
| % Chironomidae | 44 | 31 | 15 | 24 | 58 | 9 | 49 | 33 | 9 | 18 | 32 | 63 | 64 |
| % Coleoptera | 1 | 3 | 0 | 0 | 0 | 16 | 5 | 0 | 3 | 7 | 0 | 1 | 0 |
| % Oligochaeta | 21 | 17 | 3 | 10 | 5 | 0 | 0 | 4 | 1 | 0 | 3 | 4 | 3 |
| % Non-insect | 25 | 35 | 3 | 37 | 35 | 2 | 2 | 9 | 7 | 10 | 18 | 12 | 3 |
| | | | | | | | | | | | | | |
| % Intolerant | 0 | 0 | 10 | 28 | 0 | 25 | 7 | 21 | 31 | 24 | 17 | 1 | 2 |
| % Tolerant | 4 | 18 | 0 | 26 | 30 | 1 | 1 | 3 | 2 | 2 | 4 | 7 | 0 |
| Tolerance Value | 5.8 | 5.9 | 5.1 | 5.1 | 6.5 | 3.7 | 5.2 | 4.7 | 3.6 | 3.6 | 4.8 | 6.0 | 5.8 |
| | | | | | | | | | | | | | |
| % Predator | 3 | 4 | 2 | 18 | 0 | 14 | 4 | 14 | 4 | 25 | 5 | 11 | 2 |
| % Collector-filterer | 26 | 8 | 40 | 1 | 7 | 13 | 21 | 2 | 11 | 1 | 6 | 18 | 23 |
| %Collector-gatherer | 71 | 86 | 53 | 79 | 93 | 41 | 58 | 65 | 37 | 26 | 58 | 70 | 74 |
| % Scraper | 0 | 0 | 1 | 2 | 0 | 25 | 7 | 2 | 15 | 13 | 14 | 1 | 1 |
| % Shredder | 0 | 0 | 5 | 0 | 0 | 8 | 6 | 17 | 33 | 35 | 17 | 0 | 0 |
| % Other | 0 | 2 | 0 | 0 | 0 | 0 | 3 | 1 | 0 | 0 | 0 | 1 | 0 |

Appendix F: Discrete field measures (by watershed and site)

| Watershed/Station | Analyte | Specific Conductance (mS/cm) | | | Oxygen, Dissolved (calculated, as mg/L) | | | Oxygen, Dissolved (% saturation) | | | pH | | |
|---------------------------------|---------|------------------------------|--------|-------|---|---------|---------|----------------------------------|--------|-------|------|--------|------|
| | Season | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet |
| Walker Creek | | | | | | | | | | | | | |
| 201WLK030 | | | 0.395 | 0.254 | | 14.3832 | 12.1464 | | 138.3 | 96.4 | | 7.43 | 7.39 |
| 201WLK090 | | 1.804 | 0.24 | 0.193 | 8.9143 | 9.9498 | 10.992 | 91.9 | 96.6 | 91.6 | 7.69 | 7.47 | 7.1 |
| 201WLK100 | | | 0.324 | 0.234 | | 10.815 | 11.78 | | 105 | 95 | | 7.76 | 7.47 |
| 201WLK140 | | 2.33 | 0.183 | 0.148 | 9.3534 | 13.335 | 11.685 | 91.7 | 127 | 95 | 7.71 | 7.66 | 7.42 |
| Lagunitas Creek | | | | | | | | | | | | | |
| 201LAG040 | | | 0.237 | 0.196 | | 10.4133 | 10.887 | | 101.1 | 95.5 | | 7.57 | 6.9 |
| 201LAG120 | | 0.226 | | | 13.2594 | | | 135.3 | | | 7.92 | | |
| 201LAG130 | | 0.246 | 0.218 | 0.18 | 9.741 | 12.0432 | 10.5881 | 95.5 | 115.8 | 93.7 | 7.63 | 7.95 | 6.96 |
| 201LAG150 | | 0.228 | 0.221 | 0.176 | 10.6485 | 11.08 | 11.0288 | 114.5 | 110.8 | 97.6 | | 7.87 | 7.23 |
| 201LAG160 | | 0.302 | 0.265 | 0.193 | 12.9948 | 10.812 | 11.4192 | 132.6 | 106 | 97.6 | 7.53 | 7.8 | 7.16 |
| 201LAG190 | | 0.228 | 0.272 | 0.214 | 11.1201 | 12.4416 | 11.4695 | 110.1 | 115.2 | 101.5 | 7.84 | 8.25 | 6.9 |
| 201LAG210 | | 0.233 | 0.204 | 0.181 | 10.9304 | 9.7337 | 11.1983 | 105.1 | 89.3 | 99.1 | 7.64 | 8.25 | 6.9 |
| 201LAG270 | | 0.596 | 0.391 | 0.293 | 5.151 | 10.6272 | 11.4777 | 51 | 98.4 | 98.1 | 7.54 | 7 | 7 |
| 201LAG320 | | 0.232 | 0.155 | 0.138 | 10.8562 | 12.639 | 10.791 | 105.4 | 114.9 | 98.1 | 7.29 | 7.06 | 6.9 |
| San Leandro Creek | | | | | | | | | | | | | |
| 204SLE030 | | 0.844 | 0.598 | 0.541 | 11.9192 | 19.3812 | 8.6205 | 126.8 | 208.4 | 82.1 | 7.77 | 8.33 | 7.27 |
| 204SLE190 | | 1.313 | 0.732 | 0.677 | 8.8173 | 11.286 | 12.9821 | 87.3 | 104.5 | 99.1 | 8.19 | 8.2 | 7.2 |
| 204SLE210 | | 0.634 | 0.642 | 0.567 | 2.3129 | 10.3464 | 12.9559 | 22.9 | 95.8 | 98.9 | 7.38 | 8.17 | 7.38 |
| 204SLE230 | | 0.742 | 0.524 | 0.478 | 8.6828 | 11.024 | 13.3196 | 88.6 | 104 | 99.4 | 8.17 | 8.44 | 7.36 |
| Wildcat/San Pablo Creeks | | | | | | | | | | | | | |
| 206SPA020 | | 1.546 | 1.275 | 1.131 | 5.4843 | 8.484 | 13.5888 | 54.3 | 80.8 | 119.2 | 7.87 | 8.05 | 7.08 |
| 206SPA070 | | 1.54 | 1.47 | 1.303 | 4.4238 | 9.24 | 11.2218 | 43.8 | 88 | 95.1 | 7.61 | 8.08 | 7.04 |
| 206SPA150 | | 0.914 | 0.915 | 0.904 | 12.6582 | 11.473 | 13.3824 | 124.1 | 104.3 | 98.4 | 8.34 | 8.4 | 7.21 |
| 206SPA200 | | 1.995 | 1.369 | 1.175 | 8.5162 | 13.413 | 12.2636 | 86.9 | 131.5 | 98.9 | 8.12 | 8.44 | 6.95 |
| 206SPA220 | | 0.95 | 0.664 | 1.131 | 8.3323 | 11.3925 | 13.5888 | 85.9 | 108.5 | 119.2 | 8.49 | 8.6 | 7.08 |
| 206WIL020 | | 0.879 | 0.705 | 0.63 | 4.3068 | 10.5644 | 12.7512 | 44.4 | 107.8 | 101.2 | 7.74 | 8.23 | 7.23 |
| 206WIL180 | | | 0.36 | 0.301 | | 8.72 | 11.7475 | | 80 | 92.5 | | 7.59 | 7.19 |
| Suisun Creek | | | | | | | | | | | | | |
| 207SUI010 | | 0.462 | 0.473 | 0.385 | 8.0444 | 8.0219 | 11.5128 | 88.4 | 82.7 | 98.4 | 7.86 | 8.29 | 7.12 |
| 207SUI020 | | 0.43 | | | 12.9317 | | | 145.3 | | | 7.57 | | |
| 207SUI060 | | 0.557 | 0.456 | 0.367 | 8.8179 | 13.493 | 11.4 | 96.9 | 131 | 95 | 7.68 | 7.95 | 7.36 |

Appendix F: Discrete field measures (by watershed and site)

| Watershed/Station | Analyte | Specific Conductance (mS/cm) | | | Oxygen, Dissolved (calculated, as mg/L) | | | Oxygen, Dissolved (% saturation) | | | pH | | |
|----------------------------------|---------|------------------------------|--------|-------|---|----------|---------|----------------------------------|--------|-------|------|--------|------|
| | Season | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet |
| 207SUI110 | | 0.545 | 0.437 | 0.354 | 17.5933 | 9.4668 | 13.2678 | 222.7 | 96.6 | 105.3 | 8.36 | 8.26 | 7.45 |
| 207SUI125 | | | 0.455 | 0.401 | | 9.4395 | 12.2616 | | 89.9 | 104.8 | | 7.96 | 7.14 |
| 207SUI130 | | 0.5 | 0.455 | | 8.334 | 9.4395 | | 92.6 | 89.9 | | 7.56 | 7.96 | |
| 207SUI180 | | | 0.224 | 0.193 | | 9.4233 | 12.5517 | | 93.3 | 97.3 | | 8.3 | 7.63 |
| 207SUI210 | | 0.599 | 0.562 | 0.468 | 8.109 | 9.2288 | 11.3256 | 90.1 | 89.6 | 93.6 | 7.46 | 7.79 | 7.34 |
| Arroyo las Positas | | | | | | | | | | | | | |
| 204ALP010 | | 1.641 | 1.781 | 1.755 | 7.5051 | 10.99612 | 13.5255 | 80.7 | 116.98 | 106.5 | 8.28 | 8.49 | 7.19 |
| 204ALP100 | | 1.519 | 2.799 | 2.661 | 9.4913 | 9.1304 | 12.1412 | 104.3 | 90.4 | 95.6 | 7.94 | 7.94 | 7.19 |
| 204ALP110 | | 1.426 | 1.183 | 1.055 | 7.7952 | 12.2576 | 11.844 | 89.6 | 130.4 | 112.8 | 8.03 | 8.04 | 7.12 |
| 204ALP140 | | 0.92 | 1.895 | 1.578 | 9.03 | 9.7344 | 12.4614 | 105 | 93.6 | 98.9 | 8.41 | 8.03 | 7.16 |
| Pescadero/Butano Creeks | | | | | | | | | | | | | |
| 202BUT010 | | 0.368 | 0.346 | 0.369 | 11.342 | 10.6488 | 10.857 | 107 | 104.4 | 98.7 | 8.1 | 7.86 | 7.46 |
| 202BUT020 | | 0.393 | 0.333 | 0.332 | 12.524 | 11.7393 | 10.9074 | 124 | 107.7 | 102.9 | 8.06 | 8.11 | 7.49 |
| 202PES050 | | 0.703 | 0.597 | 0.603 | 10.4958 | 11.0268 | 11.748 | 102.9 | 102.1 | 106.8 | 8.03 | 8.14 | 7.72 |
| 202PES070 | | 0.703 | 0.578 | 0.55 | 11.6756 | 10.8 | 10.9436 | 115.6 | 108 | 100.4 | 8.01 | 8.18 | 7.67 |
| 202PES140 | | 0.677 | 0.561 | 0.533 | 9.7566 | 11.3526 | 11.4663 | 96.6 | 107.1 | 103.3 | 8.4 | 8.44 | 7.42 |
| 202PES190 | | 0.622 | 0.555 | 0.636 | 9.4276 | 11.1996 | 11.033 | 96.2 | 103.7 | 100.3 | 8.33 | 8.35 | 7.81 |
| San Gregorio Creek | | | | | | | | | | | | | |
| 202SGR010 | | 0.903 | 0.773 | 0.723 | 11.0192 | 11.227 | 10.7147 | 113.6 | 103 | 98.3 | 8.3 | 8.32 | |
| 202SGR040 | | 0.899 | 0.777 | 0.763 | 10.2554 | 9.646 | 11.066 | 109.1 | 91 | 100.6 | 8.47 | 8.37 | |
| 202SGR080 | | 0.925 | 0.81 | 0.776 | 9.8159 | 10.918 | 10.901 | 95.3 | 103 | 99.1 | 8.27 | 8.51 | 7.55 |
| 202SGR090 | | 0.845 | 0.782 | 0.763 | 10.2752 | 10.1304 | 10.945 | 98.8 | 93.8 | 99.5 | 8.45 | 8.48 | 7.86 |
| Stevens/Permanente Creeks | | | | | | | | | | | | | |
| 205PER010 | | 1.296 | 0.73 | 1.457 | 9.88 | 16.2657 | 14.9856 | 104 | 174.9 | 156.1 | 7.9 | 8.91 | 7.62 |
| 205PER070 | | 1.02 | 1.012 | 1.142 | 8.924 | 11.2112 | 10.7841 | 92 | 107.8 | 104.7 | 8.18 | 8.33 | 7.5 |
| 205STE020 | | 1.005 | 1.037 | 0.741 | 10.137 | 16.6098 | 13.9536 | 109 | 178.6 | 136.8 | 7.83 | 8.26 | 8.37 |
| 205STE060 | | 0.497 | 0.502 | 0.602 | 11.0797 | 11.352 | 12.18 | 109.7 | 103.2 | 116 | 8.54 | 8.36 | 7.9 |
| 205STE100 | | 0.563 | 0.515 | 0.446 | 9.3728 | 12.0213 | 11.6208 | 92.8 | 108.3 | 107.6 | 8.43 | 8.51 | 7.77 |

Appendix F: Discrete field measures (by watershed and site)

| Watershed/Station | Analyte | Salinity (ppt) | | | Temperature (C) | | | Turbidity (NTU) | | | Velocity (ft/s) | | |
|---------------------------------|---------|----------------|--------|-----|-----------------|--------|-------|-----------------|--------|-------|-----------------|--------|-------|
| | Season | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet |
| Walker Creek | | | | | | | | | | | | | |
| 201WLK030 | | | | | | 13.57 | 5.5 | | 4.45 | 11.6 | | 1.75 | 0.772 |
| 201WLK090 | 0.92 | | | | 16.48 | 13.83 | 7.6 | 4 | 0.84 | 11.1 | 0.591 | 0.671 | 0.634 |
| 201WLK100 | | | | | | 13.98 | 6.02 | | 4.64 | 23.7 | | 0.611 | 0.578 |
| 201WLK140 | 0.11 | | | | 14.72 | 12.96 | 6.28 | 1.3 | 2.67 | 9.96 | 2.18 | 5.28 | 0.857 |
| Lagunitas Creek | | | | | | | | | | | | | |
| 201LAG040 | | | | | | 13.94 | 9.41 | | 2.9 | 4.9 | | 0.594 | 0 |
| 201LAG120 | | | | | 15.99 | | | 1.4 | | | 0 | | |
| 201LAG130 | 0.12 | | | | 14.7 | 13.29 | 9.77 | 0.6 | 1.46 | 10.75 | 0.171 | 1.77 | 9.32 |
| 201LAG150 | | | | | 18.56 | 15.71 | 9.84 | 0.1 | 0.55 | 1.55 | 1.91 | 2.89 | 1.18 |
| 201LAG160 | | | | | 15.8 | 14.39 | 8.56 | 0.45 | 1.46 | 2.52 | 0 | 0.166 | 0.731 |
| 201LAG190 | 0.11 | | | | 15.12 | 12.09 | 9.93 | 0.4 | 0.65 | 1.7 | 0.621 | 0.993 | 2.95 |
| 201LAG210 | 0.11 | | | | 13.61 | 11.45 | 9.86 | 0.7 | 3.34 | 2.3 | 0.699 | 2.09 | 0.923 |
| 201LAG270 | 0.29 | | | | 15.37 | 12.2 | 8.32 | 0.55 | 0.42 | 0.9 | 0.849 | 2.4 | 3.23 |
| 201LAG320 | 0.11 | | | | 14.02 | 10.89 | 10.8 | 0.5 | 1.1 | 2.7 | 0.994 | 2.1 | 3.55 |
| San Leandro Creek | | | | | | | | | | | | | |
| 204SLE030 | | | | | 17.9 | 18.63 | 12.78 | 1.5 | 0.37 | 0.2 | 2.91 | | 4.38 |
| 204SLE190 | 0.66 | | | | 14.83 | 11.75 | 3.8 | | 1.5 | 1.7 | 0 | 0.414 | 0.758 |
| 204SLE210 | 0.31 | | | | 15.04 | 12.06 | 3.82 | 30.2 | 2.6 | 2 | 0 | 1.78 | 1.36 |
| 204SLE230 | 0.37 | | | | 15.77 | 12.43 | 2.94 | 4.6 | 11 | 2 | 0.449 | 0.339 | 0.384 |
| Wildcat/San Pablo Creeks | | | | | | | | | | | | | |
| 206SPA020 | 0.74 | | | | 15.58 | 12.97 | 9.38 | 6.2 | 4 | 2.07 | 0.175 | 0.283 | 1.18 |
| 206SPA070 | 0.74 | | | | 15.14 | 12.9 | 8.01 | 6.5 | 4 | 3.2 | 0.523 | 0.356 | 1.2 |
| 206SPA150 | | | | | 14.66 | 11.34 | 2.52 | 7.3 | 1.7 | 0.8 | 0.24 | 1.09 | 0.932 |
| 206SPA200 | 1.03 | | | | 16.24 | 14.39 | 6.08 | 5.3 | 1.3 | 1.7 | 1.11 | 0.383 | 2.25 |
| 206SPA220 | 0.47 | | | | 16.53 | 13.13 | 9.38 | 2.6 | 1.6 | 2.07 | 0.645 | 0.797 | 1.18 |
| 206WIL020 | | | | | 16.34 | 15.84 | 5.29 | 32 | 9.7 | 8.6 | 0 | 0 | 0 |
| 206WIL180 | | | | | | 11.42 | 5.24 | | 3.6 | 2.9 | | 0.649 | 2.98 |
| Suisun Creek | | | | | | | | | | | | | |
| 207SUI010 | 0.22 | | | | 18.86 | 16.43 | 8.43 | 2.3 | 1.3 | 4.12 | 0 | 3.19 | 5.93 |
| 207SUI020 | | | | | 21.08 | | | 2.4 | | | 0.456 | | |
| 207SUI060 | 0.27 | | | | 18.9 | 13.92 | 7.37 | 1.3 | 1.9 | 3.9 | 0 | 6.07 | 7.61 |

Appendix F: Discrete field measures (by watershed and site)

| Watershed/Station | Analyte | Salinity (ppt) | | | Temperature (C) | | | Turbidity (NTU) | | | Velocity (ft/s) | | |
|----------------------------------|---------|----------------|--------|-----|-----------------|--------|-------|-----------------|--------|------|-----------------|--------|-------|
| | Season | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet |
| 207SUI110 | | | | | 27.19 | 15.82 | 5.51 | 0.45 | 5.3 | 2.82 | 1.59 | 1.52 | 2.41 |
| 207SUI125 | | | | | | 13.13 | 8.7 | | 0.65 | 3.2 | | 1.35 | 1.05 |
| 207SUI130 | | 0.24 | | | 20.72 | 13.13 | | 11 | 0.65 | | 0.857 | 1.35 | |
| 207SUI180 | | | | | | 15.17 | 4.74 | | 8.5 | 5.2 | | 1.23 | 0.762 |
| 207SUI210 | | 0.29 | | | 20.4 | 14.08 | 6.99 | 0.85 | 1.6 | 1.5 | 0.606 | 0.904 | 1.59 |
| Arroyo las Positas | | | | | | | | | | | | | |
| 204ALP010 | | | | | 18.4 | 17.74 | 5.24 | 20.2 | 3.7 | 2.6 | 1.4 | 1.92 | 2.6 |
| 204ALP100 | | 0.77 | | | 20.29 | 15.19 | 5.11 | 20.6 | 6.8 | 17.8 | 2.5 | 2.52 | 1.24 |
| 204ALP110 | | 0.72 | | | 22 | 18.06 | 12.85 | 129.1 | 1 | 0.85 | 0.128 | 0.334 | 0.286 |
| 204ALP140 | | | | | 22.85 | 13.52 | 5.5 | 85.1 | 5.5 | 3.7 | 1.88 | 0 | 0.356 |
| Pescadero/Butano Creeks | | | | | | | | | | | | | |
| 202BUT010 | | | | | 12.42 | 14.46 | 11.21 | 7.2 | 11 | 12.8 | 0.697 | 1.434 | 0.869 |
| 202BUT020 | | | 0.16 | | 15.45 | 11.39 | 12.34 | 6.7 | 8.5 | 6.5 | 0.535 | 0.724 | |
| 202PES050 | | | | | 14.54 | 11.93 | 10.99 | 5.2 | 2.7 | 5.6 | 1.36 | 1.79 | 1.21 |
| 202PES070 | | | 0.27 | | 15.56 | 15.64 | 11.4 | 3.1 | 2.2 | 4.71 | 1.56 | 1.288 | |
| 202PES140 | | | 0.26 | | 15.33 | 12.55 | 10.71 | 1.02 | 1.8 | 1.65 | 0.66 | 2.21 | |
| 202PES190 | | | 0.31 | | 16.08 | 12.2 | 10.94 | 1.1 | 1.5 | 3.2 | 0.11 | 0.331 | 1.36 |
| San Gregorio Creek | | | | | | | | | | | | | |
| 202SGR010 | | | 0.36 | | 16.39 | 11.35 | 11.41 | 1.23 | 1.2 | 3.25 | 0 | 1.16 | 1.07 |
| 202SGR040 | | | 0.38 | | 17.92 | 12.59 | 11.16 | 1.34 | 1.75 | 1.26 | 1.23 | 1.58 | 0.92 |
| 202SGR080 | | | 0.38 | | 13.89 | 12.4 | 11.21 | 0.76 | 1.2 | 1.75 | 0.828 | 0.824 | 1.02 |
| 202SGR090 | | | 0.38 | | 13.4 | 11.85 | 11.09 | 0.97 | | 1.51 | 0.447 | 2.48 | 1.76 |
| Stevens/Permanente Creeks | | | | | | | | | | | | | |
| 205PER010 | | | 0.74 | | 17.51 | 18.7 | 17 | 1.02 | 1.6 | 7.8 | 0.534 | 0.31 | 0 |
| 205PER070 | | | 0.57 | | 16.69 | 13.52 | 14.01 | 2.21 | 1.4 | 4.9 | 1.02 | 1.94 | |
| 205STE020 | | | 0.36 | | 18.67 | 18.85 | 14.46 | 1 | 0.5 | 6.71 | 1.42 | 0.648 | 0 |
| 205STE060 | | | 0.29 | | 15.32 | 11.26 | 12.99 | 1.32 | 2.7 | 11.8 | 1.91 | 1.42 | 0 |
| 205STE100 | | | 0.22 | | 15.39 | 10.62 | 12.07 | 0.65 | 0 | 2.7 | 1.11 | 0.915 | |

Appendix G-1: Conventional water quality chemistry results

| Watershed/Station Season | Alkalinity, as CaCO3 (mg/L) | | | Boron (mg/L) | | | Chloride (mg/L) | | | Hardness, as CaCO3 (mg/L) | | | SSC (mg/L) | | | Sulfate (mg/L) | | | TDS (mg/L) | | | TSS (mg/L) | | |
|---------------------------------|-----------------------------|--------|------|--------------|--------|-----|-----------------|--------|------|---------------------------|--------|------|------------|--------|------|----------------|--------|------|------------|--------|------|------------|------|------|
| | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Wet | |
| Walker Creek | | | | | | | | | | | | | | | | | | | | | | | | |
| 201WLK030 | | 111 | 69 | | | | | 52.3 | 33.2 | | 127 | 92.4 | | 13.4 | 1.5 | | 22.6 | 19.4 | | 231 | 217 | | 2.6 | |
| 201WLK090 | 70 | 75.8 | 64 | | | | | 581 | 22.8 | 18.6 | 238 | 80.8 | 77 | 17 | 0.4 | 1.9 | 70.7 | 11.6 | 12 | 1162 | 151 | 193 | 13.8 | 3.3 |
| 201WLK100 | | 108 | 78 | | | | | | 30.6 | 23.1 | | 124 | 101 | | 2.4 | 8.6 | | 14.7 | 13.2 | | 193 | 224 | | 8 |
| 201WLK140 | 59 | 58.3 | 55 | | | | | 20.4 | 15.2 | 14.6 | 59.7 | 64.6 | 66 | 0.6 | 1.4 | 1.9 | 4.39 | 10.6 | 10.7 | 103 | 113 | 132 | 1.2 | 1.7 |
| Lagunitas Creek | | | | | | | | | | | | | | | | | | | | | | | | |
| 201LAG040 | | 79.7 | 73 | | | | | | 16.9 | 17.6 | | 91.9 | 92.4 | | 0.8 | 4.2 | | 16.4 | 14.9 | | 149 | 162 | | 5.2 |
| 201LAG130 | 81 | 89.2 | 81 | | | | | 7.27 | 11.8 | 12.1 | 81.6 | 91.9 | 90.2 | 2.3 | 1.1 | 2.3 | 3.74 | 7.5 | 8.1 | 102 | 134 | 149 | 1 | 11.9 |
| 201LAG150 | 78 | 72.3 | 65 | 0.16 | | | | 19.7 | 16.3 | 13.3 | 90.9 | 81.3 | 79.2 | 3.7 | 0.6 | 0.1 | 13.6 | 15.1 | 12.4 | 128 | 142 | 144 | | 1 |
| 201LAG160 | 101 | 91.4 | 72 | 0.12 | | | | 29.8 | 21.2 | 15.5 | 106 | 104 | 92.4 | 2.4 | 1.8 | 0.6 | 15.5 | 15.6 | 12.5 | 172 | 164 | 142 | | 1 |
| 201LAG190 | 75 | 118 | 99 | | | | | 6.56 | 15.5 | 13 | 76.5 | 122 | 112 | 1.1 | 0.4 | 0.2 | 3.97 | 9.22 | 8.64 | 95 | 169 | 163 | 2.9 | 1 |
| 201LAG210 | 72 | 84.2 | 87.5 | | | | | 6.57 | 10.2 | 9.73 | 72.4 | 87.9 | 95.7 | 0.4 | ND | 0.2 | 3.89 | 6.77 | 8.13 | 92 | 134 | 144 | 1 | 1 |
| 201LAG270 | 202 | 179 | 148 | | | | | 19.7 | 16.6 | 12.7 | 221 | 194 | 174 | 4.2 | 0.4 | ND | 12.5 | 13.7 | 12.7 | 241 | 234 | 210 | 1.4 | 1 |
| 201LAG320 | 70 | 59.7 | 65 | | | | | 6.3 | 8.36 | 8.04 | 68.3 | 64.6 | 72.6 | 0.6 | 0.2 | 0.2 | 3.59 | 5.03 | 6.81 | 92 | 99 | 117 | 1 | 1 |
| San Leandro Creek | | | | | | | | | | | | | | | | | | | | | | | | |
| 204SLE030 | 224 | 228 | 232 | | | | | 25.3 | 25.8 | 21.8 | 245 | 258 | 233 | 3.4 | ND | 0.4 | 42.8 | 45.8 | 42.8 | 376 | 391 | 409 | 2.4 | 1 |
| 204SLE190 | 296 | 243 | 219 | | | | | 20.9 | 19.3 | 15.3 | 456 | 312 | 280 | 18.4 | 2.6 | | 248 | 151 | 133 | 698 | 482 | 486 | 3.4 | 1 |
| 204SLE210 | 215 | 310 | 267 | | | | | 14.1 | 23.2 | 17.8 | 186 | 299 | 245 | 22.4 | 5.6 | 0.8 | 9.48 | 29 | 27.2 | 276 | 375 | 380 | 20.5 | 1 |
| 204SLE230 | 176 | 217 | 193 | | | | | 18.1 | 18.1 | 14.8 | 199 | 217 | 180 | 1.2 | 11.7 | 0.5 | 86.1 | 53 | 45.3 | 349 | 312 | 326 | 2.6 | 1 |
| Wildcat/San Pablo Creeks | | | | | | | | | | | | | | | | | | | | | | | | |
| 206SPA020 | 312 | 396 | 396 | | | | | 70.2 | 95.8 | 77.8 | 462 | 513 | 452 | 7.5 | 6.6 | 2.3 | 210 | 207 | 200 | 824 | 842 | 856 | 8.7 | 53 |
| 206SPA070 | 345 | 426 | 418 | | | | | 85.5 | 128 | 91.2 | 449 | 584 | 524 | 11.7 | 6.6 | 2.1 | 188 | 280 | 269 | 858 | 1006 | 1008 | 10.5 | 2.1 |
| 206SPA150 | 240 | 224 | 206 | 0.5 | | | | 25.9 | 23.7 | 21.4 | 333 | 327 | 324 | 8.3 | 2.1 | 2 | 300 | 293 | 253 | 664 | 686 | 664 | | 1 |
| 206SPA200 | 282 | 320 | 317 | | | | | 346 | 173 | 134 | 398 | 481 | 417 | 4.2 | 1.5 | 0.1 | 126 | 202 | 187 | 999 | 868 | 868 | 3.2 | 1 |
| 206SPA220 | 332 | 300 | 271 | | | | | 24.6 | 25.4 | 18 | 318 | 320 | 254 | 64.2 | 1.3 | 0.6 | 49.1 | 53.5 | 43.6 | 460 | 419 | 388 | 1.9 | 1 |
| 206WIL020 | 319 | 275 | 238 | 0.44 | | | | 52.1 | 44 | 31.4 | 323 | 287 | 238 | 121.5 | 17.7 | 4.7 | 44.3 | 65 | 57.2 | 472 | 430 | 410 | | 7.3 |
| 206WIL180 | | 134 | 104 | | | | | | 17.6 | 14.9 | | 169 | 157 | | 6.5 | ND | | 35.6 | 27.3 | | 231 | 214 | | 1 |
| Suisun Creek | | | | | | | | | | | | | | | | | | | | | | | | |
| 207SUI010 | 166 | 175 | 158 | | | | | 5.4 | 17.7 | 11.7 | 175 | 192 | 232 | 6.5 | 1.5 | 1.6 | 23.3 | 77.3 | 54.3 | 197 | 299 | 276 | | 1.8 |
| 207SUI020 | 169 | | | 0.38 | | | | 12.2 | | | 202 | | | 3.1 | | | 49 | | | | 268 | | | |
| 207SUI060 | 173 | 164 | 149 | | | | | 11.6 | 16.9 | 10.4 | 188 | 188 | 157 | 0.8 | 1.2 | ND | 51.6 | 83 | 54.2 | 267 | 290 | 278 | | 1.2 |
| 207SUI110 | 183 | 151 | 138 | 0.28 | | | | 19 | 12.6 | 11.2 | 232 | 180 | 161 | 1.9 | 4 | 1.1 | 96.4 | 64.5 | 58.9 | 357 | 290 | 280 | | 1 |
| 207SUI125 | | | 135 | | | | | | | 11.1 | | | 144 | | | 1.6 | | | 60.4 | | | 265 | | 1.5 |
| 207SUI130 | 150 | 153 | | | | | | 9.95 | 17 | | 155 | 180 | | 14.3 | 0.9 | | 43 | 90.9 | | 225 | 275 | | | |
| 207SUI180 | | 85.8 | 76 | | | | | | 12.7 | 9.41 | | 85.7 | 70.4 | | 4.7 | 0.7 | | 20.6 | 18.6 | | 190 | 177 | | |
| 207SUI210 | 180 | 203 | 187 | | | | | 21.6 | 19.2 | 11.8 | 204 | 251 | 215 | 11.5 | 3.5 | 0.6 | 51.9 | 105 | 83.6 | 299 | 360 | 341 | | 1 |

Appendix G-1: Conventional water quality chemistry results

| Analyte Season | Alkalinity, as CaCO3 (mg/L) | | | Boron (mg/L) | | | Chloride (mg/L) | | | Hardness, as CaCO3 (mg/L) | | | SSC (mg/L) | | | Sulfate (mg/L) | | | TDS (mg/L) | | | TSS (mg/L) | |
|----------------------------------|-----------------------------|--------|------|--------------|--------|------|-----------------|--------|------|---------------------------|--------|-----|------------|--------|-------|----------------|--------|------|------------|--------|------|------------|------|
| | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Wet |
| Arroyo las Positas | | | | | | | | | | | | | | | | | | | | | | | |
| 204ALP010 | 261 | 355 | 367 | | | | 242 | 354 | 352 | 304 | 408 | 380 | 7.4 | 7 | 2.3 | 81.9 | 128 | 128 | 751 | 1073 | 1220 | 8.48 | 2.3 |
| 204ALP100 | 183 | 405 | 473 | | | | 265 | 763 | 610 | 207 | 427 | 457 | 9.8 | 18.4 | 24.1 | 63.6 | 151 | 146 | 705 | 1615 | 1781 | 12.5 | 22.6 |
| 204ALP110 | 342 | 343 | 360 | | | | 140 | 145 | 111 | 422 | 412 | 382 | 3.4 | 1.7 | 2.1 | 77.5 | 98.8 | 75.1 | 722 | 733 | 735 | 6.4 | 1.3 |
| 204ALP140 | 91 | 534 | 489 | | | | 158 | 281 | 212 | 131 | 311 | 309 | 33.4 | 6.7 | 1.6 | 39.2 | 156 | 137 | 415 | 1172 | 1114 | 37.2 | 3.4 |
| Pescadero/Butano Creeks | | | | | | | | | | | | | | | | | | | | | | | |
| 202BUT010 | 112 | 90.6 | 92.8 | 0.21 | 0.13 | 0.13 | 33.3 | 30.4 | 31.9 | 130 | 113 | 120 | 11.1 | 18.9 | 8.6 | 37.7 | 40.6 | 44.8 | 245 | 197 | 225 | | |
| 202BUT020 | 110 | 87.5 | 84.8 | 0.19 | 0.11 | 0.12 | 36.6 | 31.5 | 24.9 | 127 | 109 | 118 | 1.6 | 3.6 | 5.23 | 42.2 | 44.2 | 37.2 | 238 | 190 | 213 | | |
| 202PES050 | 210 | 176 | 172 | 0.54 | 0.34 | 0.27 | 50.6 | 43.1 | 39.4 | 257 | 222 | 221 | 7.3 | 3.2 | 3.9 | 82.8 | 100 | 93.5 | 417 | 365 | 375 | | |
| 202PES070 | 208 | 174 | 171 | 0.52 | 0.34 | 0.26 | 50.6 | 40.3 | 28.6 | 257 | 216 | 202 | 2 | 1.7 | 5.4 | 81.4 | 97.8 | 75 | 436 | 343 | 360 | | |
| 202PES140 | 215 | 177 | 175 | 0.44 | 0.3 | 0.29 | 45.4 | 34.6 | 22.8 | 252 | 222 | 224 | 0.3 | 0.7 | 0.93 | 86.3 | 97.3 | 72.5 | 428 | 350 | 358 | | |
| 202PES190 | 203 | 181 | 218 | 0.32 | 0.25 | 0.5 | 35.2 | 28.5 | 21.2 | 252 | 232 | 260 | ND | 0.3 | 1.48 | 82.6 | 98.2 | 91.3 | 390 | 347 | 435 | | |
| San Gregorio Creeks | | | | | | | | | | | | | | | | | | | | | | | |
| 202SGR010 | 241 | 214 | 194 | 0.4 | 0.3 | 0.31 | 74.4 | 59.4 | 43.5 | 354 | 289 | 272 | 2.4 | 1.5 | 5.23 | 139 | 152 | 112 | 560 | 463 | 475 | | |
| 202SGR040 | 258 | 229 | 221 | 0.43 | | 0.32 | 57.9 | 297 | 36.4 | 313 | 326 | 288 | 1.1 | 2.6 | 1.87 | 152 | 513 | 129 | 578 | 515 | 494 | | |
| 202SGR080 | 255 | 230 | 216 | 0.38 | 0.31 | 0.26 | 55.6 | 48.5 | 34.9 | 323 | 168 | 299 | 0.7 | 1.7 | 1.5 | 175 | 198 | 144 | 603 | 528 | 504 | | |
| 202SGR090 | 261 | 243 | 235 | 0.36 | | 0.26 | 31.5 | 77.2 | 25 | 343 | 355 | 312 | 2.3 | 2.4 | 2.79 | 165 | 296 | 145 | 558 | 534 | 520 | | |
| Stevens/Permanente Creeks | | | | | | | | | | | | | | | | | | | | | | | |
| 205PER010 | 405 | 186 | 409 | 0.25 | 0.06 | 0.26 | 74.4 | 54.1 | 99.9 | 707 | 337 | 784 | 2.9 | 2.3 | 11.85 | 311 | 155 | 254 | 902 | 452 | 1003 | | |
| 205PER070 | 202 | 189 | 185 | 0.17 | 0.06 | 0.18 | 55.8 | 49.7 | 42.3 | 424 | 498 | 533 | 3.5 | 1.5 | 9.74 | 336 | 326 | 379 | 720 | 724 | 850 | | |
| 205STE020 | 369 | 394 | 250 | 0.3 | 0.16 | 0.2 | 37.9 | 44.8 | 27.1 | 495 | 530 | 345 | 15.5 | 0.8 | 1.68 | 137 | 152 | 113 | 660 | 684 | 490 | | |
| 205STE060 | 184 | 201 | 195 | 0.26 | 0.13 | 0.16 | 21.8 | 23.2 | 27 | 212 | 232 | 290 | 1.9 | 2 | 7.71 | 47 | 47.6 | 108 | 306 | 294 | 418 | | |
| 205STE100 | 233 | 217 | 223 | 0.35 | | 0.22 | 16.6 | 14.6 | 13.4 | 252 | 265 | 239 | 10.3 | 0.6 | 1.5 | 50.8 | 56.3 | 39.9 | 344 | 294 | 314 | | |

Appendix G-2: Water nutrients chemistry results

| Watershed/Station | Analyte Season | Total Ammonia (mg/L as N) | | | Unionized Ammonia (calculated, mg/L as N) | | | Chlorophyll-a (ug/L) | | | Pheophytin-a (ug/L) | | | Dry |
|---------------------------------|-------------------|---------------------------|--------|------|---|----------|----------|----------------------|--------|------|---------------------|--------|------|-------|
| | | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | |
| Walker Creek | | | | | | | | | | | | | | |
| WLK030 | | | 0.07 | 0.07 | | 4.76E-04 | 2.42E-04 | | 2.2 | 3.6 | | 2.2 | 14 | |
| WLK090 | | 0.07 | 0.07 | 0.07 | 9.96E-04 | 5.22E-04 | 1.44E-04 | 1.9 | 20 | 3.6 | 6 | 2.3 | 9.9 | 0.05 |
| WLK100 | | | 0.07 | 0.07 | | 1.01E-03 | 2.91E-04 | | 2.3 | 10.3 | | 1.25 | 10.3 | |
| WLK140 | | 0.07 | 0.07 | 0.07 | 9.69E-04 | 7.48E-04 | 2.59E-04 | 1.25 | 1.4 | 2.3 | 4.5 | 2 | 2.3 | 0.05 |
| Lagunitas Creek | | | | | | | | | | | | | | |
| LAG040 | | | 0.07 | 0.15 | | 6.56E-04 | 2.10E-04 | | 0.9 | 1.5 | | 0.9 | 1.5 | |
| LAG130 | | 0.07 | 0.07 | 0.12 | 8.08E-04 | 1.44E-03 | 2.07E-04 | 1.25 | 1.6 | 2.3 | 3 | 0.9 | 6.2 | 0.05 |
| LAG150 | | 0.07 | 0.07 | 0.07 | | 1.50E-03 | 2.25E-04 | 2.8 | 0.75 | 1.25 | 1.25 | 0.75 | 4.3 | 0.102 |
| LAG160 | | 0.07 | 0.07 | 0.07 | 6.92E-04 | 1.11E-03 | 1.78E-04 | 0.9 | 0.75 | 1.25 | 0.9 | 0.75 | 2 | 0.113 |
| LAG190 | | 0.07 | 0.10 | 0.07 | 1.30E-03 | 3.76E-03 | 1.05E-04 | 1.5 | 0.75 | 1.25 | 5.4 | 0.75 | 3 | 0.05 |
| LAG210 | | 0.07 | 0.11 | 0.07 | 7.69E-04 | 3.85E-03 | 1.05E-04 | 1.25 | 0.75 | 1.1 | 5.8 | 1 | 1.1 | 0.05 |
| LAG270 | | 0.10 | 0.13 | 0.11 | 9.41E-04 | 2.85E-04 | 1.80E-04 | 1.25 | 0.75 | 1.1 | 3.9 | 0.75 | 1.7 | 0.094 |
| LAG320 | | 0.07 | 0.07 | 0.07 | 3.46E-04 | 1.64E-04 | 1.13E-04 | 1.25 | 0.75 | 1.25 | 4.9 | 0.75 | 1.25 | 0.05 |
| San Leandro Creek | | | | | | | | | | | | | | |
| SLE030 | | 0.46 | 0.07 | 0.07 | 9.06E-03 | 5.10E-03 | 3.07E-04 | 7.9 | 4.3 | 1.25 | 13 | 5.8 | 4.6 | 6.63 |
| SLE190 | | 0.07 | 0.12 | 0.10 | 2.85E-03 | 4.04E-03 | 1.93E-04 | 1.25 | 1.1 | 1.1 | 2.1 | 0.75 | 1.1 | 0.16 |
| SLE210 | | 0.07 | 0.17 | 0.07 | 4.57E-04 | 5.35E-03 | 2.04E-04 | 5.3 | 0.9 | 1.1 | 11 | 0.9 | 1.8 | 0.16 |
| SLE230 | | 0.07 | 0.07 | 0.07 | 2.92E-03 | 3.99E-03 | 1.81E-04 | 1.25 | 1.7 | 1.1 | 3.5 | 1.7 | 2.6 | 0.16 |
| Wildcat/San Pablo Creeks | | | | | | | | | | | | | | |
| SPA020 | | 0.14 | 0.07 | 0.07 | 2.99E-03 | 1.81E-03 | 1.48E-04 | 3.6 | 4.4 | 1.25 | 14 | 5.2 | 2.5 | 0.81 |
| SPA070 | | 0.14 | 0.14 | 0.07 | 1.54E-03 | 3.87E-03 | 1.25E-04 | 2.7 | 3.6 | 1.5 | 8.8 | 1.9 | 1.5 | 0.59 |
| SPA150 | | 0.07 | 0.13 | 0.13 | 3.96E-03 | 6.34E-03 | 2.39E-04 | 1.25 | 0.9 | 1.1 | 1.25 | 0.9 | 1.1 | 0.178 |
| SPA200 | | 0.07 | 0.07 | 0.07 | 2.62E-03 | 4.58E-03 | 8.81E-05 | 1.25 | 4.8 | 1.5 | 4.5 | 0.9 | 1.5 | 0.16 |
| SPA220 | | 0.07 | 0.16 | 0.07 | 6.25E-03 | 1.38E-02 | 1.48E-04 | 1.25 | 3.9 | 1.1 | 3.7 | 1.4 | 1.1 | 0.16 |
| WIL020 | | 0.37 | 0.20 | 0.10 | 5.90E-03 | 9.53E-03 | 2.23E-04 | 92 | 6.5 | 2.7 | 51 | 2.3 | 2.7 | 0.146 |
| WIL180 | | | 0.07 | 0.07 | | 5.52E-04 | 1.42E-04 | | 0.9 | 0.85 | | 0.9 | 0.85 | |
| Suisun Creek | | | | | | | | | | | | | | |
| SUI010 | | 0.07 | 0.07 | 0.07 | 1.81E-03 | 3.80E-03 | 1.51E-04 | 1.5 | 2 | 1.5 | 5.7 | 1.5 | 1.5 | 0.22 |
| SUI020 | | 0.11 | | | 1.71E-03 | | | 1.25 | | | 1.25 | | | 0.891 |
| SUI060 | | 0.07 | 0.07 | 0.07 | 1.21E-03 | 1.55E-03 | 2.43E-04 | 1.9 | 1.1 | 1.5 | 7.2 | 1.5 | 2.7 | 0.05 |
| SUI110 | | 0.07 | 0.07 | 0.07 | 9.21E-03 | 3.56E-03 | 2.78E-04 | 3.5 | 1.25 | 1.1 | 0.9 | 1.25 | 2.3 | 0.086 |
| SUI125 | | | | 0.07 | | | 1.70E-04 | | | 5.8 | | | 5.7 | |
| SUI130 | | 0.12 | 0.07 | | 1.83E-03 | 1.48E-03 | | 2.3 | 1.5 | | 6.1 | 1.9 | | 0.7 |
| SUI180 | | | 0.07 | 0.07 | | 3.63E-03 | 3.90E-04 | | 1.7 | 1.25 | | 3.9 | 3.4 | |
| SUI210 | | 0.12 | 0.07 | 0.12 | 1.35E-03 | 1.08E-03 | 3.98E-04 | 1.25 | 0.75 | 1.25 | 2.9 | 0.75 | 1.25 | 0.05 |

Appendix G-2: Water nutrients chemistry results

| | Nitrate (mg/L N) | | Nitrite (mg/L N) | | | Nitrate + Nitrite (calculated, mg/L N) | | | Total Kjeldahl Nitrogen (mg/L) | | | Orthophosphate as P (mg/L) | | | Phosphorus, Total as P (mg/L) | | | |
|---------------------------------|------------------|------|------------------|--------|------|--|--------|------|--------------------------------|--------|------|----------------------------|--------|-------|-------------------------------|--------|------|------|
| | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | |
| Watershed/Station | | | | | | | | | | | | | | | | | | |
| Walker Creek | | | | | | | | | | | | | | | | | | |
| WLK030 | 0.364 | 0.81 | | 0.007 | 0.02 | 0 | 0.371 | 0.83 | | | 0.68 | 0.37 | | 0.01 | 0.06 | | 0.07 | 0.06 |
| WLK090 | 0.153 | 0.84 | 0.02 | 0.007 | 0.02 | 0.07 | 0.16 | 0.86 | 0.37 | 0.37 | 0.37 | 0.1 | 0.032 | 0.08 | 0.121 | 0.06 | 0.08 | |
| WLK100 | 0.754 | 2.18 | | 0.007 | 0.03 | 0 | 0.761 | 2.21 | | | 0.64 | 0.59 | | 0.072 | 0.18 | | 0.12 | 0.22 |
| WLK140 | 0.067 | 0.36 | 0.02 | 0.007 | 0.02 | 0.07 | 0.074 | 0.38 | 0.37 | 0.54 | 0.37 | 0.093 | 0.014 | 0.04 | 0.097 | 0.04 | 0.05 | |
| Lagunitas Creek | | | | | | | | | | | | | | | | | | |
| LAG040 | 0.167 | 0.54 | | 0.007 | 0.02 | 0 | 0.174 | 0.56 | | | 0.37 | 0.37 | | 0.05 | 0.08 | | 0.09 | 0.08 |
| LAG130 | 0.165 | 0.37 | 0.02 | 0.007 | 0.02 | 0.07 | 0.172 | 0.39 | 0.37 | 0.37 | 0.37 | 0.04 | 0.018 | 0.04 | 0.04 | 0.04 | 0.04 | |
| LAG150 | 0.313 | 0.51 | 0.007 | 0.007 | 0.02 | 0.109 | 0.32 | 0.53 | 0.37 | 0.37 | 0.37 | 0.02 | 0.028 | 0.04 | 0.04 | 0.04 | 0.04 | |
| LAG160 | 0.307 | 0.37 | 0.007 | 0.007 | 0.02 | 0.12 | 0.314 | 0.39 | 0.37 | 0.37 | 0.37 | 0.025 | 0.024 | 0.04 | 0.04 | 0.04 | 0.04 | |
| LAG190 | 0.256 | 0.27 | 0.02 | 0.007 | 0.02 | 0.07 | 0.263 | 0.29 | 0.37 | 0.37 | 0.37 | 0.04 | 0.026 | 0.04 | 0.04 | 0.04 | 0.04 | |
| LAG210 | 0.181 | 0.27 | 0.02 | 0.007 | 0.02 | 0.07 | 0.188 | 0.29 | 0.37 | 0.37 | 0.37 | 0.04 | 0.015 | 0.04 | 0.04 | 0.04 | 0.04 | |
| LAG270 | 0.485 | 0.58 | 0.02 | 0.007 | 0.02 | 0.114 | 0.492 | 0.6 | 0.37 | 0.37 | 0.37 | 0.063 | 0.052 | 0.04 | 0.076 | 0.06 | 0.04 | |
| LAG320 | 0.175 | 0.14 | 0.02 | 0.007 | 0.02 | 0.07 | 0.182 | 0.16 | 0.37 | 0.37 | 0.37 | 0.04 | 0.011 | 0.04 | 0.04 | 0.04 | 0.04 | |
| San Leandro Creek | | | | | | | | | | | | | | | | | | |
| SLE030 | 7.65 | 6.13 | 0.07 | 0.007 | 0.02 | 6.7 | 7.657 | 6.15 | 0.71 | 0.37 | 0.37 | 0.17 | 0.044 | 0.06 | 0.17 | 0.07 | 0.08 | |
| SLE190 | 0.217 | 0.41 | 0.02 | 0.007 | 0.02 | 0.18 | 0.224 | 0.43 | 0.37 | 0.37 | 0.37 | 0.17 | 0.146 | 0.09 | 0.2 | 0.12 | 0.3 | |
| SLE210 | 0.096 | 0.28 | 0.02 | 0.007 | 0.02 | 0.18 | 0.103 | 0.3 | 0.71 | 0.37 | 0.37 | 0.36 | 0.144 | 0.07 | 0.46 | 0.15 | 0.1 | |
| SLE230 | 0.219 | 0.35 | 0.02 | 0.007 | 0.02 | 0.18 | 0.226 | 0.37 | 0.37 | 0.37 | 0.37 | 0.17 | 0.075 | 0.04 | 0.07 | 0.09 | 0.08 | |
| Wildcat/San Pablo Creeks | | | | | | | | | | | | | | | | | | |
| SPA020 | 1.2 | 1.72 | 0.04 | 0.015 | 0.02 | 0.85 | 1.215 | 1.74 | 1.25 | 0.37 | 0.37 | 0.34 | 0.239 | 0.21 | 0.391 | 0.3 | 0.22 | |
| SPA070 | 0.676 | 0.87 | 0.02 | 0.007 | 0.02 | 0.61 | 0.683 | 0.89 | 1.02 | 0.37 | 0.37 | 0.44 | 0.282 | 0.22 | 0.503 | 0.3 | 0.23 | |
| SPA150 | 0.027 | 0.15 | 0.007 | 0.007 | 0.02 | 0.185 | 0.034 | 0.17 | 0.37 | 0.37 | 0.37 | 0.237 | 0.181 | 0.12 | 0.23 | 0.18 | 0.17 | |
| SPA200 | 0.427 | 0.48 | 0.02 | 0.007 | 0.02 | 0.18 | 0.434 | 0.5 | 0.37 | 0.37 | 0.37 | 1.99 | 0.322 | 0.3 | 0.61 | 0.36 | 0.31 | |
| SPA220 | 0.15 | 0.27 | 0.02 | 0.01 | 0.02 | 0.18 | 0.16 | 0.29 | 0.37 | 0.37 | 0.37 | 0.17 | 0.123 | 0.07 | 0.12 | 0.14 | 0.13 | |
| WIL020 | 0.183 | 0.62 | 0.007 | 0.007 | 0.02 | 0.153 | 0.19 | 0.64 | 1.68 | 0.37 | 0.37 | 0.041 | 0.028 | 0.04 | 0.38 | 0.09 | 0.06 | |
| WIL180 | 0.06 | 0.17 | | 0.007 | 0.02 | 0 | 0.067 | 0.19 | | | 0.37 | 0.37 | | 0.039 | 0.04 | 0.04 | 0.04 | |
| Suisun Creek | | | | | | | | | | | | | | | | | | |
| SUI010 | 0.672 | 1.1 | 0.02 | 0.007 | 0.02 | 0.24 | 0.679 | 1.12 | 0.37 | 0.37 | 0.37 | 0.04 | 0.02 | 0.04 | 0.04 | 0.04 | 0.04 | |
| SUI020 | | | 0.007 | | | 0.898 | 0 | 0 | 0.37 | | | 0.024 | | | 0.04 | | | |
| SUI060 | 0.211 | 0.67 | 0.02 | 0.007 | 0.02 | 0.07 | 0.218 | 0.69 | 0.37 | 0.37 | 0.37 | 0.04 | 0.02 | 0.04 | 0.04 | 0.04 | 0.04 | |
| SUI110 | 0.24 | 0.91 | 0.007 | 0.007 | 0.02 | 0.093 | 0.247 | 0.93 | 0.37 | 0.37 | 0.37 | 0.025 | 0.056 | 0.04 | 0.04 | 0.06 | 0.04 | |
| SUI125 | | 0.16 | | | 0.02 | 0 | 0 | 0.18 | | | | | | 0.04 | | | | |
| SUI130 | 0.097 | | 0.13 | 0.007 | | 0.83 | 0.104 | 0 | 0.37 | 0.37 | | 0.16 | 0.007 | | 0.188 | 0.04 | | |
| SUI180 | 0.043 | 0.99 | | 0.007 | 0.02 | 0 | 0.05 | 1.01 | | | 0.37 | | 0.083 | 0.04 | | 0.09 | 0.04 | |
| SUI210 | 1.31 | 0.98 | 0.02 | 0.007 | 0.02 | 0.07 | 1.317 | 1 | 0.37 | 0.37 | 0.37 | 0.04 | 0.044 | 0.04 | 0.05 | 0.04 | 0.04 | |

Appendix G-2: Water nutrients chemistry results

| Analyte Season | Total Ammonia (mg/L as N) | | | Unionized Ammonia (calculated, mg/L as N) | | | Chlorophyll-a (ug/L) | | | Pheophytin-a (ug/L) | | | Dry |
|----------------------------------|---------------------------|--------|------|---|----------|----------|----------------------|--------|-------|---------------------|--------|-----|-------|
| | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | |
| Arroyo las Positas | | | | | | | | | | | | | |
| ALP010 | 0.10 | 0.07 | 0.12 | 6.11E-03 | 6.68E-03 | 2.44E-04 | 1.7 | 3.1 | 3.9 | 5.4 | 2.2 | 5.7 | 1.59 |
| ALP100 | 0.10 | 0.14 | 0.11 | 3.33E-03 | 3.26E-03 | 2.23E-04 | 7.4 | 9.8 | 5.3 | 6.9 | 8 | 5.3 | 0.48 |
| ALP110 | 0.07 | 0.07 | 0.07 | 3.27E-03 | 2.53E-03 | 2.17E-04 | 1.7 | 3.7 | 3.1 | 4.3 | 2.5 | 4.4 | 6.59 |
| ALP140 | 0.07 | 0.07 | 0.07 | 7.87E-03 | 1.86E-03 | 1.43E-04 | 3.3 | 2.6 | 1.9 | 3.3 | 2.8 | 1.9 | 0.35 |
| Pescadero/Butano Creeks | | | | | | | | | | | | | |
| BUT010 | 0.41 | 0.07 | 0.05 | 1.10E-02 | 1.27E-03 | 3.16E-04 | 4.6 | 3.6 | 0.567 | 1.9 | 1.25 | | 0.728 |
| BUT020 | 0.49 | 0.07 | 0.05 | 1.49E-02 | 1.79E-03 | 3.44E-04 | 1.25 | 1.25 | 0.17 | 1.25 | 1.25 | | 0.286 |
| PES050 | 0.56 | 0.07 | ND | 1.60E-02 | 2.06E-03 | ND | 1.5 | 0.9 | 0.27 | 1.5 | 0.9 | | 0.331 |
| PES070 | 0.48 | 0.07 | ND | 1.41E-02 | 2.99E-03 | ND | 1.25 | 0.9 | 0.12 | 1.25 | 0.9 | | 0.15 |
| PES140 | 0.07 | 0.07 | ND | 4.51E-03 | 4.28E-03 | ND | 1.25 | 0.9 | 0.08 | 1.25 | 0.9 | | 0.106 |
| PES190 | 0.07 | 0.07 | 0.05 | 4.15E-03 | 3.28E-03 | 6.89E-04 | 1.25 | 0.75 | ND | 1.25 | 0.75 | | 0.123 |
| San Gregorio Creek | | | | | | | | | | | | | |
| SGR010 | 0.07 | 0.07 | ND | 3.89E-03 | 2.86E-03 | ND | 1.25 | 2.8 | 0.18 | 1.25 | 0.9 | | 0.241 |
| SGR040 | 0.07 | 0.07 | ND | 6.41E-03 | 3.68E-03 | ND | 1.25 | 1.4 | 0.15 | 1.25 | 1.5 | | 0.107 |
| SGR080 | 0.07 | 0.07 | ND | 3.17E-03 | 4.65E-03 | ND | 1.25 | 1.2 | 0.25 | 1.25 | 1.2 | | 0.131 |
| SGR090 | 0.07 | 0.07 | ND | 4.37E-03 | 4.36E-03 | ND | 1.25 | 2.3 | 0.08 | 1.25 | 2.3 | | 0.202 |
| Stevens/Permanente Creeks | | | | | | | | | | | | | |
| PER010 | 0.07 | 0.07 | 0.06 | 1.85E-03 | 1.61E-02 | 7.31E-04 | 1.25 | 22 | 4.5 | 2 | 4.9 | | 0.879 |
| PER070 | 0.07 | 0.07 | ND | 3.21E-03 | 3.61E-03 | ND | 1.25 | 1.4 | 0.5 | 1.25 | 0.75 | | 0.929 |
| STE020 | 0.07 | 0.07 | ND | 1.70E-03 | 4.39E-03 | ND | 4.9 | 6.7 | 4.1 | 2.1 | 3.3 | | 2.55 |
| STE060 | 0.07 | 0.07 | ND | 6.07E-03 | 3.13E-03 | ND | 6 | 2 | 1.95 | 1.25 | 1.4 | | 0.276 |
| STE100 | 0.07 | 0.07 | ND | 4.81E-03 | 4.34E-03 | ND | 1.25 | 1.1 | 0.08 | 1.25 | 0.75 | | 0.187 |

Appendix G-2: Water nutrients chemistry results

| | Nitrate (mg/L N) | | Nitrite (mg/L N) | | | Nitrate + Nitrite (calculated, mg/L N) | | | Total Kjeldahl Nitrogen (mg/L) | | | Orthophosphate as P (mg/L) | | | Phosphorus, Total as P (mg/L) | | |
|----------------------------------|------------------|--------|------------------|--------|--------|--|--------|--------|--------------------------------|--------|------|----------------------------|--------|--------|-------------------------------|--------|--------|
| | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet | Dry | Spring | Wet |
| Arroyo las Positas | | | | | | | | | | | | | | | | | |
| ALP010 | 4.39 | 5.56 | 0.02 | 0.037 | 0.03 | 1.61 | 4.427 | 5.59 | 0.37 | 0.52 | 0.75 | 0.17 | 0.023 | 0.04 | 0.047 | 0.05 | 0.04 |
| ALP100 | 4.1 | 5 | 0.03 | 0.12 | 0.03 | 0.51 | 4.22 | 5.03 | 0.55 | 0.72 | 0.54 | 0.17 | 0.092 | 0.07 | 0.132 | 0.14 | 0.11 |
| ALP110 | 8.04 | 6.61 | 0.02 | 0.049 | 0.03 | 6.61 | 8.089 | 6.64 | 0.37 | 0.37 | 0.37 | 0.17 | 0.076 | 0.06 | 0.094 | 0.09 | 0.08 |
| ALP140 | 2.86 | 2.94 | 0.02 | 0.013 | 0.02 | 0.37 | 2.873 | 2.96 | 0.65 | 0.37 | 0.37 | 0.17 | 0.038 | 0.04 | 0.134 | 0.07 | 0.05 |
| Pescadero/Butano Creeks | | | | | | | | | | | | | | | | | |
| BUT010 | 0.484 | 0.574 | 0.007 | 0.007 | 0 | 0.735 | 0.491 | 0.574 | 0.37 | 0.37 | 0.18 | 0.04 | 0.02 | 0.019 | 0.068 | 0.06 | 0.081 |
| BUT020 | 0.268 | 0.314 | 0.007 | 0.007 | 0 | 0.293 | 0.275 | 0.314 | 0.37 | 0.37 | 0.22 | 0.025 | 0.023 | 0.0242 | 0.04 | 0.04 | 0.033 |
| PES050 | 0.265 | 0.291 | 0.007 | 0.007 | 0 | 0.338 | 0.272 | 0.291 | 0.37 | 0.37 | 0.21 | 0.107 | 0.092 | 0.08 | 0.14 | 0.1 | 0.115 |
| PES070 | 0.145 | 0.178 | 0.007 | 0.007 | 0 | 0.157 | 0.152 | 0.178 | 0.37 | 0.37 | 0.15 | 0.104 | 0.096 | 0.0898 | 0.12 | 0.1 | 0.0896 |
| PES140 | 0.132 | 0.0997 | 0.007 | 0.007 | 0 | 0.113 | 0.139 | 0.0997 | 0.37 | 0.37 | 0 | 0.146 | 0.125 | 0.0974 | 0.13 | 0.12 | 0.0876 |
| PES190 | 0.094 | 0.154 | 0.007 | 0.007 | 0 | 0.13 | 0.101 | 0.154 | 0.37 | 0.37 | 0.23 | 0.118 | 0.092 | 0.15 | 0.11 | 0.09 | 0.173 |
| San Gregorio Creek | | | | | | | | | | | | | | | | | |
| SGR010 | 0.137 | 0.229 | 0.007 | 0.007 | 0 | 0.248 | 0.144 | 0.229 | 0.37 | 0.37 | 0.27 | 0.165 | 0.128 | 0.127 | 0.14 | 0.12 | 0.131 |
| SGR040 | 0.059 | 0.175 | 0.007 | 0.007 | 0.0051 | 0.114 | 0.066 | 0.1801 | 0.37 | 0.37 | 0.17 | 0.204 | 0.149 | 0.143 | 0.18 | 0.14 | 0.131 |
| SGR080 | 0.127 | 0.186 | 0.007 | 0.007 | 0 | 0.138 | 0.134 | 0.186 | 0.37 | 0.37 | 0.26 | 0.175 | 0.132 | 0.112 | 0.15 | 0.14 | 0.109 |
| SGR090 | 0.13 | 0.21 | 0.007 | 0.007 | 0 | 0.209 | 0.137 | 0.21 | 0.37 | 0.37 | 0.28 | 0.221 | 0.178 | 0.146 | 0.19 | 0.17 | 0.165 |
| Stevens/Permanente Creeks | | | | | | | | | | | | | | | | | |
| PER010 | 0.218 | 1.27 | 0.06 | 0.007 | 0.0141 | 0.939 | 0.225 | 1.2841 | 0.37 | 0.37 | 0.72 | 0.334 | 0.074 | 0.167 | 0.31 | 0.1 | 0.262 |
| PER070 | 1.54 | 2.11 | 0.007 | 0.007 | 0.0207 | 0.936 | 1.547 | 2.1307 | 0.37 | 0.37 | 0 | 0.025 | 0.012 | 0.014 | 0.04 | 0.04 | 0.056 |
| STE020 | 3 | 1.26 | 0.04 | 0.02 | 0.0074 | 2.59 | 3.02 | 1.2674 | 0.37 | 0.37 | 0 | 0.272 | 0.145 | 0.0646 | 0.52 | 0.32 | 0.186 |
| STE060 | 0.373 | 0.688 | 0.007 | 0.007 | 0.0057 | 0.283 | 0.38 | 0.6937 | 0.37 | 0.37 | 0.22 | 0.062 | 0.036 | 0.0296 | 0.05 | 0.05 | 0.066 |
| STE100 | 0.042 | 0.115 | 0.007 | 0.007 | 0 | 0.194 | 0.049 | 0.115 | 0.37 | 0.37 | 0 | 0.085 | 0.065 | 0.062 | 0.07 | 0.06 | 0.081 |

Appendix G-3: Concentrations of hardness-dependent metals in water

| Station | Season | Year | Hardness (mg/L) | Metal Name | Total Metal (ug/L) | Dissolved Metal (ug/L) | Acute WQO (ug/L) | Acute Exceedance Factor | Chronic WQO (ug/L) | Chronic Exceedance Factor | WQO Fraction |
|-----------|--------|------|-----------------|------------|--------------------|------------------------|------------------|-------------------------|--------------------|---------------------------|--------------|
| 201LAG130 | Dry | 1 | 81.6 | cadmium | ND | ND | 3.118223372 | | 0.966877315 | | Total |
| 201LAG270 | Dry | 1 | 221 | cadmium | ND | ND | 9.593900455 | | 2.114117258 | | Total |
| 204ALP010 | Dry | 1 | 304 | cadmium | 0.012 | 0.011 | 13.74681619 | 0.000872929 | 2.715592472 | 0.004418925 | Total |
| 204ALP100 | Dry | 1 | 207 | cadmium | 0.018 | 0.014 | 8.911181164 | 0.002019934 | 2.008224095 | 0.008963143 | Total |
| 204ALP110 | Dry | 1 | 422 | cadmium | 0.01 | 0.011 | 19.90092065 | 0.000502489 | 3.513238187 | 0.002846377 | Total |
| 204SLE030 | Dry | 1 | 245 | cadmium | 0.023 | 0.019 | 10.77705419 | 0.002134164 | 2.292373966 | 0.010033267 | Total |
| 204SLE230 | Dry | 1 | 199 | cadmium | ND | ND | 8.523677186 | | 1.947025743 | | Total |
| 206SPA020 | Dry | 1 | 462 | cadmium | 0.088 | 0.05 | 22.04128305 | 0.003992508 | 3.772151591 | 0.023328861 | Total |
| 206SPA070 | Dry | 1 | 449 | cadmium | 0.087 | 0.051 | 21.34295732 | 0.004076286 | 3.688553484 | 0.023586482 | Total |
| 206SPA200 | Dry | 1 | 398 | cadmium | 0.065 | 0.039 | 18.62896937 | 0.003489189 | 3.355370225 | 0.019371931 | Total |
| 206SPA220 | Dry | 1 | 318 | cadmium | 0.016 | 0.01 | 14.46300428 | 0.001106271 | 2.813312869 | 0.005687245 | Total |
| 206WIL020 | Dry | 1 | 323 | cadmium | 0.074 | 0.011 | 14.71977489 | 0.005027251 | 2.847987499 | 0.02598326 | Total |
| 207SUI010 | Dry | 1 | 175 | cadmium | ND | ND | 7.373397412 | | 1.760134007 | | Total |
| 207SUI020 | Dry | 1 | 202 | cadmium | ND | ND | 8.668761769 | | 1.970035959 | | Total |
| 207SUI060 | Dry | 1 | 188 | cadmium | ND | ND | 7.9941221 | | 1.86200568 | | Total |
| 207SUI110 | Dry | 1 | 232 | cadmium | 0.012 | 0.004 | 10.13423965 | 0.001184105 | 2.196308912 | 0.005463712 | Total |
| 201LAG040 | Spring | 1 | 91.9 | cadmium | 0.08 | 0.05 | 3.565665674 | 0.022436203 | 1.06146969 | 0.075367201 | Total |
| 201LAG130 | Spring | 1 | 91.9 | cadmium | ND | ND | 3.565665674 | | 1.06146969 | | Total |
| 201LAG270 | Spring | 1 | 194 | cadmium | ND | ND | 8.282492945 | | 1.908508843 | | Total |
| 204ALP010 | Spring | 1 | 408 | cadmium | 0.02 | 0.02 | 19.15778907 | 0.001043962 | 3.421390288 | 0.005845577 | Total |
| 204ALP100 | Spring | 1 | 427 | cadmium | 0.02 | 0.02 | 20.16709603 | 0.000991714 | 3.545881564 | 0.005640346 | Total |
| 204ALP110 | Spring | 1 | 412 | cadmium | 0.02 | 0.02 | 19.36978423 | 0.001032536 | 3.447700661 | 0.005800968 | Total |
| 204SLE030 | Spring | 1 | 258 | cadmium | 0.02 | 0.02 | 11.4242514 | 0.001750662 | 2.387349766 | 0.008377491 | Total |
| 204SLE230 | Spring | 1 | 217 | cadmium | ND | ND | 9.398256702 | | 2.084013084 | | Total |
| 206SPA020 | Spring | 1 | 513 | cadmium | 0.05 | 0.03 | 24.80465044 | 0.002015751 | 4.095400823 | 0.012208817 | Total |
| 206SPA070 | Spring | 1 | 584 | cadmium | 0.05 | 0.03 | 28.7100808 | 0.001741549 | 4.53418907 | 0.01102733 | Total |
| 206SPA200 | Spring | 1 | 481 | cadmium | 0.02 | 0.01 | 23.06642924 | 0.000867061 | 3.893431804 | 0.005136856 | Total |
| 206SPA220 | Spring | 1 | 320 | cadmium | 0.01 | 0.01 | 14.56565097 | 0.000686547 | 2.827196673 | 0.003537073 | Total |
| 206WIL020 | Spring | 1 | 287 | cadmium | ND | 0.02 | 12.88283673 | | 2.595620069 | | Total |
| 207SUI010 | Spring | 1 | 192 | cadmium | ND | ND | 8.186240685 | | 1.893042581 | | Total |
| 207SUI060 | Spring | 1 | 188 | cadmium | ND | ND | 7.9941221 | | 1.86200568 | | Total |
| 207SUI110 | Spring | 1 | 180 | cadmium | ND | ND | 7.611462447 | | 1.799501567 | | Total |
| | | | | | | | | | | | |
| 202BUT010 | Dry | 2 | 130 | cadmium | 0.034 | 0.021 | 5.272891375 | 0.006448075 | 1.393736194 | 0.02439486 | Total |
| 202PES050 | Dry | 2 | 257 | cadmium | 0.04 | 0.02 | 11.37431592 | 0.003516695 | 2.38008105 | 0.01680615 | Total |
| 202PES070 | Dry | 2 | 257 | cadmium | 0.03 | 0.021 | 11.37431592 | 0.002637521 | 2.38008105 | 0.012604613 | Total |
| 202SGR010 | Dry | 2 | 354 | cadmium | 0.023 | 0.019 | 16.32286553 | 0.001409066 | 3.06048052 | 0.00751516 | Total |
| 202SGR080 | Dry | 2 | 323 | cadmium | 0.02 | 0.018 | 14.71977489 | 0.001358716 | 2.847987499 | 0.007022503 | Total |
| 205PER010 | Dry | 2 | 707 | cadmium | 0.016 | 0.009 | 35.61769593 | 0.000449215 | 5.268371381 | 0.003036992 | Total |
| 205PER070 | Dry | 2 | 424 | cadmium | 0.074 | 0.071 | 20.00734263 | 0.003698642 | 3.526305453 | 0.020985136 | Total |
| 205STE020 | Dry | 2 | 495 | cadmium | 0.015 | 0.003 | 23.82513596 | 0.000629587 | 3.982137579 | 0.003766821 | Total |
| 205STE060 | Dry | 2 | 212 | cadmium | 0.006 | 0.014 | 9.154351297 | 0.000655426 | 2.046214586 | 0.002932244 | Total |
| 202BUT010 | Spring | 2 | 113 | cadmium | 0.02 | ND | 4.501872672 | 0.004442596 | 1.248502603 | 0.01601919 | Total |
| 202PES050 | Spring | 2 | 222 | cadmium | 0.01 | ND | 9.64288258 | 0.001037034 | 2.121624949 | 0.004713368 | Total |
| 202PES070 | Spring | 2 | 216 | cadmium | 0.02 | 0.01 | 9.349417518 | 0.002139171 | 2.076468482 | 0.009631738 | Total |
| 202SGR010 | Spring | 2 | 289 | cadmium | 0.02 | ND | 12.98414898 | 0.00154034 | 2.609812124 | 0.007663387 | Total |
| 202SGR080 | Spring | 2 | 168 | cadmium | ND | 0.01 | 7.041571493 | | 1.704610292 | | Total |
| 205PER010 | Spring | 2 | 337 | cadmium | 0.01 | ND | 15.44142056 | 0.000647609 | 2.94447083 | 0.003396196 | Total |
| 205PER070 | Spring | 2 | 498 | cadmium | 0.38 | 0.37 | 23.98807632 | 0.015841204 | 4.001075422 | 0.094974466 | Total |

Appendix G-3: Concentrations of hardness-dependent metals in water

| Station | Season | Year | Hardness (mg/L) | Metal Name | Total Metal (ug/L) | Dissolved Metal (ug/L) | Acute WQO (ug/L) | Acute Exceedance Factor | Chronic WQO (ug/L) | Chronic Exceedance Factor | WQO Fraction |
|-----------|--------|------|-----------------|------------|--------------------|------------------------|------------------|-------------------------|--------------------|---------------------------|--------------|
| 205STE020 | Spring | 2 | 530 | cadmium | ND | ND | 25.73379898 | | 4.201590002 | | Total |
| 205STE060 | Spring | 2 | 232 | cadmium | ND | ND | 10.13423965 | | 2.196308912 | | Total |
| 202BUT010 | Wet | 2 | 120 | cadmium | 0.035 | 0.024 | 4.817671335 | 0.007264921 | 1.308836424 | 0.02674131 | Total |
| 202PES050 | Wet | 2 | 221 | cadmium | 0.03 | 0.021 | 9.593900455 | 0.003126987 | 2.114117258 | 0.01419032 | Total |
| 202PES070 | Wet | 2 | 202 | cadmium | 0.031 | 0.021 | 8.668761769 | 0.003576059 | 1.970035959 | 0.015735753 | Total |
| 202SGR010 | Wet | 2 | 272 | cadmium | 0.028 | 0.019 | 12.12591297 | 0.002309104 | 2.488489208 | 0.011251807 | Total |
| 202SGR080 | Wet | 2 | 299 | cadmium | 0.025 | 0.026 | 13.49204633 | 0.001852944 | 2.680459658 | 0.009326759 | Total |
| 205PER010 | Wet | 2 | 784 | cadmium | 0.044 | 0.023 | 40.02296342 | 0.001099369 | 5.713855203 | 0.00770058 | Total |
| 205PER070 | Wet | 2 | 533 | cadmium | 0.95 | 1 | 25.89816629 | 0.036682134 | 4.220252761 | 0.225105001 | Total |
| 205STE020 | Wet | 2 | 345 | cadmium | 0.139 | 0.145 | 15.85552631 | 0.008766659 | 2.999216465 | 0.046345438 | Total |
| 205STE060 | Wet | 2 | 290 | cadmium | 0.181 | 0.179 | 13.03483881 | 0.013885864 | 2.616900234 | 0.069165801 | Total |
| | | | | | | | | | | | |
| 201LAG130 | Dry | 1 | 81.6 | copper | 0.929 | 0.796 | 11.55829696 | 0.068868277 | 7.840974337 | 0.101517996 | Dissolved |
| 201LAG270 | Dry | 1 | 221 | copper | 1.44 | 1.17 | 29.55192535 | 0.039591329 | 18.37020444 | 0.063690091 | Dissolved |
| 204ALP010 | Dry | 1 | 304 | copper | 2.1 | 1.64 | 39.90826719 | 0.041094242 | 24.12383054 | 0.06798257 | Dissolved |
| 204ALP100 | Dry | 1 | 207 | copper | 2.5 | 1.54 | 27.78475939 | 0.055426069 | 17.37110526 | 0.088652966 | Dissolved |
| 204ALP110 | Dry | 1 | 422 | copper | 1.25 | 0.902 | 54.35866372 | 0.016593491 | 31.92715945 | 0.028251809 | Dissolved |
| 204SLE030 | Dry | 1 | 245 | copper | 1.07 | 0.688 | 32.56654342 | 0.021125976 | 20.0619531 | 0.03429377 | Dissolved |
| 204SLE230 | Dry | 1 | 199 | copper | 0.774 | 0.826 | 26.77187258 | 0.030853277 | 16.79580197 | 0.049178956 | Dissolved |
| 206SPA020 | Dry | 1 | 462 | copper | 4.32 | 3.74 | 59.20045575 | 0.063175189 | 34.49589045 | 0.108418712 | Dissolved |
| 206SPA070 | Dry | 1 | 449 | copper | 4.17 | 3.13 | 57.6296369 | 0.054312332 | 33.66474198 | 0.092975612 | Dissolved |
| 206SPA200 | Dry | 1 | 398 | copper | 3.85 | 3.12 | 51.44097734 | 0.060652036 | 30.36902694 | 0.102736252 | Dissolved |
| 206SPA220 | Dry | 1 | 318 | copper | 2.74 | 2 | 41.63765028 | 0.04803345 | 25.07002493 | 0.079776546 | Dissolved |
| 206WIL020 | Dry | 1 | 323 | copper | 5.74 | 0.78 | 42.25421106 | 0.018459699 | 25.40647112 | 0.03070084 | Dissolved |
| 207SUI010 | Dry | 1 | 175 | copper | 6.65 | 6.77 | 23.71864256 | 0.285429488 | 15.04897097 | 0.449864646 | Dissolved |
| 207SUI020 | Dry | 1 | 202 | copper | 2.43 | 2.15 | 27.151976 | 0.079183924 | 17.01192804 | 0.126381912 | Dissolved |
| 207SUI060 | Dry | 1 | 188 | copper | 1.05 | 0.893 | 25.37528369 | 0.035191725 | 15.99921498 | 0.055815239 | Dissolved |
| 207SUI110 | Dry | 1 | 232 | copper | 1.07 | 0.98 | 30.93585761 | 0.031678449 | 19.1487432 | 0.051178294 | Dissolved |
| 201LAG040 | Spring | 1 | 91.9 | copper | 0.88 | 0.78 | 12.92811646 | 0.060333615 | 8.679283937 | 0.089869165 | Dissolved |
| 201LAG130 | Spring | 1 | 91.9 | copper | 1.54 | 1.29 | 12.92811646 | 0.099782517 | 8.679283937 | 0.148629773 | Dissolved |
| 201LAG270 | Spring | 1 | 194 | copper | 1.13 | 0.95 | 26.13762788 | 0.036346068 | 16.43453309 | 0.05780511 | Dissolved |
| 204ALP010 | Spring | 1 | 408 | copper | 2.69 | 2.3 | 52.65788193 | 0.043678172 | 31.01986484 | 0.074146036 | Dissolved |
| 204ALP100 | Spring | 1 | 427 | copper | 2.87 | 2.27 | 54.96529012 | 0.04129879 | 32.25012565 | 0.070387323 | Dissolved |
| 204ALP110 | Spring | 1 | 412 | copper | 1.43 | 1.29 | 53.14415887 | 0.024273599 | 31.27954758 | 0.041241006 | Dissolved |
| 204SLE030 | Spring | 1 | 258 | copper | 1.47 | 1.36 | 34.19223327 | 0.039775115 | 20.96813628 | 0.064860319 | Dissolved |
| 204SLE230 | Spring | 1 | 217 | copper | 1.31 | 0.76 | 29.04769942 | 0.026163862 | 18.08571298 | 0.04202212 | Dissolved |
| 206SPA020 | Spring | 1 | 513 | copper | 2.13 | 1.68 | 65.33892158 | 0.025712086 | 37.7247248 | 0.044533128 | Dissolved |
| 206SPA070 | Spring | 1 | 584 | copper | 2.44 | 1.96 | 73.82671855 | 0.026548654 | 42.14349698 | 0.046507768 | Dissolved |
| 206SPA200 | Spring | 1 | 481 | copper | 2.26 | 2.07 | 61.49169604 | 0.033663082 | 35.70456609 | 0.057975778 | Dissolved |
| 206SPA220 | Spring | 1 | 320 | copper | 1.83 | 1.62 | 41.88434135 | 0.038677939 | 25.20469511 | 0.064273739 | Dissolved |
| 206WIL020 | Spring | 1 | 287 | copper | 0.05 | 1.88 | 37.80208061 | 0.049732712 | 22.96629149 | 0.081859102 | Dissolved |
| 207SUI010 | Spring | 1 | 192 | copper | 1.77 | 1.56 | 25.88366668 | 0.060269668 | 16.2896477 | 0.095766344 | Dissolved |
| 207SUI060 | Spring | 1 | 188 | copper | 1.49 | 1.33 | 25.37528369 | 0.052413207 | 15.99921498 | 0.083129079 | Dissolved |
| 207SUI110 | Spring | 1 | 180 | copper | 1.34 | 1.1 | 24.35662642 | 0.045162248 | 15.41562536 | 0.071356171 | Dissolved |
| | | | | | | | | | | | |
| 202BUT010 | Dry | 2 | 130 | copper | 1.01 | 0.85 | 17.92490272 | 0.047420062 | 11.67334675 | 0.07281545 | Dissolved |
| 202PES050 | Dry | 2 | 257 | copper | 1.59 | 1.27 | 34.06735135 | 0.037279094 | 20.89866987 | 0.060769418 | Dissolved |
| 202PES070 | Dry | 2 | 257 | copper | 1.4 | 1.23 | 34.06735135 | 0.036104949 | 20.89866987 | 0.05885542 | Dissolved |
| 202SGR010 | Dry | 2 | 354 | copper | 1.67 | 1.59 | 46.06491289 | 0.034516509 | 27.47603616 | 0.057868609 | Dissolved |

Appendix G-3: Concentrations of hardness-dependent metals in water

| Station | Season | Year | Hardness (mg/L) | Metal Name | Total Metal (ug/L) | Dissolved Metal (ug/L) | Acute WQO (ug/L) | Acute Exceedance Factor | Chronic WQO (ug/L) | Chronic Exceedance Factor | WQO Fraction |
|-----------|--------|------|-----------------|------------|--------------------|------------------------|------------------|-------------------------|--------------------|---------------------------|--------------|
| 202SGR080 | Dry | 2 | 323 | copper | 1.69 | 1.66 | 42.25421106 | 0.039286025 | 25.40647112 | 0.065337685 | Dissolved |
| 205PER010 | Dry | 2 | 707 | copper | 2.89 | 2.4 | 88.39391155 | 0.027151191 | 49.62033373 | 0.048367268 | Dissolved |
| 205PER070 | Dry | 2 | 424 | copper | 1.85 | 1.74 | 54.60136386 | 0.031867336 | 32.05641238 | 0.054279312 | Dissolved |
| 205STE020 | Dry | 2 | 495 | copper | 1.69 | 1.54 | 63.17662176 | 0.024376106 | 36.59071824 | 0.042087176 | Dissolved |
| 205STE060 | Dry | 2 | 212 | copper | 1.29 | 1.1 | 28.41665986 | 0.038709687 | 17.72902225 | 0.062045159 | Dissolved |
| 202BUT010 | Spring | 2 | 113 | copper | 1.11 | 0.76 | 15.70760206 | 0.048384215 | 10.3558636 | 0.073388375 | Dissolved |
| 202PES050 | Spring | 2 | 222 | copper | 1.2 | 1.01 | 29.67789905 | 0.034032059 | 18.44120979 | 0.054768641 | Dissolved |
| 202PES070 | Spring | 2 | 216 | copper | 1.13 | 1 | 28.92155936 | 0.034576282 | 18.01447136 | 0.055510927 | Dissolved |
| 202SGR010 | Spring | 2 | 289 | copper | 1.28 | 1.19 | 38.05023367 | 0.031274447 | 23.1029798 | 0.051508507 | Dissolved |
| 202SGR080 | Spring | 2 | 168 | copper | 1.28 | 1.24 | 22.82368611 | 0.054329524 | 14.53307695 | 0.085322606 | Dissolved |
| 205PER010 | Spring | 2 | 337 | copper | 2.97 | 2.58 | 43.9776758 | 0.058666129 | 26.34453529 | 0.097933024 | Dissolved |
| 205PER070 | Spring | 2 | 498 | copper | 1.69 | 1.55 | 63.53731636 | 0.024395113 | 36.78013045 | 0.042142319 | Dissolved |
| 205STE020 | Spring | 2 | 530 | copper | 1.33 | 1.22 | 67.37706752 | 0.018107051 | 38.79042361 | 0.031451062 | Dissolved |
| 205STE060 | Spring | 2 | 232 | copper | 1.36 | 1.07 | 30.93585761 | 0.034587695 | 19.1487432 | 0.055878341 | Dissolved |
| 202BUT010 | Wet | 2 | 120 | copper | 0.96 | 0.647 | 16.62279124 | 0.038922464 | 10.90162319 | 0.05934896 | Dissolved |
| 202PES050 | Wet | 2 | 221 | copper | 1.17 | 1.01 | 29.55192535 | 0.03417713 | 18.37020444 | 0.054980335 | Dissolved |
| 202PES070 | Wet | 2 | 202 | copper | 1.16 | 0.977 | 27.151976 | 0.035982648 | 17.01192804 | 0.057430292 | Dissolved |
| 202SGR010 | Wet | 2 | 272 | copper | 1.54 | 1.32 | 35.93769364 | 0.036730237 | 21.93663051 | 0.060173325 | Dissolved |
| 202SGR080 | Wet | 2 | 299 | copper | 1.49 | 1.41 | 39.28952464 | 0.035887428 | 23.7843793 | 0.059282607 | Dissolved |
| 205PER010 | Wet | 2 | 784 | copper | 3.84 | 2.33 | 97.43701595 | 0.023912883 | 54.20306778 | 0.042986497 | Dissolved |
| 205PER070 | Wet | 2 | 533 | copper | 2.26 | 1.68 | 67.73634474 | 0.024802047 | 38.97796775 | 0.043101272 | Dissolved |
| 205STE020 | Wet | 2 | 345 | copper | 1.98 | 1.45 | 44.96064492 | 0.032250427 | 26.87801603 | 0.053947434 | Dissolved |
| 205STE060 | Wet | 2 | 290 | copper | 1.93 | 1.4 | 38.17427295 | 0.036673914 | 23.17127231 | 0.060419643 | Dissolved |
| 201LAG130 | Dry | 1 | 81.6 | lead | 0.062 | 0.015 | 63.02482211 | 0.000238001 | 2.455986836 | 0.006107525 | Dissolved |
| 201LAG270 | Dry | 1 | 221 | lead | 0.076 | ND | 224.0481802 | | 8.73083593 | | Dissolved |
| 204ALP010 | Dry | 1 | 304 | lead | 0.247 | 0.053 | 336.2235493 | 0.000157633 | 13.10214902 | 0.004045138 | Dissolved |
| 204ALP100 | Dry | 1 | 207 | lead | 0.513 | 0.043 | 206.1390671 | 0.000208597 | 8.032943501 | 0.005352957 | Dissolved |
| 204ALP110 | Dry | 1 | 422 | lead | 0.14 | 0.049 | 510.4495279 | 9.59938E-05 | 19.89148528 | 0.002463366 | Dissolved |
| 204SLE030 | Dry | 1 | 245 | lead | 0.483 | 0.107 | 255.4691611 | 0.000418837 | 9.955266444 | 0.01074808 | Dissolved |
| 204SLE230 | Dry | 1 | 199 | lead | 0.074 | 0.021 | 196.0514337 | 0.000107115 | 7.639842913 | 0.002748748 | Dissolved |
| 206SPA020 | Dry | 1 | 462 | lead | 0.832 | 0.13 | 572.8214751 | 0.000226947 | 22.32203052 | 0.005823843 | Dissolved |
| 206SPA070 | Dry | 1 | 449 | lead | 0.949 | 0.088 | 552.3821658 | 0.00015931 | 21.52553998 | 0.004088167 | Dissolved |
| 206SPA200 | Dry | 1 | 398 | lead | 0.504 | 0.021 | 473.7848912 | 4.43239E-05 | 18.46271703 | 0.001137427 | Dissolved |
| 206SPA220 | Dry | 1 | 318 | lead | 0.361 | 0.11 | 356.0572065 | 0.000308939 | 13.87503817 | 0.007927906 | Dissolved |
| 206WIL020 | Dry | 1 | 323 | lead | 3.61 | 0.086 | 363.1991867 | 0.000236785 | 14.15335088 | 0.0060763 | Dissolved |
| 207SUI010 | Dry | 1 | 175 | lead | 0.239 | 0.237 | 166.4629138 | 0.001423741 | 6.48682077 | 0.036535617 | Dissolved |
| 207SUI020 | Dry | 1 | 202 | lead | 0.053 | 0.003 | 199.8215621 | 1.50134E-05 | 7.786759403 | 0.000385269 | Dissolved |
| 207SUI060 | Dry | 1 | 188 | lead | 0.062 | 0.023 | 182.361434 | 0.000126123 | 7.106363276 | 0.003236536 | Dissolved |
| 207SUI110 | Dry | 1 | 232 | lead | 0.044 | 0.006 | 238.3396302 | 2.51742E-05 | 9.287753219 | 0.000646012 | Dissolved |
| 201LAG040 | Spring | 1 | 91.9 | lead | 0.03 | ND | 73.32139217 | | 2.85722939 | | Dissolved |
| 201LAG130 | Spring | 1 | 91.9 | lead | 0.09 | ND | 73.32139217 | | 2.85722939 | | Dissolved |
| 201LAG270 | Spring | 1 | 194 | lead | 0.02 | ND | 189.8023821 | | 7.396326341 | | Dissolved |
| 204ALP010 | Spring | 1 | 408 | lead | 0.22 | 0.05 | 488.990525 | 0.000102251 | 19.05525874 | 0.002623948 | Dissolved |
| 204ALP100 | Spring | 1 | 427 | lead | 0.33 | 0.05 | 518.1610228 | 9.64951E-05 | 20.19199117 | 0.002476229 | Dissolved |
| 204ALP110 | Spring | 1 | 412 | lead | 0.05 | 0.02 | 495.1014662 | 4.03958E-05 | 19.29339335 | 0.001036624 | Dissolved |
| 204SLE030 | Spring | 1 | 258 | lead | 0.13 | 0.08 | 272.8487334 | 0.000293203 | 10.63252343 | 0.007524084 | Dissolved |
| 204SLE230 | Spring | 1 | 217 | lead | 0.32 | ND | 218.8987581 | | 8.530170342 | | Dissolved |
| 206SPA020 | Spring | 1 | 513 | lead | 0.38 | 0.05 | 654.4997157 | 7.63942E-05 | 25.50491429 | 0.001960407 | Dissolved |

Appendix G-3: Concentrations of hardness-dependent metals in water

| Station | Season | Year | Hardness (mg/L) | Metal Name | Total Metal (ug/L) | Dissolved Metal (ug/L) | Acute WQO (ug/L) | Acute Exceedance Factor | Chronic WQO (ug/L) | Chronic Exceedance Factor | WQO Fraction |
|-----------|--------|------|-----------------|------------|--------------------|------------------------|------------------|-------------------------|--------------------|---------------------------|--------------|
| 206SPA070 | Spring | 1 | 584 | lead | 0.31 | 0.02 | 771.9223416 | 2.59093E-05 | 30.08070545 | 0.000664878 | Dissolved |
| 206SPA200 | Spring | 1 | 481 | lead | 0.05 | ND | 602.9769888 | | 23.49714759 | | Dissolved |
| 206SPA220 | Spring | 1 | 320 | lead | 0.15 | 0.04 | 358.9103471 | 0.000111448 | 13.98622096 | 0.002859958 | Dissolved |
| 206WIL020 | Spring | 1 | 287 | lead | ND | 0.05 | 312.4738734 | 0.000160013 | 12.17665825 | 0.004106217 | Dissolved |
| 207SUI010 | Spring | 1 | 192 | lead | 0.06 | 0.02 | 187.3149844 | 0.000106772 | 7.299396026 | 0.002739953 | Dissolved |
| 207SUI060 | Spring | 1 | 188 | lead | 0.05 | ND | 182.361434 | | 7.106363276 | | Dissolved |
| 207SUI110 | Spring | 1 | 180 | lead | 0.12 | 0.02 | 172.5408588 | 0.000115915 | 6.723669561 | 0.002974566 | Dissolved |
| | | | | | | | | | | | |
| 202BUT010 | Dry | 2 | 130 | lead | 0.059 | 0.003 | 114.0197131 | 2.63112E-05 | 4.443184528 | 0.000675191 | Dissolved |
| 202PES050 | Dry | 2 | 257 | lead | 0.11 | 0.003 | 271.5031811 | 1.10496E-05 | 10.58008919 | 0.000283551 | Dissolved |
| 202PES070 | Dry | 2 | 257 | lead | 0.035 | ND | 271.5031811 | | 10.58008919 | | Dissolved |
| 202SGR010 | Dry | 2 | 354 | lead | 0.043 | 0.006 | 408.1419255 | 1.47008E-05 | 15.90470489 | 0.000377247 | Dissolved |
| 202SGR080 | Dry | 2 | 323 | lead | 0.022 | 0.006 | 363.1991867 | 1.65199E-05 | 14.15335088 | 0.000423928 | Dissolved |
| 205PER010 | Dry | 2 | 707 | lead | 0.423 | 0.084 | 984.5572401 | 8.53175E-05 | 38.36678217 | 0.002189394 | Dissolved |
| 205PER070 | Dry | 2 | 424 | lead | 0.019 | 0.008 | 513.5311494 | 1.55784E-05 | 20.01157165 | 0.000399769 | Dissolved |
| 205STE020 | Dry | 2 | 495 | lead | 0.142 | 0.076 | 625.406621 | 0.000121521 | 24.37119816 | 0.003118435 | Dissolved |
| 205STE060 | Dry | 2 | 212 | lead | 0.066 | 0.007 | 212.4983721 | 3.29414E-05 | 8.280756489 | 0.000845333 | Dissolved |
| 202BUT010 | Spring | 2 | 113 | lead | 0.15 | ND | 95.38913597 | | 3.717177684 | | Dissolved |
| 202PES050 | Spring | 2 | 222 | lead | 0.03 | ND | 225.3395345 | | 8.781158154 | | Dissolved |
| 202PES070 | Spring | 2 | 216 | lead | 0.02 | ND | 217.615428 | | 8.480160809 | | Dissolved |
| 202SGR010 | Spring | 2 | 289 | lead | 0.024 | ND | 315.2484865 | | 12.28478094 | | Dissolved |
| 202SGR080 | Spring | 2 | 168 | lead | 0.02 | ND | 158.0333593 | | 6.158333134 | | Dissolved |
| 205PER010 | Spring | 2 | 337 | lead | 0.55 | 0.25 | 383.3565785 | 0.000652134 | 14.9388555 | 0.016734883 | Dissolved |
| 205PER070 | Spring | 2 | 498 | lead | ND | 0.02 | 630.2357136 | 3.17342E-05 | 24.55938097 | 0.000814353 | Dissolved |
| 205STE020 | Spring | 2 | 530 | lead | 0.05 | ND | 682.2338102 | | 26.58567214 | | Dissolved |
| 205STE060 | Spring | 2 | 232 | lead | 0.1 | ND | 238.3396302 | | 9.287753219 | | Dissolved |
| 202BUT010 | Wet | 2 | 120 | lead | 0.124 | 0.01955 | 102.9740473 | 0.000189854 | 4.01275079 | 0.00487197 | Dissolved |
| 202PES050 | Wet | 2 | 221 | lead | 0.057 | 0.00247 | 224.0481802 | 1.10244E-05 | 8.73083593 | 0.000282905 | Dissolved |
| 202PES070 | Wet | 2 | 202 | lead | 0.063 | 0.00219 | 199.8215621 | 1.09598E-05 | 7.786759403 | 0.000281247 | Dissolved |
| 202SGR010 | Wet | 2 | 272 | lead | 0.056 | 0.00614 | 291.8342577 | 2.10393E-05 | 11.37236205 | 0.000539905 | Dissolved |
| 202SGR080 | Wet | 2 | 299 | lead | 0.056 | 0.00604 | 329.1997367 | 1.83475E-05 | 12.82844113 | 0.000470829 | Dissolved |
| 205PER010 | Wet | 2 | 784 | lead | 1.31 | 0.141 | 1123.037877 | 0.000125552 | 43.7631738 | 0.003221887 | Dissolved |
| 205PER070 | Wet | 2 | 533 | lead | 0.108 | 0.00478 | 687.1535484 | 6.95623E-06 | 26.77738728 | 0.000178509 | Dissolved |
| 205STE020 | Wet | 2 | 345 | lead | 0.165 | 0.03967 | 394.9787804 | 0.000100436 | 15.39175602 | 0.002577354 | Dissolved |
| 205STE060 | Wet | 2 | 290 | lead | 0.205 | 0.01514 | 316.6377623 | 4.78149E-05 | 12.33891902 | 0.001227012 | Dissolved |
| | | | | | | | | | | | |
| 201LAG130 | Dry | 1 | 81.6 | nickel | 1.03 | 0.648 | 395.0244081 | 0.001640405 | 43.91902316 | 0.014754427 | Dissolved |
| 201LAG270 | Dry | 1 | 221 | nickel | 3.92 | 2.98 | 917.6771547 | 0.003247329 | 102.0278327 | 0.029207716 | Dissolved |
| 204ALP010 | Dry | 1 | 304 | nickel | 1.92 | 1.19 | 1201.835722 | 0.000990152 | 133.6207329 | 0.008905804 | Dissolved |
| 204ALP100 | Dry | 1 | 207 | nickel | 1.63 | 0.948 | 868.2503538 | 0.001091851 | 96.53253474 | 0.009820523 | Dissolved |
| 204ALP110 | Dry | 1 | 422 | nickel | 0.723 | 0.174 | 1586.165101 | 0.000109699 | 176.350677 | 0.00098667 | Dissolved |
| 204SLE030 | Dry | 1 | 245 | nickel | 2.48 | 2.42 | 1001.310024 | 0.002416834 | 111.3261795 | 0.021737924 | Dissolved |
| 204SLE230 | Dry | 1 | 199 | nickel | 0.135 | 1.55 | 839.7765812 | 0.001845729 | 93.36680561 | 0.016601189 | Dissolved |
| 206SPA020 | Dry | 1 | 462 | nickel | 8.91 | 8.42 | 1712.462887 | 0.004916895 | 190.3925318 | 0.044224424 | Dissolved |
| 206SPA070 | Dry | 1 | 449 | nickel | 6.81 | 6.04 | 1671.608078 | 0.003613287 | 185.8502725 | 0.03249928 | Dissolved |
| 206SPA200 | Dry | 1 | 398 | nickel | 2.41 | ND | 1509.507015 | | 167.8277904 | | Dissolved |
| 206SPA220 | Dry | 1 | 318 | nickel | 0.985 | 0.616 | 1248.496704 | 0.000493393 | 138.8085256 | 0.004437768 | Dissolved |
| 206WIL020 | Dry | 1 | 323 | nickel | 16.6 | 8.12 | 1265.08408 | 0.006418546 | 140.6527189 | 0.057730843 | Dissolved |
| 207SUI010 | Dry | 1 | 175 | nickel | 1.03 | 0.88 | 753.2588552 | 0.001168257 | 83.74771895 | 0.010507749 | Dissolved |

Appendix G-3: Concentrations of hardness-dependent metals in water

| Station | Season | Year | Hardness (mg/L) | Metal Name | Total Metal (ug/L) | Dissolved Metal (ug/L) | Acute WQO (ug/L) | Acute Exceedance Factor | Chronic WQO (ug/L) | Chronic Exceedance Factor | WQO Fraction |
|-----------|--------|------|-----------------|------------|--------------------|------------------------|------------------|-------------------------|--------------------|---------------------------|--------------|
| 207SUI020 | Dry | 1 | 202 | nickel | 1.04 | 0.89 | 850.4745366 | 0.001046475 | 94.55620995 | 0.009412391 | Dissolved |
| 207SUI060 | Dry | 1 | 188 | nickel | ND | ND | 800.3346083 | | 88.98162614 | | Dissolved |
| 207SUI110 | Dry | 1 | 232 | nickel | 0.51 | 0.43 | 956.1739154 | 0.000449709 | 106.3079229 | 0.004044854 | Dissolved |
| 201LAG040 | Spring | 1 | 91.9 | nickel | 2.11 | 4.39 | 436.8164443 | 0.010049988 | 48.5654839 | 0.090393416 | Dissolved |
| 201LAG130 | Spring | 1 | 91.9 | nickel | 2.82 | 1.69 | 436.8164443 | 0.003868902 | 48.5654839 | 0.034798377 | Dissolved |
| 201LAG270 | Spring | 1 | 194 | nickel | 4.71 | 1.97 | 821.8911806 | 0.002396911 | 91.37829728 | 0.02155873 | Dissolved |
| 204ALP010 | Spring | 1 | 408 | nickel | 2.09 | 1.54 | 1541.532036 | 0.000999006 | 171.3883492 | 0.008985442 | Dissolved |
| 204ALP100 | Spring | 1 | 427 | nickel | 2.5 | 1.87 | 1602.049899 | 0.001167255 | 178.1167573 | 0.010498731 | Dissolved |
| 204ALP110 | Spring | 1 | 412 | nickel | 1.46 | 1.18 | 1554.308068 | 0.00075918 | 172.8087953 | 0.006828356 | Dissolved |
| 204SLE030 | Spring | 1 | 258 | nickel | 2.48 | 2.51 | 1046.078625 | 0.002399437 | 116.3035762 | 0.021581452 | Dissolved |
| 204SLE230 | Spring | 1 | 217 | nickel | 2.5 | 0.63 | 903.6057618 | 0.000697207 | 100.4633678 | 0.006270942 | Dissolved |
| 206SPA020 | Spring | 1 | 513 | nickel | 5.04 | 4.19 | 1871.084269 | 0.002239343 | 208.0281412 | 0.020141506 | Dissolved |
| 206SPA070 | Spring | 1 | 584 | nickel | 4.15 | 3.65 | 2087.946302 | 0.001748129 | 232.1389771 | 0.01572334 | Dissolved |
| 206SPA200 | Spring | 1 | 481 | nickel | 4.05 | 3.63 | 1771.8575 | 0.002048697 | 196.9960564 | 0.018426765 | Dissolved |
| 206SPA220 | Spring | 1 | 320 | nickel | 1.47 | 1.37 | 1255.13644 | 0.001091515 | 139.5467349 | 0.009817499 | Dissolved |
| 206WIL020 | Spring | 1 | 287 | nickel | ND | 3.65 | 1144.727567 | 0.003188532 | 127.2714179 | 0.028678866 | Dissolved |
| 207SUI010 | Spring | 1 | 192 | nickel | 0.71 | 0.47 | 814.7172208 | 0.000576887 | 90.58069263 | 0.005188744 | Dissolved |
| 207SUI060 | Spring | 1 | 188 | nickel | 0.52 | 0.48 | 800.3346083 | 0.000599749 | 88.98162614 | 0.005394372 | Dissolved |
| 207SUI110 | Spring | 1 | 180 | nickel | 1.38 | 0.94 | 771.4265752 | 0.001218522 | 85.7676157 | 0.010959848 | Dissolved |
| | | | | | | | | | | | |
| 202BUT010 | Dry | 2 | 130 | nickel | 3.05 | 2.78 | 585.7740559 | 0.004745857 | 65.12667015 | 0.042686045 | Dissolved |
| 202PES050 | Dry | 2 | 257 | nickel | 2.92 | 2.52 | 1042.647435 | 0.002416924 | 115.9220947 | 0.021738738 | Dissolved |
| 202PES070 | Dry | 2 | 257 | nickel | 2.42 | 2.09 | 1042.647435 | 0.002004513 | 115.9220947 | 0.01802935 | Dissolved |
| 202SGR010 | Dry | 2 | 354 | nickel | 2.43 | 2.22 | 1367.0702 | 0.001623911 | 151.9915897 | 0.014606071 | Dissolved |
| 202SGR080 | Dry | 2 | 323 | nickel | 2.31 | 2.28 | 1265.08408 | 0.001802252 | 140.6527189 | 0.016210138 | Dissolved |
| 205PER010 | Dry | 2 | 707 | nickel | 3.34 | 2.53 | 2454.386089 | 0.001030808 | 272.8799469 | 0.009271476 | Dissolved |
| 205PER070 | Dry | 2 | 424 | nickel | 2.17 | 1.6 | 1592.522479 | 0.001004695 | 177.0574937 | 0.009036613 | Dissolved |
| 205STE020 | Dry | 2 | 495 | nickel | 0.83 | 0.63 | 1815.390486 | 0.000347033 | 201.8360768 | 0.003121345 | Dissolved |
| 205STE060 | Dry | 2 | 212 | nickel | 1.32 | 1.29 | 885.9601644 | 0.001456047 | 98.50152087 | 0.013096244 | Dissolved |
| 202BUT010 | Spring | 2 | 113 | nickel | 3.09 | 2.63 | 520.2815434 | 0.005054955 | 57.84517788 | 0.045466193 | Dissolved |
| 202PES050 | Spring | 2 | 222 | nickel | 2.32 | 1.71 | 921.1888509 | 0.001856297 | 102.4182649 | 0.016696241 | Dissolved |
| 202PES070 | Spring | 2 | 216 | nickel | 1.67 | 1.53 | 900.0816963 | 0.001699846 | 100.0715603 | 0.015289059 | Dissolved |
| 202SGR010 | Spring | 2 | 289 | nickel | 2.2 | 2.14 | 1151.472662 | 0.00185849 | 128.0213412 | 0.016715963 | Dissolved |
| 202SGR080 | Spring | 2 | 168 | nickel | 2.29 | 2.23 | 727.6888308 | 0.003064497 | 80.90483008 | 0.027563249 | Dissolved |
| 205PER010 | Spring | 2 | 337 | nickel | 1.68 | 1.56 | 1311.320809 | 0.00118964 | 145.7933429 | 0.010700077 | Dissolved |
| 205PER070 | Spring | 2 | 498 | nickel | 8.71 | 7.86 | 1824.694154 | 0.004307571 | 202.8704636 | 0.038743935 | Dissolved |
| 205STE020 | Spring | 2 | 530 | nickel | 0.96 | 0.98 | 1923.408108 | 0.000509512 | 213.8455333 | 0.004582747 | Dissolved |
| 205STE060 | Spring | 2 | 232 | nickel | 1.06 | 0.51 | 956.1739154 | 0.000533376 | 106.3079229 | 0.004797385 | Dissolved |
| 202BUT010 | Wet | 2 | 120 | nickel | 3.42 | 3.06 | 547.4209208 | 0.005589848 | 60.86254826 | 0.050277224 | Dissolved |
| 202PES050 | Wet | 2 | 221 | nickel | 3.01 | 2.73 | 917.6771547 | 0.002974902 | 102.0278327 | 0.026757405 | Dissolved |
| 202PES070 | Wet | 2 | 202 | nickel | 3.09 | 2.9 | 850.4745366 | 0.003409861 | 94.55620995 | 0.030669588 | Dissolved |
| 202SGR010 | Wet | 2 | 272 | nickel | 3.96 | 3.48 | 1093.90435 | 0.003181265 | 121.6208657 | 0.028613511 | Dissolved |
| 202SGR080 | Wet | 2 | 299 | nickel | 3.96 | 3.72 | 1185.091497 | 0.003138998 | 131.759101 | 0.028233344 | Dissolved |
| 205PER010 | Wet | 2 | 784 | nickel | 5.03 | 4.08 | 2678.708438 | 0.001523122 | 297.8202246 | 0.01369954 | Dissolved |
| 205PER070 | Wet | 2 | 533 | nickel | 33.7 | 30.9 | 1932.614686 | 0.015988702 | 214.8691256 | 0.143808469 | Dissolved |
| 205STE020 | Wet | 2 | 345 | nickel | 3.55 | 2.49 | 1337.608473 | 0.001861531 | 148.7160193 | 0.016743321 | Dissolved |
| 205STE060 | Wet | 2 | 290 | nickel | 4.88 | 3.25 | 1154.842512 | 0.002814237 | 128.3960029 | 0.025312314 | Dissolved |

Appendix G-3: Concentrations of hardness-dependent metals in water

| Station | Season | Year | Hardness (mg/L) | Metal Name | Total Metal (ug/L) | Dissolved Metal (ug/L) | Acute WQO (ug/L) | Acute Exceedance Factor | Chronic WQO (ug/L) | Chronic Exceedance Factor | WQO Fraction |
|-----------|--------|------|-----------------|------------|--------------------|------------------------|------------------|-------------------------|--------------------|---------------------------|--------------|
| 201LAG130 | Dry | 1 | 81.6 | silver | ND | ND | 2.860928809 | | | | Dissolved |
| 201LAG270 | Dry | 1 | 221 | silver | ND | ND | 15.87649711 | | | | Dissolved |
| 204ALP010 | Dry | 1 | 304 | silver | ND | ND | 27.47530624 | | | | Dissolved |
| 204ALP100 | Dry | 1 | 207 | silver | ND | ND | 14.18629453 | | | | Dissolved |
| 204ALP110 | Dry | 1 | 422 | silver | ND | ND | 48.2989171 | | | | Dissolved |
| 204SLE030 | Dry | 1 | 245 | silver | ND | ND | 18.95682684 | | | | Dissolved |
| 204SLE230 | Dry | 1 | 199 | silver | 0.013 | ND | 13.25645096 | | | | Dissolved |
| 206SPA020 | Dry | 1 | 462 | silver | 0.011 | ND | 56.43963233 | | | | Dissolved |
| 206SPA070 | Dry | 1 | 449 | silver | 0.012 | ND | 53.73579575 | | | | Dissolved |
| 206SPA200 | Dry | 1 | 398 | silver | ND | ND | 43.67157571 | | | | Dissolved |
| 206SPA220 | Dry | 1 | 318 | silver | 0.03 | ND | 29.68756818 | | | | Dissolved |
| 206WIL020 | Dry | 1 | 323 | silver | 0.009 | ND | 30.49497744 | | | | Dissolved |
| 207SUI010 | Dry | 1 | 175 | silver | 0.02 | ND | 10.62736072 | | | | Dissolved |
| 207SUI020 | Dry | 1 | 202 | silver | ND | ND | 13.60204895 | | | | Dissolved |
| 207SUI060 | Dry | 1 | 188 | silver | ND | ND | 12.0213012 | | | | Dissolved |
| 207SUI110 | Dry | 1 | 232 | silver | ND | ND | 17.25994121 | | | | Dissolved |
| 201LAG040 | Spring | 1 | 91.9 | silver | ND | ND | 3.509963963 | | | | Dissolved |
| 201LAG130 | Spring | 1 | 91.9 | silver | ND | ND | 3.509963963 | | | | Dissolved |
| 201LAG270 | Spring | 1 | 194 | silver | ND | ND | 12.68875329 | | | | Dissolved |
| 204ALP010 | Spring | 1 | 408 | silver | ND | 0.01 | 45.57592151 | 0.000219414 | | | Dissolved |
| 204ALP100 | Spring | 1 | 427 | silver | ND | ND | 49.2874016 | | | | Dissolved |
| 204ALP110 | Spring | 1 | 412 | silver | ND | 0.08 | 46.34716666 | 0.001726103 | | | Dissolved |
| 204SLE030 | Spring | 1 | 258 | silver | ND | ND | 20.71981494 | | | | Dissolved |
| 204SLE230 | Spring | 1 | 217 | silver | ND | ND | 15.38546832 | | | | Dissolved |
| 206SPA020 | Spring | 1 | 513 | silver | ND | ND | 67.57745718 | | | | Dissolved |
| 206SPA070 | Spring | 1 | 584 | silver | ND | ND | 84.45591569 | | | | Dissolved |
| 206SPA200 | Spring | 1 | 481 | silver | ND | ND | 60.49082524 | | | | Dissolved |
| 206SPA220 | Spring | 1 | 320 | silver | ND | ND | 30.00944342 | | | | Dissolved |
| 206WIL020 | Spring | 1 | 287 | silver | ND | ND | 24.88609991 | | | | Dissolved |
| 207SUI010 | Spring | 1 | 192 | silver | ND | ND | 12.46459269 | | | | Dissolved |
| 207SUI060 | Spring | 1 | 188 | silver | ND | 0.01 | 12.0213012 | 0.000831857 | | | Dissolved |
| 207SUI110 | Spring | 1 | 180 | silver | ND | ND | 11.1549772 | | | | Dissolved |
| | | | | | | | | | | | |
| 202BUT010 | Dry | 2 | 130 | silver | 0.011 | ND | 6.37356683 | | | | Dissolved |
| 202PES050 | Dry | 2 | 257 | silver | ND | ND | 20.58187566 | | | | Dissolved |
| 202PES070 | Dry | 2 | 257 | silver | ND | ND | 20.58187566 | | | | Dissolved |
| 202SGR010 | Dry | 2 | 354 | silver | ND | 0.015 | 35.70142809 | 0.000420151 | | | Dissolved |
| 202SGR080 | Dry | 2 | 323 | silver | ND | ND | 30.49497744 | | | | Dissolved |
| 205PER010 | Dry | 2 | 707 | silver | ND | ND | 117.3279195 | | | | Dissolved |
| 205PER070 | Dry | 2 | 424 | silver | ND | ND | 48.69330484 | | | | Dissolved |
| 205STE020 | Dry | 2 | 495 | silver | ND | ND | 63.5507851 | | | | Dissolved |
| 205STE060 | Dry | 2 | 212 | silver | ND | ND | 14.78079041 | | | | Dissolved |
| 202BUT010 | Spring | 2 | 113 | silver | ND | ND | 5.008352648 | | | | Dissolved |
| 202PES050 | Spring | 2 | 222 | silver | ND | ND | 16.000262 | | | | Dissolved |
| 202PES070 | Spring | 2 | 216 | silver | ND | ND | 15.26372138 | | | | Dissolved |
| 202SGR010 | Spring | 2 | 289 | silver | ND | ND | 25.18513409 | | | | Dissolved |
| 202SGR080 | Spring | 2 | 168 | silver | ND | ND | 9.906766857 | | | | Dissolved |
| 205PER010 | Spring | 2 | 337 | silver | ND | ND | 32.80374351 | | | | Dissolved |
| 205PER070 | Spring | 2 | 498 | silver | ND | ND | 64.21469846 | | | | Dissolved |
| 205STE020 | Spring | 2 | 530 | silver | ND | ND | 71.47505097 | | | | Dissolved |

Appendix G-3: Concentrations of hardness-dependent metals in water

| Station | Season | Year | Hardness (mg/L) | Metal Name | Total Metal (ug/L) | Dissolved Metal (ug/L) | Acute WQO (ug/L) | Acute Exceedance Factor | Chronic WQO (ug/L) | Chronic Exceedance Factor | WQO Fraction |
|-----------|--------|------|-----------------|------------|--------------------|------------------------|------------------|-------------------------|--------------------|---------------------------|--------------|
| 205STE060 | Spring | 2 | 232 | silver | ND | ND | 17.25994121 | | | | Dissolved |
| 202BUT010 | Wet | 2 | 120 | silver | ND | ND | 5.553818991 | | | | Dissolved |
| 202PES050 | Wet | 2 | 221 | silver | ND | ND | 15.87649711 | | | | Dissolved |
| 202PES070 | Wet | 2 | 202 | silver | ND | ND | 13.60204895 | | | | Dissolved |
| 202SGR010 | Wet | 2 | 272 | silver | ND | ND | 22.69125378 | | | | Dissolved |
| 202SGR080 | Wet | 2 | 299 | silver | ND | ND | 26.7026536 | | | | Dissolved |
| 205PER010 | Wet | 2 | 784 | silver | ND | ND | 140.1598441 | | | | Dissolved |
| 205PER070 | Wet | 2 | 533 | silver | ND | ND | 72.17233852 | | | | Dissolved |
| 205STE020 | Wet | 2 | 345 | silver | ND | ND | 34.15456996 | | | | Dissolved |
| 205STE060 | Wet | 2 | 290 | silver | ND | ND | 25.33521151 | | | | Dissolved |
| | | | | | | | | | | | |
| 201LAG130 | Dry | 1 | 81.6 | zinc | 0.37 | 0.097 | 100.8535973 | 0.00096179 | 100.8535973 | 0.00096179 | Dissolved |
| 201LAG270 | Dry | 1 | 221 | zinc | 0.694 | 0.511 | 234.5956179 | 0.002178216 | 234.5956179 | 0.002178216 | Dissolved |
| 204ALP010 | Dry | 1 | 304 | zinc | 2.57 | 1.28 | 307.3654933 | 0.004164423 | 307.3654933 | 0.004164423 | Dissolved |
| 204ALP100 | Dry | 1 | 207 | zinc | 3.59 | 1.02 | 221.9412336 | 0.004595811 | 221.9412336 | 0.004595811 | Dissolved |
| 204ALP110 | Dry | 1 | 422 | zinc | 3 | 2.18 | 405.8294521 | 0.005371715 | 405.8294521 | 0.005371715 | Dissolved |
| 204SLE030 | Dry | 1 | 245 | zinc | 3.99 | 2.29 | 256.0098913 | 0.008944967 | 256.0098913 | 0.008944967 | Dissolved |
| 204SLE230 | Dry | 1 | 199 | zinc | 0.728 | 0.49 | 214.6517995 | 0.002282767 | 214.6517995 | 0.002282767 | Dissolved |
| 206SPA020 | Dry | 1 | 462 | zinc | 9.96 | 7.52 | 438.1950506 | 0.017161307 | 438.1950506 | 0.017161307 | Dissolved |
| 206SPA070 | Dry | 1 | 449 | zinc | 9.86 | 5.74 | 427.7250116 | 0.013419837 | 427.7250116 | 0.013419837 | Dissolved |
| 206SPA200 | Dry | 1 | 398 | zinc | 3.42 | 1.26 | 386.1866431 | 0.003262671 | 386.1866431 | 0.003262671 | Dissolved |
| 206SPA220 | Dry | 1 | 318 | zinc | 2.68 | 1.28 | 319.3175738 | 0.004008549 | 319.3175738 | 0.004008549 | Dissolved |
| 206WIL020 | Dry | 1 | 323 | zinc | 18.2 | 0.67 | 323.5665506 | 0.002070671 | 323.5665506 | 0.002070671 | Dissolved |
| 207SUI010 | Dry | 1 | 175 | zinc | 1.27 | 1.08 | 192.505198 | 0.005610238 | 192.505198 | 0.005610238 | Dissolved |
| 207SUI020 | Dry | 1 | 202 | zinc | 0.57 | 0.26 | 217.3904879 | 0.001196004 | 217.3904879 | 0.001196004 | Dissolved |
| 207SUI060 | Dry | 1 | 188 | zinc | 0.854 | 2.3 | 204.5550792 | 0.011243915 | 204.5550792 | 0.011243915 | Dissolved |
| 207SUI110 | Dry | 1 | 232 | zinc | 1.01 | 0.6 | 244.4523922 | 0.002454466 | 244.4523922 | 0.002454466 | Dissolved |
| 201LAG040 | Spring | 1 | 91.9 | zinc | 0.39 | 0.3 | 111.5407486 | 0.0026896 | 111.5407486 | 0.0026896 | Dissolved |
| 201LAG130 | Spring | 1 | 91.9 | zinc | 0.59 | 1.58 | 111.5407486 | 0.014165227 | 111.5407486 | 0.014165227 | Dissolved |
| 201LAG270 | Spring | 1 | 194 | zinc | 0.66 | 0.87 | 210.0732371 | 0.004141413 | 210.0732371 | 0.004141413 | Dissolved |
| 204ALP010 | Spring | 1 | 408 | zinc | 2.89 | 1.86 | 394.3925278 | 0.004716114 | 394.3925278 | 0.004716114 | Dissolved |
| 204ALP100 | Spring | 1 | 427 | zinc | 4.07 | 0.16 | 409.8999454 | 0.000390339 | 409.8999454 | 0.000390339 | Dissolved |
| 204ALP110 | Spring | 1 | 412 | zinc | 2.32 | 1.99 | 397.6662493 | 0.005004196 | 397.6662493 | 0.005004196 | Dissolved |
| 204SLE030 | Spring | 1 | 258 | zinc | 1.74 | 1.23 | 267.4740778 | 0.004598576 | 267.4740778 | 0.004598576 | Dissolved |
| 204SLE230 | Spring | 1 | 217 | zinc | 1.47 | ND | 230.9929124 | | 230.9929124 | | Dissolved |
| 206SPA020 | Spring | 1 | 513 | zinc | 3.38 | 1.76 | 478.8491992 | 0.003675479 | 478.8491992 | 0.003675479 | Dissolved |
| 206SPA070 | Spring | 1 | 584 | zinc | 3.31 | 1.9 | 534.438735 | 0.003555132 | 534.438735 | 0.003555132 | Dissolved |
| 206SPA200 | Spring | 1 | 481 | zinc | 1.87 | 1.31 | 453.4170481 | 0.002889172 | 453.4170481 | 0.002889172 | Dissolved |
| 206SPA220 | Spring | 1 | 320 | zinc | 3.19 | 1.94 | 321.0183799 | 0.006043268 | 321.0183799 | 0.006043268 | Dissolved |
| 206WIL020 | Spring | 1 | 287 | zinc | ND | ND | 292.7383719 | | 292.7383719 | | Dissolved |
| 207SUI010 | Spring | 1 | 192 | zinc | 0.56 | 0.29 | 208.2367865 | 0.001392645 | 208.2367865 | 0.001392645 | Dissolved |
| 207SUI060 | Spring | 1 | 188 | zinc | 0.54 | 0.32 | 204.5550792 | 0.001564371 | 204.5550792 | 0.001564371 | Dissolved |
| 207SUI110 | Spring | 1 | 180 | zinc | 0.93 | 0.29 | 197.1554176 | 0.001470921 | 197.1554176 | 0.001470921 | Dissolved |
| | | | | | | | | | | | |
| 202BUT010 | Dry | 2 | 130 | zinc | 1.51 | 0.61 | 149.644419 | 0.00407633 | 149.644419 | 0.00407633 | Dissolved |
| 202PES050 | Dry | 2 | 257 | zinc | 1.49 | 0.57 | 266.5954036 | 0.002138071 | 266.5954036 | 0.002138071 | Dissolved |
| 202PES070 | Dry | 2 | 257 | zinc | 0.81 | 0.48 | 266.5954036 | 0.001800481 | 266.5954036 | 0.001800481 | Dissolved |
| 202SGR010 | Dry | 2 | 354 | zinc | 0.88 | 0.56 | 349.6928768 | 0.001601405 | 349.6928768 | 0.001601405 | Dissolved |
| 202SGR080 | Dry | 2 | 323 | zinc | 0.87 | 0.76 | 323.5665506 | 0.002348821 | 323.5665506 | 0.002348821 | Dissolved |

Appendix G-3: Concentrations of hardness-dependent metals in water

| Station | Season | Year | Hardness (mg/L) | Metal Name | Total Metal (ug/L) | Dissolved Metal (ug/L) | Acute WQO (ug/L) | Acute Exceedance Factor | Chronic WQO (ug/L) | Chronic Exceedance Factor | WQO Fraction |
|-----------|--------|------|-----------------|------------|--------------------|------------------------|------------------|-------------------------|--------------------|---------------------------|--------------|
| 205PER010 | Dry | 2 | 707 | zinc | 5.68 | 3.86 | 628.3901825 | 0.00614268 | 628.3901825 | 0.00614268 | Dissolved |
| 205PER070 | Dry | 2 | 424 | zinc | 1.88 | 1.25 | 407.4585284 | 0.003067797 | 407.4585284 | 0.003067797 | Dissolved |
| 205STE020 | Dry | 2 | 495 | zinc | 4.57 | 4.08 | 464.5744346 | 0.008782231 | 464.5744346 | 0.008782231 | Dissolved |
| 205STE060 | Dry | 2 | 212 | zinc | 1.09 | 0.44 | 226.4752235 | 0.001942817 | 226.4752235 | 0.001942817 | Dissolved |
| 202BUT010 | Spring | 2 | 113 | zinc | 1.86 | 0.25 | 132.8892001 | 0.001881266 | 132.8892001 | 0.001881266 | Dissolved |
| 202PES050 | Spring | 2 | 222 | zinc | 0.49 | 0.22 | 235.4947324 | 0.000934203 | 235.4947324 | 0.000934203 | Dissolved |
| 202PES070 | Spring | 2 | 216 | zinc | 0.55 | 0.32 | 230.0906576 | 0.001390756 | 230.0906576 | 0.001390756 | Dissolved |
| 202SGR010 | Spring | 2 | 289 | zinc | 0.7 | 0.4 | 294.4659371 | 0.001358391 | 294.4659371 | 0.001358391 | Dissolved |
| 202SGR080 | Spring | 2 | 168 | zinc | 0.82 | 0.56 | 185.9605733 | 0.003011391 | 185.9605733 | 0.003011391 | Dissolved |
| 205PER010 | Spring | 2 | 337 | zinc | 3.57 | 2.19 | 335.410873 | 0.006529305 | 335.410873 | 0.006529305 | Dissolved |
| 205PER070 | Spring | 2 | 498 | zinc | 1.42 | 1.11 | 466.9589934 | 0.002377082 | 466.9589934 | 0.002377082 | Dissolved |
| 205STE020 | Spring | 2 | 530 | zinc | 1.89 | 1.7 | 492.2608149 | 0.003453454 | 492.2608149 | 0.003453454 | Dissolved |
| 205STE060 | Spring | 2 | 232 | zinc | 1.34 | 0.31 | 244.4523922 | 0.001268141 | 244.4523922 | 0.001268141 | Dissolved |
| 202BUT010 | Wet | 2 | 120 | zinc | 2.03 | 1.06 | 139.8320072 | 0.007580525 | 139.8320072 | 0.007580525 | Dissolved |
| 202PES050 | Wet | 2 | 221 | zinc | 1.59 | 1.1 | 234.5956179 | 0.00468892 | 234.5956179 | 0.00468892 | Dissolved |
| 202PES070 | Wet | 2 | 202 | zinc | 1.5 | 0.949 | 217.3904879 | 0.004365416 | 217.3904879 | 0.004365416 | Dissolved |
| 202SGR010 | Wet | 2 | 272 | zinc | 1.51 | 0.947 | 279.7219547 | 0.003385505 | 279.7219547 | 0.003385505 | Dissolved |
| 202SGR080 | Wet | 2 | 299 | zinc | 1.98 | 1.52 | 303.0766791 | 0.005015232 | 303.0766791 | 0.005015232 | Dissolved |
| 205PER010 | Wet | 2 | 784 | zinc | 9.62 | 3.17 | 685.915034 | 0.004621564 | 685.915034 | 0.004621564 | Dissolved |
| 205PER070 | Wet | 2 | 533 | zinc | 5.27 | 2.64 | 494.6206983 | 0.005337423 | 494.6206983 | 0.005337423 | Dissolved |
| 205STE020 | Wet | 2 | 345 | zinc | 3.83 | 2.73 | 342.1451919 | 0.007979069 | 342.1451919 | 0.007979069 | Dissolved |
| 205STE060 | Wet | 2 | 290 | zinc | 2.46 | 0.986 | 295.3290345 | 0.003338649 | 295.3290345 | 0.003338649 | Dissolved |

Appendix G-4: Concentrations of non-hardness-dependent metals in water compared to acute and chronic water quality objectives (WQO's)

| Station | Season | Year | Metal Name | Metal, Total (ug/L) | Metal, Dissolved (ug/L) | Acute WQO (ug/L) | Acute Exceedance Factor | Chronic WQO (ug/L) | Chronic Exceedance Factor | WQO Fraction |
|-----------|--------|------|------------|---------------------|-------------------------|------------------|-------------------------|--------------------|---------------------------|--------------|
| 201LAG130 | Dry | 1 | arsenic | 0.439 | 0.327 | 340 | 0.000961765 | 150 | 0.00218 | Dissolved |
| 201LAG270 | Dry | 1 | arsenic | 0.854 | 0.737 | 340 | 0.002167647 | 150 | 0.004913333 | Dissolved |
| 204ALP010 | Dry | 1 | arsenic | 3.15 | 3.06 | 340 | 0.009 | 150 | 0.0204 | Dissolved |
| 204ALP100 | Dry | 1 | arsenic | 3.26 | 3 | 340 | 0.008823529 | 150 | 0.02 | Dissolved |
| 204ALP110 | Dry | 1 | arsenic | 2.2 | 1.96 | 340 | 0.005764706 | 150 | 0.013066667 | Dissolved |
| 204SLE030 | Dry | 1 | arsenic | 0.95 | 5.93 | 340 | 0.017441176 | 150 | 0.039533333 | Dissolved |
| 204SLE230 | Dry | 1 | arsenic | 1.45 | 1 | 340 | 0.002941176 | 150 | 0.006666667 | Dissolved |
| 206SPA020 | Dry | 1 | arsenic | 6.42 | 5.59 | 340 | 0.016441176 | 150 | 0.037266667 | Dissolved |
| 206SPA070 | Dry | 1 | arsenic | 7.5 | 6.6 | 340 | 0.019411765 | 150 | 0.044 | Dissolved |
| 206SPA200 | Dry | 1 | arsenic | 6.21 | 1.4 | 340 | 0.004117647 | 150 | 0.009333333 | Dissolved |
| 206SPA220 | Dry | 1 | arsenic | 3.44 | 2.93 | 340 | 0.008617647 | 150 | 0.019533333 | Dissolved |
| 206WIL020 | Dry | 1 | arsenic | 6.18 | 4.57 | 340 | 0.013441176 | 150 | 0.030466667 | Dissolved |
| 207SUI010 | Dry | 1 | arsenic | 1.41 | 1.21 | 340 | 0.003558824 | 150 | 0.008066667 | Dissolved |
| 207SUI020 | Dry | 1 | arsenic | 1.3 | 1.23 | 340 | 0.003617647 | 150 | 0.0082 | Dissolved |
| 207SUI060 | Dry | 1 | arsenic | 0.892 | 0.776 | 340 | 0.002282353 | 150 | 0.005173333 | Dissolved |
| 207SUI110 | Dry | 1 | arsenic | 0.77 | 0.77 | 340 | 0.002264706 | 150 | 0.005133333 | Dissolved |
| 201LAG040 | Spring | 1 | arsenic | 1.01 | 0.59 | 340 | 0.001735294 | 150 | 0.003933333 | Dissolved |
| 201LAG130 | Spring | 1 | arsenic | 0.47 | 0.46 | 340 | 0.001352941 | 150 | 0.003066667 | Dissolved |
| 201LAG270 | Spring | 1 | arsenic | 0.62 | 0.96 | 340 | 0.002823529 | 150 | 0.0064 | Dissolved |
| 204ALP010 | Spring | 1 | arsenic | 3.65 | 3.61 | 340 | 0.010617647 | 150 | 0.024066667 | Dissolved |
| 204ALP100 | Spring | 1 | arsenic | 4.69 | 4.49 | 340 | 0.013205882 | 150 | 0.029933333 | Dissolved |
| 204ALP110 | Spring | 1 | arsenic | 2.12 | 2.04 | 340 | 0.006 | 150 | 0.0136 | Dissolved |
| 204SLE030 | Spring | 1 | arsenic | 1.03 | 1.09 | 340 | 0.003205882 | 150 | 0.007266667 | Dissolved |
| 204SLE230 | Spring | 1 | arsenic | 1.05 | 0.98 | 340 | 0.002882353 | 150 | 0.006533333 | Dissolved |
| 206SPA020 | Spring | 1 | arsenic | 4.33 | 4.07 | 340 | 0.011970588 | 150 | 0.027133333 | Dissolved |
| 206SPA070 | Spring | 1 | arsenic | 4.7 | 4.37 | 340 | 0.012852941 | 150 | 0.029133333 | Dissolved |
| 206SPA200 | Spring | 1 | arsenic | 4 | 3.87 | 340 | 0.011382353 | 150 | 0.0258 | Dissolved |
| 206SPA220 | Spring | 1 | arsenic | 3.09 | 2.9 | 340 | 0.008529412 | 150 | 0.019333333 | Dissolved |
| 206WIL020 | Spring | 1 | arsenic | 0.12 | 2.4 | 340 | 0.007058824 | 150 | 0.016 | Dissolved |
| 207SUI010 | Spring | 1 | arsenic | 0.93 | 0.96 | 340 | 0.002823529 | 150 | 0.0064 | Dissolved |
| 207SUI060 | Spring | 1 | arsenic | 0.88 | 0.79 | 340 | 0.002323529 | 150 | 0.005266667 | Dissolved |
| 207SUI110 | Spring | 1 | arsenic | 0.96 | 1 | 340 | 0.002941176 | 150 | 0.006666667 | Dissolved |
| | | | | | | | | | | |
| 202BUT010 | Dry | 2 | arsenic | 0.71 | 0.49 | 340 | 0.001441176 | 150 | 0.003266667 | Dissolved |
| 202PES050 | Dry | 2 | arsenic | 1.38 | 1.2 | 340 | 0.003529412 | 150 | 0.008 | Dissolved |
| 202PES070 | Dry | 2 | arsenic | 1.32 | 1.2 | 340 | 0.003529412 | 150 | 0.008 | Dissolved |
| 202SGR010 | Dry | 2 | arsenic | 1.62 | 1.66 | 340 | 0.004882353 | 150 | 0.011066667 | Dissolved |
| 202SGR080 | Dry | 2 | arsenic | 1.15 | 1.36 | 340 | 0.004 | 150 | 0.009066667 | Dissolved |
| 205PER010 | Dry | 2 | arsenic | 0.88 | 0.89 | 340 | 0.002617647 | 150 | 0.005933333 | Dissolved |
| 205PER070 | Dry | 2 | arsenic | 0.92 | 0.86 | 340 | 0.002529412 | 150 | 0.005733333 | Dissolved |
| 205STE020 | Dry | 2 | arsenic | 0.85 | 1.09 | 340 | 0.003205882 | 150 | 0.007266667 | Dissolved |
| 205STE060 | Dry | 2 | arsenic | 1.15 | 0.9 | 340 | 0.002647059 | 150 | 0.006 | Dissolved |
| 202BUT010 | Spring | 2 | arsenic | 0.42 | 0.41 | 340 | 0.001205882 | 150 | 0.002733333 | Dissolved |
| 202PES050 | Spring | 2 | arsenic | 0.99 | 0.9 | 340 | 0.002647059 | 150 | 0.006 | Dissolved |
| 202PES070 | Spring | 2 | arsenic | 0.66 | 0.71 | 340 | 0.002088235 | 150 | 0.004733333 | Dissolved |
| 202SGR010 | Spring | 2 | arsenic | 1.26 | 1.01 | 340 | 0.002970588 | 150 | 0.006733333 | Dissolved |

Appendix G-4: Concentrations of non-hardness-dependent metals in water compared to acute and chronic water quality objectives (WQO's)

| Station | Season | Year | Metal Name | Metal, Total (ug/L) | Metal, Dissolved (ug/L) | Acute WQO (ug/L) | Acute Exceedance Factor | Chronic WQO (ug/L) | Chronic Exceedance Factor | WQO Fraction |
|-----------|--------|------|------------|---------------------|-------------------------|------------------|-------------------------|--------------------|---------------------------|--------------|
| 202SGR080 | Spring | 2 | arsenic | 1.03 | 0.73 | 340 | 0.002147059 | 150 | 0.004866667 | Dissolved |
| 205PER010 | Spring | 2 | arsenic | 1.01 | 1.03 | 340 | 0.003029412 | 150 | 0.006866667 | Dissolved |
| 205PER070 | Spring | 2 | arsenic | 1 | 1.04 | 340 | 0.003058824 | 150 | 0.006933333 | Dissolved |
| 205STE020 | Spring | 2 | arsenic | 0.97 | 1.02 | 340 | 0.003 | 150 | 0.0068 | Dissolved |
| 205STE060 | Spring | 2 | arsenic | 0.89 | 0.93 | 340 | 0.002735294 | 150 | 0.0062 | Dissolved |
| 202BUT010 | Wet | 2 | arsenic | 0.667 | 0.52 | 340 | 0.001529412 | 150 | 0.003466667 | Dissolved |
| 202PES050 | Wet | 2 | arsenic | 0.883 | 0.747 | 340 | 0.002197059 | 150 | 0.00498 | Dissolved |
| 202PES070 | Wet | 2 | arsenic | 0.797 | 0.717 | 340 | 0.002108824 | 150 | 0.00478 | Dissolved |
| 202SGR010 | Wet | 2 | arsenic | 1.13 | 1.05 | 340 | 0.003088235 | 150 | 0.007 | Dissolved |
| 202SGR080 | Wet | 2 | arsenic | 0.895 | 0.889 | 340 | 0.002614706 | 150 | 0.005926667 | Dissolved |
| 205PER010 | Wet | 2 | arsenic | 1.54 | 1.58 | 340 | 0.004647059 | 150 | 0.010533333 | Dissolved |
| 205PER070 | Wet | 2 | arsenic | 1.95 | 1.94 | 340 | 0.005705882 | 150 | 0.012933333 | Dissolved |
| 205STE020 | Wet | 2 | arsenic | 1.05 | 1.09 | 340 | 0.003205882 | 150 | 0.007266667 | Dissolved |
| 205STE060 | Wet | 2 | arsenic | 1.08 | 1.12 | 340 | 0.003294118 | 150 | 0.007466667 | Dissolved |
| 201LAG130 | Dry | 1 | chromium | 1.81 | 0.58 | 16 | 0.03625 | 11 | 0.052727273 | Dissolved |
| 201LAG270 | Dry | 1 | chromium | 4.7 | 1.35 | 16 | 0.084375 | 11 | 0.122727273 | Dissolved |
| 204ALP010 | Dry | 1 | chromium | 3.57 | 2.1 | 16 | 0.13125 | 11 | 0.190909091 | Dissolved |
| 204ALP100 | Dry | 1 | chromium | 2.58 | 1.04 | 16 | 0.065 | 11 | 0.094545455 | Dissolved |
| 204ALP110 | Dry | 1 | chromium | 11.5 | 8.96 | 16 | 0.56 | 11 | 0.814545455 | Dissolved |
| 204SLE030 | Dry | 1 | chromium | 15.8 | 14.5 | 16 | 0.90625 | 11 | 1.318181818 | Dissolved |
| 204SLE230 | Dry | 1 | chromium | 1.44 | 1.01 | 16 | 0.063125 | 11 | 0.091818182 | Dissolved |
| 206SPA020 | Dry | 1 | chromium | 4.59 | 1.66 | 16 | 0.10375 | 11 | 0.150909091 | Dissolved |
| 206SPA070 | Dry | 1 | chromium | 4.58 | 1.8 | 16 | 0.1125 | 11 | 0.163636364 | Dissolved |
| 206SPA200 | Dry | 1 | chromium | 2.77 | 0.383 | 16 | 0.0239375 | 11 | 0.034818182 | Dissolved |
| 206SPA220 | Dry | 1 | chromium | 2.69 | 0.851 | 16 | 0.0531875 | 11 | 0.077363636 | Dissolved |
| 206WIL020 | Dry | 1 | chromium | 5.83 | 0.15 | 16 | 0.009375 | 11 | 0.013636364 | Dissolved |
| 207SUI010 | Dry | 1 | chromium | 3.47 | 1.21 | 16 | 0.075625 | 11 | 0.11 | Dissolved |
| 207SUI020 | Dry | 1 | chromium | 0.34 | 0.17 | 16 | 0.010625 | 11 | 0.015454545 | Dissolved |
| 207SUI060 | Dry | 1 | chromium | 3.02 | 0.806 | 16 | 0.050375 | 11 | 0.073272727 | Dissolved |
| 207SUI110 | Dry | 1 | chromium | 0.12 | 0.05 | 16 | 0.003125 | 11 | 0.004545455 | Dissolved |
| 201LAG040 | Spring | 1 | chromium | 0.32 | 0.23 | 16 | 0.014375 | 11 | 0.020909091 | Dissolved |
| 201LAG130 | Spring | 1 | chromium | 1.25 | 0.59 | 16 | 0.036875 | 11 | 0.053636364 | Dissolved |
| 201LAG270 | Spring | 1 | chromium | 2.65 | 2.33 | 16 | 0.145625 | 11 | 0.211818182 | Dissolved |
| 204ALP010 | Spring | 1 | chromium | 4.88 | 4.16 | 16 | 0.26 | 11 | 0.378181818 | Dissolved |
| 204ALP100 | Spring | 1 | chromium | 2.98 | 2.31 | 16 | 0.144375 | 11 | 0.21 | Dissolved |
| 204ALP110 | Spring | 1 | chromium | 9.38 | 8.93 | 16 | 0.558125 | 11 | 0.811818182 | Dissolved |
| 204SLE030 | Spring | 1 | chromium | 21.9 | 30.6 | 16 | 1.9125 | 11 | 2.781818182 | Dissolved |
| 204SLE230 | Spring | 1 | chromium | 1.37 | 0.16 | 16 | 0.01 | 11 | 0.014545455 | Dissolved |
| 206SPA020 | Spring | 1 | chromium | 0.84 | 0.3 | 16 | 0.01875 | 11 | 0.027272727 | Dissolved |
| 206SPA070 | Spring | 1 | chromium | 0.91 | 0.32 | 16 | 0.02 | 11 | 0.029090909 | Dissolved |
| 206SPA200 | Spring | 1 | chromium | 0.62 | 0.2 | 16 | 0.0125 | 11 | 0.018181818 | Dissolved |
| 206SPA220 | Spring | 1 | chromium | 0.63 | 0.3 | 16 | 0.01875 | 11 | 0.027272727 | Dissolved |
| 206WIL020 | Spring | 1 | chromium | 0.17 | 0.28 | 16 | 0.0175 | 11 | 0.025454545 | Dissolved |
| 207SUI010 | Spring | 1 | chromium | 0.18 | 0.07 | 16 | 0.004375 | 11 | 0.006363636 | Dissolved |
| 207SUI060 | Spring | 1 | chromium | 0.14 | 0.09 | 16 | 0.005625 | 11 | 0.008181818 | Dissolved |
| 207SUI110 | Spring | 1 | chromium | 0.42 | 0.17 | 16 | 0.010625 | 11 | 0.015454545 | Dissolved |

Appendix G-4: Concentrations of non-hardness-dependent metals in water compared to acute and chronic water quality objectives (WQO's)

| Station | Season | Year | Metal Name | Metal, Total (ug/L) | Metal, Dissolved (ug/L) | Acute WQO (ug/L) | Acute Exceedance Factor | Chronic WQO (ug/L) | Chronic Exceedance Factor | WQO Fraction |
|-----------|--------|------|------------|---------------------|-------------------------|------------------|-------------------------|--------------------|---------------------------|--------------|
| 202BUT010 | Dry | 2 | chromium | 0.17 | -0.03 | 16 | -0.001875 | 11 | -0.002727273 | Dissolved |
| 202PES050 | Dry | 2 | chromium | 0.32 | 0.12 | 16 | 0.0075 | 11 | 0.010909091 | Dissolved |
| 202PES070 | Dry | 2 | chromium | 0.14 | 0.05 | 16 | 0.003125 | 11 | 0.004545455 | Dissolved |
| 202SGR010 | Dry | 2 | chromium | 0.27 | 0.13 | 16 | 0.008125 | 11 | 0.011818182 | Dissolved |
| 202SGR080 | Dry | 2 | chromium | 0.21 | 0.22 | 16 | 0.01375 | 11 | 0.02 | Dissolved |
| 205PER010 | Dry | 2 | chromium | 0.45 | 0.25 | 16 | 0.015625 | 11 | 0.022727273 | Dissolved |
| 205PER070 | Dry | 2 | chromium | 0.87 | 0.46 | 16 | 0.02875 | 11 | 0.041818182 | Dissolved |
| 205STE020 | Dry | 2 | chromium | 0.47 | 0.06 | 16 | 0.00375 | 11 | 0.005454545 | Dissolved |
| 205STE060 | Dry | 2 | chromium | 0.32 | 0.33 | 16 | 0.020625 | 11 | 0.03 | Dissolved |
| 202BUT010 | Spring | 2 | chromium | 0.24 | 0.06 | 16 | 0.00375 | 11 | 0.005454545 | Dissolved |
| 202PES050 | Spring | 2 | chromium | 0.23 | 0.1 | 16 | 0.00625 | 11 | 0.009090909 | Dissolved |
| 202PES070 | Spring | 2 | chromium | 0.08 | 0.08 | 16 | 0.005 | 11 | 0.007272727 | Dissolved |
| 202SGR010 | Spring | 2 | chromium | 0.22 | 0.09 | 16 | 0.005625 | 11 | 0.008181818 | Dissolved |
| 202SGR080 | Spring | 2 | chromium | 0.2 | 0.14 | 16 | 0.00875 | 11 | 0.012727273 | Dissolved |
| 205PER010 | Spring | 2 | chromium | 0.52 | 0.35 | 16 | 0.021875 | 11 | 0.031818182 | Dissolved |
| 205PER070 | Spring | 2 | chromium | 2.72 | 2.31 | 16 | 0.144375 | 11 | 0.21 | Dissolved |
| 205STE020 | Spring | 2 | chromium | 0.47 | 0.44 | 16 | 0.0275 | 11 | 0.04 | Dissolved |
| 205STE060 | Spring | 2 | chromium | 0.45 | 0.19 | 16 | 0.011875 | 11 | 0.017272727 | Dissolved |
| 202BUT010 | Wet | 2 | chromium | 0.087 | -0.03 | 16 | -0.001875 | 11 | -0.002727273 | Dissolved |
| 202PES050 | Wet | 2 | chromium | 0.085 | -0.03 | 16 | -0.001875 | 11 | -0.002727273 | Dissolved |
| 202PES070 | Wet | 2 | chromium | 0.107 | -0.03 | 16 | -0.001875 | 11 | -0.002727273 | Dissolved |
| 202SGR010 | Wet | 2 | chromium | 0.284 | 0.039 | 16 | 0.0024375 | 11 | 0.003545455 | Dissolved |
| 202SGR080 | Wet | 2 | chromium | 0.322 | 0.0511 | 16 | 0.00319375 | 11 | 0.004645455 | Dissolved |
| 205PER010 | Wet | 2 | chromium | 1.56 | 0.543 | 16 | 0.0339375 | 11 | 0.049363636 | Dissolved |
| 205PER070 | Wet | 2 | chromium | 8.12 | 6.8 | 16 | 0.425 | 11 | 0.618181818 | Dissolved |
| 205STE020 | Wet | 2 | chromium | 1.72 | 1.15 | 16 | 0.071875 | 11 | 0.104545455 | Dissolved |
| 205STE060 | Wet | 2 | chromium | 2.46 | 1.33 | 16 | 0.083125 | 11 | 0.120909091 | Dissolved |
| | | | | | | | | | | |
| 201LAG130 | Dry | 1 | mercury | 0.00109 | | 2.4 | 0.000454167 | 0.025 | 0.0436 | Total |
| 201LAG270 | Dry | 1 | mercury | 0.0000933 | | 2.4 | 0.000038875 | 0.025 | 0.003732 | Total |
| 201WLK090 | Dry | 1 | mercury | 0.015 | | 2.4 | 0.00625 | 0.025 | 0.6 | Total |
| 204ALP010 | Dry | 1 | mercury | 0.00304 | | 2.4 | 0.001266667 | 0.025 | 0.1216 | Total |
| 204SLE030 | Dry | 1 | mercury | 0.00149 | | 2.4 | 0.000620833 | 0.025 | 0.0596 | Total |
| 206SPA020 | Dry | 1 | mercury | 0.00581 | | 2.4 | 0.002420833 | 0.025 | 0.2324 | Total |
| 206WIL020 | Dry | 1 | mercury | 0.0231 | | 2.4 | 0.009625 | 0.025 | 0.924 | Total |
| 207SUI010 | Dry | 1 | mercury | 0.00099 | | 2.4 | 0.0004125 | 0.025 | 0.0396 | Total |
| 207SUI020 | Dry | 1 | mercury | 0.00119 | | 2.4 | 0.000495833 | 0.025 | 0.0476 | Total |
| 201LAG040 | Spring | 1 | mercury | 0.00129 | | 2.4 | 0.0005375 | 0.025 | 0.0516 | Total |
| 201LAG130 | Spring | 1 | mercury | 0.0015 | | 2.4 | 0.000625 | 0.025 | 0.06 | Total |
| 201WLK090 | Spring | 1 | mercury | 0.00463 | | 2.4 | 0.001929167 | 0.025 | 0.1852 | Total |
| 204ALP010 | Spring | 1 | mercury | 0.00445 | | 2.4 | 0.001854167 | 0.025 | 0.178 | Total |
| 204SLE030 | Spring | 1 | mercury | 0.00051 | | 2.4 | 0.0002125 | 0.025 | 0.0204 | Total |
| 206SPA020 | Spring | 1 | mercury | 0.00224 | | 2.4 | 0.000933333 | 0.025 | 0.0896 | Total |
| 206WIL020 | Spring | 1 | mercury | 0.00545 | | 2.4 | 0.002270833 | 0.025 | 0.218 | Total |
| 207SUI010 | Spring | 1 | mercury | 0.00064 | | 2.4 | 0.000266667 | 0.025 | 0.0256 | Total |

Appendix G-4: Concentrations of non-hardness-dependent metals in water compared to acute and chronic water quality objectives (WQO's)

| Station | Season | Year | Metal Name | Metal, Total (ug/L) | Metal, Dissolved (ug/L) | Acute WQO (ug/L) | Acute Exceedance Factor | Chronic WQO (ug/L) | Chronic Exceedance Factor | WQO Fraction |
|-----------|--------|------|------------|---------------------|-------------------------|------------------|-------------------------|--------------------|---------------------------|--------------|
| 202BUT010 | Dry | 2 | mercury | 0.00102 | | 2.4 | 0.000425 | 0.025 | 0.0408 | Total |
| 202PES050 | Dry | 2 | mercury | 0.00176 | | 2.4 | 0.000733333 | 0.025 | 0.0704 | Total |
| 202PES070 | Dry | 2 | mercury | 0.00157 | | 2.4 | 0.000654167 | 0.025 | 0.0628 | Total |
| 202SGR010 | Dry | 2 | mercury | 0.000904 | | 2.4 | 0.000376667 | 0.025 | 0.03616 | Total |
| 202SGR080 | Dry | 2 | mercury | 0.00061 | | 2.4 | 0.000254167 | 0.025 | 0.0244 | Total |
| 205PER010 | Dry | 2 | mercury | 0.00192 | | 2.4 | 0.0008 | 0.025 | 0.0768 | Total |
| 205PER070 | Dry | 2 | mercury | 0.0024 | | 2.4 | 0.001 | 0.025 | 0.096 | Total |
| 205STE020 | Dry | 2 | mercury | 0.003 | | 2.4 | 0.00125 | 0.025 | 0.12 | Total |
| 205STE060 | Dry | 2 | mercury | 0.00166 | | 2.4 | 0.000691667 | 0.025 | 0.0664 | Total |
| 202BUT010 | Spring | 2 | mercury | 0.00098 | | 2.4 | 0.000408333 | 0.025 | 0.0392 | Total |
| 202PES050 | Spring | 2 | mercury | 0.00077 | | 2.4 | 0.000320833 | 0.025 | 0.0308 | Total |
| 202PES070 | Spring | 2 | mercury | 0.00096 | | 2.4 | 0.0004 | 0.025 | 0.0384 | Total |
| 202SGR010 | Spring | 2 | mercury | -0.00009 | | 2.4 | -0.0000375 | 0.025 | -0.0036 | Total |
| 202SGR080 | Spring | 2 | mercury | 0.00068 | | 2.4 | 0.000283333 | 0.025 | 0.0272 | Total |
| 205PER010 | Spring | 2 | mercury | 0.00207 | | 2.4 | 0.0008625 | 0.025 | 0.0828 | Total |
| 205PER070 | Spring | 2 | mercury | 0.00137 | | 2.4 | 0.000570833 | 0.025 | 0.0548 | Total |
| 205STE020 | Spring | 2 | mercury | 0.00062 | | 2.4 | 0.000258333 | 0.025 | 0.0248 | Total |
| 205STE060 | Spring | 2 | mercury | 0.00131 | | 2.4 | 0.000545833 | 0.025 | 0.0524 | Total |
| 202BUT010 | Wet | 2 | mercury | 0.00909 | | 2.4 | 0.0037875 | 0.025 | 0.3636 | Total |
| 202PES050 | Wet | 2 | mercury | 0.00778 | | 2.4 | 0.003241667 | 0.025 | 0.3112 | Total |
| 202PES070 | Wet | 2 | mercury | 0.0133 | | 2.4 | 0.005541667 | 0.025 | 0.532 | Total |
| 202SGR010 | Wet | 2 | mercury | 0.0145 | | 2.4 | 0.006041667 | 0.025 | 0.58 | Total |
| 202SGR080 | Wet | 2 | mercury | 0.0139 | | 2.4 | 0.005791667 | 0.025 | 0.556 | Total |
| 205PER010 | Wet | 2 | mercury | 0.0125 | | 2.4 | 0.005208333 | 0.025 | 0.5 | Total |
| 205PER070 | Wet | 2 | mercury | 0.0156 | | 2.4 | 0.0065 | 0.025 | 0.624 | Total |
| 205STE020 | Wet | 2 | mercury | 0.011 | | 2.4 | 0.004583333 | 0.025 | 0.44 | Total |
| 205STE060 | Wet | 2 | mercury | 0.0109 | | 2.4 | 0.004541667 | 0.025 | 0.436 | Total |
| | | | | | | | | | | |
| 201LAG130 | Dry | 1 | selenium | -0.05 | -0.05 | 20 | -0.0025 | 5 | -0.01 | Total |
| 201LAG270 | Dry | 1 | selenium | 0.232 | 0.291 | 20 | 0.0116 | 5 | 0.0464 | Total |
| 204ALP010 | Dry | 1 | selenium | 2.27 | 2.35 | 20 | 0.1135 | 5 | 0.454 | Total |
| 204ALP100 | Dry | 1 | selenium | 2.58 | 2.42 | 20 | 0.129 | 5 | 0.516 | Total |
| 204ALP110 | Dry | 1 | selenium | 2.56 | 2.19 | 20 | 0.128 | 5 | 0.512 | Total |
| 204SLE030 | Dry | 1 | selenium | 1.65 | 1.66 | 20 | 0.0825 | 5 | 0.33 | Total |
| 204SLE230 | Dry | 1 | selenium | 0.178 | 3.4 | 20 | 0.0089 | 5 | 0.0356 | Total |
| 206SPA020 | Dry | 1 | selenium | 1.6 | 1.53 | 20 | 0.08 | 5 | 0.32 | Total |
| 206SPA070 | Dry | 1 | selenium | 1.79 | 1.53 | 20 | 0.0895 | 5 | 0.358 | Total |
| 206SPA200 | Dry | 1 | selenium | 3.39 | 0.173 | 20 | 0.1695 | 5 | 0.678 | Total |
| 206SPA220 | Dry | 1 | selenium | 0.942 | 0.869 | 20 | 0.0471 | 5 | 0.1884 | Total |
| 206WIL020 | Dry | 1 | selenium | 1.33 | 1.11 | 20 | 0.0665 | 5 | 0.266 | Total |
| 207SUI010 | Dry | 1 | selenium | 0.318 | 0.165 | 20 | 0.0159 | 5 | 0.0636 | Total |
| 207SUI020 | Dry | 1 | selenium | 1.71 | 1.44 | 20 | 0.0855 | 5 | 0.342 | Total |
| 207SUI060 | Dry | 1 | selenium | 0.188 | 0.175 | 20 | 0.0094 | 5 | 0.0376 | Total |
| 207SUI110 | Dry | 1 | selenium | 0.63 | 0.64 | 20 | 0.0315 | 5 | 0.126 | Total |
| 201LAG040 | Spring | 1 | selenium | 1.08 | 0.66 | 20 | 0.054 | 5 | 0.216 | Total |
| 201LAG130 | Spring | 1 | selenium | 0.5 | 0.57 | 20 | 0.025 | 5 | 0.1 | Total |
| 201LAG270 | Spring | 1 | selenium | 0.68 | 1.17 | 20 | 0.034 | 5 | 0.136 | Total |

Appendix G-4: Concentrations of non-hardness-dependent metals in water compared to acute and chronic water quality objectives (WQO's)

| Station | Season | Year | Metal Name | Metal, Total (ug/L) | Metal, Dissolved (ug/L) | Acute WQO (ug/L) | Acute Exceedance Factor | Chronic WQO (ug/L) | Chronic Exceedance Factor | WQO Fraction |
|-----------|--------|------|------------|---------------------|-------------------------|------------------|-------------------------|--------------------|---------------------------|--------------|
| 204ALP010 | Spring | 1 | selenium | 5.13 | 5.13 | 20 | 0.2565 | 5 | 1.026 | Total |
| 204ALP100 | Spring | 1 | selenium | 8.6 | 8.58 | 20 | 0.43 | 5 | 1.72 | Total |
| 204ALP110 | Spring | 1 | selenium | 3.47 | 3.41 | 20 | 0.1735 | 5 | 0.694 | Total |
| 204SLE030 | Spring | 1 | selenium | 2.16 | 2.3 | 20 | 0.108 | 5 | 0.432 | Total |
| 204SLE230 | Spring | 1 | selenium | 0.62 | 0.57 | 20 | 0.031 | 5 | 0.124 | Total |
| 206SPA020 | Spring | 1 | selenium | 2.74 | 2.57 | 20 | 0.137 | 5 | 0.548 | Total |
| 206SPA070 | Spring | 1 | selenium | 3.14 | 2.78 | 20 | 0.157 | 5 | 0.628 | Total |
| 206SPA200 | Spring | 1 | selenium | 2.82 | 2.79 | 20 | 0.141 | 5 | 0.564 | Total |
| 206SPA220 | Spring | 1 | selenium | 1.37 | 1.2 | 20 | 0.0685 | 5 | 0.274 | Total |
| 206WIL020 | Spring | 1 | selenium | 0.39 | 1.41 | 20 | 0.0195 | 5 | 0.078 | Total |
| 207SUI010 | Spring | 1 | selenium | 0.9 | 1.04 | 20 | 0.045 | 5 | 0.18 | Total |
| 207SUI060 | Spring | 1 | selenium | 0.9 | 0.67 | 20 | 0.045 | 5 | 0.18 | Total |
| 207SUI110 | Spring | 1 | selenium | 0.85 | 0.95 | 20 | 0.0425 | 5 | 0.17 | Total |
| | | | | | | | | | | |
| 202BUT010 | Dry | 2 | selenium | 0.96 | 0.83 | 20 | 0.048 | 5 | 0.192 | Total |
| 202PES050 | Dry | 2 | selenium | 1.56 | 1.45 | 20 | 0.078 | 5 | 0.312 | Total |
| 202PES070 | Dry | 2 | selenium | 1.46 | 1.49 | 20 | 0.073 | 5 | 0.292 | Total |
| 202SGR010 | Dry | 2 | selenium | 1.56 | 1.64 | 20 | 0.078 | 5 | 0.312 | Total |
| 202SGR080 | Dry | 2 | selenium | 1.02 | 1.79 | 20 | 0.051 | 5 | 0.204 | Total |
| 205PER010 | Dry | 2 | selenium | 1.97 | 2.06 | 20 | 0.0985 | 5 | 0.394 | Total |
| 205PER070 | Dry | 2 | selenium | 5.84 | 5.84 | 20 | 0.292 | 5 | 1.168 | Total |
| 205STE020 | Dry | 2 | selenium | 4.26 | 0.54 | 20 | 0.213 | 5 | 0.852 | Total |
| 205STE060 | Dry | 2 | selenium | 0.7 | 4.23 | 20 | 0.035 | 5 | 0.14 | Total |
| 202BUT010 | Spring | 2 | selenium | 0.04 | 0.7 | 20 | 0.002 | 5 | 0.008 | Total |
| 202PES050 | Spring | 2 | selenium | 1.37 | 1.26 | 20 | 0.0685 | 5 | 0.274 | Total |
| 202PES070 | Spring | 2 | selenium | 0.2 | 0.59 | 20 | 0.01 | 5 | 0.04 | Total |
| 202SGR010 | Spring | 2 | selenium | 1.3 | 0.63 | 20 | 0.065 | 5 | 0.26 | Total |
| 202SGR080 | Spring | 2 | selenium | 1.23 | 0.34 | 20 | 0.0615 | 5 | 0.246 | Total |
| 205PER010 | Spring | 2 | selenium | 1.74 | 1.83 | 20 | 0.087 | 5 | 0.348 | Total |
| 205PER070 | Spring | 2 | selenium | 10.3 | 5.09 | 20 | 0.515 | 5 | 2.06 | Total |
| 205STE020 | Spring | 2 | selenium | 4.92 | 10.1 | 20 | 0.246 | 5 | 0.984 | Total |
| 205STE060 | Spring | 2 | selenium | 0.59 | 0.72 | 20 | 0.0295 | 5 | 0.118 | Total |
| 202BUT010 | Wet | 2 | selenium | 0.791 | 1.11 | 20 | 0.03955 | 5 | 0.1582 | Total |
| 202PES050 | Wet | 2 | selenium | 1.22 | 1.16 | 20 | 0.061 | 5 | 0.244 | Total |
| 202PES070 | Wet | 2 | selenium | 1.12 | 1.16 | 20 | 0.056 | 5 | 0.224 | Total |
| 202SGR010 | Wet | 2 | selenium | 1.04 | 1.01 | 20 | 0.052 | 5 | 0.208 | Total |
| 202SGR080 | Wet | 2 | selenium | 0.901 | 0.981 | 20 | 0.04505 | 5 | 0.1802 | Total |
| 205PER010 | Wet | 2 | selenium | 3.9 | 3.95 | 20 | 0.195 | 5 | 0.78 | Total |
| 205PER070 | Wet | 2 | selenium | 18.7 | 18.8 | 20 | 0.935 | 5 | 3.74 | Total |
| 205STE020 | Wet | 2 | selenium | 4.71 | 4.9 | 20 | 0.2355 | 5 | 0.942 | Total |
| 205STE060 | Wet | 2 | selenium | 4.67 | 4.84 | 20 | 0.2335 | 5 | 0.934 | Total |

Appendix G-5: Organic analytes detected in water samples (detected values in bold font)

Note: See Appendix G-6 for all organic analytes not detected.

| Station ID | Season | Organic Carbon, Dissolved | Organic Carbon, Total | Anthracene | Benz(a)anthracene | Benzo(a)pyrene | Benzo(b)fluoranthene | Benzo(e)pyrene | Benzo(g,h,i)perylene |
|------------|--------|---------------------------|-----------------------|--------------|-------------------|----------------|----------------------|----------------|----------------------|
| | | mg/L | mg/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L |
| 201LAG130 | Dry | 3.4 | 4 | -0.02 | -0.02 | 0.035 | -0.02 | -0.02 | -0.02 |
| 201LAG270 | Dry | 2.8 | 5 | -0.02 | 0.035 | 0.035 | -0.02 | -0.02 | -0.02 |
| 204SLE030 | Dry | 5.1 | 7.7 | -0.02 | 0.035 | -0.02 | -0.02 | -0.02 | -0.02 |
| 206SPA020 | Dry | 23.9 | 27.7 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 206WIL020 | Dry | 17 | 58 | 0.035 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 207SUI010 | Dry | -0.6 | 3 | -0.02 | -0.02 | 0.035 | 0.035 | 0.035 | -0.02 |
| 207SUI020 | Dry | 11.9 | 28 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 207SUI110 | Dry | 19.8 | 24.7 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 204ALP010 | Dry | 3.8 | 6.6 | -0.02 | 0.035 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202BUT010 | Dry | 10.9 | 11.8 | -0.02 | -0.02 | -0.02 | 0.035 | -0.02 | -0.02 |
| 202PES050 | Dry | 18.4 | 19.3 | -0.02 | -0.02 | -0.02 | 0.035 | -0.02 | -0.02 |
| 202PES070 | Dry | 10.2 | 11.1 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202SGR010 | Dry | 8.2 | 12.3 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202SGR080 | Dry | 5.5 | 6 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 205PER010 | Dry | 6.1 | 7.2 | -0.02 | -0.02 | -0.02 | 0.035 | -0.02 | 0.057 |
| 205PER070 | Dry | 1.4 | 2 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 205STE020 | Dry | 2.9 | 4.5 | -0.02 | -0.02 | -0.02 | 0.035 | -0.02 | 0.035 |
| 205STE060 | Dry | 3.3 | 3.4 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | 0.035 |
| 201LAG040 | Spring | 1.1 | 5 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 201LAG130 | Spring | -1 | -1 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 204SLE030 | Spring | 4.5 | 6.9 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 206SPA020 | Spring | 11 | 21.2 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 206WIL020 | Spring | 2.2 | 2.4 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 207SUI010 | Spring | 3.5 | 8.5 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 204ALP010 | Spring | 5.9 | 6.1 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202BUT010 | Spring | 4.3 | 8.7 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202PES050 | Spring | 2.4 | 4.4 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202PES070 | Spring | 4.9 | 6.1 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202SGR010 | Spring | 3.1 | 5.1 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202SGR080 | Spring | -1 | 4.1 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 205PER010 | Spring | 2.7 | 10.9 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 205PER070 | Spring | 5.1 | 12.9 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 205STE020 | Spring | 8.3 | 13.2 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 205STE060 | Spring | 6.1 | 8.3 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202BUT010 | Wet | 0.5 | 2.6 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 202PES050 | Wet | 1.4 | 2.4 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 202PES070 | Wet | 1.3 | 2.4 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 202SGR010 | Wet | 1.7 | 3.1 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 202SGR080 | Wet | 1.6 | 2.7 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 205PER010 | Wet | 1.7 | 2.9 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 205PER070 | Wet | 0.7 | 0.9 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 205STE020 | Wet | 1.3 | 1.9 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 205STE060 | Wet | 1.4 | 2.4 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |

Appendix G-5: Organic analytes detected in water samples (detected values in bold font)

| Station ID | Season | Benzo(k)fluoranthene | Chrysene | Dibenz(a,h)anthracene | Fluorene | Methylnaphthalene, 1- | Naphthalene | Perylene | Phenanthrene | Pyrene |
|------------|--------|----------------------|--------------|-----------------------|--------------|-----------------------|--------------|--------------|--------------|--------------|
| | | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L |
| 201LAG130 | Dry | -0.02 | 0.035 | -0.02 | -0.02 | -0.02 | -0.02 | | -0.02 | -0.02 |
| 201LAG270 | Dry | -0.02 | 0.035 | -0.02 | -0.02 | -0.02 | -0.02 | | -0.02 | -0.02 |
| 204SLE030 | Dry | -0.02 | 0.035 | -0.02 | 0.035 | 0.043 | -0.02 | | -0.02 | -0.02 |
| 206SPA020 | Dry | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | | -0.02 | -0.02 |
| 206WIL020 | Dry | -0.02 | -0.02 | 0.035 | -0.02 | -0.02 | -0.02 | 0.08 | 0.035 | -0.02 |
| 207SUI010 | Dry | 0.035 | 0.035 | -0.02 | -0.02 | -0.02 | -0.02 | | -0.02 | -0.02 |
| 207SUI020 | Dry | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 207SUI110 | Dry | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 204ALP010 | Dry | -0.02 | -0.02 | -0.02 | -0.02 | 0.045 | -0.02 | | -0.02 | -0.02 |
| 202BUT010 | Dry | 0.035 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202PES050 | Dry | 0.035 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202PES070 | Dry | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202SGR010 | Dry | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202SGR080 | Dry | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | 0.035 | -0.02 | -0.02 |
| 205PER010 | Dry | 0.035 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | 0.12 | -0.02 | 0.035 |
| 205PER070 | Dry | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 205STE020 | Dry | 0.035 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | 0.035 | -0.02 | -0.02 |
| 205STE060 | Dry | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | 0.035 | -0.02 | -0.02 |
| 201LAG040 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 201LAG130 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | 0.035 | -0.02 | -0.02 | -0.02 |
| 204SLE030 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 206SPA020 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 206WIL020 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 207SUI010 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 204ALP010 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202BUT010 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202PES050 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202PES070 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202SGR010 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202SGR080 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 205PER010 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 205PER070 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 205STE020 | Spring | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 205STE060 | Spring | -0.02 | -0.02 | 0.035 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 | -0.02 |
| 202BUT010 | Wet | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 202PES050 | Wet | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 202PES070 | Wet | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 202SGR010 | Wet | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 202SGR080 | Wet | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 205PER010 | Wet | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 205PER070 | Wet | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 205STE020 | Wet | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |
| 205STE060 | Wet | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 | -0.025 |

Appendix G-5: Organic analytes detected in water samples (detected values in bold font)

| Station ID | Season | Trimethylnaphthalene, 2,3,5- | Sum PAHs | PCB 005 | PCB 018 | PCB 101 | Sum PCBs | Carbophenothion | Chlordene, gamma | Chlorpyrifos (ELISA) | Dacthal | DDE(p,p') | DDMU(p,p') |
|------------|--------|------------------------------|--------------|--------------|--------------|--------------|--------------|-----------------|------------------|----------------------|---------------|---------------|---------------|
| | | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L |
| 201LAG130 | Dry | -0.02 | 0.07 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 201LAG270 | Dry | -0.02 | 0.105 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 204SLE030 | Dry | -0.02 | 0.07 | -0.001 | 0.002 | 0.003 | 0.005 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 206SPA020 | Dry | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | 0.054 | -0.001 | 0.079 | 0.005 | -0.001 | -0.001 |
| 206WIL020 | Dry | -0.02 | 0.035 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | 0.0015 | -0.05 | -0.001 | -0.001 | -0.001 |
| 207SUI010 | Dry | -0.02 | 0.14 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 207SUI020 | Dry | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | 0.002 | -0.001 |
| 207SUI110 | Dry | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 204ALP010 | Dry | 0.035 | 0.035 | 0.003 | -0.001 | -0.001 | 0.003 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | 0.002 |
| 202BUT010 | Dry | -0.02 | 0.07 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 202PES050 | Dry | -0.02 | 0.07 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 202PES070 | Dry | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 202SGR010 | Dry | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | 0.004 | -0.001 | 0.0015 |
| 202SGR080 | Dry | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 205PER010 | Dry | -0.02 | 0.07 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 205PER070 | Dry | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 205STE020 | Dry | -0.02 | 0.07 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 205STE060 | Dry | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 201LAG040 | Spring | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 201LAG130 | Spring | -0.02 | 0.035 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 204SLE030 | Spring | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 206SPA020 | Spring | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | 0.0015 | -0.001 |
| 206WIL020 | Spring | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | 0.057 | -0.001 | -0.001 | -0.001 |
| 207SUI010 | Spring | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 204ALP010 | Spring | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | 0.0015 | -0.001 |
| 202BUT010 | Spring | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | 0.0015 | 0.0015 |
| 202PES050 | Spring | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | 0.0015 |
| 202PES070 | Spring | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 202SGR010 | Spring | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | 0.0015 | -0.001 | -0.001 |
| 202SGR080 | Spring | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 205PER010 | Spring | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | 0.0015 | -0.001 |
| 205PER070 | Spring | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 205STE020 | Spring | -0.02 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 205STE060 | Spring | -0.02 | 0.035 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 202BUT010 | Wet | -0.025 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 202PES050 | Wet | -0.025 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 202PES070 | Wet | -0.025 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 202SGR010 | Wet | -0.025 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 202SGR080 | Wet | -0.025 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 205PER010 | Wet | -0.025 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 205PER070 | Wet | -0.025 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 205STE020 | Wet | -0.025 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |
| 205STE060 | Wet | -0.025 | 0 | -0.001 | -0.001 | -0.001 | 0 | -0.03 | -0.001 | -0.05 | -0.001 | -0.001 | -0.001 |

Appendix G-5: Organic analytes detected in water samples (detected values in bold font)

| Station ID | Season | DDT(p,p') | Diazinon (ELISA) | Diazinon (DFG) | Dieldrin | Dioxathion | Disulfoton | Endosulfan I | Fonofos (Dyfonate) | HCH, delta | HCH, gamma | Hexachlorobenzene | Mevinphos |
|------------|--------|---------------|------------------|----------------|---------------|-------------|-------------|--------------|--------------------|---------------|---------------|-------------------|--------------|
| | | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L |
| 201LAG130 | Dry | -0.002 | -0.03 | 0.013 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | 0.002 | -0.0005 | -0.03 |
| 201LAG270 | Dry | -0.002 | -0.03 | 0.013 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 204SLE030 | Dry | -0.002 | -0.03 | 0.013 | -0.001 | -0.03 | -0.01 | 0.008 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 206SPA020 | Dry | -0.002 | 0.174 | 0.12 | -0.001 | 0.04 | 0.03 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | 0.056 |
| 206WIL020 | Dry | -0.002 | -0.03 | 0.03 | -0.001 | -0.03 | 0.03 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 207SUI010 | Dry | -0.002 | -0.03 | 0.013 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | 0.003 | -0.0005 | -0.03 |
| 207SUI020 | Dry | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 207SUI110 | Dry | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 204ALP010 | Dry | -0.002 | 0.0383 | 0.03 | -0.001 | -0.03 | -0.01 | 0.007 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 202BUT010 | Dry | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 202PES050 | Dry | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 202PES070 | Dry | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 202SGR010 | Dry | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | 0.03 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 202SGR080 | Dry | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 205PER010 | Dry | -0.002 | -0.03 | 0.026 | 0.0015 | -0.03 | 0.03 | -0.001 | 0.03 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 205PER070 | Dry | 0.0035 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | 0.03 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 205STE020 | Dry | -0.002 | -0.03 | 0.013 | 0.0015 | -0.03 | -0.01 | -0.001 | 0.03 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 205STE060 | Dry | -0.002 | -0.03 | -0.005 | 0.0015 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 201LAG040 | Spring | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 201LAG130 | Spring | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 204SLE030 | Spring | -0.002 | 0.096 | 0.118 | -0.001 | -0.03 | 0.03 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 206SPA020 | Spring | -0.002 | -0.03 | 0.021 | -0.001 | -0.03 | 0.03 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 206WIL020 | Spring | 0.0035 | -0.03 | 0.013 | -0.001 | -0.03 | 0.03 | -0.001 | -0.02 | -0.001 | -0.001 | 0.00075 | -0.03 |
| 207SUI010 | Spring | -0.002 | -0.03 | 0.013 | -0.001 | -0.03 | 0.03 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 204ALP010 | Spring | -0.002 | -0.03 | 0.02 | -0.001 | -0.03 | 0.03 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 202BUT010 | Spring | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | 0.0015 | 0.0015 | -0.0005 | -0.03 |
| 202PES050 | Spring | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 202PES070 | Spring | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 202SGR010 | Spring | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 202SGR080 | Spring | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 205PER010 | Spring | -0.002 | -0.03 | 0.013 | -0.001 | -0.03 | 0.03 | -0.001 | -0.02 | -0.001 | -0.001 | 0.00075 | -0.03 |
| 205PER070 | Spring | 0.0035 | -0.03 | 0.013 | -0.001 | -0.03 | 0.03 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 205STE020 | Spring | -0.002 | -0.03 | 0.013 | -0.001 | -0.03 | 0.03 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 205STE060 | Spring | -0.002 | -0.03 | 0.013 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 202BUT010 | Wet | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 202PES050 | Wet | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 202PES070 | Wet | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 202SGR010 | Wet | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 202SGR080 | Wet | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 205PER010 | Wet | -0.002 | -0.03 | 0.01 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 205PER070 | Wet | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 205STE020 | Wet | -0.002 | -0.03 | 0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |
| 205STE060 | Wet | -0.002 | -0.03 | -0.005 | -0.001 | -0.03 | -0.01 | -0.001 | -0.02 | -0.001 | -0.001 | -0.0005 | -0.03 |

Appendix G-5: Organic analytes detected in water samples (detected values in bold font)

| Station ID | Season | Oxadiazon | Parathion, Methyl | Propazine | Secbumeton | Terbutylazine | Thiobencarb |
|------------|--------|---------------|-------------------|--------------|--------------|---------------|-------------|
| | | ug/L | ug/L | ug/L | ug/L | ug/L | ug/L |
| 201LAG130 | Dry | -0.001 | -0.01 | | | | -0.1 |
| 201LAG270 | Dry | -0.001 | -0.01 | | | | -0.1 |
| 204SLE030 | Dry | -0.001 | -0.01 | | | | -0.1 |
| 206SPA020 | Dry | 0.018 | 0.03 | | | | 0.21 |
| 206WIL020 | Dry | 0.033 | -0.01 | 0.035 | 0.035 | -0.02 | -0.1 |
| 207SUI010 | Dry | -0.001 | -0.01 | | | | -0.1 |
| 207SUI020 | Dry | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 207SUI110 | Dry | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 204ALP010 | Dry | 0.005 | -0.01 | | | | -0.1 |
| 202BUT010 | Dry | 0.0015 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 202PES050 | Dry | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 202PES070 | Dry | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 202SGR010 | Dry | 0.0015 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 202SGR080 | Dry | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 205PER010 | Dry | 0.011 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 205PER070 | Dry | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 205STE020 | Dry | 0.014 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 205STE060 | Dry | 0.0015 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 201LAG040 | Spring | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 201LAG130 | Spring | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 204SLE030 | Spring | 0.0015 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 206SPA020 | Spring | 0.042 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 206WIL020 | Spring | 0.0015 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 207SUI010 | Spring | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 204ALP010 | Spring | 0.062 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 202BUT010 | Spring | -0.001 | -0.01 | -0.02 | -0.02 | 0.035 | -0.1 |
| 202PES050 | Spring | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 202PES070 | Spring | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 202SGR010 | Spring | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 202SGR080 | Spring | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 205PER010 | Spring | 0.0015 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 205PER070 | Spring | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 205STE020 | Spring | 0.0015 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 205STE060 | Spring | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 202BUT010 | Wet | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 202PES050 | Wet | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 202PES070 | Wet | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 202SGR010 | Wet | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 202SGR080 | Wet | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 205PER010 | Wet | 0.009 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 205PER070 | Wet | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 205STE020 | Wet | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |
| 205STE060 | Wet | -0.001 | -0.01 | -0.02 | -0.02 | -0.02 | -0.1 |

Appendix G-6: List of organic analytes not detected in chemical analysis of water samples

| | | |
|------------------------------|----------------------|--------------------|
| Acenaphthene | PCB 114 | Dichlorvos |
| Acenaphthylene | PCB 118 | Dicrotophos |
| Biphenyl | PCB 128 | Dimethoate |
| Chrysenes, C1- | PCB 137 | Endosulfan II |
| Chrysenes, C2- | PCB 138 | Endosulfan sulfate |
| Chrysenes, C3 | PCB 141 | Endrin |
| Dibenzothiophene | PCB 149 | Endrin Aldehyde |
| Dibenzothiophenes, C1- | PCB 151 | Endrin Ketone |
| Dibenzothiophenes, C2- | PCB 153 | Ethion |
| Dibenzothiophenes, C3- | PCB 156 | Ethoprop |
| Dimethylnaphthalene, 2,6- | PCB 157 | Famphur |
| Fluoranthene | PCB 158 | Fenchlorphos |
| Fluoranthene/Pyrenes, C1- | PCB 170 | Fenitrothion |
| Fluorenes, C1- | PCB 174 | Fensulfothion |
| Fluorenes, C2- | PCB 177 | Fenthion |
| Fluorenes, C3- | PCB 180 | HCH, alpha |
| Indeno(1,2,3-c,d)pyrene | PCB 183 | HCH, beta |
| Methylnaphthalene, 2- | PCB 187 | Heptachlor |
| Methylphenanthrene, 1- | PCB 189 | Heptachlor epoxide |
| Naphthalenes, C1- | PCB 194 | Leptophos |
| Naphthalenes, C2- | PCB 195 | Malathion |
| Naphthalenes, C3- | PCB 200 | Merphos |
| Naphthalenes, C4- | PCB 201 | Methidathion |
| Phenanthrene/Anthracene, C1- | PCB 203 | Methoxychlor |
| Phenanthrene/Anthracene, C2- | PCB 206 | Mirex |
| Phenanthrene/Anthracene, C3- | PCB 209 | Molinate |
| Phenanthrene/Anthracene, C4- | Aldrin | Naled(Dibrom) |
| PCB 008 | Ametryn | Nonachlor, cis |
| PCB 015 | Aspon | Nonachlor, trans |
| PCB 027 | Atraton | Oxychlorane |
| PCB 028 | Atrazine | Parathion, Ethyl |
| PCB 029 | Azinphos ethyl | Phorate |
| PCB 031 | Azinphos methyl | Phosmet |
| PCB 033 | Bolstar | Phosphamidon |
| PCB 044 | Chlordane, cis | Prometon |
| PCB 049 | Chlordane, trans | Prometryn |
| PCB 052 | Chlordene, alpha | Simazine |
| PCB 056 | Chlorfenvinphos | Simetryn |
| PCB 060 | Chlorpyrifos (DFG) | Tedion |
| PCB 066 | Chlorpyrifos methyl | Terbufos |
| PCB 070 | Ciodrin(Crotoxyphos) | Terbutryn |
| PCB 074 | Coumaphos | Tetrachlorvinphos |
| PCB 087 | DDD(o,p') | Thionazin |
| PCB 095 | DDD(p,p') | Tokuthion |
| PCB 097 | DDE(o,p') | Trichlorfon |
| PCB 099 | DDT(o,p') | Trichloronate |
| PCB 105 | Demeton-s | |
| PCB 110 | Dichlofenthion | |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | Organic Carbon,Total % | Fine-ASTM,Clay % | Fine-ASTM,Silt % | Sand-ASTM,Fine % | Sand-ASTM,Medium % | Sand-ASTM,Coarse % | Gravel-ASTM % |
|-------------|------------------------|-------------|---------------------------|---------------------|---------------------|---------------------|-----------------------|-----------------------|------------------|
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 0.6 | 1.3 | 2.11 | 15.16 | 78.47 | 2.96 | 0 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 1 | 11 | 21.48 | 33.07 | 22.96 | 10.03 | 1.46 |
| 204ALP010 | El Charro | 18/Sep/2001 | 0.85 | 16.5 | 36.65 | 36.5 | 7.94 | 2.41 | 0 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 9.42 | 7 | 68.41 | 16.06 | 8.53 | 0 | 0 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 2.92 | 29 | 53.95 | 8.66 | 7.3 | 1.09 | 0 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 0.45 | 4.25 | 3.63 | 63.22 | 11.61 | 12.16 | 5.13 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 1.79 | 25 | 42.68 | 31.46 | 0.85 | 0 | 0 |
| 207SUI020 | Rockville | 17/Jun/2002 | 0.26 | 2 | 0.72 | 7.34 | 42.86 | 18.64 | 28.45 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 0.66 | 7 | 9.48 | 57.12 | 23.97 | 2.42 | 0 |
| 202PES050 | Water Lane | 19/Jun/2002 | 0.58 | 8 | 17.3 | 56.11 | 14.83 | 2.1 | 1.66 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 0.73 | 10.5 | 23.38 | 54.52 | 10.51 | 1.09 | 0 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 2.22 | 10 | 8.87 | 22.73 | 45.18 | 10.43 | 2.79 |
| 205STE020 | La Avenida | 17/Jun/2002 | 0.72 | 1.9 | 1.07 | 1.63 | 10.11 | 11.37 | 73.92 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | Aluminum mg/kg | Arsenic mg/kg | Cadmium mg/kg | Chromium mg/kg | Copper mg/kg | Lead mg/kg | Manganese mg/kg | Mercury mg/kg | Nickel mg/kg | Silver mg/kg | Zinc mg/kg | Metals MPECQ | PCB 008 ng/g | PCB 018 ng/g |
|-------------|------------------------|-------------|-------------------|------------------|------------------|-------------------|-----------------|---------------|--------------------|------------------|-----------------|-----------------|---------------|-----------------|-----------------|-----------------|
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 20694 | 10.5 | 0.173 | 263 | 36.6 | 13.2 | 616 | 0.185 | 429 | 0.303 | 86.8 | 1.73 | 0 | 0 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 30966 | 6.87 | 0.112 | 114 | 20.1 | 7.96 | 629 | 0.728 | 73.7 | 0.224 | 50.7 | 0.44 | 0 | 0 |
| 204ALP010 | El Charro | 18/Sep/2001 | 41298 | 7.53 | 0.199 | 133 | 52.5 | 11.5 | 411 | 0.331 | 109 | 0.378 | 77.1 | 0.62 | 0 | 0 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 38724 | 10.1 | 1.24 | 475 | 73.4 | 130 | 882 | 0.69 | 74 | 0.435 | 320 | 1.22 | 0 | 0 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 22977 | 5.97 | 0.515 | 47.5 | 25.5 | 16.2 | 4445 | 0.19 | 38.8 | 0.244 | 70.8 | 0.28 | 0 | 0 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 33671 | 5.12 | 0.069 | 192 | 15.3 | 4.39 | 479 | 0 | 77.3 | 0.171 | 36.1 | 0.53 | 0 | 0 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 42718 | 7.16 | 0.209 | 72.3 | 28.2 | 13.1 | 450 | 0 | 51.3 | 0.283 | 67.6 | 0.34 | 0.508 | 0.876 |
| 207SUI020 | Rockville | 17/Jun/2002 | 39627 | 11.5 | 0.149 | 58 | 32.8 | 12.3 | 1120 | 0 | 45.1 | 0.425 | 65.9 | 0.33 | 0 | 0 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 37765 | 3.43 | 0.163 | 20.5 | 11.3 | 7.24 | 371 | 0 | 10.6 | 0.143 | 23.2 | 0.10 | 0 | 0 |
| 202PES050 | Water Lane | 19/Jun/2002 | 31744 | 5.65 | 0.282 | 54.8 | 13.9 | 8.1 | 293 | 0 | 22.2 | 0.294 | 5.78 | 0.19 | 0 | 0 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 39779 | 4.86 | 0.219 | 71.4 | 17.3 | 6.41 | 443 | 0 | 34.5 | 0.281 | 36.5 | 0.26 | 0 | 0 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 10150 | 1.04 | 0.202 | 26.8 | 13.6 | 13.6 | 113 | 0 | 13.7 | 0.0963 | 137 | 0.16 | 0 | 0.366 |
| 205STE020 | La Avenida | 17/Jun/2002 | 29394 | 3.12 | 0.171 | 47.2 | 22 | 20.7 | 840 | 0 | 44.3 | 0.187 | 61 | 0.27 | 0 | 0 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | PCB 027 | PCB 028 | PCB 029 | PCB 031 | PCB 033 | PCB 044 | PCB 049 | PCB 052 | PCB 056 | PCB 060 | PCB 066 | PCB 070 | PCB 074 | PCB 087 |
|-------------|------------------------|-------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| | | | ng/g | ng/g | ng/g | ng/g | ng/g | ng/g | ng/g | ng/g | ng/g | ng/g | ng/g | ng/g | ng/g | ng/g |
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.189 | 0 | 0 | 0.129 | 0.136 | 0 | 0 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 204ALP010 | El Charro | 18/Sep/2001 | 0.212 | 0 | 0 | 0.233 | 1.05 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.198 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 0 | 0.771 | 0 | 0 | 0.671 | 0.85 | 0 | 1.01 | 0 | 0 | 1.72 | 1.2 | 0 | 1.24 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 0 | 0.43 | 0 | 0.362 | 0 | 0.418 | 0 | 0.746 | 0 | 0 | 0.752 | 0.531 | 0 | 0.438 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.229 | 0 | 0 | 0.14 | 0.198 | 0 | 0 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 0 | 1.15 | 0 | 1.22 | 0.588 | 1.28 | 0.929 | 1.76 | 0.428 | 0.315 | 1.05 | 1.2 | 0.472 | 1.02 |
| 207SUI020 | Rockville | 17/Jun/2002 | 0 | 0 | 0 | 0.163 | 0 | 0 | 0 | 0.243 | 0 | 0 | 0 | 0.152 | 0 | 0 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 0 | 0.173 | 0 | 0.169 | 0 | 0.197 | 0 | 0.189 | 0 | 0 | 0.215 | 0.36 | 0 | 0.213 |
| 202PES050 | Water Lane | 19/Jun/2002 | 0 | 0.219 | 0 | 0.238 | 0.154 | 0.191 | 0 | 0.254 | 0 | 0 | 0.251 | 0.279 | 0 | 0 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.272 | 0 | 0 | 0 | 0.236 | 0 | 0 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 0 | 1.97 | 0 | 1.07 | 0 | 1.76 | 0.606 | 1.76 | 0 | 0.368 | 4.36 | 0.772 | 0.337 | 1.09 |
| 205STE020 | La Avenida | 17/Jun/2002 | 0 | 0.207 | 0 | 0.179 | 0 | 0.351 | 0.221 | 0.705 | 0.159 | 0.118 | 0.402 | 0.566 | 0.182 | 0.773 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | PCB 095 ng/g | PCB 097 ng/g | PCB 099 ng/g | PCB 101 ng/g | PCB 105 ng/g | PCB 110 ng/g | PCB 114 ng/g | PCB 118 ng/g | PCB 128 ng/g | PCB 137 ng/g | PCB 138 ng/g | PCB 141 ng/g | PCB 149 ng/g | PCB 151 ng/g |
|-------------|------------------------|-------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 0.144 | 0 | 0 | 0.22 | 0 | 0.254 | 0 | 0.217 | 0 | 0 | 0.215 | 0 | 0 | 0 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 0 | 0 | 0 | 0.144 | 0 | 0.166 | 0 | 0.171 | 0 | 0 | 0 | 0 | 0 | 0 |
| 204ALP010 | El Charro | 18/Sep/2001 | 0 | 0 | 0 | 0.277 | 0 | 0.374 | 0 | 0.321 | 0 | 0 | 0.306 | 0 | 0 | 0 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 1.46 | 0 | 0.74 | 2.55 | 0 | 3.07 | 0 | 2.48 | 0 | 0 | 3.57 | 0 | 1.98 | 0 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 0.787 | 0 | 0.343 | 0.93 | 0.265 | 1.23 | 0 | 0.862 | 0 | 0 | 1.63 | 0 | 0.863 | 0.27 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 0.18 | 0 | 0 | 0.214 | 0 | 0.239 | 0 | 0.183 | 0 | 0 | 0 | 0 | 0 | 0.37 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 1.61 | 0.676 | 0.865 | 2.36 | 0.966 | 2.71 | 0 | 2.2 | 0.543 | 0 | 3.47 | 0.665 | 2.29 | 1.08 |
| 207SUI020 | Rockville | 17/Jun/2002 | 0.177 | 0 | 0 | 0.239 | 0 | 0.26 | 0 | 0.224 | 0 | 0 | 0.163 | 0 | 0 | 0 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 0.284 | 0 | 0 | 0.311 | 0.236 | 0.591 | 0 | 0.425 | 0 | 0 | 0.347 | 0 | 0 | 0 |
| 202PES050 | Water Lane | 19/Jun/2002 | 0.211 | 0 | 0 | 0.256 | 0.17 | 0.398 | 0 | 0.324 | 0 | 0 | 0.206 | 0 | 0 | 0 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 0.209 | 0 | 0 | 0.247 | 0.178 | 0.386 | 0 | 0.283 | 0 | 0 | 0.242 | 0 | 0 | 0 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 3.43 | 0.593 | 1.8 | 4.32 | 0.892 | 4.45 | 0.673 | 2.41 | 1.12 | 0.627 | 11.9 | 2.12 | 10.5 | 3.96 |
| 205STE020 | La Avenida | 17/Jun/2002 | 1.73 | 0.506 | 0.567 | 2.42 | 0.615 | 2.23 | 0 | 1.57 | 0.615 | 0.155 | 6.11 | 1.48 | 4.64 | 1.96 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | PCB 153 ng/g | PCB 156 ng/g | PCB 157 ng/g | PCB 158 ng/g | PCB 170 ng/g | PCB 174 ng/g | PCB 177 ng/g | PCB 180 ng/g | PCB 183 ng/g | PCB 187 ng/g | PCB 189 ng/g | PCB 194 ng/g | PCB 195 ng/g | PCB 200 ng/g |
|-------------|------------------------|-------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|-----------------|
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 0.148 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 204ALP010 | El Charro | 18/Sep/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 2.28 | 0 | 0 | 0 | 0 | 0 | 0 | 1.75 | 0 | 0.7 | 0 | 0 | 0 | 0 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 1.15 | 0 | 0 | 0 | 0.324 | 0.301 | 0 | 0.86 | 0 | 0.564 | 0 | 0.294 | 0 | 0 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 2.51 | 0.314 | 0 | 0.334 | 0.843 | 1.04 | 0.552 | 2.23 | 0.537 | 1.2 | 0 | 0.628 | 0.206 | 0 |
| 207SUI020 | Rockville | 17/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 202PES050 | Water Lane | 19/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 0.181 | 0 | 0 | 0 | 0 | 0 | 0 | 0.308 | 0 | 0.209 | 0 | 0 | 0 | 0 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 12.1 | 0.731 | 0 | 0.956 | 3.71 | 4.46 | 2.96 | 9.2 | 2.52 | 5.69 | 0 | 2.36 | 0.94 | 0.454 |
| 205STE020 | La Avenida | 17/Jun/2002 | 5.25 | 0.495 | 0.117 | 0.622 | 2.17 | 2.71 | 1.52 | 5.48 | 1.42 | 2.9 | 0 | 1.37 | 0.59 | 0.172 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | PCB 201 ng/g | PCB 203 ng/g | PCB 206 ng/g | PCB 209 ng/g | PCB AROCLOR 1248 ng/g | PCB AROCLOR 1254 ng/g | PCB AROCLOR 1260 ng/g | Sum congeners ng/g |
|-------------|------------------------|-------------|-----------------|-----------------|-----------------|-----------------|--------------------------|--------------------------|--------------------------|-----------------------|
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.652 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.481 |
| 204ALP010 | El Charro | 18/Sep/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2.971 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 28.042 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 0.483 | 0.422 | 0.381 | 0.348 | 0 | 0 | 0 | 15.984 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.753 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 0.723 | 0.523 | 0.425 | 0.287 | 38 | 21 | 23 | 44.583 |
| 207SUI020 | Rockville | 17/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1.621 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3.71 |
| 202PES050 | Water Lane | 19/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3.151 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 0.272 | 0.261 | 0 | 0 | 0 | 0 | 7 | 3.284 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 2.53 | 1.82 | 0.894 | 0 | 37 | 42 | 86 | 114.579 |
| 205STE020 | La Avenida | 17/Jun/2002 | 1.4 | 0.929 | 0.356 | 0 | 0 | 23 | 49 | 55.962 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | Sum of 18 PCB | PCB mean PEC-Q | Aldrin | Chlordane, cis | Chlordane, trans | Chlordene, alpha | Chlordene, gamma |
|-------------|------------------------|-------------|------------------|----------------|--------|----------------|------------------|------------------|------------------|
| | | | Congeners x 2.01 | | ng/g | ng/g | ng/g | ng/g | ng/g |
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 3.791 | 0.006 | 0 | 0 | 0 | 0 | 0 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 2.836 | 0.004 | 0 | 0 | 0 | 0 | 0 |
| 204ALP010 | El Charro | 18/Sep/2001 | 4.319 | 0.006 | 0 | 0 | 0 | 0 | 0 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 40.508 | 0.060 | 0 | 4.72 | 5.38 | 0 | 0 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 20.474 | 0.030 | 0 | 2.37 | 2.17 | 0 | 0 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 3.509 | 0.005 | 0 | 0 | 0 | 0 | 0 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 47.967 | 0.071 | 0 | 1.56 | 1.72 | 0 | 0 |
| 207SUI020 | Rockville | 17/Jun/2002 | 3.857 | 0.006 | 0 | 0 | 0 | 0 | 0 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 5.714 | 0.008 | 0 | 0 | 0 | 0 | 0 |
| 202PES050 | Water Lane | 19/Jun/2002 | 5.168 | 0.008 | 0 | 0 | 0 | 0 | 0 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 5.367 | 0.008 | 0 | 0 | 0 | 0 | 0 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 127.820 | 0.189 | 0 | 10 | 16.5 | 2.74 | 2.16 |
| 205STE020 | La Avenida | 17/Jun/2002 | 60.111 | 0.089 | 0 | 0 | 0.745 | 0 | 0 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | sum Chlordanes ng/g | Chlorpyrifos ng/g | Dacthal ng/g | DCBP(p,p') ng/g | DDD(o,p') ng/g | DDD(p,p') ng/g | sum DDD ng/g | DDE(o,p') ng/g | DDE(p,p') ng/g | sum DDE ng/g | DDMU(p,p') ng/g |
|-------------|------------------------|-------------|------------------------|----------------------|-----------------|--------------------|-------------------|-------------------|-----------------|-------------------|-------------------|-----------------|--------------------|
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 204ALP010 | El Charro | 18/Sep/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 15.08 | 0 | 5.35 | 0 | 0 | 15.9 | 15.9 | 0 | 16.2 | 16.2 | 0 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 6.74 | 0 | 0 | 0 | 0 | 2.84 | 2.84 | 0 | 2.34 | 2.34 | 0 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 4.91 | 2.95 | 0 | 0 | 0 | 1.75 | 1.75 | 0 | 1.52 | 1.52 | 0 |
| 207SUI020 | Rockville | 17/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2.01 | 2.01 | 0 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 202PES050 | Water Lane | 19/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 0 | 0 | 1.12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 55.1 | 0 | 0 | 0 | 3.4 | 12.7 | 16.1 | 0 | 30 | 30 | 0 |
| 205STE020 | La Avenida | 17/Jun/2002 | 1.366 | 0 | 0 | 0 | 0 | 1.05 | 1.05 | 0 | 3.23 | 3.23 | 0 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | DDT(o,p') | DDT(p,p') | sum DDT op,pp | sum DDT all | sum DDT all | Diazinon | Dieldrin | Endosulfan I | Endosulfan II | Endosulfan sulfate |
|-------------|------------------------|-------------|-----------|-----------|---------------|-------------|-------------|----------|----------|--------------|---------------|--------------------|
| | | | ng/g | ng/g | ng/g | ng/g | ng/g OC | ng/g | ng/g | ng/g | ng/g | ng/g |
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 0 | 0 | 0 | 0 | 0.00 | 0 | 0 | 0 | 0 | 0 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 0 | 0 | 0 | 0 | 0.00 | 0 | 0 | 0 | 0 | 0 |
| 204ALP010 | El Charro | 18/Sep/2001 | 0 | 0 | 0 | 0 | 0.00 | 0 | 0 | 0 | 0 | 0 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 0 | 0 | 0 | 32.1 | 340.76 | 0 | 0 | 0 | 0 | 0 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 0 | 0 | 0 | 5.18 | 177.40 | 0 | 1.85 | 0 | 0 | 0 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 0 | 0 | 0 | 0 | 0.00 | 0 | 0 | 0 | 0 | 0 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 0 | 0 | 0 | 3.27 | 182.68 | 0 | 0 | 0 | 0 | 0 |
| 207SUI020 | Rockville | 17/Jun/2002 | 0 | 0 | 0 | 2.01 | 773.08 | 0 | 0 | 0 | 0 | 0 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 0 | 0 | 0 | 0 | 0.00 | 0 | 0 | 0 | 0 | 0 |
| 202PES050 | Water Lane | 19/Jun/2002 | 0 | 0 | 0 | 0 | 0.00 | 0 | 0 | 0 | 0 | 0 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 0 | 0 | 0 | 0 | 0.00 | 0 | 0 | 0 | 0 | 0 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 4.42 | 22.4 | 26.82 | 72.92 | 3284.68 | 0 | 5.79 | 0 | 0 | 0 |
| 205STE020 | La Avenida | 17/Jun/2002 | 0 | 0 | 0 | 4.28 | 594.44 | 0 | 0 | 0 | 0 | 0 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | Endrin ng/g | HCH, alpha ng/g | HCH, beta ng/g | HCH, delta ng/g | HCH, gamma ng/g | Heptachlor ng/g | Heptachlor epoxide ng/g | Hexachlorobenzene ng/g | Methoxychlor ng/g | Mirex ng/g |
|-------------|------------------------|-------------|----------------|--------------------|-------------------|--------------------|--------------------|--------------------|----------------------------|---------------------------|----------------------|---------------|
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 204ALP010 | El Charro | 18/Sep/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 1.29 | 0 | 0 | 0 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.672 | 0 | 0 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 207SUI020 | Rockville | 17/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 202PES050 | Water Lane | 19/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 1.7 | 0.918 | 0 | 0 |
| 205STE020 | La Avenida | 17/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | Nonachlor, cis ng/g | Nonachlor, trans ng/g | Oxadiazon ng/g | Oxychlorthane ng/g | Parathion, Ethyl ng/g | Parathion, Methyl ng/g | Tedion ng/g | Toxaphene ng/g | Acenaphthene ng/g |
|-------------|------------------------|-------------|------------------------|--------------------------|-------------------|-----------------------|--------------------------|---------------------------|----------------|-------------------|----------------------|
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2.74 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 204ALP010 | El Charro | 18/Sep/2001 | 0 | 0 | 3.64 | 0 | 0 | 0 | 0 | 0 | 0 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 0 | 4.98 | 5.93 | 0 | 6.17 | 0 | 44.4 | 0 | 7.05 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 0 | 2.2 | 4.08 | 0 | 0 | 0 | 0 | 0 | 0 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 0 | 1.63 | 267 | 0 | 0 | 0 | 0 | 0 | 3.13 |
| 207SUI020 | Rockville | 17/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 202PES050 | Water Lane | 19/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 5.36 | 21.7 | 18.4 | 1.54 | 0 | 0 | 8.65 | 0 | 8.2 |
| 205STE020 | La Avenida | 17/Jun/2002 | 0 | 0.621 | 6.6 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | Acenaphthylene ng/g | Anthracene ng/g | Benz[a]anthracene ng/g | Benzo(a)pyrene ng/g | Benzo(b)fluoranthene ng/g | Benzo(e)pyrene ng/g | Benzo(g,h,i)perylene ng/g |
|-------------|------------------------|-------------|------------------------|--------------------|---------------------------|------------------------|------------------------------|------------------------|------------------------------|
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 0 | 0 | 2.35 | 2.57 | 13.8 | 6.56 | 5.61 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 0 | 0 | 0 | 1.69 | 6.19 | 3.16 | 3.35 |
| 204ALP010 | El Charro | 18/Sep/2001 | 0 | 0 | 2.38 | 6.72 | 9.24 | 4.86 | 7.7 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 6.54 | 12.45 | 53 | 0 | 106.67 | 0 | 61.89 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 0 | 5.06 | 14.4 | 29.6 | 33.3 | 27.9 | 33.3 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 0 | 0 | 0 | 0 | 2.60 | 0 | 6.56 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 0 | 5.95 | 58.8 | 40.4 | 69.0 | 33.6 | 41.7 |
| 207SUI020 | Rockville | 17/Jun/2002 | 0 | 0 | 2.53 | 1.71 | 4.45 | 1.93 | 5.33 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 0 | 0 | 0 | 0 | 1.46 | 0 | 3.22 |
| 202PES050 | Water Lane | 19/Jun/2002 | 0 | 0 | 0 | 0 | 1.36 | 0 | 2.27 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 0 | 0 | 0 | 1.79 | 2.73 | 1.73 | 4.51 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 3.95 | 32.3 | 149 | 278 | 351 | 212 | 293 |
| 205STE020 | La Avenida | 17/Jun/2002 | 0 | 0 | 8.43 | 15.2 | 31.0 | 24.6 | 26.7 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | Benzo(k)fluoranthene ng/g | Biphenyl ng/g | Chrysene ng/g | Chrysenes, C1- ng/g | Chrysenes, C2- ng/g | Chrysenes, C3- ng/g | Dibenz(a,h)anthracene ng/g |
|-------------|------------------------|-------------|------------------------------|------------------|------------------|------------------------|------------------------|------------------------|-------------------------------|
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 0 | 23.3 | 11.6 | 9.27 | 11.8 | 16.1 | 4.23 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 0 | 4.91 | 4.55 | 4.74 | 3.93 | 3.15 | 0 |
| 204ALP010 | El Charro | 18/Sep/2001 | 0 | 0 | 5.4 | 4.43 | 5.83 | 6.7 | 4.07 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 59.95 | 3.53 | 50.39 | 119 | 100 | 123 | 9.29 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 12.4 | 0 | 27.6 | 24.5 | 26.2 | 36.8 | 11.2 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 0 | 4.65 | 3.83 | 2.62 | 1.65 | 0 | 14.1 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 20.4 | 2.16 | 28.2 | 40.9 | 47.5 | 51.8 | 21.3 |
| 207SUI020 | Rockville | 17/Jun/2002 | 0 | 0 | 2.99 | 1.8 | 0 | 0 | 5.37 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 0 | 0 | 1.81 | 1.82 | 1.49 | 0 | 5.68 |
| 202PES050 | Water Lane | 19/Jun/2002 | 0 | 0 | 2.14 | 2.88 | 4.04 | 4.43 | 3.18 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 0 | 0 | 2.49 | 2.83 | 3.08 | 3.32 | 2.90 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 131 | 4.8 | 204 | 107 | 75.9 | 74.8 | 86.8 |
| 205STE020 | La Avenida | 17/Jun/2002 | 8.21 | 3.22 | 26.5 | 28.3 | 48.3 | 67.1 | 14.8 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | Dibenzothiophene | Dibenzothiophenes, C1- | Dibenzothiophenes, C2- | Dibenzothiophenes, C3- | Dimethylnaphthalene, 2,6- |
|-------------|------------------------|-------------|------------------|------------------------|------------------------|------------------------|---------------------------|
| | | | ng/g | ng/g | ng/g | ng/g | ng/g |
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 0 | 14.9 | 10.4 | 10.7 | 8.02 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 0 | 7.34 | 5.93 | 6.32 | 7.15 |
| 204ALP010 | El Charro | 18/Sep/2001 | 4.49 | 12.3 | 12.1 | 13.3 | 5.81 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 38.34 | 142 | 418 | 694 | 3.57 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 0 | 70.6 | 73.8 | 65.8 | 4.64 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 0 | 0 | 0 | 0 | 4.26 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 5.60 | 10 | 23.2 | 22.1 | 4.36 |
| 207SUI020 | Rockville | 17/Jun/2002 | 0 | 0 | 0 | 0 | 2.73 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 0 | 0 | 1.42 | 0 | 0 |
| 202PES050 | Water Lane | 19/Jun/2002 | 0 | 0 | 2.3 | 1.87 | 0 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 0 | 0 | 1.57 | 0 | 0 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 19.7 | 15.3 | 29 | 27.1 | 9.17 |
| 205STE020 | La Avenida | 17/Jun/2002 | 3.6 | 7.04 | 14.4 | 10.1 | 4.54 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | Fluoranthene ng/g | Fluoranthene/Pyrenes, C1- ng/g | Fluorene ng/g | Fluorenes, C1- ng/g | Fluorenes, C2- ng/g | Fluorenes, C3- ng/g | Indeno(1,2,3-c,d)pyrene ng/g |
|-------------|------------------------|-------------|----------------------|-----------------------------------|------------------|------------------------|------------------------|------------------------|---------------------------------|
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 5.86 | 7.55 | 3.48 | 4.47 | 11.8 | 7.88 | 3.4 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 2.94 | 7.56 | 2.16 | 4.56 | 8.41 | 7.53 | 2.12 |
| 204ALP010 | El Charro | 18/Sep/2001 | 6.97 | 3.56 | 0 | 0 | 4.37 | 0 | 7.22 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 104.54 | 303 | 7.92 | 7.1 | 0 | 0 | 71.89 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 44.7 | 26.8 | 0 | 4.92 | 7.44 | 7.28 | 30.8 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 1.86 | 3.02 | 0 | 1.92 | 0 | 5.41 | 9.56 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 66.0 | 36.5 | 5.70 | 4.91 | 0 | 17.4 | 60.5 |
| 207SUI020 | Rockville | 17/Jun/2002 | 4.55 | 2.47 | 0 | 0 | 0 | 2.54 | 7.10 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 1.97 | 1.83 | 0 | 0 | 0 | 1.98 | 4.94 |
| 202PES050 | Water Lane | 19/Jun/2002 | 1.43 | 2.34 | 0 | 0 | 0 | 3.03 | 2.81 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 2.50 | 2.45 | 0 | 0 | 0 | 2.25 | 5.22 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 468 | 113 | 10.5 | 5.12 | 0 | 35 | 377 |
| 205STE020 | La Avenida | 17/Jun/2002 | 18.0 | 32.7 | 2.18 | 3.46 | 0 | 11.8 | 29.6 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | Methylnaphthalene, 1- | Methylnaphthalene, 2- | Methylphenanthrene, 1- | Naphthalene | Naphthalenes, C1- | Naphthalenes, C2- |
|-------------|------------------------|-------------|-----------------------|-----------------------|------------------------|-------------|-------------------|-------------------|
| | | | ng/g | ng/g | ng/g | ng/g | ng/g | ng/g |
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 5.79 | 20.7 | 3.42 | 13.8 | 27.8 | 20.1 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 5.67 | 9.4 | 4.07 | 6.2 | 17.1 | 23.6 |
| 204ALP010 | El Charro | 18/Sep/2001 | 0 | 0 | 0 | 3.08 | 0 | 8.52 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 5.6 | 10.47 | 49.36 | 16.81 | 17.3 | 12.9 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 4.61 | 10.6 | 4.35 | 15.1 | 19.4 | 17.6 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 4.03 | 6.11 | 2.01 | 4.98 | 11.4 | 14 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 3.58 | 7.10 | 6.99 | 11.6 | 11.3 | 16.9 |
| 207SUI020 | Rockville | 17/Jun/2002 | 2.64 | 4.28 | 0 | 3.04 | 8 | 10.4 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 0 | 0 | 0 | 1.46 | 2.34 | 3.38 |
| 202PES050 | Water Lane | 19/Jun/2002 | 0 | 0 | 0 | 1.76 | 2.41 | 3.82 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 0 | 0 | 0 | 0 | 2.2 | 3.74 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 4.86 | 11.7 | 20.0 | 17.7 | 17.1 | 23 |
| 205STE020 | La Avenida | 17/Jun/2002 | 2.44 | 7.46 | 1.94 | 5.97 | 10.6 | 12.6 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | Naphthalenes, C3- | Naphthalenes, C4- | Perylene | Phenanthrene | Phenanthrene-Anthracene, C1- | Phenanthrene-Anthracene, C2- |
|-------------|------------------------|-------------|-------------------|-------------------|----------|--------------|------------------------------|------------------------------|
| | | | ng/g | ng/g | ng/g | ng/g | ng/g | ng/g |
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 8.29 | 3.59 | 4.82 | 38.1 | 27.9 | 13.3 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 12.7 | 2.01 | 8.33 | 16.5 | 4.07 | 15.3 |
| 204ALP010 | El Charro | 18/Sep/2001 | 0 | 0 | 2.94 | 5.33 | 5.6 | 8.65 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 25 | 6.75 | 0 | 68.87 | 142 | 379 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 11.6 | 0 | 12.2 | 27.7 | 22.5 | 24.1 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 10.6 | 2.08 | 3.14 | 14.8 | 16.2 | 10.4 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 21.2 | 7.87 | 17.1 | 46.8 | 40.8 | 62 |
| 207SUI020 | Rockville | 17/Jun/2002 | 9.68 | 2.77 | 4.34 | 4.63 | 5.19 | 3.9 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 5.02 | 1.97 | 11.6 | 2.47 | 3.69 | 5.05 |
| 202PES050 | Water Lane | 19/Jun/2002 | 6.15 | 2.49 | 8.44 | 2.63 | 5.47 | 7.16 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 5.16 | 1.58 | 11.9 | 2.52 | 3.5 | 4.23 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 17.1 | 5.77 | 39.3 | 206 | 106 | 73.2 |
| 205STE020 | La Avenida | 17/Jun/2002 | 14.3 | 3.81 | 10.5 | 12.7 | 14.3 | 18.4 |

Appendix H: Sediment chemistry results

| StationCode | StationName | SampleDate | Phenanthrene-Anthracene, C3- ng/g | Phenanthrene-Anthracene, C4- ng/g | Pyrene ng/g | Trimethylnaphthalene, 2,3,5- ng/g |
|-------------|------------------------|-------------|--------------------------------------|--------------------------------------|----------------|--------------------------------------|
| 201LAG270 | Creamery Gulch | 01/Oct/2001 | 6 | 9.25 | 5.77 | 0 |
| 201WLK090 | Walker Creek | 02/Oct/2001 | 8.52 | 5.68 | 3.99 | 1.85 |
| 204ALP010 | El Charro | 18/Sep/2001 | 4.92 | 3.67 | 7.65 | 0 |
| 204SLE030 | Empire Road | 19/Sep/2001 | 521 | 388 | 176.18 | 0 |
| 206SPA020 | 3rd Ave Bridge | 26/Sep/2001 | 13.4 | 23.9 | 50.3 | 0 |
| 201LAG120 | Green Bridge (Hwy 1) | 18/Jun/2002 | 3.67 | 0 | 2.58 | 1.86 |
| 206WIL020 | Richmond Parkway | 17/Jun/2002 | 56 | 19.2 | 56.1 | 4.34 |
| 207SUI020 | Rockville | 17/Jun/2002 | 2.17 | 0 | 4.46 | 0 |
| 202BUT010 | Bean Hollow | 19/Jun/2002 | 3.53 | 0 | 2.40 | 0 |
| 202PES050 | Water Lane | 19/Jun/2002 | 6.72 | 2.47 | 2.12 | 0 |
| 202SGR010 | San Gregorio USGS Gage | 18/Jun/2002 | 3.81 | 1.6 | 3.29 | 0 |
| 205PER010 | Charleston Rd | 17/Jun/2002 | 56.3 | 19.4 | 395 | 3.76 |
| 205STE020 | La Avenida | 17/Jun/2002 | 23.5 | 16.6 | 36.3 | 2.36 |

Appendix I-1: Summary of metals found in clam tissue samples

| Station ID | Analyte | wet weight | units | SMW 85% EDL |
|----------------|-----------------|--------------|-------------|---------------|
| <i>Control</i> | <i>Aluminum</i> | <i>60.12</i> | <i>ug/g</i> | <i>206.33</i> |
| 201LAG040 | Aluminum | 5.17 | ug/g | |
| 201LAG130 | Aluminum | 10.24 | ug/g | |
| 201LAG270 | Aluminum | 28.74 | ug/g | |
| 201WLK090 | Aluminum | 25.92 | ug/g | |
| 202BUT010 | Aluminum | 53.22 | ug/g | |
| 202PES050 | Aluminum | 72.79 | ug/g | |
| 202SGR010 | Aluminum | 55.42 | ug/g | |
| 204ALP010 | Aluminum | 7.85 | ug/g | |
| 205PER010 | Aluminum | 14.15 | ug/g | |
| 205STE020 | Aluminum | 16.04 | ug/g | |
| 206SPA020 | Aluminum | 43.25 | ug/g | |
| 207SUI020 | Aluminum | 97.83 | ug/g | |
| <i>Control</i> | <i>Arsenic</i> | <i>0.80</i> | <i>ug/g</i> | <i>0.9</i> |
| 201LAG040 | Arsenic | 0.69 | ug/g | |
| 201LAG130 | Arsenic | 0.61 | ug/g | |
| 201LAG270 | Arsenic | 0.65 | ug/g | |
| 201WLK090 | Arsenic | 0.59 | ug/g | |
| 202BUT010 | Arsenic | 0.59 | ug/g | |
| 202PES050 | Arsenic | 0.53 | ug/g | |
| 202SGR010 | Arsenic | 0.56 | ug/g | |
| 204ALP010 | Arsenic | 0.66 | ug/g | |
| 205PER010 | Arsenic | 0.70 | ug/g | |
| 205STE020 | Arsenic | 0.79 | ug/g | |
| 206SPA020 | Arsenic | 0.67 | ug/g | |
| 207SUI020 | Arsenic | 0.78 | ug/g | |
| <i>Control</i> | <i>Cadmium</i> | <i>0.10</i> | <i>ug/g</i> | <i>0.92</i> |
| 201LAG040 | Cadmium | 0.12 | ug/g | |
| 201LAG130 | Cadmium | 0.08 | ug/g | |
| 201LAG270 | Cadmium | 0.10 | ug/g | |
| 201WLK090 | Cadmium | 0.07 | ug/g | |
| 202BUT010 | Cadmium | 0.11 | ug/g | |
| 202PES050 | Cadmium | 0.10 | ug/g | |
| 202SGR010 | Cadmium | 0.11 | ug/g | |
| 204ALP010 | Cadmium | 0.08 | ug/g | |
| 205PER010 | Cadmium | 0.08 | ug/g | |
| 205STE020 | Cadmium | 0.12 | ug/g | |
| 206SPA020 | Cadmium | 0.08 | ug/g | |
| 207SUI020 | Cadmium | 0.11 | ug/g | |
| <i>Control</i> | <i>Chromium</i> | <i>0.50</i> | <i>ug/g</i> | <i>2</i> |
| 201LAG040 | Chromium | 0.30 | ug/g | |
| 201LAG130 | Chromium | 0.28 | ug/g | |
| 201LAG270 | Chromium | 0.51 | ug/g | |
| 201WLK090 | Chromium | 0.30 | ug/g | |
| 202BUT010 | Chromium | 0.40 | ug/g | |
| 202PES050 | Chromium | 0.36 | ug/g | |
| 202SGR010 | Chromium | 0.39 | ug/g | |
| 204ALP010 | Chromium | 0.28 | ug/g | |
| 205PER010 | Chromium | 0.43 | ug/g | |
| 205STE020 | Chromium | 0.48 | ug/g | |
| 206SPA020 | Chromium | 0.34 | ug/g | |
| 207SUI020 | Chromium | 0.53 | ug/g | |

Appendix I-1: Summary of metals found in clam tissue samples

| Station ID | Analyte | wet weight | units | SMW 85% EDL |
|----------------|------------------------|------------|-------|-------------|
| <i>Control</i> | <i>Copper</i> | 8.46 | ug/g | 8.78 |
| 201LAG040 | Copper | 7.72 | ug/g | |
| 201LAG130 | Copper | 6.72 | ug/g | |
| 201LAG270 | Copper | 7.68 | ug/g | |
| 201WLK090 | Copper | 5.22 | ug/g | |
| 202BUT010 | Copper | 6.88 | ug/g | |
| 202PES050 | Copper | 6.63 | ug/g | |
| 202SGR010 | Copper | 7.10 | ug/g | |
| 204ALP010 | Copper | 7.34 | ug/g | |
| 205PER010 | Copper | 9.54 | ug/g | |
| 205STE020 | Copper | 9.95 | ug/g | |
| 206SPA020 | Copper | 6.96 | ug/g | |
| 207SUI020 | Copper | 15.33 | ug/g | |
| <i>Control</i> | <i>Lead</i> | 0.05 | ug/g | 0.21 |
| 201LAG040 | Lead | 0.01 | ug/g | |
| 201LAG130 | Lead | 0.01 | ug/g | |
| 201LAG270 | Lead | 0.03 | ug/g | |
| 201WLK090 | Lead | 0.01 | ug/g | |
| 202BUT010 | Lead | 0.02 | ug/g | |
| 202PES050 | Lead | 0.02 | ug/g | |
| 202SGR010 | Lead | 0.02 | ug/g | |
| 204ALP010 | Lead | 0.03 | ug/g | |
| 205PER010 | Lead | 0.06 | ug/g | |
| 205STE020 | Lead | 0.04 | ug/g | |
| 206SPA020 | Lead | 0.08 | ug/g | |
| 207SUI020 | Lead | 0.05 | ug/g | |
| <i>Control</i> | <i>Manganese</i> | 3.61 | ug/g | 9.55 |
| 201LAG040 | Manganese | 2.21 | ug/g | |
| 201LAG130 | Manganese | 2.32 | ug/g | |
| 201LAG270 | Manganese | 2.55 | ug/g | |
| 201WLK090 | Manganese | 5.09 | ug/g | |
| 202BUT010 | Manganese | 6.64 | ug/g | |
| 202PES050 | Manganese | 6.53 | ug/g | |
| 202SGR010 | Manganese | 5.76 | ug/g | |
| 204ALP010 | Manganese | 0.79 | ug/g | |
| 205PER010 | Manganese | 1.22 | ug/g | |
| 205STE020 | Manganese | 2.16 | ug/g | |
| 206SPA020 | Manganese | 5.80 | ug/g | |
| 207SUI020 | Manganese | 6.16 | ug/g | |
| <i>Control</i> | <i>Mercury (total)</i> | 0.02 | ug/g | 0.04 |
| 201LAG040 | Mercury (total) | 0.02 | ug/g | |
| 201LAG130 | Mercury (total) | 0.02 | ug/g | |
| 201LAG270 | Mercury (total) | 0.03 | ug/g | |
| 201WLK090 | Mercury (total) | 0.01 | ug/g | |
| 202BUT010 | Mercury (total) | 0.02 | ug/g | |
| 202PES050 | Mercury (total) | 0.01 | ug/g | |
| 202SGR010 | Mercury (total) | 0.02 | ug/g | |
| 204ALP010 | Mercury (total) | 0.02 | ug/g | |
| 205PER010 | Mercury (total) | 0.02 | ug/g | |
| 205STE020 | Mercury (total) | 0.02 | ug/g | |
| 206SPA020 | Mercury (total) | 0.02 | ug/g | |
| 207SUI020 | Mercury (total) | 0.03 | ug/g | |
| <i>Control</i> | <i>Nickel</i> | 0.34 | ug/g | 1 |
| 201LAG040 | Nickel | 0.14 | ug/g | |
| 201LAG130 | Nickel | 0.16 | ug/g | |
| 201LAG270 | Nickel | 0.51 | ug/g | |
| 201WLK090 | Nickel | 0.25 | ug/g | |
| 202BUT010 | Nickel | 0.36 | ug/g | |
| 202PES050 | Nickel | 0.34 | ug/g | |

Appendix I-1: Summary of metals found in clam tissue samples

| Station ID | Analyte | wet weight | units | SMW 85% EDL |
|----------------|-----------------|-------------|-------------|--------------|
| <i>Control</i> | <i>Nickel</i> | <i>0.34</i> | <i>ug/g</i> | <i>1</i> |
| 202SGR010 | Nickel | 0.43 | ug/g | |
| 204ALP010 | Nickel | 0.16 | ug/g | |
| 205PER010 | Nickel | 0.15 | ug/g | |
| 205STE020 | Nickel | 0.21 | ug/g | |
| 206SPA020 | Nickel | 0.32 | ug/g | |
| 207SUI020 | Nickel | 0.32 | ug/g | |
| <i>Control</i> | <i>Selenium</i> | <i>0.27</i> | <i>ug/g</i> | <i>0.43</i> |
| 201LAG040 | Selenium | 0.19 | ug/g | |
| 201LAG130 | Selenium | 0.18 | ug/g | |
| 201LAG270 | Selenium | 0.20 | ug/g | |
| 201WLK090 | Selenium | 0.16 | ug/g | |
| 202BUT010 | Selenium | 0.23 | ug/g | |
| 202PES050 | Selenium | 0.21 | ug/g | |
| 202SGR010 | Selenium | 0.23 | ug/g | |
| 204ALP010 | Selenium | 0.21 | ug/g | |
| 205PER010 | Selenium | 0.33 | ug/g | |
| 205STE020 | Selenium | 0.40 | ug/g | |
| 206SPA020 | Selenium | 0.27 | ug/g | |
| 207SUI020 | Selenium | 0.32 | ug/g | |
| <i>Control</i> | <i>Silver</i> | <i>0.01</i> | <i>ug/g</i> | <i>0.03</i> |
| 201LAG040 | Silver | 0.01 | ug/g | |
| 201LAG130 | Silver | 0.01 | ug/g | |
| 201LAG270 | Silver | 0.01 | ug/g | |
| 201WLK090 | Silver | 0.01 | ug/g | |
| 202BUT010 | Silver | 0.01 | ug/g | |
| 202PES050 | Silver | 0.01 | ug/g | |
| 202SGR010 | Silver | 0.01 | ug/g | |
| 204ALP010 | Silver | 0.01 | ug/g | |
| 205PER010 | Silver | 0.01 | ug/g | |
| 205STE020 | Silver | 0.02 | ug/g | |
| 206SPA020 | Silver | 0.01 | ug/g | |
| 207SUI020 | Silver | 0.01 | ug/g | |
| <i>Control</i> | <i>Zinc</i> | <i>9.90</i> | <i>ug/g</i> | <i>19.39</i> |
| 201LAG040 | Zinc | 7.26 | ug/g | |
| 201LAG130 | Zinc | 5.49 | ug/g | |
| 201LAG270 | Zinc | 7.69 | ug/g | |
| 201WLK090 | Zinc | 5.67 | ug/g | |
| 202BUT010 | Zinc | 7.54 | ug/g | |
| 202PES050 | Zinc | 7.31 | ug/g | |
| 202SGR010 | Zinc | 6.94 | ug/g | |
| 204ALP010 | Zinc | 6.90 | ug/g | |
| 205PER010 | Zinc | 7.76 | ug/g | |
| 205STE020 | Zinc | 8.88 | ug/g | |
| 206SPA020 | Zinc | 8.39 | ug/g | |
| 207SUI020 | Zinc | 9.54 | ug/g | |

Appendix I-2: PAH's detected in clam tissue samples (bold font indicates detected value)

| StationCode | SampleDate | Season | LabBatch | AnalysisDate | MethodName | AnalyteName | Unit | Result | ResultQualCode |
|-------------|-------------|--------|--------------|--------------|------------|-------------------------------|------|-------------|----------------|
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benz[a]anthracene | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(b)fluoranthene | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(e)pyrene | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(g,h,i)perylene | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysene | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C1 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C2 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C3 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenz(a,h)anthracene | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C1 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C2 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C3 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluoranthene | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluoranthene/Pyrenes, C1 - | ng/g | 26.0 | |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C1 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C2 - | ng/g | 19.3 | |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C3 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Indeno(1,2,3-c,d)pyrene | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalene | ng/g | 14.1 | |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C4 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | 44.2 | |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C2 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C3 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C4 - | ng/g | -13.9 | ND |
| 201LAG040 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Pyrene | ng/g | -13.9 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benz[a]anthracene | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(b)fluoranthene | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(e)pyrene | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(g,h,i)perylene | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysene | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C1 - | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C2 - | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C3 - | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenz(a,h)anthracene | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C1 - | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C2 - | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C3 - | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluoranthene | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluoranthene/Pyrenes, C1 - | ng/g | 35.0 | |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C1 - | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C2 - | ng/g | 26.5 | |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C3 - | ng/g | 73.4 | |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Indeno(1,2,3-c,d)pyrene | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalene | ng/g | 18.9 | |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C4 - | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | 68.6 | |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C2 - | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C3 - | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C4 - | ng/g | -17.5 | ND |
| 201LAG130 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Pyrene | ng/g | -17.5 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benz[a]anthracene | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(b)fluoranthene | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(e)pyrene | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(g,h,i)perylene | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysene | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C1 - | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C2 - | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C3 - | ng/g | -15.3 | ND |

| | | | | | | | | | |
|-----------|-------------|-----|--------------|-------------|-----------|-----------------------|------|-------|----|
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenz(a,h)anthracene | ng/g | -15.3 | ND |
|-----------|-------------|-----|--------------|-------------|-----------|-----------------------|------|-------|----|

Appendix I-2: PAH's detected in clam tissue samples (bold font indicates detected value)

| StationCode | SampleDate | Season | LabBatch | AnalysisDate | MethodName | AnalyteName | Unit | Result | ResultQualCode |
|-------------|-------------|--------|--------------|--------------|------------|-------------------------------|------|-------------|----------------|
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C1 - | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C2 - | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C3 - | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluoranthene | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluoranthene/Pyrenes, C1 - | ng/g | 25.5 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C1 - | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C2 - | ng/g | 22.5 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C3 - | ng/g | 75.5 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Indeno(1,2,3-c,d)pyrene | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalene | ng/g | 15.8 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C4 - | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | 77.2 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C2 - | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C3 - | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C4 - | ng/g | -15.3 | ND |
| 201LAG270 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Pyrene | ng/g | -15.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benz[a]anthracene | ng/g | 14.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(b)fluoranthene | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(e)pyrene | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(g,h,i)perylene | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysene | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C1 - | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C2 - | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C3 - | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenz(a,h)anthracene | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C1 - | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C2 - | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C3 - | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluoranthene | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluoranthene/Pyrenes, C1 - | ng/g | 30.0 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C1 - | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C2 - | ng/g | 25.1 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C3 - | ng/g | 71.4 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Indeno(1,2,3-c,d)pyrene | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalene | ng/g | 15.2 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C4 - | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | 39.7 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C2 - | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C3 - | ng/g | -13.3 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C4 - | ng/g | 61.6 | ND |
| 201WLK090 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Pyrene | ng/g | -13.3 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benz[a]anthracene | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(b)fluoranthene | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(e)pyrene | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(g,h,i)perylene | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysene | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C1 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C2 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C3 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenz(a,h)anthracene | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C1 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C2 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C3 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluoranthene | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluoranthene/Pyrenes, C1 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C1 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C2 - | ng/g | 17.4 | ND |

| | | | | | | | | | |
|-----------|-------------|-----|--------------|-------------|-----------|-------------------------|------|-------------|----|
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C3 - | ng/g | 54.8 | |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Indeno(1,2,3-c,d)pyrene | ng/g | -14.0 | ND |

Appendix I-2: PAH's detected in clam tissue samples (bold font indicates detected value)

| StationCode | SampleDate | Season | LabBatch | AnalysisDate | MethodName | AnalyteName | Unit | Result | ResultQualCode |
|-------------|-------------|--------|--------------|--------------|------------|-------------------------------|------|-------------|----------------|
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalene | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C4 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C2 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C3 - | ng/g | -14.0 | ND |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C4 - | ng/g | 21.6 | |
| 202BUT010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Pyrene | ng/g | -14.0 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benz[a]anthracene | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(b)fluoranthene | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(e)pyrene | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(g,h,i)perylene | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysene | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C1 - | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C2 - | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C3 - | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenz(a,h)anthracene | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C1 - | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C2 - | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C3 - | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluoranthene | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluoranthene/Pyrenes, C1 - | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C1 - | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C2 - | ng/g | 15.7 | |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C3 - | ng/g | 53.2 | |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Indeno(1,2,3-c,d)pyrene | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalene | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C4 - | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene | ng/g | -14.4 | ND |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | 24.6 | |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C2 - | ng/g | 26.8 | |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C3 - | ng/g | 15.0 | |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C4 - | ng/g | 25.6 | |
| 202PES050 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Pyrene | ng/g | -14.4 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benz[a]anthracene | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(b)fluoranthene | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(e)pyrene | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(g,h,i)perylene | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysene | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C1 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C2 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C3 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenz(a,h)anthracene | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C1 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C2 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C3 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluoranthene | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluoranthene/Pyrenes, C1 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C1 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C2 - | ng/g | 18.1 | |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C3 - | ng/g | 55.5 | |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Indeno(1,2,3-c,d)pyrene | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalene | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | 311 | |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | 152 | |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | -16.3 | ND |

| | | | | | | | | | |
|-----------|-------------|-----|--------------|-------------|-----------|--------------------|------|------------|----|
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | 161 | |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C4 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene | ng/g | -16.3 | ND |

Appendix I-2: PAH's detected in clam tissue samples (bold font indicates detected value)

| StationCode | SampleDate | Season | LabBatch | AnalysisDate | MethodName | AnalyteName | Unit | Result | ResultQualCode |
|-------------|-------------|--------|--------------|--------------|------------|-------------------------------|------|-------------|----------------|
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | 148 | |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C2 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C3 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C4 - | ng/g | -16.3 | ND |
| 202SGR010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Pyrene | ng/g | -16.3 | ND |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benz[a]anthracene | ng/g | -16.4 | ND |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(b)fluoranthene | ng/g | -16.4 | ND |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(e)pyrene | ng/g | -16.4 | ND |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(g,h,i)perylene | ng/g | -16.4 | ND |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysene | ng/g | 29.4 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C1 - | ng/g | 40.3 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C2 - | ng/g | -16.4 | ND |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C3 - | ng/g | -16.4 | ND |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenz(a,h)anthracene | ng/g | -16.4 | ND |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C1 - | ng/g | 38.4 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C2 - | ng/g | 42.6 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C3 - | ng/g | 20.2 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluoranthene | ng/g | 35.7 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluoranthene/Pyrenes, C1 - | ng/g | 36.5 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C1 - | ng/g | -16.4 | ND |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C2 - | ng/g | 30.6 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C3 - | ng/g | 63.4 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Indeno(1,2,3-c,d)pyrene | ng/g | -16.4 | ND |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalene | ng/g | 17.2 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | -16.4 | ND |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | -16.4 | ND |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | 23.7 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C4 - | ng/g | 16.7 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene | ng/g | 34.0 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | 57.7 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C2 - | ng/g | 32.9 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C3 - | ng/g | 24.8 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C4 - | ng/g | 64.5 | |
| 204ALP010 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Pyrene | ng/g | 33.1 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benz[a]anthracene | ng/g | 46.1 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(b)fluoranthene | ng/g | 17.0 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(e)pyrene | ng/g | 15.5 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(g,h,i)perylene | ng/g | 16.5 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysene | ng/g | 44.2 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C1 - | ng/g | 45.3 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C2 - | ng/g | 15.2 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C3 - | ng/g | -13.7 | ND |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenz(a,h)anthracene | ng/g | -13.7 | ND |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C1 - | ng/g | -13.7 | ND |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C2 - | ng/g | -13.7 | ND |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C3 - | ng/g | -13.7 | ND |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluoranthene | ng/g | 35.2 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluoranthene/Pyrenes, C1 - | ng/g | -13.7 | ND |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C1 - | ng/g | -13.7 | ND |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C2 - | ng/g | 24.9 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C3 - | ng/g | 93.0 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Indeno(1,2,3-c,d)pyrene | ng/g | -13.7 | ND |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalene | ng/g | 15.1 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | -13.7 | ND |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | -13.7 | ND |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | 13.8 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C4 - | ng/g | -13.7 | ND |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene | ng/g | 16.1 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | -13.7 | ND |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C2 - | ng/g | 30.7 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C3 - | ng/g | 23.9 | |
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C4 - | ng/g | 19.1 | |

| | | | | | | | | | |
|-----------|-------------|-----|--------------|-------------|-----------|----------------------|------|-------------|----|
| 205PER010 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Pyrene | ng/g | 41.6 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benz[a]anthracene | ng/g | 28.7 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(b)fluoranthene | ng/g | -12.5 | ND |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(e)pyrene | ng/g | -12.5 | ND |

Appendix I-2: PAH's detected in clam tissue samples (bold font indicates detected value)

| StationCode | SampleDate | Season | LabBatch | AnalysisDate | MethodName | AnalyteName | Unit | Result | ResultQualCode |
|-------------|-------------|--------|--------------|--------------|------------|-------------------------------|------|-------------|----------------|
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(g,h,i)perylene | ng/g | 14.7 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysene | ng/g | 24.7 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C1 - | ng/g | 32.8 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C2 - | ng/g | -12.5 | ND |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C3 - | ng/g | -12.5 | ND |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenz(a,h)anthracene | ng/g | 17.5 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C1 - | ng/g | -12.5 | ND |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C2 - | ng/g | -12.5 | ND |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C3 - | ng/g | -12.5 | ND |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluoranthene | ng/g | 22.2 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluoranthene/Pyrenes, C1 - | ng/g | -12.5 | ND |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C1 - | ng/g | -12.5 | ND |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C2 - | ng/g | 22.4 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C3 - | ng/g | 117 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Indeno(1,2,3-c,d)pyrene | ng/g | 15.3 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalene | ng/g | -12.5 | ND |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | -12.5 | ND |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | -12.5 | ND |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | -12.5 | ND |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C4 - | ng/g | -12.5 | ND |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene | ng/g | 12.7 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | 94.5 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C2 - | ng/g | 30.3 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C3 - | ng/g | 31.5 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C4 - | ng/g | 31.7 | |
| 205STE020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Pyrene | ng/g | 23.6 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benz[a]anthracene | ng/g | -14.4 | ND |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(b)fluoranthene | ng/g | -14.4 | ND |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(e)pyrene | ng/g | 17.5 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(g,h,i)perylene | ng/g | -14.4 | ND |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysene | ng/g | 69.8 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C1 - | ng/g | 75.3 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C2 - | ng/g | 28.9 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C3 - | ng/g | 17.2 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenz(a,h)anthracene | ng/g | -14.4 | ND |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C1 - | ng/g | 30.8 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C2 - | ng/g | 51.1 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C3 - | ng/g | 37.8 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluoranthene | ng/g | 96.2 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluoranthene/Pyrenes, C1 - | ng/g | 80.5 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C1 - | ng/g | -14.4 | ND |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C2 - | ng/g | 28.5 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C3 - | ng/g | 110 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Indeno(1,2,3-c,d)pyrene | ng/g | -14.4 | ND |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalene | ng/g | 17.7 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | -14.4 | ND |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | 283 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | 15.1 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | 146 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | 25.4 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | 169 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C4 - | ng/g | 17.8 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene | ng/g | 41.7 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | 72.5 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | 173 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C2 - | ng/g | 89.7 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C3 - | ng/g | 82.8 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C4 - | ng/g | 66.2 | |
| 206SPA020 | 28/Sep/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Pyrene | ng/g | 112 | |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benz[a]anthracene | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(b)fluoranthene | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(e)pyrene | ng/g | -13.1 | ND |

| | | | | | | | | | |
|-----------|-------------|-----|--------------|-------------|-----------|----------------------|------|-------|----|
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Benzo(g,h,i)perylene | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysene | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C1 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C2 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Chrysenes, C3 - | ng/g | -13.1 | ND |

Appendix I-2: PAH's detected in clam tissue samples (bold font indicates detected value)

| StationCode | SampleDate | Season | LabBatch | AnalysisDate | MethodName | AnalyteName | Unit | Result | ResultQualCode |
|-------------|-------------|--------|--------------|--------------|------------|-------------------------------|------|-------------|----------------|
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenz(a,h)anthracene | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C1 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C2 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Dibenzothiophenes, C3 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluoranthene | ng/g | 20.2 | |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluoranthene/Pyrenes, C1 - | ng/g | 18.4 | |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C1 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C2 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Fluorenes, C3 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Indeno(1,2,3-c,d)pyrene | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalene | ng/g | 19.5 | |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Naphthalenes, C4 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C2 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C3 - | ng/g | -13.1 | ND |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Phenanthrene/Anthracene, C4 - | ng/g | 25.3 | |
| 207BIGBRK | 28/Aug/2001 | Dry | L-040803-PAH | 16/Jul/2003 | EPA 8270M | Pyrene | ng/g | 19.4 | |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benz[a]anthracene | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(b)fluoranthene | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(e)pyrene | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Benzo(g,h,i)perylene | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysene | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C1 - | ng/g | 30.8 | |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C2 - | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Chrysenes, C3 - | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenz(a,h)anthracene | ng/g | 19.8 | |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C1 - | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C2 - | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Dibenzothiophenes, C3 - | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluoranthene | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluoranthene/Pyrenes, C1 - | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C1 - | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C2 - | ng/g | 32.1 | |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Fluorenes, C3 - | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Indeno(1,2,3-c,d)pyrene | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalene | ng/g | 15.1 | |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C1 - | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C2 - | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C3 - | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Naphthalenes, C4 - | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C1 - | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C2 - | ng/g | 30.0 | |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C3 - | ng/g | 19.9 | |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Phenanthrene/Anthracene, C4 - | ng/g | -12.7 | ND |
| 207SUI020 | 17/Jul/2002 | Dry | L-032803-PAH | 23/Jun/2003 | EPA 8270M | Pyrene | ng/g | -12.7 | ND |

Appendix I-3: Summary of pesticides found in clam tissue (by site)

| Analyte | Units | 28/Aug/2001 | 28/Sep/2001 | 28/Sep/2001 | 28/Sep/2001 | 28/Sep/2001 | 28/Sep/2001 | 17/Jul/2002 | 28/Sep/2001 | 17/Jul/2002 | 17/Jul/2002 | 17/Jul/2002 | 17/Jul/2002 |
|--------------------|---------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|
| | | Control | 201WLK090 | 201LAG040 | 201LAG130 | 201LAG270 | 206SPA020 | 207SUI020 | 204ALP010 | 202BUT010 | 202PES050 | 202SGR010 | 205PER010 |
| chlordane, cis | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.90 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.61 |
| chlordane, trans | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.76 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1.81 |
| chlordene, alpha | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| chlordene, gamma | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| oxychlordane | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.60 |
| nonachlor, cis | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| nonachlor, trans | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.40 | 0.97 | 0.00 | 0.39 | 0.00 | 0.00 | 0.00 | 2.83 |
| total chlordanes | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.40 | 2.63 | 0.00 | 0.39 | 0.00 | 0.00 | 0.00 | 7.85 |
| DDD, o,p' | ng/g ww | 0.82 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| DDD, p,p' | ng/g ww | 2.20 | 0.00 | 0.91 | 0.96 | 0.00 | 1.60 | 1.81 | 0.00 | 0.00 | 0.00 | 0.00 | 2.04 |
| DDE, o,p' | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| DDE, p,p' | ng/g ww | 19.78 | 12.18 | 14.33 | 14.58 | 10.13 | 8.36 | 17.24 | 9.91 | 5.80 | 5.32 | 7.02 | 11.01 |
| DDT, o,p' | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| DDT, p,p' | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.46 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| DDMU, p,p' | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Total DDTs | ng/g ww | 22.81 | 12.18 | 15.24 | 15.53 | 10.13 | 9.96 | 21.51 | 9.91 | 5.80 | 5.32 | 7.02 | 13.05 |
| chlorpyrifos | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| diazinon | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| aldrin | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| dacthal | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| DCBP, p,p' | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| dieldrin | ng/g ww | 0.52 | 0.00 | 0.00 | 0.00 | 0.47 | 1.14 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.72 |
| endosulfan I | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| endosulfan II | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| endosulfan sulfate | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| total endosulfan | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| endrin | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| HCH, alpha | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| HCH, beta | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| HCH, delta | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| HCH, gamma | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| heptachlor | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| heptachlor epoxide | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| hexachlorobenzene | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| methoxychlor | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| mirex | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| oxadiazon | ng/g ww | 2.81 | 0.00 | 0.00 | 0.00 | 0.00 | 6.03 | 3.00 | 24.53 | 0.00 | 0.00 | 3.35 | 1.22 |
| parathion, ethyl | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| parathion, methyl | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| tedion | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| toxaphene | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| PCB 1248 | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 13.10 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| PCB 1254 | ng/g ww | 6.35 | 5.10 | 5.51 | 6.11 | 5.19 | 7.55 | 4.68 | 7.11 | 5.81 | 4.14 | 0.00 | 17.82 |
| PCB 1260 | ng/g ww | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 5.25 |
| Total PCBs | ng/g ww | 6.35 | 5.10 | 5.51 | 6.11 | 5.19 | 7.55 | 17.78 | 7.11 | 5.81 | 4.14 | 0.00 | 23.08 |
| % Moisture | % | 92.70 | 94.00 | 94.20 | 94.50 | 93.90 | 93.20 | 92.20 | 94.40 | 93.00 | 93.10 | 94.00 | 92.90 |
| % Lipid | % | 0.54 | 0.40 | 0.43 | 0.39 | 0.40 | 0.50 | 0.74 | 0.42 | 0.61 | 0.62 | 0.52 | 0.68 |

| 17/Jul/2002 |
|-------------|
| 205STE020 |
| 1.09 |
| 1.02 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 1.20 |
| 3.31 |
| 0.00 |
| 1.14 |
| 0.00 |
| 11.76 |
| 0.00 |
| 0.00 |
| 0.00 |
| 12.90 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.42 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 2.71 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 0.00 |
| 18.32 |
| 0.00 |
| 18.32 |
| 92.00 |
| 0.75 |

Appendix I-4: Summary of detections of PCB's in clam tissue (bold font indicates detection)

| Analyte (PCB Congener) | Control | 201WLK090 | 210LAG040 | 201LAG130 | 201LAG270 | 206SPA020 | 204ALP010 | 202BUT010 | 202PES050 | 202SGR010 | 205PER010 | 205STE020 |
|------------------------|--------------|--------------|--------------|--------------|-------------|-------------|-------------|--------------|-------------|--------------|--------------|--------------|
| 8 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | -1.25175 |
| 18 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | 3.12 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | 1.38 |
| 27 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | -1.25175 |
| 28 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | 2.96 | 1.93 | -1.4115 | -1.43025 | -1.606 | 1.82 | 2.32 |
| 29 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | -1.25175 |
| 31 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | 3.35 | -1.767 | -1.4115 | -1.43025 | -1.606 | 1.82 | 2.57 |
| 33 | -1.3395 | -1.627 | -1.64475 | -1.73875 | 1.43 | 2.16 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | -1.25175 |
| 44 | 3.45 | 3.91 | 4.01 | 4.14 | 3.51 | 7.14 | 4.76 | 3.55 | 3.09 | 3.41 | 5.97 | 12.4 |
| 49 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | 2.56 | -1.767 | -1.4115 | -1.43025 | -1.606 | 2.17 | 4.52 |
| 52 | 2.1 | 1.75 | 2.6 | 2.31 | 1.96 | 4.47 | 3.1 | 1.77 | -1.43025 | 2.34 | 4.9 | 10.1 |
| 56 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | 2.05 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | 2.92 |
| 60 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | 1.46 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | 2.09 |
| 66 | 2.26 | 2.01 | 2.1 | 2.73 | 2.81 | 5.55 | 3.34 | 2.13 | 2.11 | -1.606 | 7.94 | 8.6 |
| 70 | 2.64 | 2.28 | 2.21 | 3.1 | 2.16 | 4.87 | 2.45 | 2.61 | -1.43025 | -1.606 | 2.68 | 7.5 |
| 74 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | 2.09 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | 4.02 |
| 87 | 1.74 | 1.71 | -1.64475 | 2.43 | 1.51 | 3.12 | 2.85 | 1.65 | -1.43025 | -1.606 | 2.79 | 6.16 |
| 95 | 5.05 | 5.06 | 5.26 | 6.01 | 4.67 | 7.64 | 7.57 | 4.84 | 4.12 | 4.32 | 13.2 | 12.8 |
| 97 | 2.38 | 2.73 | 2.76 | 3.1 | 2.26 | 3.76 | 4.26 | 2.49 | 2.02 | 2.04 | 3.58 | 6.69 |
| 99 | 2.85 | 2.43 | 2.62 | 2.98 | 2.21 | 3.68 | 3.47 | 1.76 | -1.43025 | -1.606 | 7.84 | 6.37 |
| 101 | 5.81 | 5.34 | 5.7 | 6.59 | 5.23 | 8.5 | 8.6 | 4.35 | 2.79 | 3.19 | 16 | 15.2 |
| 105 | 2.31 | 2.04 | 2.37 | 3.32 | 2.15 | 3.4 | 3.5 | 2.09 | 1.55 | -1.606 | 2.67 | 6.43 |
| 110 | 6.17 | 5.8 | 6.24 | 7.32 | 5.98 | 9.36 | 9.42 | 4.98 | 3.43 | 3.47 | 12.8 | 15.4 |
| 114 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | -1.25175 |
| 118 | 7.16 | 7.6 | 7.85 | 9.89 | 6.95 | 10.7 | 10.3 | 7.81 | 5.68 | 5.37 | 10.6 | 18 |
| 128 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | 1.71 | -1.767 | -1.4115 | -1.43025 | -1.606 | 2.9 | 3.25 |
| 137 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | -1.25175 |
| 138 | 9.81 | 9.79 | 10.9 | 11 | 9.27 | 12.6 | 14.4 | 9.13 | 7.18 | 7.05 | 33.4 | 19 |
| 141 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | -1.25175 |
| 149 | 5.64 | 4.96 | 5.62 | 5.81 | 5 | 6.96 | 7.82 | 3.77 | 2.76 | 2.96 | 27.7 | 10.5 |
| 151 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | 1.57 | 1.83 | -1.4115 | -1.43025 | -1.606 | 12.2 | 3.56 |
| 153 | 16.9 | 19.4 | 20.3 | 19.3 | 16.9 | 20.5 | 23.9 | 21.9 | 22.5 | 20.6 | 64.7 | 32 |
| 156 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | 1.85 | 1.28 |
| 157 | 1.94 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | -1.25175 |
| 158 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | 2.04 | 1.42 |
| 170 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | 3.16 | -1.25175 |
| 174 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | 4.03 | -1.25175 |
| 177 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | 9.04 | 1.71 |
| 180 | 2.94 | 3.27 | 3.56 | 3.31 | 2.66 | 3.8 | 3.89 | 3.56 | 3.59 | 3.17 | 11.6 | 4.66 |
| 183 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | 5.78 | 1.42 |
| 187 | 3.33 | 3.16 | 3.62 | 3.1 | 2.67 | 4.41 | 3.99 | 2.26 | 2.13 | 2.25 | 18.4 | 5.22 |
| 189 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | -1.25175 |
| 194 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | -1.25175 |
| 195 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | -1.25175 |
| 200 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | -1.25175 |
| 201 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | 1.45 | -1.25175 |
| 203 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | 3.59 | 1.32 |
| 206 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | -1.25175 |
| 207 | 109.7 | 119.1 | 125.9 | 122.2 | 92.8 | 96.2 | 91.8 | 101.1 | 87.4 | 114.8 | 109.7 | 111.8 |
| 209 | -1.3395 | -1.627 | -1.64475 | -1.73875 | -1.31925 | -1.44025 | -1.767 | -1.4115 | -1.43025 | -1.606 | -1.3665 | -1.25175 |

Appendix J: Water and sediment toxicity

| ToxSpeciesName | StationCode | StationName | MatrixName | Unit | % of control-- Dry | % of control-- Spring | % of control-- Wet |
|------------------------------|-------------|------------------------|-------------|----------|-----------------------|--------------------------|-----------------------|
| H. azteca % survival | 201WLK090 | Walker Creek | sediment | % | 90.8 | | |
| H. azteca growth | 201WLK090 | Walker Creek | sediment | mg/ind | 120 | | |
| C. dubia % survival | 201LAG040 | Olema Low | samplewater | % | | 100 | |
| C. dubia young/female | 201LAG040 | Olema Low | samplewater | Num/Rep | | 137 | |
| P. promelas % survival | 201LAG040 | Olema Low | samplewater | % | | 108 | |
| P. promelas growth | 201LAG040 | Olema Low | samplewater | mg/ind | | 106 | |
| S. capricornutum cell growth | 201LAG040 | Olema Low | samplewater | cells/ml | | 126 | |
| H. azteca % survival | 201LAG120 | Green Bridge (Hwy 1) | sediment | % | 104 | | |
| H. azteca growth | 201LAG120 | Green Bridge (Hwy 1) | sediment | mg/ind | 84.3 | | |
| C. dubia % survival | 201LAG130 | Gallagher's Ranch | samplewater | % | 100 | 100 | |
| C. dubia young/female | 201LAG130 | Gallagher's Ranch | samplewater | Num/Rep | 132 | 139 | |
| P. promelas % survival | 201LAG130 | Gallagher's Ranch | samplewater | % | 103 | 108 | |
| P. promelas growth | 201LAG130 | Gallagher's Ranch | samplewater | mg/ind | 90 | 114 | |
| S. capricornutum cell growth | 201LAG130 | Gallagher's Ranch | samplewater | cells/ml | 106 | 121 | |
| C. dubia % survival | 201LAG270 | Creamery Gulch | samplewater | % | 100 | 100 | |
| C. dubia young/female | 201LAG270 | Creamery Gulch | samplewater | Num/Rep | 111 | 132 | |
| H. azteca % survival | 201LAG270 | Creamery Gulch | sediment | % | 96.1 | | |
| H. azteca growth | 201LAG270 | Creamery Gulch | sediment | mg/ind | 115 | | |
| P. promelas % survival | 201LAG270 | Creamery Gulch | samplewater | % | 97 | 103 | |
| P. promelas growth | 201LAG270 | Creamery Gulch | samplewater | mg/ind | 98 | 111 | |
| S. capricornutum cell growth | 201LAG270 | Creamery Gulch | samplewater | cells/ml | 109 | 132 | |
| C. dubia % survival | 202BUT010 | Bean Hollow | samplewater | % | 100 | 100 | 100 |
| C. dubia young/female | 202BUT010 | Bean Hollow | samplewater | Num/Rep | 95.9 | 89.1 | 118 |
| H. azteca % survival | 202BUT010 | Bean Hollow | sediment | % | 103 | | |
| H. azteca growth | 202BUT010 | Bean Hollow | sediment | mg/ind | 62.9 | | |
| P. promelas % survival | 202BUT010 | Bean Hollow | samplewater | % | 108 | 100 | 103 |
| P. promelas growth | 202BUT010 | Bean Hollow | samplewater | mg/ind | 109 | 106 | 103 |
| S. capricornutum cell growth | 202BUT010 | Bean Hollow | samplewater | cells/ml | 130 | 109 | 111 |
| C. dubia % survival | 202PES050 | Water Lane | samplewater | % | 100 | 100 | 100 |
| C. dubia young/female | 202PES050 | Water Lane | samplewater | Num/Rep | 87.7 | 110 | 147 |
| H. azteca % survival | 202PES050 | Water Lane | sediment | % | 86.5 | | |
| H. azteca growth | 202PES050 | Water Lane | sediment | mg/ind | 63 | | |
| P. promelas % survival | 202PES050 | Water Lane | samplewater | % | 105 | 109 | 103 |
| P. promelas growth | 202PES050 | Water Lane | samplewater | mg/ind | 106 | 111 | 96 |
| S. capricornutum cell growth | 202PES050 | Water Lane | samplewater | cells/ml | 132 | 116 | 105 |
| C. dubia % survival | 202PES070 | Cloverdale Rd | samplewater | % | 100 | 100 | 100 |
| C. dubia young/female | 202PES070 | Cloverdale Rd | samplewater | Num/Rep | 87.4 | 97.7 | 155 |
| P. promelas % survival | 202PES070 | Cloverdale Rd | samplewater | % | 105 | 100 | 92 |
| P. promelas growth | 202PES070 | Cloverdale Rd | samplewater | mg/ind | 108 | 111 | 91 |
| S. capricornutum cell growth | 202PES070 | Cloverdale Rd | samplewater | cells/ml | 116 | 105 | 91.6 |
| C. dubia % survival | 202SGR010 | San Gregorio USGS Gage | samplewater | % | 142.9 | 100 | 100 |
| C. dubia young/female | 202SGR010 | San Gregorio USGS Gage | samplewater | Num/Rep | 118 | 95.2 | 119 |
| H. azteca % survival | 202SGR010 | San Gregorio USGS Gage | sediment | % | 104 | | |
| H. azteca growth | 202SGR010 | San Gregorio USGS Gage | sediment | mg/ind | 70.6 | | |
| P. promelas % survival | 202SGR010 | San Gregorio USGS Gage | samplewater | % | 95 | 97 | 100 |
| P. promelas growth | 202SGR010 | San Gregorio USGS Gage | samplewater | mg/ind | 109 | 106 | 88 |
| S. capricornutum cell growth | 202SGR010 | San Gregorio USGS Gage | samplewater | cells/ml | 107 | 115 | 116 |
| C. dubia % survival | 202SGR080 | La Honda at Confluence | samplewater | % | 142.9 | 100 | 100 |
| C. dubia young/female | 202SGR080 | La Honda at Confluence | samplewater | Num/Rep | 94.9 | 96.1 | 130 |
| P. promelas % survival | 202SGR080 | La Honda at Confluence | samplewater | % | 94 | 81 | 103 |
| P. promelas growth | 202SGR080 | La Honda at Confluence | samplewater | mg/ind | 100 | 113 | 72 |
| S. capricornutum cell growth | 202SGR080 | La Honda at Confluence | samplewater | cells/ml | 111 | 112 | 106 |
| C. dubia % survival | 204ALP010 | El Charro | samplewater | % | 90 | 100 | |
| C. dubia young/female | 204ALP010 | El Charro | samplewater | Num/Rep | 105 | 128 | |
| H. azteca % survival | 204ALP010 | El Charro | sediment | % | 83.8 | | |
| H. azteca growth | 204ALP010 | El Charro | sediment | mg/ind | 104 | | |
| P. promelas % survival | 204ALP010 | El Charro | samplewater | % | 105 | 105 | |
| P. promelas growth | 204ALP010 | El Charro | samplewater | mg/ind | 110 | 109 | |
| S. capricornutum cell growth | 204ALP010 | El Charro | samplewater | cells/ml | 101 | 61.8 | |
| C. dubia % survival | 204ALP100 | Altamont Creek | samplewater | % | 100 | 100 | |
| C. dubia young/female | 204ALP100 | Altamont Creek | samplewater | Num/Rep | 117 | 115 | |
| P. promelas % survival | 204ALP100 | Altamont Creek | samplewater | % | 105 | 103 | |

Appendix J: Water and sediment toxicity

| ToxSpeciesName | StationCode | StationName | MatrixName | Unit | % of control-- Dry | % of control-- Spring | % of control-- Wet |
|-------------------------------------|-------------|--------------------------|-------------|----------|-----------------------|--------------------------|-----------------------|
| <i>P. promelas</i> growth | 204ALP100 | Altamont Creek | samplewater | mg/ind | 104 | 96 | |
| <i>S. capricornutum</i> cell growth | 204ALP100 | Altamont Creek | samplewater | cells/ml | 91.4 | 79.5 | |
| <i>C. dubia</i> % survival | 204ALP110 | Arroyo las Positas | samplewater | % | 100 | 100 | |
| <i>C. dubia</i> young/female | 204ALP110 | Arroyo las Positas | samplewater | Num/Rep | 104 | 116 | |
| <i>P. promelas</i> % survival | 204ALP110 | Arroyo las Positas | samplewater | % | 105 | 103 | |
| <i>P. promelas</i> growth | 204ALP110 | Arroyo las Positas | samplewater | mg/ind | 105 | 124 | |
| <i>S. capricornutum</i> cell growth | 204ALP110 | Arroyo las Positas | samplewater | cells/ml | 64.7 | 55.6 | |
| <i>C. dubia</i> % survival | 204SLE030 | Empire Road | samplewater | % | 100 | 90 | |
| <i>C. dubia</i> young/female | 204SLE030 | Empire Road | samplewater | Num/Rep | 63 | 141 | |
| <i>H. azteca</i> % survival | 204SLE030 | Empire Road | sediment | % | 0 | | |
| <i>H. azteca</i> growth | 204SLE030 | Empire Road | sediment | mg/ind | | | |
| <i>P. promelas</i> % survival | 204SLE030 | Empire Road | samplewater | % | 100 | 108 | |
| <i>P. promelas</i> growth | 204SLE030 | Empire Road | samplewater | mg/ind | 99 | 120 | |
| <i>S. capricornutum</i> cell growth | 204SLE030 | Empire Road | samplewater | cells/ml | 107 | 86.7 | |
| <i>C. dubia</i> % survival | 204SLE230 | Kaiser Creek at Callahan | samplewater | % | 90 | 100 | |
| <i>C. dubia</i> young/female | 204SLE230 | Kaiser Creek at Callahan | samplewater | Num/Rep | 121 | 91 | |
| <i>P. promelas</i> % survival | 204SLE230 | Kaiser Creek at Callahan | samplewater | % | 100 | 94 | |
| <i>P. promelas</i> growth | 204SLE230 | Kaiser Creek at Callahan | samplewater | mg/ind | 94 | 110 | |
| <i>S. capricornutum</i> cell growth | 204SLE230 | Kaiser Creek at Callahan | samplewater | cells/ml | 117 | 135 | |
| <i>C. dubia</i> % survival | 205PER010 | Charleston Rd | samplewater | % | 142.9 | 100 | 80 |
| <i>C. dubia</i> young/female | 205PER010 | Charleston Rd | samplewater | Num/Rep | 101 | 93.8 | 72.5 |
| <i>H. azteca</i> % survival | 205PER010 | Charleston Rd | sediment | % | 96 | | |
| <i>H. azteca</i> growth | 205PER010 | Charleston Rd | sediment | mg/ind | 72.1 | | |
| <i>P. promelas</i> % survival | 205PER010 | Charleston Rd | samplewater | % | 97 | 98 | 92 |
| <i>P. promelas</i> growth | 205PER010 | Charleston Rd | samplewater | mg/ind | 89 | 115 | 78 |
| <i>S. capricornutum</i> cell growth | 205PER010 | Charleston Rd | samplewater | cells/ml | 57.3 | 59.4 | 44.6 |
| <i>C. dubia</i> % survival | 205PER070 | Rancho San Antonio | samplewater | % | 128.6 | 100 | 70 |
| <i>C. dubia</i> young/female | 205PER070 | Rancho San Antonio | samplewater | Num/Rep | 78.8 | 86.7 | 80 |
| <i>P. promelas</i> % survival | 205PER070 | Rancho San Antonio | samplewater | % | 100 | 87 | 84 |
| <i>P. promelas</i> growth | 205PER070 | Rancho San Antonio | samplewater | mg/ind | 80 | 109 | 76 |
| <i>S. capricornutum</i> cell growth | 205PER070 | Rancho San Antonio | samplewater | cells/ml | 59.1 | 60.9 | 55.6 |
| <i>C. dubia</i> % survival | 205STE020 | La Avenida | samplewater | % | 142.9 | 100 | 100 |
| <i>C. dubia</i> young/female | 205STE020 | La Avenida | samplewater | Num/Rep | 99 | 80.5 | 127 |
| <i>H. azteca</i> % survival | 205STE020 | La Avenida | sediment | % | 86.5 | | |
| <i>H. azteca</i> growth | 205STE020 | La Avenida | sediment | mg/ind | 71.3 | | |
| <i>P. promelas</i> % survival | 205STE020 | La Avenida | samplewater | % | 91 | 95 | 95 |
| <i>P. promelas</i> growth | 205STE020 | La Avenida | samplewater | mg/ind | 87 | 121 | 92 |
| <i>S. capricornutum</i> cell growth | 205STE020 | La Avenida | samplewater | cells/ml | 93 | 29 | 76.6 |
| <i>C. dubia</i> % survival | 205STE060 | "Belleville"/Barranca | samplewater | % | 142.93 | 100 | 70 |
| <i>C. dubia</i> young/female | 205STE060 | "Belleville"/Barranca | samplewater | Num/Rep | 96.7 | 92.2 | 129 |
| <i>P. promelas</i> % survival | 205STE060 | "Belleville"/Barranca | samplewater | % | 98 | 73 | 97 |
| <i>P. promelas</i> growth | 205STE060 | "Belleville"/Barranca | samplewater | mg/ind | 101 | 112 | 90 |
| <i>S. capricornutum</i> cell growth | 205STE060 | "Belleville"/Barranca | samplewater | cells/ml | 107 | 132 | 104 |
| <i>C. dubia</i> % survival | 206SPA020 | 3rd St Bridge | samplewater | % | 100 | 100 | |
| <i>C. dubia</i> young/female | 206SPA020 | 3rd St Bridge | samplewater | Num/Rep | 121 | 103 | |
| <i>H. azteca</i> % survival | 206SPA020 | 3rd St Bridge | sediment | % | 80.9 | | |
| <i>H. azteca</i> growth | 206SPA020 | 3rd St Bridge | sediment | mg/ind | 85 | | |
| <i>P. promelas</i> % survival | 206SPA020 | 3rd St Bridge | samplewater | % | 93 | 98 | |
| <i>P. promelas</i> growth | 206SPA020 | 3rd St Bridge | samplewater | mg/ind | 92 | 117 | |
| <i>S. capricornutum</i> cell growth | 206SPA020 | 3rd St Bridge | samplewater | cells/ml | 55.1 | 65.7 | |
| <i>C. dubia</i> % survival | 206SPA070 | Cemetery Bridge | samplewater | % | 100 | 100 | |
| <i>C. dubia</i> young/female | 206SPA070 | Cemetery Bridge | samplewater | Num/Rep | 125 | 134 | |
| <i>P. promelas</i> % survival | 206SPA070 | Cemetery Bridge | samplewater | % | 95 | 108 | |
| <i>P. promelas</i> growth | 206SPA070 | Cemetery Bridge | samplewater | mg/ind | 104 | 129 | |
| <i>S. capricornutum</i> cell growth | 206SPA070 | Cemetery Bridge | samplewater | cells/ml | 103 | 53.6 | |
| <i>C. dubia</i> % survival | 206SPA200 | Lauterwasser Creek | samplewater | % | 100 | 100 | |
| <i>C. dubia</i> young/female | 206SPA200 | Lauterwasser Creek | samplewater | Num/Rep | 99 | 102 | |
| <i>P. promelas</i> % survival | 206SPA200 | Lauterwasser Creek | samplewater | % | 95 | 100 | |
| <i>P. promelas</i> growth | 206SPA200 | Lauterwasser Creek | samplewater | mg/ind | 93 | 116 | |
| <i>S. capricornutum</i> cell growth | 206SPA200 | Lauterwasser Creek | samplewater | cells/ml | 128 | 118 | |
| <i>C. dubia</i> % survival | 206SPA220 | Orinda Village | samplewater | % | 90 | 100 | |
| <i>C. dubia</i> young/female | 206SPA220 | Orinda Village | samplewater | Num/Rep | 93 | 137 | |

Appendix J: Water and sediment toxicity

| ToxSpeciesName | StationCode | StationName | MatrixName | Unit | % of control-- Dry | % of control-- Spring | % of control-- Wet |
|-------------------------------------|-------------|------------------------------|-------------|----------|-----------------------|--------------------------|-----------------------|
| <i>P. promelas</i> % survival | 206SPA220 | Orinda Village | samplewater | % | 89 | 108 | |
| <i>P. promelas</i> growth | 206SPA220 | Orinda Village | samplewater | mg/ind | 95 | 113 | |
| <i>S. capricornutum</i> cell growth | 206SPA220 | Orinda Village | samplewater | cells/ml | 107 | 125 | |
| <i>C. dubia</i> % survival | 206WIL020 | Richmond Parkway | samplewater | % | 143 | 100 | |
| <i>C. dubia</i> young/female | 206WIL020 | Richmond Parkway | samplewater | Num/Rep | 145 | 162 | |
| <i>H. azteca</i> % survival | 206WIL020 | Richmond Parkway | sediment | % | 81.1 | | |
| <i>H. azteca</i> growth | 206WIL020 | Richmond Parkway | sediment | mg/ind | 51 | | |
| <i>P. promelas</i> % survival | 206WIL020 | Richmond Parkway | samplewater | % | 97 | 100 | |
| <i>P. promelas</i> growth | 206WIL020 | Richmond Parkway | samplewater | mg/ind | 103 | 121 | |
| <i>S. capricornutum</i> cell growth | 206WIL020 | Richmond Parkway | samplewater | cells/ml | 98.9 | 84.4 | |
| <i>C. dubia</i> % survival | 207SUI010 | Cordelia | samplewater | % | 100 | 100 | |
| <i>C. dubia</i> young/female | 207SUI010 | Cordelia | samplewater | Num/Rep | 118 | 97 | |
| <i>P. promelas</i> % survival | 207SUI010 | Cordelia | samplewater | % | 100 | 95 | |
| <i>P. promelas</i> growth | 207SUI010 | Cordelia | samplewater | mg/ind | 87 | 116 | |
| <i>S. capricornutum</i> cell growth | 207SUI010 | Cordelia | samplewater | cells/ml | 121 | 129 | |
| <i>C. dubia</i> % survival | 207SUI020 | Rockville | samplewater | % | 143 | | |
| <i>C. dubia</i> young/female | 207SUI020 | Rockville | samplewater | Num/Rep | 132 | | |
| <i>H. azteca</i> % survival | 207SUI020 | Rockville | sediment | % | 89.2 | | |
| <i>H. azteca</i> growth | 207SUI020 | Rockville | sediment | mg/ind | 97.5 | | |
| <i>P. promelas</i> % survival | 207SUI020 | Rockville | samplewater | % | 105 | | |
| <i>P. promelas</i> growth | 207SUI020 | Rockville | samplewater | mg/ind | 87 | | |
| <i>S. capricornutum</i> cell growth | 207SUI020 | Rockville | samplewater | cells/ml | 106 | | |
| <i>C. dubia</i> % survival | 207SUI060 | Putah South Canal (upstream) | samplewater | % | 100 | 100 | |
| <i>C. dubia</i> young/female | 207SUI060 | Putah South Canal (upstream) | samplewater | Num/Rep | 104 | 118 | |
| <i>P. promelas</i> % survival | 207SUI060 | Putah South Canal (upstream) | samplewater | % | 97 | 95 | |
| <i>P. promelas</i> growth | 207SUI060 | Putah South Canal (upstream) | samplewater | mg/ind | 99 | 127 | |
| <i>S. capricornutum</i> cell growth | 207SUI060 | Putah South Canal (upstream) | samplewater | cells/ml | 110 | 129 | |
| <i>C. dubia</i> % survival | 207SUI110 | Wooden Valley | samplewater | % | 143 | 100 | |
| <i>C. dubia</i> young/female | 207SUI110 | Wooden Valley | samplewater | Num/Rep | 116 | 96.1 | |
| <i>P. promelas</i> % survival | 207SUI110 | Wooden Valley | samplewater | % | 97 | 95 | |
| <i>P. promelas</i> growth | 207SUI110 | Wooden Valley | samplewater | mg/ind | 90 | 104 | |
| <i>S. capricornutum</i> cell growth | 207SUI110 | Wooden Valley | samplewater | cells/ml | 80.2 | 135 | |

Appendix K: Summary coliform data (including 2005 re-sample of some sites)

| Station ID | Station Name | Estimated Use | E. coli (MPN/100 ml) | | | | | Geometric Mean |
|---|--------------------------|---------------|----------------------|---------|---------|---------|---------|----------------|
| 2001 Sites | | | 8/7/01 | 8/14/01 | 8/21/01 | 8/28/01 | 9/4/01 | |
| Walker Creek Watershed | | | | | | | | |
| 201WLK162 | Turtle Pond | beach | 70 | 40 | 30 | 170 | 80 | 65 |
| Lagunitas Creek Watershed | | | | | | | | |
| 201LAG390 | Cataract | | ND (2) | ND (2) | 9 | 20 | 40 | 8 |
| 201LAG115 | Hagmaier Pond | | 8 | 20 | 1600 | 80 | 230 | 86 |
| 201LAG120 | Green Bridge | | 300 | 30 | 1600 | 800 | 1100 | 417 |
| 201LAG185 | Swimming hole @ SPTaylor | beach | 50 | 70 | 130 | 40 | 900 | 110 |
| 201LAG230 | Inkwells | | 70 | 50 | 240 | 110 | 40 | 82 |
| San Leandro Creek Watershed | | | | | | | | |
| 204SLE190 | Canyon School | | 300 | 300 | 110 | 130 | 700 | 246 |
| 204SLE200 | Huckleberry Preserve | | 170 | 3000 | 110 | 20 | ND (20) | 118 |
| 204SLE170 | Redwood Park (West Fork) | | 220 | 80 | 13 | 20 | 130 | 57 |
| 204SLE090 | Chabot City Park | | 110 | 230 | 23 | 40 | 80 | 71 |
| 204SLE070 | Root Park | | 1600 | 170 | 170 | >16000 | 220 | 696 |
| Wildcat/San Pablo Creeks Watershed | | | | | | | | |
| 206WIL130 | Lone Oak | | 500 | 120 | 50 | 40 | 70 | 97 |
| 206WIL070 | Alvarado Park | | 170 | 500 | 500 | 230 | 40 | 208 |
| 206SPA060 | San Pablo City Park | | 900 | 5000 | 1600 | 500 | 140 | 872 |
| 206SPA150 | Briones 1 | | 130 | 170 | 900 | 230 | NA | 260 |

| 2002 Sites | | | 7/22/02 | 7/29/02 | 8/5/02 | 8/12/02 | 8/19/02 | Geometric Mean |
|--|--------------------------|-------|---------|---------|--------|---------|---------|----------------|
| Pescadero/Butano Creeks Watershed | | | | | | | | |
| 202BUT040 | Butano Falls | | 30 | 17 | 27 | 500 | 30 | 46 |
| 202BUT050 | Butano State Park | | 7 | ND (2) | 4 | 4 | 4 | 4 |
| 202PES060 | Community Church | | 1600 | 240 | 130 | 280 | 500 | 371 |
| 202PES134 | Memorial Park Swim 1 | beach | 110 | 30 | 80 | 13 | 17 | 36 |
| 202PES135 | Memorial Park Swim 2 | beach | 30 | 50 | 50 | 13 | 30 | 31 |
| 202PES150 | Jones Gulch | | 2 | ND (2) | 23 | 4 | ND (2) | 4 |
| 202PES193 | Iverson Trail | | 23 | 80 | 13 | 8 | 23 | 21 |
| 202PES194 | Sequoia Nature Trail | | 70 | 4 | 22 | 8 | 23 | 16 |
| San Gregorio Creek Watershed | | | | | | | | |
| 202SGR079 | San Gregorio confluence | | 300 | 300 | 240 | 240 | 240 | 262 |
| 202SGR100 | Playbowl | | 130 | 240 | 170 | 170 | 130 | 164 |
| Stevens/Permanente Creeks Watershed | | | | | | | | |
| 205STE020 | L'Avenida | | >1600 | >1600 | 500 | 500 | 500 | 796 |
| 205STE070 | Chestnut Picnic Area | | 7 | 80 | 50 | >1600 | 7 | 50 |
| 205STE080 | Above the Reservoir | | 130 | 140 | 30 | 240 | 50 | 92 |
| 205STE090 | Camp Cooley | | 220 | 23 | 50 | >1600 | 500 | 182 |
| 205PER080 | Lower Meadow/West Branch | | 300 | 240 | 80 | 50 | 220 | 145 |

| 2001 Sites | | | 7/12/05 | 7/19/05 | 7/26/05 | 8/2/05 | 8/9/05 | 8/16/05 | Geo Mean |
|---|----------------------|--|---------|---------|---------|--------|--------|---------|----------|
| Lagunitas Creek Watershed | | | | | | | | | |
| 201LAG115 | Hagmaier Pond | | ND | 10 | ND | ND | 10 | | 10 |
| San Leandro Creek Watershed | | | | | | | | | |
| 204SLE190 | Canyon School | | 180 | 160 | 98 | 150 | 120 | | 138 |
| 204SLE200 | Huckleberry Preserve | | 52 | 98 | 63 | 4600 | 400 | | 226 |
| 204SLE070 | Root Park | | 63 | 10 | 110 | 30 | 41 | | 39 |
| Wildcat/San Pablo Creeks Watershed | | | | | | | | | |
| 206WIL070 | Alvarado Park | | 85 | 120 | 520 | 74 | 110 | | 134 |
| 206SPA060 | San Pablo City Park | | | 74 | 220 | 150 | 310 | 680 | 220 |
| 206SPA150 | Briones 1 | | 680 | 990 | 730 | 1100 | 1400 | | 946 |

Appendix K: Summary coliform data (including 2005 re-sample of some sites)

| Station ID | Fecal Coliform (MPN/100 ml) | | | | | | | |
|---|-----------------------------|---------|---------|---------|---------|----------------|-----------------|---------|
| 2001 Sites | 8/7/01 | 8/14/01 | 8/21/01 | 8/28/01 | 9/4/01 | Geometric Mean | 90th Percentile | Average |
| Walker Creek Watershed | | | | | | | | |
| 201WLK162 | 70 | 40 | 30 | 170 | 80 | 65 | 134 | 78 |
| Lagunitas Creek Watershed | | | | | | | | |
| 201LAG390 | ND (2) | ND (2) | 9 | 20 | 40 | 8 | 32 | 15 |
| 201LAG115 | 20 | 20 | 1600 | 80 | 230 | 103 | 1052 | 390 |
| 201LAG120 | 300 | 30 | 1600 | 800 | 1100 | 417 | 1400 | 766 |
| 201LAG185 | 50 | 70 | 130 | 40 | 900 | 110 | 592 | 238 |
| 201LAG230 | 70 | 50 | 240 | 110 | 40 | 82 | 188 | 102 |
| San Leandro Creek Watershed | | | | | | | | |
| 204SLE190 | 300 | 300 | 110 | 130 | 700 | 246 | 540 | 308 |
| 204SLE200 | 170 | 3000 | 110 | 20 | ND (20) | 118 | 1868 | 664 |
| 204SLE170 | 220 | 80 | 13 | 20 | 130 | 57 | 184 | 93 |
| 204SLE090 | 110 | 230 | 23 | 40 | 80 | 71 | 182 | 97 |
| 204SLE070 | 1600 | 170 | 170 | >16000 | 220 | 696 | 10240 | 3632 |
| Wildcat/San Pablo Creeks Watershed | | | | | | | | |
| 206WIL150 | 500 | 120 | 50 | 40 | 70 | 97 | 348 | 156 |
| 206WIL070 | 170 | 500 | 500 | 230 | 40 | 208 | 500 | 288 |
| 206SPA060 | 900 | 5000 | 1600 | 500 | 140 | 872 | 3640 | 1628 |
| 206SPA150 | 130 | 170 | 900 | 230 | NA | 260 | 699 | 358 |

| 2002 Sites | 7/22/02 | 7/29/02 | 8/5/02 | 8/12/02 | 8/19/02 | Geometric Mean | 90th Percentile | Average |
|--|---------|---------|--------|---------|---------|----------------|-----------------|---------|
| Pescadero/Butano Creeks Watershed | | | | | | | | |
| 202BUT040 | 30 | 17 | 27 | 500 | 30 | 46 | 312 | 121 |
| 202BUT050 | 7 | ND (2) | 4 | 4 | 4 | 4 | 6 | 4 |
| 202PES060 | 1600 | 240 | 130 | 280 | 500 | 371 | 1160 | 550 |
| 202PES134 | 110 | 30 | 80 | 13 | 17 | 36 | 98 | 50 |
| 202PES135 | 30 | 50 | 50 | 13 | 30 | 31 | 50 | 35 |
| 202PES150 | 2 | 1 | 23 | 4 | ND (2) | 3 | 15 | 6 |
| 202PES193 | 23 | 80 | 13 | 8 | 23 | 21 | 57 | 29 |
| 202PES194 | 70 | 7 | 22 | 8 | 23 | 18 | 51 | 26 |
| San Gregorio Creek Watershed | | | | | | | | |
| 202SGR079 | 300 | 300 | 240 | 240 | 240 | 262 | 300 | 264 |
| 202SGR100 | 130 | 240 | 170 | 170 | 130 | 164 | 212 | 168 |
| Stevens/Permanente Creeks Watershed | | | | | | | | |
| 205STE020 | >1600 | >1600 | 500 | 500 | 500 | 796 | 1600 | 940 |
| 205STE070 | 7 | 80 | 50 | >1600 | 7 | 50 | 992 | 349 |
| 205STE080 | 130 | 140 | 30 | 240 | 50 | 92 | 200 | 118 |
| 205STE090 | 220 | 23 | 50 | >1600 | 500 | 182 | 1160 | 479 |
| 205PER090 | 300 | 240 | 80 | 50 | 220 | 145 | 276 | 178 |

Appendix K: Summary coliform data (including 2005 re-sample of some sites)

| Station ID | Total Coliform (MPN/100 ml) | | | | | | Geometric | |
|---|-----------------------------|---------|---------|---------|---------|--------|-----------|--|
| 2001 Sites | 8/7/01 | 8/14/01 | 8/21/01 | 8/28/01 | 9/4/01 | Median | Mean | |
| Walker Creek Watershed | | | | | | | | |
| 201WLK162 | 70 | 130 | 1600 | 170 | 700 | 170 | 280 | |
| Lagunitas Creek Watershed | | | | | | | | |
| 201LAG390 | ND (2) | ND (2) | 1600 | 20 | 500 | 20 | 36 | |
| 201LAG115 | 20 | 20 | 1600 | 110 | 500 | 110 | 129 | |
| 201LAG120 | 1600 | 500 | 1600 | 1700 | 5000 | 1600 | 1612 | |
| 201LAG185 | 170 | 40 | 1600 | 80 | 9000 | 170 | 379 | |
| 201LAG230 | 140 | 1100 | 1600 | 110 | 300 | 300 | 382 | |
| San Leandro Creek Watershed | | | | | | | | |
| 204SLE190 | 300 | 700 | 1600 | 500 | 700 | 700 | 652 | |
| 204SLE200 | 170 | 3000 | 1600 | 20 | ND (20) | 170 | 201 | |
| 204SLE170 | 1600 | 230 | 1600 | 70 | 300 | 300 | 415 | |
| 204SLE090 | 170 | 500 | 1600 | 700 | 130 | 500 | 415 | |
| 204SLE070 | 1600 | 500 | 1600 | >16000 | 500 | 1600 | 1592 | |
| Wildcat/San Pablo Creeks Watershed | | | | | | | | |
| 206WIL150 | 500 | 800 | 1600 | 40 | 140 | 500 | 324 | |
| 206WIL070 | 900 | 500 | 1600 | 300 | 40 | 500 | 387 | |
| 206SPA060 | 1600 | 9000 | 1600 | 2400 | 140 | 1600 | 1506 | |
| 206SPA150 | 1600 | 2200 | 1600 | 2400 | NA | 1900 | 1917 | |

| 2002 Sites | 7/22/02 | 7/29/02 | 8/5/02 | 8/12/02 | 7/19/02 | Median | Geometric Mean |
|--|---------|---------|--------|---------|---------|--------|----------------|
| Pescadero/Butano Creeks Watershed | | | | | | | |
| 202BUT040 | 900 | 900 | 1600 | >1600 | 9000 | 1600 | 1796 |
| 202BUT050 | 1600 | 900 | >1600 | 900 | 1600 | 1600 | 1271 |
| 202PES060 | >1600 | >1600 | 13000 | 7000 | 13000 | 7000 | 4969 |
| 202PES134 | 1600 | >1600 | 8000 | 1600 | 2200 | 1600 | 2353 |
| 202PES135 | 1600 | 1600 | 8000 | >1600 | 3000 | 1600 | 2503 |
| 202PES150 | 300 | 500 | 1600 | 500 | 1600 | 500 | 719 |
| 202PES193 | >1600 | 900 | 900 | 500 | 5000 | 900 | 1265 |
| 202PES194 | 1600 | 900 | 1600 | 1600 | 5000 | 1600 | 1791 |
| San Gregorio Creek Watershed | | | | | | | |
| 202SGR079 | 1600 | 1600 | 13000 | 1600 | >1600 | 1600 | 2433 |
| 202SGR100 | 1600 | 1600 | 5000 | 1600 | 9000 | 1600 | 2839 |
| Stevens/Permanente Creeks Watershed | | | | | | | |
| 205STE020 | >1600 | >1600 | 17000 | 30000 | 30000 | 17000 | 8291 |
| 205STE070 | >1600 | >1600 | 5000 | 11000 | 8000 | 5000 | 4077 |
| 205STE080 | >1600 | >1600 | 13000 | 7000 | 17000 | 7000 | 5243 |
| 205STE090 | >1600 | >1600 | 11000 | 30000 | 16000 | 11000 | 6702 |
| 205PER090 | >1600 | 1600 | 3000 | 17000 | 5000 | 3000 | 3656 |

| 2001 Sites | 7/12/05 | 7/19/05 | 7/26/05 | 8/2/05 | 8/9/05 | 8/16/05 | Median | Geo Mean |
|---|---------|---------|---------|--------|--------|---------|--------|----------|
| Lagunitas Creek Watershed | | | | | | | | |
| 201LAG115 | 2500 | 3300 | 1500 | 3700 | 6100 | | 3300 | 3085 |
| San Leandro Creek Watershed | | | | | | | | |
| 204SLE190 | 2100 | 1800 | 1700 | 2400 | 3900 | | 2100 | 2269 |
| 204SLE200 | 5200 | 12000 | 8300 | 17000 | 11000 | | 11000 | 9936 |
| 204SLE070 | 3700 | 9200 | 24000 | 11000 | 2900 | | 9200 | 7642 |
| Wildcat/San Pablo Creeks Watershed | | | | | | | | |
| 206WIL070 | 8700 | 8200 | 24000 | 6900 | 12000 | | 8700 | 10723 |
| 206SPA060 | | 3900 | 7700 | 6900 | 6100 | 7300 | 6900 | 6209 |
| 206SPA150 | 11000 | 13000 | 13000 | 8200 | 1800 | | 11000 | 7721 |